

1 **Modeling Seasonal Effects of River Flow on Water Temperatures in an Agriculturally**
2 **Dominated California River**

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11 **Key Points:**

- 12 • In this snowmelt and groundwater-influenced river, cool water temperatures lasted longer
13 into summer in high-flow years than low-flow years
- 14 • Statistical water temperature model predictions became more accurate when the influence
15 of river flow was allowed to vary seasonally
- 16 • These accessible methods can be used to model any river or stream with daily long-term
17 flow and water temperature measurements
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20

21 **Abstract**

22 Low summer river flows can increase vulnerability to warming, impacting coldwater fish. Water
23 managers need tools to quantify the complex linkages between flow and water temperature, yet
24 statistical models often assume a constant relationship between these variables. In California's
25 snowmelt and groundwater-influenced Scott River where agricultural irrigation consumes most
26 summer river flow, flow variation had stronger effects on water temperature in April–July than
27 other months. Using 24 years of daily air temperature and flow data as predictors, we compared
28 multiple statistical methods for modeling daily Scott River water temperatures, including
29 generalized additive models with non-linear interactions between flow and day of the year.
30 Models with seasonally varying flow effects performed better than those assuming a constant
31 relationship between water temperature and flow. Cross-validation root mean squared errors of
32 the selected models were ≤ 1 °C. We applied the models to several instream flow scenarios
33 currently being considered by stakeholders and regulatory agencies. Relative to historic
34 conditions, the most protective flow scenario would reduce average annual maximum
35 temperature from 25.9 °C to 24.6 °C, reduce average annual degree-days exceedance of 22 °C (a
36 cumulative thermal stress metric) from 107 to 54, and delay the onset of water temperatures
37 greater than 22 °C during some drought years. Withdrawal of river water after 1 June, including
38 for groundwater management purposes, could contribute to additional exceedances of 22 °C.
39 These methods can be applied to model any stream with long-term flow and water temperature
40 measurements, with applications including scenario prediction and infilling data gaps.

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43 **Plain Language Summary**

44 Warm water threatens culturally and economically important salmon in Pacific Northwest rivers,
45 including our Scott River study area, causing chronic stress or even mortality. Climate change
46 and agricultural water use have reduced summer river flow in recent decades, intensifying water
47 scarcity. Years with deep mountain snowpack and resulting high groundwater levels extend the
48 high flow season and keep water temperatures cool through the end of July, whereas in drought
49 years the river warms sooner. We used 24 years of river flow and air temperature data to create
50 computer models that simulate water temperatures, provide a tool for assessing the effects of
51 water management. Our models allow the effect of flow on water temperatures to vary by season
52 (i.e., stronger cooling in spring and summer), improving accuracy of the simulated temperatures.
53 We used these models to simulate water temperatures under two alternative flow scenarios being
54 considered in local water management plans. Our simulations indicate that relative to current
55 conditions, the higher flow scenario would reduce the summer's hottest temperatures. Diverting
56 additional water from the river after 1 June could increase the number of days with warm river
57 temperatures that are detrimental to fish. Our model is freely available for public use.

58

59 **1 Introduction**

60 Water temperature in rivers and streams affects everything from water chemistry and physics to
61 inter-species interactions (Wenger et al., 2011), food webs (Power and Dietrich, 2002), and
62 whole-community metabolism (Bernhardt et al., 2017). Effects on individual species include

63 development (Steel et al., 2012), thermal tolerances (Dahlke et al., 2020), bioenergetics (Gibeau
64 and Palen, 2020), and behavior (Sutton and Soto, 2012).

