

1 **Towards emulating an explicit organic chemistry**
2 **mechanism with random forest models**

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7 **Key Points:**

- 8 • Random forests were trained to reproduce toluene and dodecane SOA formation
9 calculated by an explicit organic chemistry mechanism
10 • The random forests performances are highly sensitive to the chemical regime
11 • The number of predictors cannot be too high and their selection depends on the
12 precursor

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Abstract

Predicting secondary organic aerosol (SOA) formation relies either on extremely detailed, numerically expensive models accounting for the condensation of individual species or on extremely simplified, numerically affordable models parameterizing SOA formation for large-scale simulations. In this work, we explore the possibility of creating a random forest to reproduce the behavior of a detailed atmospheric organic chemistry model at a fraction of the numerical cost. A comprehensive dataset was created based on thousands of individual detailed simulations, randomly initialized to account for the variety of atmospheric chemical environments. Recurrent random forests were trained to predict organic matter formation from dodecane and toluene precursors, and the partitioning between gas and particle phases. Validation tests show that the random forests perform well without any divergence over 10 days of simulations. The distribution of errors shows that the sampling of initial conditions for the training simulations needs to focus on chemical regimes where SOA production is the most sensitive. Sensitivity tests show that specializing multiple random forests for a specific chemical regime is not more efficient than training a single general random forest for the entire dataset. The most important predictors are those providing information about the chemical regime, oxidants levels and existing organic mass. The choice of predictors is crucial as using too many unimportant predictors reduces the performances of the random forests.

Plain Language Summary

Organic compounds constitute a significant fraction of atmospheric particles and thus have an impact on health and climate. Predicting the contribution of organic compounds to atmospheric particles is extremely complex because of the very large number of different chemical species potentially condensing into the aerosol phase. Air quality and climate models usually rely on simplified, empirical approaches to predict organic aerosol mass concentrations, based on laboratory experiments. In this work, we apply a machine learning approach to construct a tool that behaves like the most detailed organic chemistry model, for a numerical cost affordable by air quality and climate models. Building upon this method, it will be possible to bring the complexities of organic chemistry to large-scale models.

1 Introduction

Secondary organic aerosol (SOA) constitutes a major fraction of atmospheric particles worldwide. It is composed of a multitude of organic compounds (*e.g.* Kourtchev et al., 2016). Our current understanding and modeling of SOA formation processes are highly uncertain (Pai et al., 2020) and involve representing the complex interplay between gas-phase oxidation and condensation of semi- and low-volatile organic species. SOA models need to include processes such as (i) the multi-step oxidation of the large variety of organic compounds emitted naturally and by human activities, (ii) the condensation of semi-volatile species to the particle phase and (iii) the heterogeneous and in-particle reactivity of condensed species. This complexity can only be represented in models that explicitly account for aerosol physico-chemical processes. In these so-called explicit models, the aim is to represent the fate of each individual chemical species through individual reactions, which can number in the 10^9 range (*e.g.* Aumont et al., 2005). The Generator of Explicit Chemistry and Kinetics for Organics in the Atmosphere (Aumont et al., 2005) is an example of such a model able to generate chemical mechanisms that explicitly describe the oxidation of organic compounds in the atmosphere, as well as their condensation into the particle phase (Camredon et al., 2007). It has previously been used to study SOA formation in various settings such as atmospheric chamber experiments (La et al., 2016), sensitivity studies (Valorso et al., 2011; Aumont et al., 2012; Hodzic

62 et al., 2015) and urban plume modeling (Lee-Taylor et al., 2015; Mouchel-Vallon et al.,
63 2020).

64 Because of their size, explicit mechanisms like those generated by GECKO-A can-
65 not be used in 3D air quality models, which rely on empirical SOA parameterizations.
66 The volatility basis set (VBS, Donahue et al., 2006) and its derivatives (*e.g.*, Donahue
67 et al., 2011; Cappa & Wilson, 2012) are the most prevalent representations of SOA chem-
68 istry in this field. In such models, simplifications are made to represent SOA formation
69 by grouping organic species of similar properties into discretized bins, *i.e.*, volatility or
70 oxidation state are the typically chosen properties for the VBS bins. This approach has
71 been presented by Pankow et al. (2015) as the *anonymized* view of SOA modeling, as
72 opposed to the *molecular* view of SOA modeling used by explicit models.

73 Previous attempts have been made to bring the molecular view of SOA modeling
74 to 3D models. Li et al. (2015) included the near explicit Master Chemical Mechanism
75 (MCM v3.2, Saunders et al., 2003; Jenkin et al., 2003), in the Community Multiscale Air
76 Quality model (CMAQ, Foley et al., 2010). MCM is a *near* explicit mechanism, as some
77 simplifications are made to simplify its development such as removing unlikely reaction
78 channels and simplifying the oxidation of minor or unknown products. At the time MCM
79 therefore used approximately 17 000 reactions involving approximately 6000 species to
80 represent the progressive oxidation of 142 primary hydrocarbons. Although the imple-
81 mentation of MCM in CMAQ was able to reproduce reasonably well the observed SOA
82 surface concentrations over eastern US for a case study, this approach did not have fur-
83 ther applications to our knowledge, and was limited by the considerable computational
84 cost required to run regional scale simulations.

85 Lannuque et al. (2018) created VBS-GECKO, an empirical VBS parameterization
86 where the stoichiometric coefficients were optimized to fit data produced from GECKO-
87 A runs instead of being fitted to empirical data. Their method had the advantage of pa-
88 rameterizing the model over the multi-day simulated aging of SOA, which cannot be ob-
89 tained from the shorter chamber studies used to derive traditional VBS parameteriza-
90 tions. Lannuque et al. (2020) ran VBS-GECKO in an air quality model (Menuet et al.,
91 2013) and showed that VBS-GECKO was producing more SOA in the summer over Eu-
92 rope compared to the traditional SOA parameterization that is based on laboratory data
93 (Couvidat et al., 2012). Because, in essence, their resulting model was a linear combi-
94 nation of multiple VBS produced for different levels of pollution, it relied on the assump-
95 tion that atmospheric chemistry behaves linearly between the selected chemical regimes.
96 As a result, VBS-GECKO may have been applied outside of its application domain.

97 Here, we propose to use a machine learning (ML) approach to bring the molecu-
98 lar view to 3D chemistry-climate models across a range of chemical regimes represen-
99 tative of tropospheric conditions. Machine learning techniques have been applied pre-
100 viously for air quality forecasts (Liao et al., 2020) demonstrating that it is possible to
101 run a trained artificial intelligence in a 3D model. Keller and Evans (2019) used the GEOS-
102 Chem chemical mechanism solver to train multiple random forests that were then able
103 to emulate the chemical solver behavior in that same model for various pollutant, for a
104 fraction of the computational cost of the default GEOS-Chem model. Kelp et al. (2020)
105 improved on this method by using a unique neural network to predict 20 chemical species.
106 They implemented it in GEOS-Chem (Kelp et al., 2022), achieving stable one-year simu-
107 lations for ozone prediction with less than 10 % bias compared to the reference and re-
108 ducing computational times by a factor of five. The motivation of these previous stud-
109 ies stems from reducing the costs of calculating chemistry, that is usually taking from
110 50 % to 90 % of the computational costs of running global chemistry models such as GEOS-
111 Chem (Keller & Evans, 2019).