65 The net balance of surface and streambed heat fluxes determine stream temperatures. These
66 energy fluxes include shortwave radiation (primarily from the visible light spectrum), longwave
67 radiation (i.e., heat radiated from objects including clouds, land, and vegetation), latent heat (i.e.,
68 evaporation), sensible heat (i.e., convection of heat from air to water), conduction of heat
69 between the water and stream bed, and advection (i.e., movement of water) (Caissie, 2006;
70 Moore et al., 2005a; Webb et al., 2008; Dugdale et al., 2017). Humans affect stream
71 temperatures through water diversions (Bartholow, 1991; Dymond J., 1984; Folegot et al., 2018;
72 Gibeau and Palen, 2020; Meier et al. 2003, Null et al.; 2017), discharge of industrial wastewater
73 and sewage (Erickson and Stefan, 2000), reservoir impoundments (Webb and Walling, 1993;
74 Chandesris et al., 2019), removal or enhancement of riparian vegetation (Johnson, 2004; Moore
75 et al. 2005a, Wondzell et al., 2019), and alteration of channel and floodplain morphology (Gu
76 and Li, 2002) including urbanization (Tan and Cherkauer, 2013). Stream temperatures have
77 warmed in recent decades in response to rising air temperatures resulting from anthropogenic
78 greenhouse gas emissions, a trend that is expected to continue (Isaak et al., 2017, 2018; Liu et
79 al., 2020; Wanders et al., 2019).

80 River flow rates (i.e., discharge) can affect stream temperatures. Higher flows increase a
81 stream's ability to store heat, reducing the temperature increase resulting from an equivalent
82 amount of solar radiation (Brown, 1969; Meier et al., 2003; Sinokrot and Gulliver, 2000). Higher
83 flow rates reduce daily temperature maximums and ranges (Folegot et al., 2018). Summer stream
84 temperatures are typically negatively correlated with flow (Arora et al., 2016; Isaak et al., 2017;
85 Luce et al., 2014; Mayer, 2012; McGrath et al., 2017; Moore, et al. 2005b; Neumann et al, 2003;
86 Webb et al., 2003), with flow affecting daily maximum temperatures more strongly than daily
87 mean temperatures (Asarian et al., 2020; Gu et al., 1998; Gu and Li, 2002). Stream temperature
88 model fit often increases when flow is included as a predictor (Hilderbrand et al., 2014;
89 Piotrowski and Napiorkowski, 2019; Rahmani et al., 2020; Sohrabi et al., 2017; van Vliet et al.
90 2011; Webb et al., 2003), although not always (Benyahya et al., 2008; Toffolon and Piccolroaz,
91 2015). Cooling effects of high flows are due to faster downstream transport of cold water
92 (Bartholow, 1991; Dymond J., 1984; Folegot et al., 2018), greater depth and thermal mass which
93 is more resistant to heating (Gu and Li, 2002; Meier et al., 2003; Sinokrot and Gulliver, 2000),
94 and greater accretion of cool groundwater (Kelleher et al., 2012; Mayer, 2012; Isaak et al.,
95 2017).

96 The relationship between water temperature and flow varies seasonally. The source and flow
97 paths of river water vary seasonally according to precipitation form (i.e., snow and rain)(Siegel
98 and Volk, 2019), groundwater dynamics of hillslope (Hahm et al., 2019) and alluvial (Foglia et
99 al., 2013) aquifers, and irrigation management (i.e., water withdrawals and subsequent return
100 flows back to the river via surface or groundwater)(Tolley et al., 2019). Flow effects on water
101 temperature are also seasonally mediated by variables that affect the amount of solar radiation
102 striking the water, including day length, solar angle (Piotrowski and Napiorkowski, 2019; Yard
103 et al., 2005), cloud cover (Dugdale et al., 2017), wildfire smoke (Asarian et al., 2020; David et
104 al., 2018), and leaf out and leaf fall of deciduous riparian vegetation (Dugdale et al., 2018). Some
105 of these variables follow exactly the same seasonal trajectory each year while others fluctuate
106 among years.