112 Schreck et al. (2022) recently presented a neural network approach to emulate the
113 behavior of idealized GECKO-A simulations for the SOA formation following the oxi-

Table 1. Environmental and chemistry parameters used for generating the GECKO-A box-model dataset.

	Parameter	Range	Parameter	Range
	Latitude [°]	−80–80	Relative humidity [%]	3–102
	Temperature [K]	216–313	Atmospheric pressure [atm]	0.5–1.02
Preexisting aerosol seed [$\mu\text{g m}^{-3}$]		0.03–340	Initial NO_x [ppb]	10^{-4} –42
	Initial precursor [ppb]	0–16	Initial CO [ppb]	33–1012
	Initial O_3 [ppb]	1–100	NO Emission [$\text{molec cm}^{-2} \text{s}^{-1}$]	10^7 – 10^9

114 dation of three individual precursors reacting with OH under varied environmental con-
 115 ditions. While this work showed the ability of neural networks to reproduce idealized ox-
 116 idation situations, it showed their limitations when extrapolated to realistic simulations
 117 with diurnally varying conditions. The results indicate that this type of system needs
 118 to be trained with a dataset representative of the conditions in which it will be applied.

119 In this paper, we train a random forest on a dataset constructed with multiple GECKO-
 120 A simulations, with the primary aim of predicting SOA mass from the oxidation of toluene
 121 and dodecane for realistic atmospheric conditions over a range of chemical regimes cov-
 122 ering daytime and nighttime oxidation by the main oxidants (OH, O_3 and NO_3). Our
 123 objective is to build an empirical SOA model that is able to reproduce the aerosol mass
 124 that a complex, explicit mechanism would predict, at a numerical cost that is compa-
 125 rable to that of reduced chemical mechanisms currently used in large-scale models.

126 2 Methods

127 2.1 Reference Organic Chemistry Mechanisms

128 The Generator for Explicit Chemistry of Organics in the Atmosphere (GECKO-
 129 A, Aumont et al., 2005) is a software tool allowing the automatic generation of detailed
 130 multi-generational organic chemistry mechanisms. It is based on state-of-the-art knowl-
 131 edge of atmospheric organic chemistry and structure activity relationships (SAR) (*e.g.*,
 132 Atkinson, 1997; Raventos-Duran et al., 2010) to estimate unknown reaction kinetics and
 133 thermodynamics. It calculates the gas-particle partitioning of individual organic species
 134 based on estimates of their volatility (Valorso et al., 2011). In the present study, chem-
 135 ical mechanisms were generated for the oxidation by OH, NO_3 and O_3 of toluene (C_7H_8)
 136 and dodecane ($\text{C}_{12}\text{H}_{26}$), two compounds emitted by anthropogenic activities. Currently,
 137 GECKO-A only includes gas-phase oxidation and condensation of semi-volatile organic
 138 compounds. There are no heterogeneous processes and no aerosol phase processes (*e.g.*,
 139 oligomerization) included in the model. The resulting toluene full oxidation mechanism
 140 contains 8560 species involved in 47 349 chemical reactions, including 4536 gas-aerosol
 141 equilibrium reactions. The oxidation mechanism for dodecane was completed to the 4th
 142 generation. This was done in order to reduce the generated mechanism to a manageable
 143 size for the purpose of this work. The resulting dodecane oxidation mechanism contains
 144 75 745 species involved in 465 751 chemical reactions, including 20 968 gas-aerosol equi-
 145 librium reactions.

146 2.2 Dataset Construction

147 To create the dataset used to train and evaluate the random forest model, we ran
 148 a large set of simulations for toluene and dodecane. Each simulation is performed for 230
 149 hours with a uniformly sampled set of randomly chosen initial conditions and external
 150 forcing (see Tab. 1). Temperature, relative humidity and atmospheric pressure are se-

lected in ranges typical of values found in the lower troposphere. This ensures that the SOA emulator will have the correct sensitivity to changes in these parameters through the effects of (i) temperature on reaction rates and SOA evaporation, (ii) relative humidity on OH formation and (iii) pressure on third-body reaction rates. Initial concentrations of precursors, NO_x and CO are randomly picked to cover a wide range of chemical regimes. The latitude is varied from 80°S to 80°N to ensure that the model does not fit to a specific diurnal cycle. All model simulations start at 10:00am UTC and simulate a diurnal light cycle defined by the chosen latitude. The model time-step length is 5 minutes. After initialization, the precursor, NO_x , O_3 , and CO freely react without constraints. The other external forcing (temperature, relative humidity, pressure, latitude, NO emissions, seed) are maintained constant for the whole simulation. The simulated photochemistry leads to the multi-generational formation of semi- and low-volatile secondary organic compounds that can condense to form SOA.

2.2.1 Outcomes

In this work, the aim is to predict the distribution of organic species between gas and aerosol phases. In order to build a flexible approach that will allow future developments such as adding additional phases (*e.g.*, aqueous phase) and predicting organic matters properties (*e.g.*, solubility for deposition of organic vapors and particles), the first chosen outcome is the total organic mass m_t [$\mu\text{g m}^{-3}$]:

$$m_t = m_g + m_a \quad (1)$$

m_g and m_a are respectively the total gas- and particle-phase organic mass. The goal is to have only one outcome m_t in mass concentration units ($\mu\text{g m}^{-3}$) and predict the contributing phases to the total mass as fractions of this m_t . We arbitrarily chose to predict gaseous mass fraction γ ($m_g = \gamma \times m_t$). m_a can then be derived as $m_a = (1 - \gamma) \times m_t$.

Following the method of Keller and Evans (2019), we also established a variance criterion to decide whether the random forest should predict the value of the predictor or its trend. For stability and better performances, stable and unstable outcomes are identified. This classification is based on the standard deviation of the ratio between post- and pre- numerical time-step solve value. This ratio was calculated for each outcome and each time-step on the whole training dataset. The standard deviation of these ratios was calculated and if the value of this standard deviation is below a threshold of 0.07, the outcome is classified as stable and its trend is predicted. Otherwise, the outcome is unstable and its direct value is predicted. For the two outcomes used in this work, we found that the value of the total mass outcome m_t is unstable and needs to be directly predicted, while the gas phase fraction γ is stable and its trend is predicted.

2.2.2 Predictors

We selected the predictors based on parameters relevant to SOA formation (see Tab. 2). The concentration of the precursor as well as the main daytime oxidants (OH, O_3) and the aerosol seed concentrations have been chosen for their key role in SOA formation. The pressure and temperature modulate the kinetics that control gas-phase oxidation. Temperature is also very important for the condensation of vapors. The solar zenith angle and the photolysis rate of NO_2 represent the influence of the diurnal cycle on gas kinetics.

At each time-step of the GECKO-A simulations, predictors are computed before chemistry is integrated, and outcomes are computed after the solver has finished. 4000 simulations ($\approx 1 \times 10^6$ time-steps) were produced for each precursor, resulting in a total of 8000 simulations ($\approx 2 \times 10^6$ time-steps).

Table 2. List of predictors and outcomes used for training the random forests.

	Predictors [units]	Outcomes [units] (prediction method)
	Temperature [K]	Total organic mass [$\mu\text{g m}^{-3}$] (direct)
	Water vapor concentration [molec/cm^3]	Organic gaseous fraction [dimensionless] (trend)
	Pressure [atm]	
	Solar zenith angle [deg]	
	J_{NO_2} [s^{-1}]	
	NO [molec/cm^3]	
	NO ₂ [molec/cm^3]	
s	O ₃ [molec/cm^3]	
	OH [molec/cm^3]	
	H ₂ O ₂ [molec/cm^3]	
	CH ₂ O [molec/cm^3]	
	Aerosol seed mass [$\mu\text{g m}^{-3}$]	
	Total organic mass [$\mu\text{g m}^{-3}$]	
	Organic gaseous fraction [dimensionless]	
	Organic aerosol fraction [dimensionless]	
	Precursor [molec/cm^3]	

193 The collection of all predictor and outcome values at every time-step therefore con-
 194 stitutes a dataset representative of what the integration of an explicit chemical scheme
 195 would produce for a given predictor over one time-step. The random forest has to be re-
 196 current to autonomously predict chemistry over multiple time-steps as both outcomes
 197 (*i.e.*, total organic mass and its gas phase fraction) are also predictors.

198 2.3 Random Forest Regression

199 We use Python libraries *scikit-learn* (Pedregosa et al., 2011) to fit the random forests,
 200 and *dask* (Dask Development Team, 2016) to handle parallelization of the code on the
 201 NCAR CISL supercomputers (Computational and Information Systems Laboratory, 2017).

202 Keller and Evans (2019) trained individual random forests for each of their out-
 203 comes, and integrated the random forest model within GEOS-Chem at each model time-
 204 step. We chose to train only one random forest that predicts all outcomes simultaneously,
 205 *i.e.*, each tree predicts a vector of values instead of predicting single values. This approach
 206 has the advantage of reducing model bias by implicitly accounting for correlations be-
 207 tween outcomes. As the dataset contains values spanning many orders of magnitude, it
 208 was successively log-transformed, power-transformed and normalized to map the data
 209 as close as possible to a Gaussian distribution. The training dataset was shuffled prior
 210 to the regression procedure to avoid bias related to the random forest learning a specific
 211 diurnal cycle.

212 Hyperparameters for the random forest were tuned automatically during the ran-
 213 dom forest training with the *scikit-learn* library. The number of decision trees is the most
 214 important hyperparameter because it impacts both the numerical cost of running the
 215 random forest as well as the quality of the random forest. Random forest’s training con-
 216 figurations were tested with 10, 50, 75 and 100 decision trees. The random forest hyper-
 217 parameters optimization consistently selected 50 trees, which is the same number of de-
 218 cision trees that were selected in Keller and Evans (2019).

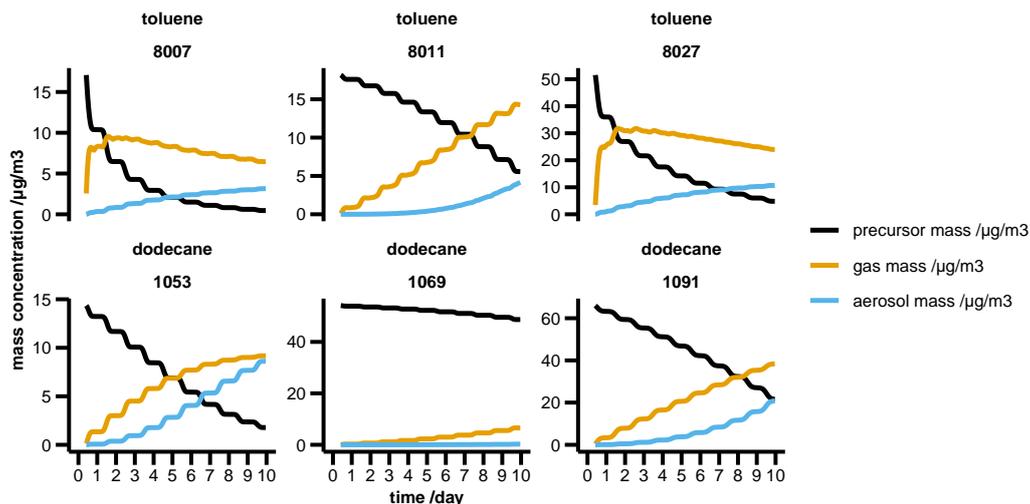


Figure 1. Time evolution of precursor oxidation, organic gas and organic aerosol formation from toluene and dodecane oxidation for representative training simulations.

3 Results

3.1 Training Dataset Characterization

Fig. 1 depicts the typical time evolution of a few randomly selected simulations from the validation dataset. As expected, the precursor is progressively oxidized during the ten days of the simulation. The kinetics of this decay depends on the concentration of the precursors' oxidants: OH (daytime) and NO_3 (nighttime) for dodecane, OH, O_3 (daytime) and NO_3 (nighttime) for toluene. Given that concentrations of these oxidants on randomly selected initial and environmental conditions in each simulation, the decay kinetics vary for each simulation. The oxidation of the precursor leads to the progressive formation of gaseous organic compounds. Depending on the availability of oxidants, the formation of these secondary organic compounds can peak early in the simulation as in simulations 8007 and 8027 in Fig. 1. On the other hand, the evolution displays a characteristic stepwise diurnal profile, with the organic mass increasing during daytime when photochemistry can take place. After the peak of the quicker oxidation simulations, the total organic mass decreases because the oxidation products are ultimately lost to the terminal CO_2 formation step. As their oxidation progresses, the secondary gaseous compounds become more oxidized and are able to condense onto the pre-existing aerosol seed, forming secondary organic aerosol (SOA). SOA formation therefore highly depends on the availability of oxidants. For instance, simulation 1069 displays a typical case of a slow precursor decay causing the slow formation of secondary organic compounds with almost no SOA production.

As shown in Fig. 1 on a few sample simulations, this work is aimed at reproducing a large variety of situations, with the SOA formation behavior that non-linearly depends on multiple parameters. Figure 2 depicts the distribution of SOA mass yield as a function of key parameters describing the chemical regimes controlling SOA formation.