107 Given stream temperature's importance and vulnerability to human alterations of river flow,
108 water managers need predictive tools. Stream temperature models are often grouped into two
109 categories: process-based and statistical (Caissie, 2006). Process-based (i.e., deterministic)
110 models simulate stream energy budgets using physically based equations representing energy
111 fluxes such as shortwave radiation, longwave radiation, latent heat, sensible heat, conduction and
112 advection (Brown, 1969; Caissie, 2006; Dugdale et al., 2017). Statistical models use empirical
113 relationships between stream temperature and predictor variables, and typically require many
114 fewer variables as data inputs than process-based models do, so are often much simpler to
115 develop (Benyahya et al., 2007; Caissie, 2006; Gallice et al., 2015; Ouellet et al., 2020;
116 Piotrowski and Napiorkowski, 2019). Mohseni et al.'s (1998) non-linear regression of stream
117 temperature and air temperature has been widely replicated (Arismendi et al. 2014; Jones et al.
118 2016) and adapted (Piotrowski and Napiorkowski, 2019; Santiago et al., 2017; Segura et al.
119 2015; van Vliet et al. 2011). Recent advances in statistical models of stream temperature include
120 spatial stream network models (FitzGerald et al., 2021; Isaak et al., 2017), generalized additive
121 models (GAM) (Arora et al., 2016; Jackson et al., 2018; Laanaya et al., 2017; Siegel and Volk,
122 2019; Yang and Moyer, 2020), Least Absolute Shrinkage and Selection Operator regression (St-
123 Hilaire et al., 2018), functional data analysis (Boudreault et al., 2019), and machine learning
124 (Rahmani et al., 2020; Zhu et al., 2018, 2020). Daily stream temperatures are highly correlated
125 with adjacent days' temperatures. For measurements such as daily stream temperature that are
126 not independent, it is best to use a model that explicitly includes the correlation structure (Steel
127 et al., 2013). For example, some stream temperature models include a first-order (AR-1)
128 (Benyahya, 2007b; David et al., 2018; Letcher et al., 2016; Jackson et al., 2018; Sohrabi et al.,
129 2017), second-order, periodic (Benyahya et al., 2007a, 2007b, 2008), or moving average
130 autoregressive error structures (Yang and Moyer, 2020).

131 Process-based models account for the seasonal effects of flow by explicitly modeling energy
132 fluxes, but it is infeasible to include all these individual fluxes in statistical model. However,
133 statistical models can represent the implicit aggregation of these fluxes by allowing the
134 coefficients of hydroclimatic predictors to vary seasonally. One approach is to divide the year
135 into multiple seasons and develop separate models for each (Mohseni et al., 1998, Sohrabi et al.,
136 2017), but this may create abrupt changes at seasonal transitions. Recent approaches that allow
137 smooth variation across seasons are time-varying coefficient models (Li et al., 2014), and GAMs
138 that interact day of the year with predictor variables (Jackson et al., 2018; Siegel and Volk, 2019;
139 Yang and Moyer, 2020).

140 To test the hypothesis that statistical models with seasonally varying effects of river flow would
141 perform better than models with a constant relationship between stream temperature and flow,
142 we modeled daily stream temperatures in the Scott River of Northern California where low flows
143 and high temperatures are limiting factors for culturally and economically important coldwater
144 fish. We compared multiple statistical approaches that: 1) include all days in a single model
145 rather than dividing the year, 2) use interactions to allow the influence of predictors to vary
146 smoothly by day of the year, 3) allow non-linear relationships, 4) have error structures that
147 include temporal autocorrelation, and 5) are all implemented within the R software environment
148 with simple, publicly accessible code. After model selection and validation which confirmed our
149 hypothesis, we applied our final model to predict daily stream temperatures under flow scenarios
150 being considered by local water managers. Results indicated that stream temperatures under
151 these flow scenarios would be more favorable for coldwater fish than the historic flow scenario.
152 Our accessible modeling approach could be widely replicated in other geographic areas to

153 provide accurate stream temperature predictions to inform river management. Paired with air
154 temperatures from a nearby weather station, our methods can be applied in any river or stream
155 with long-term measurements of flow and stream temperature. Other potential applications
156 include imputing missing measurements for analyses that require continuous temperature time
157 series.