The precursor controls the total amount of organic carbon that is available to form SOA. Both dodecane and toluene SOA yields are not constant as a function of the precursor concentration. This is a typical illustration of the non-linearity of SOA formation and atmospheric chemistry in general. As precursor mixing ratios increase, the precursor becomes a significant competitor for oxidants, slowing the formation of later gener-

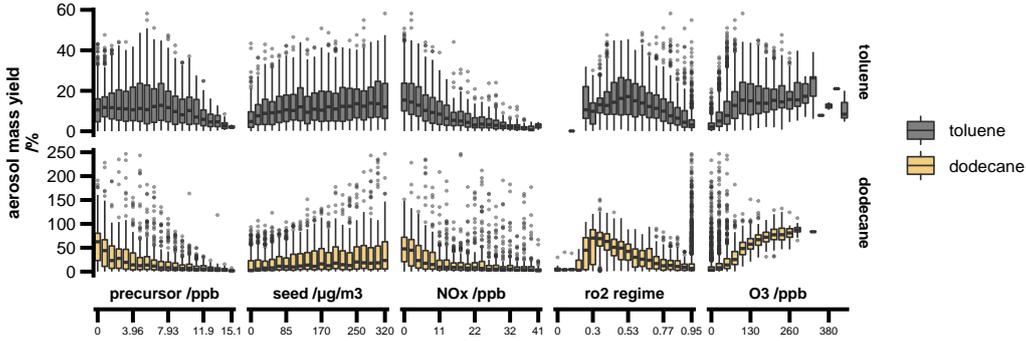


Figure 2. Dependence of average organic aerosol mass and aerosol mass yield as a function of the chemical environment.

249 ations organic compounds that are more likely to efficiently contribute to SOA. This hy-
 250 pothesis is confirmed by the fact that the dodecane SOA yield decay starts for lower pre-
 251 cursor mixing ratios, which is explained by the faster reaction rate of dodecane with the
 252 main oxidant OH compared to toluene ($k_{\text{dodecane} + \text{OH},298} = 1.32 \times 10^{-11} \text{ cm}^3/\text{molec}/\text{s}$
 253 vs. $k_{\text{toluene} + \text{OH},298} = 5.6 \times 10^{-12} \text{ cm}^3/\text{molec}/\text{s}$, Mellouki et al., 2021).

254 As the seed concentrations were selected over a wide range (see Tab. 1), the lim-
 255 iting effect of low pre-existing seeds can only be seen in the lowest chosen concentrations,
 256 below $40 \mu\text{g m}^{-3}$. In the rest of the range of seed concentrations, SOA yield has iden-
 257 tical distribution probability. In our simulations, as the nature of this seed is not accounted
 258 for, the fact that seed is present is enough to trigger SOA condensation and it is rarely
 259 limiting.

260 The NO_x mixing ratios control the formation of ozone and OH through the pho-
 261 tolysis of NO_2 and the reaction of ozone with NO. However the relationship of NO_x with
 262 oxidants levels is not trivial and also depends on the concentrations of organic compounds.
 263 Here, the simulated higher SOA yields at lower NO_x levels could be explained by the role
 264 of NO_x on the oxidation of organic compounds. After the initial oxidation step form-
 265 ing a peroxy radical ($\text{RH} + \text{OH} \xrightarrow{+\text{O}_2} \text{RO}_2 + \text{H}_2\text{O}$), the peroxy radical can react with NO
 266 to form an alkoxy radical that can fragment, leading to more volatile compounds that
 267 are less likely to form SOA. If NO concentration is low enough, peroxy radicals are more
 268 likely to react with HO_2 and other peroxy radicals to form more oxidized species that
 269 are more likely to form SOA. This can explain the higher SOA yields at lower NO_x shown
 270 on Fig. 2. This effect is better seen after defining the RO_2 regime R as:

$$R = \frac{k_{\text{RO}_2 + \text{NO}} \times \text{NO}}{k_{\text{RO}_2 + \text{NO}} \times \text{NO} + k_{\text{RO}_2 + \text{HO}_2} \times \text{HO}_2} \quad (2)$$

271 where $k_{\text{RO}_2 + \text{NO}} = 7.7 \times 10^{-12} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$, $k_{\text{RO}_2 + \text{HO}_2} = 5.1 \times 10^{-12} \text{ cm}^3 \text{ molec}^{-1} \text{ s}^{-1}$.
 272 This ratio indicates which pathway is favored for RO_2 radicals: when $R = 1$, they only
 273 react with NO and when $R = 0$ they never react with NO. For dodecane, Fig. 2 shows
 274 that from $R = 0.4$ to $R = 1$, the median value of the SOA yield decreases from 60%
 275 to less than 1%. Below $R = 0.4$, the SOA yield decreases down to less than 1% at $R =$
 276 0.15 . This low yield for low R values can be explained by low levels of oxidants lim-
 277 iting SOA production in very low NO_x situations. For toluene, this peak in SOA yield hap-
 278 pens around $R = 0.5$ with a median value of 19 %. It is less marked than for dodecane
 279 because the impact of the $\text{RO}_2 + \text{NO}$ reaction pathway on fragmentation is lower on cyclic
 280 and shorter molecules like toluene and its oxidation products (Aumont et al., 2013).

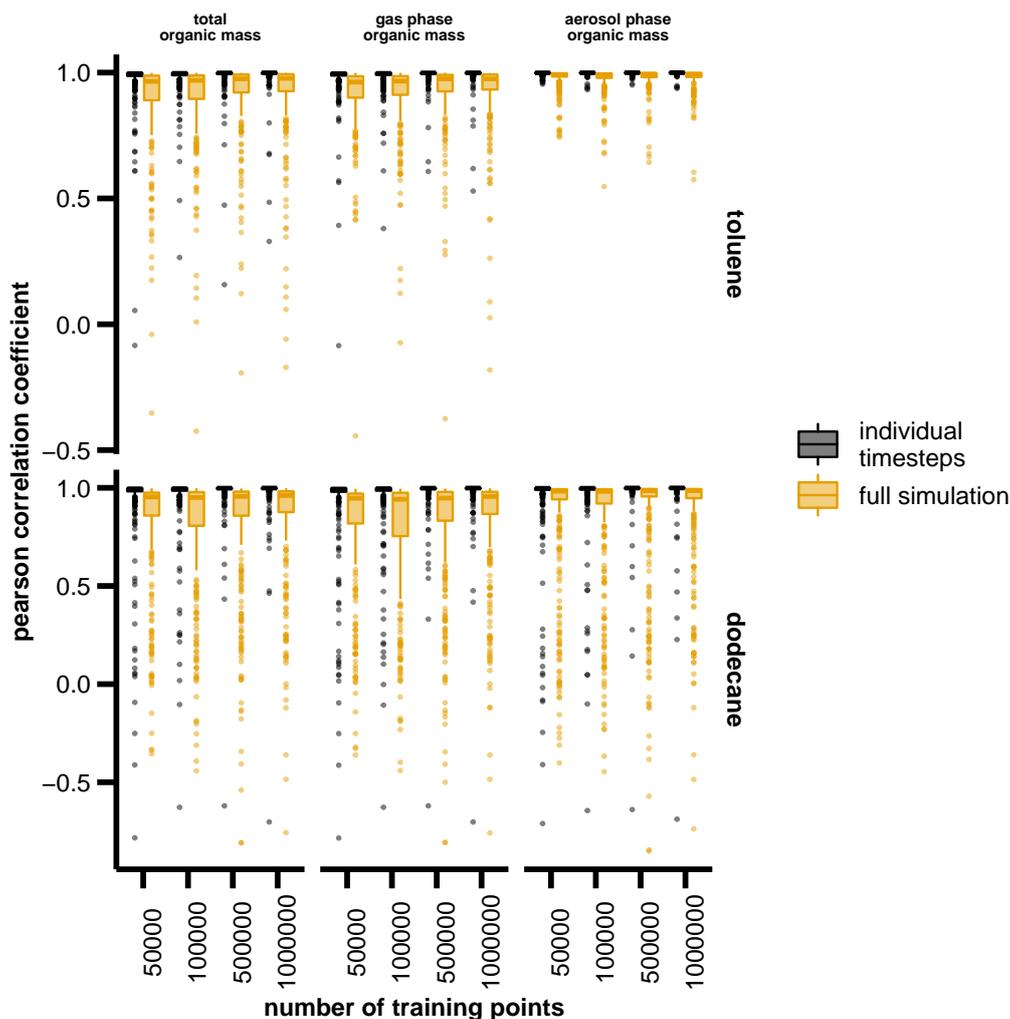


Figure 3. Boxplot distribution of Pearson correlation coefficients for testing individual time-step predictions (black) and full simulation runs (yellow) for toluene and dodecane simulations. The middle lines of the boxplots are the median, the top and bottom of the boxes denote the 1st and 3rd quartiles and the whiskers extend to the 5th and 95th percentiles of the distribution.

281 For both precursors, the SOA yield increases with ozone mixing ratios. Higher ozone
 282 mixing ratios are associated with higher OH concentrations, which can explain higher
 283 SOA yields.

284 3.2 Training Dataset Size

285 Two kinds of tests are presented for each validation simulation. First, the random
 286 forest was tested on each time-step individually by the reference predictors as inputs and
 287 comparing the random forest output with the reference outcomes. The resulting Pear-
 288 son correlation coefficients (r) distributions are shown on Fig. 3 (black boxplots), as a
 289 function of the number of points used for training the random forest. With the excep-
 290 tion of a few outliers, the random forest is very accurate to predict outcomes on a sin-
 291 gle time-step, even with only 50 000 training points. The Pearson correlation coefficient
 292 displays median values around $r = 0.99$ for both toluene and dodecane.

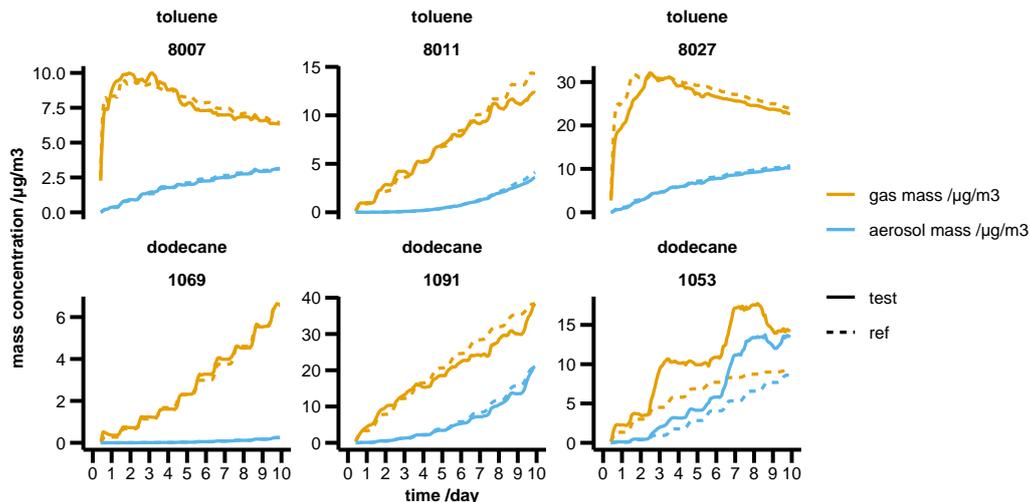


Figure 4. Time-series of three toluene and three dodecane sample experiments, comparing the reference (dashed lines) and the random forest recurrent simulations (continuous lines) for the predicted organic gas (black) and aerosol (orange) mass.

293 However, for practical application the random forest needs to be run recurrently.
 294 This means that for the second test in Fig. 3 (orange boxplots), as well as the rest of
 295 this paper, the random forest is constrained with the initial conditions and the environ-
 296 mental conditions from the reference simulation. At the end of each time-step, the pre-
 297 dicted outcomes are used in the input predictors for the next time-step. In these vali-
 298 dation tests, the recurrent random forest model accumulates errors with time. This is
 299 reflected in the median r values decreasing compared to individual time-steps tests to
 300 approximately $r = 0.98$ for toluene and $r = 0.96$ for dodecane. The r values are also
 301 more spread with the interquartile range for 100 000 points increasing from 0.002 to 0.05
 302 for toluene and from 0.003 to 0.09 for dodecane. Increasing the number of points used
 303 to train the random forest slightly improve the r scores. For toluene, using 500 000 or
 304 1 000 000 points provides similar performances while for dodecane, using 1 000 000 points
 305 still increases r and reduces the interquartile range. We could not test higher numbers
 306 of training points due to the limited size of the created dataset, but it seems that above
 307 500 000 training points, the gains are marginal at best for the considered precursors.

308 3.3 Sample Simulations Tests

309 To illustrate the behavior of the recurrent random forest model, Fig. 4 displays the
 310 recurrent random forest results on the same sample simulations that were shown on Fig.
 311 1. The associated relative errors on predicted aerosol mass are shown on Fig. 5. The ran-
 312 dom forest is able to reproduce the timeseries of gas and aerosol mass in all but one of
 313 the examples (simulation 1053). For these simulations, the random forest can reproduce
 314 the typical stepwise daytime growth of organic mass of the slower oxidation simulations
 315 (simulation 1069, 1091 and 8011), as well as reproducing the peaking growth of organic
 316 mass for faster oxidation simulations (8007 and 8027). For all simulations (except 1053),
 317 the relative error tends to be the highest for the first five days of the simulations, con-
 318 verging towards errors lower than 10% for the last five days. Finally, for the worst ran-
 319 dom forest simulation in this sample (1053), the model exhibits errors around $\pm 100\%$
 320 after 2 days and cannot recover from the accumulated errors. The relative error remains
 321 between 50 % and 100 % but not producing any unrealistic mass concentrations.

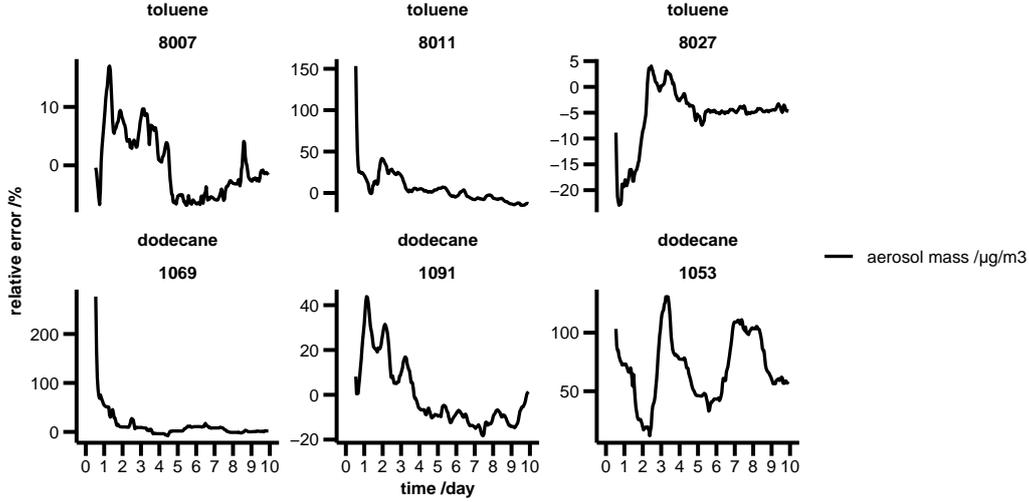


Figure 5. Time-series of the relative error on predicted aerosol mass for three toluene and three dodecane sample experiments shown on Fig. 4.

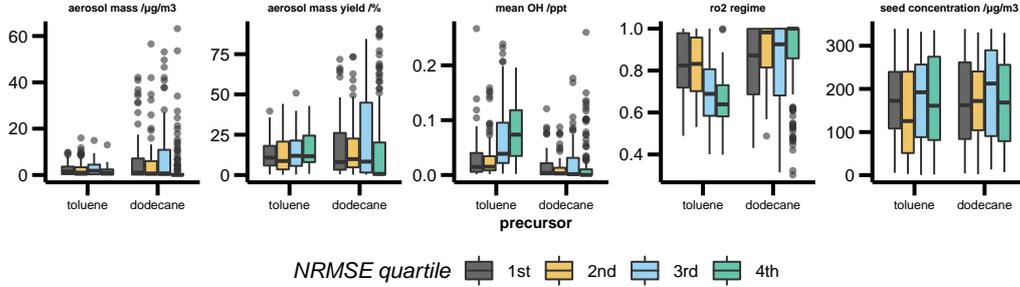


Figure 6. Boxplot distribution of simulations properties, according to their NRMSE quartile.

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3.4 Errors Distribution

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In order to identify the type of situations where the random forest is not able to reproduce the explicit model behaviour, we examine the distribution of the normalized root mean square error (NRMSE) defined below:

$$\text{NRMSE} = \frac{\sqrt{\frac{\sum_{i=1}^N (y_i - \hat{y}_i)^2}{N}}}{Q_3 - Q_1} \quad (3)$$

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y_i and \hat{y}_i are the i^{th} reference and predicted aerosol mass respectively, N is the number of time-steps and $Q_3 - Q_1$ is the difference between the first and the third quartiles. The validations random forest simulations were split in four categories, depending on their NRMSE. The distributions of environmental conditions according to this split are displayed on Fig. 6.

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First the existing seed concentration distribution does not vary with the quality of the random forest simulations. As was shown above on Fig. 2, the aerosol mass yields dependence on seed concentration is low, which explains why the random forest perfor-

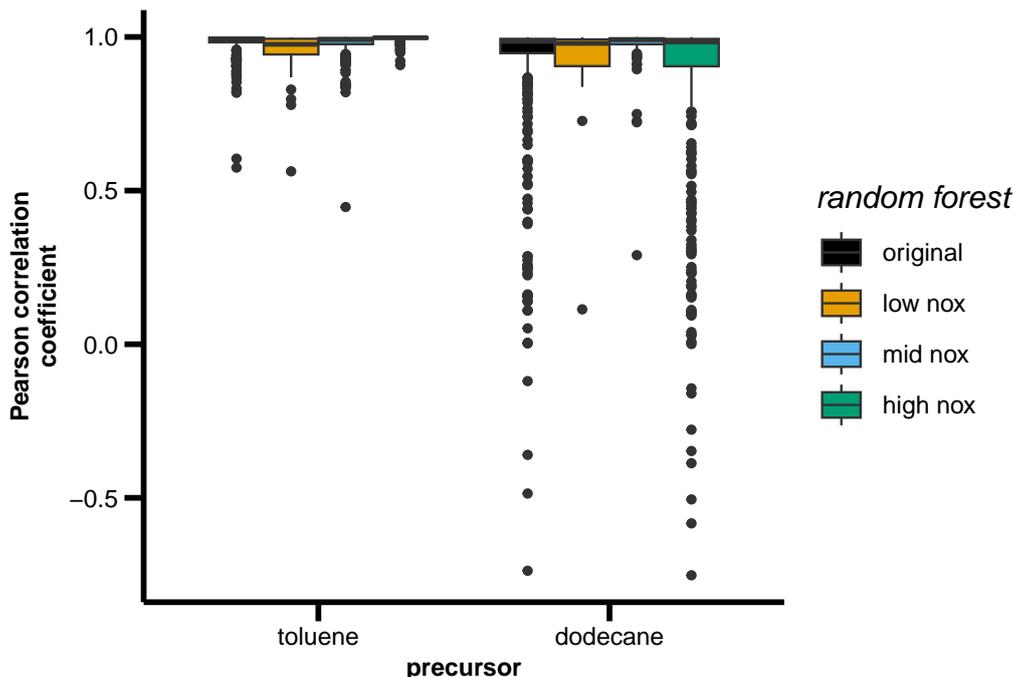


Figure 7. Boxplots of the Pearson correlation coefficients distribution for the toluene and dodecane validation simulations for the original random forests (black) and the three specialized random forests: low NO_x (orange), mid NO_x (blue) and high NO_x (green).

334 mances are not sensitive to this variable. As long as some seed aerosol is available, the
 335 system is not sensitive to its mass concentrations.

336 For toluene, the lower quality random forest simulations (third and fourth NRMSE
 337 quartiles) are typically described with lower NO_x regimes: on average $R \approx 0.82$ for the
 338 first two NRMSE quartiles, compared to $R = 0.7$ and $R = 0.65$ for the last two NRMSE
 339 quartiles. The third and fourth NRMSE quartiles simulations also exhibit higher OH mix-
 340 ing ratios: $\text{OH}_{\text{median}}^{3^{\text{rd}} \text{quartile}} = 0.04$ ppt and $\text{OH}_{\text{median}}^{4^{\text{th}} \text{quartile}} = 0.07$ ppt compared to $\text{OH}_{\text{median}}^{1^{\text{st}} \text{quartile}} =$
 341 0.01 ppt for the first quartile. The aerosol mass yields ($\approx 12 - 13$ %) are similar for
 342 all quartiles. The non-linear dependence of the SOA yield on the RO_2 regime (Fig. 2)
 343 seems to be the determining factor for toluene simulations. Under-representing lower RO_2
 344 regimes in the training dataset therefore has a strong impact on the random forest per-
 345 formances. In our case, this under-representation is the likely the consequence of the sim-
 346 ple random selection of the training simulations environmental conditions.

347 Dodecane lower quality simulations are heavily skewed towards simulations with
 348 high NO_x regimes ($R_{\text{median}}^{4^{\text{th}} \text{quartile}} = 1$ vs $R_{\text{median}}^{1^{\text{st}} \text{quartile}} = 0.87$) and lower OH mixing ratio
 349 ($\text{OH}_{\text{median}}^{4^{\text{th}} \text{quartile}} = 8.8 \times 10^{-5}$ ppt vs $\text{OH}_{\text{median}}^{1^{\text{st}} \text{quartile}} = 3.7 \times 10^{-3}$ ppt) and lower aerosol
 350 mass yields ($Y_{\text{median}}^{4^{\text{th}} \text{quartile}} = 0.68$ % vs $Y_{\text{median}}^{1^{\text{st}} \text{quartile}} = 8.1$ %). This behaviour difference
 351 between toluene and dodecane may be explained by the number of dodecane training
 352 simulations available for higher NO_x regimes, which include many outliers (1236 outliers
 353 out of 2568 point for the highest RO_2 regime bin on Fig. 2) in terms of aerosol yield. The
 354 random forest is therefore not able to reproduce the complex behaviour of dodecane SOA
 355 formation in this specific regime. In this case, the complexity of the dodecane oxidation
 356 for very high NO_x situations cannot be properly reproduced by the random forest with
 357 the given training dataset.

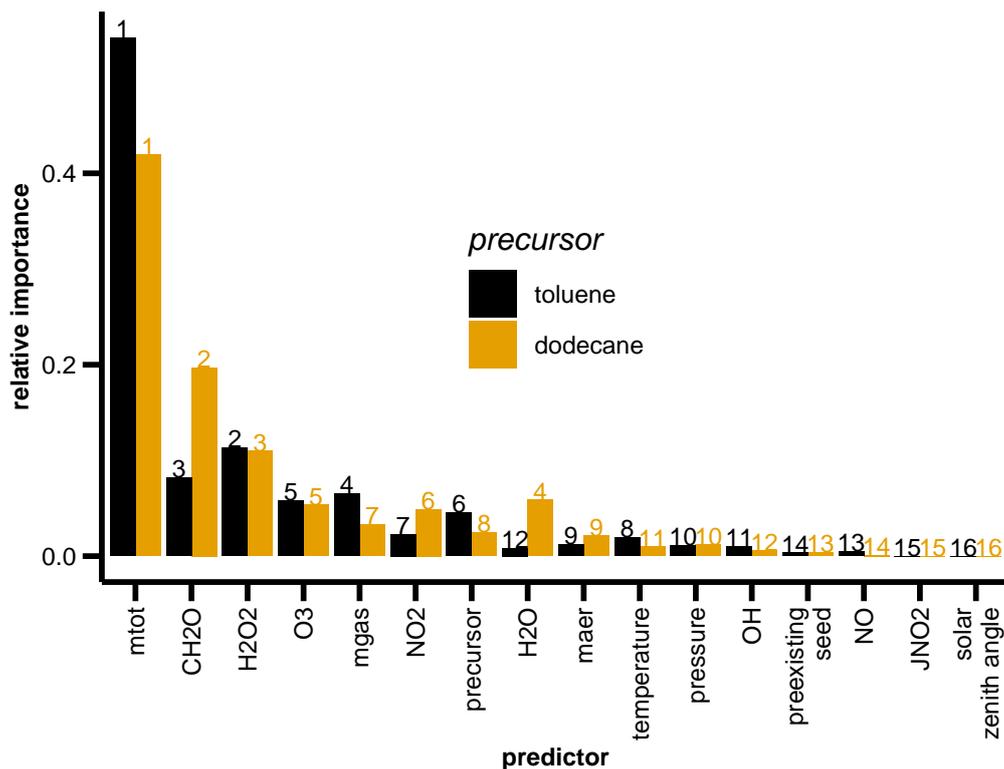


Figure 8. Predictors relative importance for the toluene (black) and dodecane (orange) random forests. The rank of each predictor is indicated at the top of the bars.

358 A possible way to improve the ability of the system to reproduce the complex relationship between R and SOA formation is to create independent random forests specialized for specific RO_2 regimes. To test this hypothesis, the training dataset was split in three separate sets according to the initial RO_2 regimes: a low NO_x set ($R < 0.3$, 100 280 training points for toluene, 72 200 points for dodecane), a mid NO_x set ($0.3 < R < 0.7$, 594 780 points for toluene, 268 640 points for dodecane) and a high NO_x set ($R > 0.7$, 363 170 points for toluene, 650 900 points for dodecane). Three specialized random forests were therefore trained on these three datasets for each precursor.

366 Figure 7 displays the distribution of Pearson correlation coefficients for the toluene and dodecane validation simulations for each specialized random forest compared to the original random forest trained with 1 000 000 points. For toluene simulations, all the specialized random forests display similar performances to the original random forest: for low NO_x , $r_{\text{median}} = 0.98$, for mid NO_x , $r_{\text{median}} = 0.99$ and for high NO_x , $r_{\text{median}} = 1.0$. Similarly for dodecane simulations, all specialized random forests exhibit performances similar to the original one, with median r values ranging from 0.98 to 0.99.

373 It is likely that the potential improvement caused by specializing the random forests over different RO_2 regimes is negatively compensated by performance reduction caused by a lower number of training points. Furthermore, in the dodecane case, specializing a random forest for high NO_x does not have a significant impact on the number of validation outliers in this regime.

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3.5 Predictors Importance

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After training the random forest, it is possible to estimate the relative importance of the chosen predictors by counting their occurrences as threshold criteria in individual nodes (Fig. 8). For both toluene and dodecane, the most important predictor for organic mass is the total organic mass in the previous time-step. This finding is consistent with the fact that total organic mass is one of the predicted outcomes by the random forests. CH_2O and H_2O_2 are the second and third (resp. third and second) most important predictors for dodecane (resp. toluene). Since higher H_2O_2 concentrations are indicative of low NO_x situations, H_2O_2 can be considered as a proxy for the RO_2 regime. CH_2O is the only predictor related to secondary organic gaseous species and could be interpreted as a proxy for organic gases formation.

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The water vapor concentration and organic gaseous fraction (m_g) are the fourth most important predictors for dodecane and toluene respectively. Since the random forests are predicting the trend of m_g , it is logical that its previous step value is a significant predictor. The importance of water vapor is likely related to its role in OH production.

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O_3 is the fifth most important predictor for both precursors. The information brought by the ozone predictor is related to general oxidants levels, the diurnal cycle as well as the RO_2 regime. NO_2 is the sixth most important predictor for dodecane while it is the precursor's concentration for toluene. These two predictors provide information about the diurnal cycle, the oxidants levels as well as the potential for secondary organic matter production. The seventh is the organic gaseous mass fraction for dodecane and NO_2 for toluene. The precursor's concentration and temperature are the eighth most important predictors for dodecane and toluene respectively. For both precursors, the organic aerosol fraction (m_a) is the ninth most important predictor. Because the gaseous mass fraction is a more important predictor for both precursors, m_a only provides complementary information. The remaining predictors only have negligible contributions to the random forests. Since we've shown that SOA formation is not sensitive to pre-existing particle seed (see Fig. 2) it is logical that this predictor is not important. Predictors directly related to the diurnal cycle (J_{NO_2} and solar zenith angle) are unimportant here, meaning that there is enough information provided by the time evolution of ozone and precursor concentrations to control for daytime vs. nighttime organic matter production. Similarly, the precursor decay as well as ozone concentrations give enough information related to oxidant concentrations and RO_2 regime, explaining the weak importance of OH and NO predictors.

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Since the contributions of the various predictors are dominated by only a few of them, we trained new random forests only using the 8 most important predictors for each precursor: m_t , CH_2O , H_2O_2 , H_2O , O_3 , NO_2 , m_g and the precursor concentration for dodecane and m_t , H_2O_2 , CH_2O , m_g , O_3 , the precursor concentration, NO_2 and temperature for toluene. Figure 9 compares the Pearson correlation coefficients (r) and NRMSE scores calculated for each validation simulation (359 for each precursor) of the random forest trained with the 16 original predictors with the random forest trained with the eight most important identified predictors (see Fig. 8).

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Reducing the number of predictors for dodecane improves r for 239 (67%) validation simulations, with an average r increase of 272%. The NRMSE decreased for 212 (59%) dodecane validation simulations, with an average NRMSE reduction of 59%. For 160 (45%) of the dodecane validation simulations, both r and NRMSE are improved. Reducing the number of predictors leads to an improvement of r for 212 (62%) toluene validation simulations, with an average r increase of 2%. The NRMSE decreased for 190 (53%) toluene validation simulations, with an average NRMSE reduction of 40%. For 148 (41%) of the toluene validation simulations, both r and NRMSE are improved.

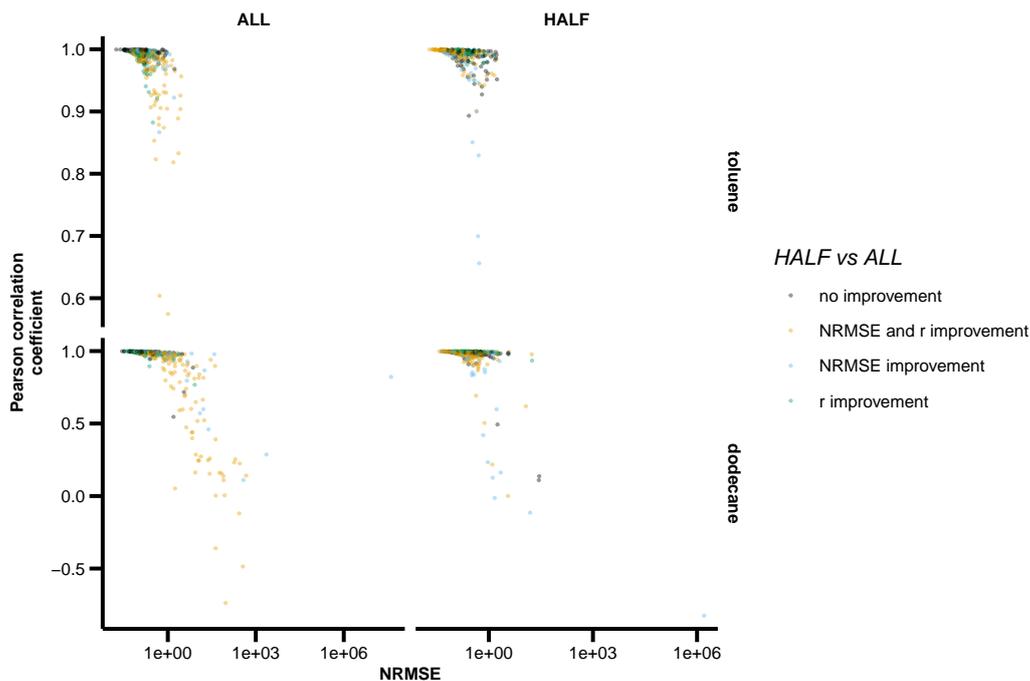


Figure 9. Scatterplot of the validation simulations' Pearson correlation coefficients (r) as a function of their NRMSE for toluene (top row) and dodecane (bottom row), for the random forests trained with the 16 originally selected predictors (ALL, left column) and the random forests trained with only the eight most important predictors (HALF, left column). The colors depict whether r , NRMSE or both scores are improved when reducing the number of predictors.

428 As shown in Fig. 9, reducing the number of predictors is beneficial for the worst
 429 performing simulations, especially for dodecane. For both toluene and dodecane, the ma-
 430 jority of validation simulations are improved when halving the number of predictors. How-
 431 ever, 7 out of the 8 selected predictors are shared by both random forests. The relative
 432 importance and ranks of the predictors differ between both random forests. There is there-
 433 fore no guarantee that different precursors have the same optimal number of predictors
 434 and that they share the same predictors.

435 4 Conclusions

436 In this work we trained two recurrent random forests to predict organic mass produc-
 437 tion in both gas and aerosol phases resulting from toluene and dodecane oxidation.
 438 The random forests were trained on a dataset created with the GECKO-A explicit or-
 439 ganic chemistry box model. The dataset contains a series of single box-model simula-
 440 tions covering a wide range of environmental conditions to ensure that the resulting ran-
 441 dom forests are able to reproduce the complex relations between organic aerosol produc-
 442 tion and the chemical environment. The resulting random forests show very good per-
 443 formances in predicting organic mass evolution in varied conditions when tested on a sim-
 444 ilar random set of box-model simulations. The distribution of errors in testing simula-
 445 tions highlights however the importance of carefully preparing the training dataset. Our
 446 results suggest that random sampling over a range of possible environmental conditions
 447 is insufficient to build a robust training dataset, and that it is more important to prop-
 448 erly sample a range of more complex chemical parameters such as the RO_2 regimes (R).

449 We have shown that the range of R that needs focus depends on the precursors. For in-
450 stance, the dodecane random forest has weaker performance for high R whereas the toluene
451 random forest has lower performance for medium R values. However, creating multiple
452 random forests each trained over a smaller range of R does not lead to more robust re-
453 sults than a single random forest. It seems more efficient to add additional training data
454 points for the poorly performing RO₂ regimes.

455 The selection of predictors is also a crucial step. We have shown that it is possi-
456 ble to increase the random forest performance by reducing the number of predictors to
457 the most important ones. However, there is no reason to think that these predictors have
458 to be the same for different precursors, highlighting the care that must be taken in their
459 selection. In this work, we selected the most important predictors by first training a ran-
460 dom forests with a wide selection of predictors, and then training a new random forest
461 with only the most important predictors identified in the first random forest.

462 In this work, we have therefore shown the feasibility of building random forests that
463 behave like a detailed chemical mechanism for predicting secondary organic mass and
464 its partitioning between gas and particle phases. There are still some limitations to over-
465 come before the implementation of the random forest SOA emulator within a chemistry-
466 climate model. First, even if the random forests are performing well, there are still some
467 critical outliers at the validation stage (*e.g.*, high NO_x dodecane). More work needs to
468 be focused at removing these outliers because falling in one of these bad cases in a 3D
469 model run would likely make the full simulation diverge. Second, the random forests were
470 each trained to reproduce the oxidation of a single precursor. Additional studies are re-
471 quired to quantify whether it is important to represent the interactions of multiple pri-
472 mary hydrocarbons, their competition for oxidants, and the impact on the resulting SOA
473 formation.

474 Open Research Section

475 The dataset that was constructed to train and test the random forests and the Python
476 code implementing the random forest training and testing have been uploaded on Zen-
477 do (Mouchel-Vallon & Hodzic, 2022).

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