158

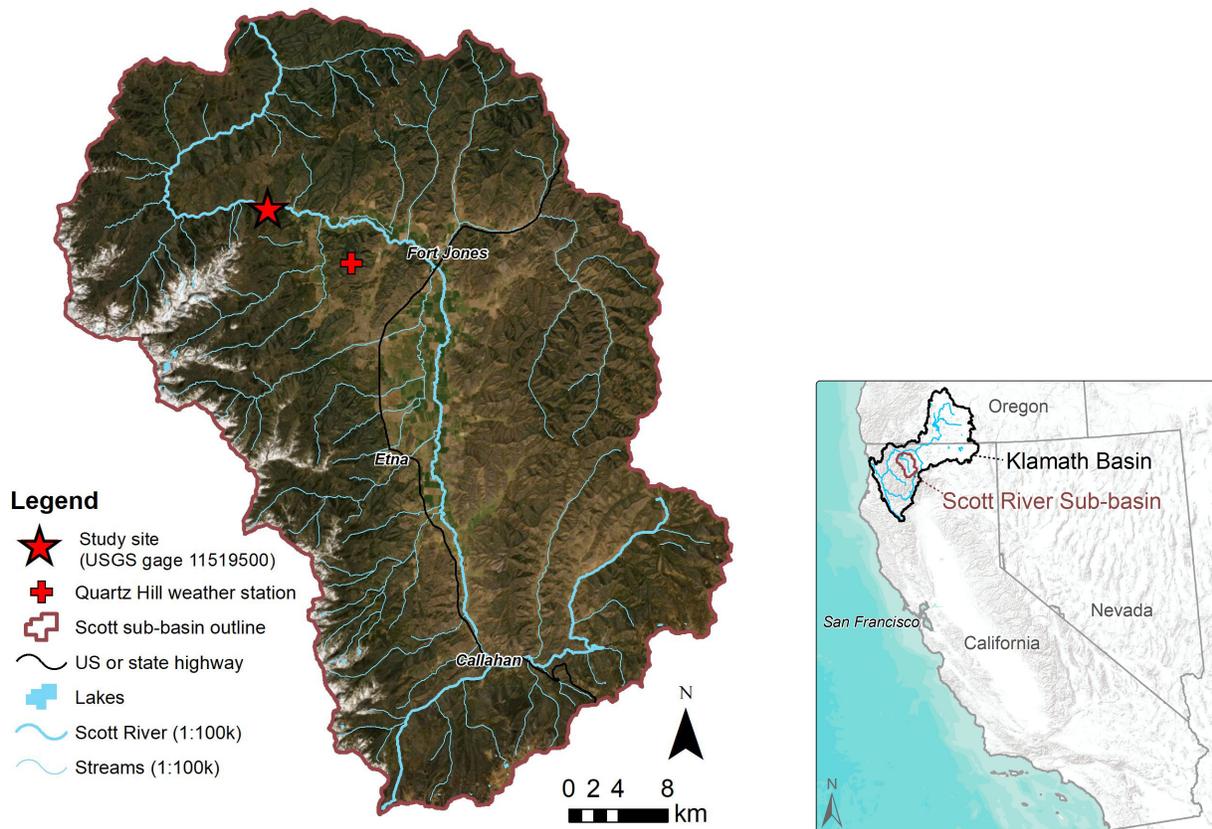
159 **2 Study Area**

160 The Scott River is a tributary of the Klamath River in Siskiyou County, California, USA (Figure
161 1). The climate is Mediterranean with precipitation occurring primarily in winter and spring as
162 rain at low elevations and snow at higher elevations. The mountainous headwaters are primarily
163 National Forest, with elevations exceeding 2500m (Foglia et al., 2013). The human population
164 lives primarily on private land in the alluvial Scott Valley, where irrigated agriculture is the
165 dominant land use, utilizing groundwater and surface water (Foglia et al., 2018). Other land uses
166 include timber harvest and mining. There are many water diversions but no major dams or
167 reservoirs. The valley aquifer fills during the high flows of winter rainstorms and spring
168 snowmelt-driven runoff. As runoff recedes through the summer, most surface water is diverted
169 for irrigation and river water at the valley outlet becomes increasingly composed of groundwater
170 from valley alluvium. Minimum flows occur in early September before rising due to fall rains
171 (Figure 2). In late summer of drought years, portions of the river have no surface flow (Tolley et
172 al. 2019). Summer and fall river flows have declined in recent decades (Kim and Jain, 2010;
173 Asarian and Walker, 2016) due to a combination of climate change (Drake et al., 2000) and
174 increased withdrawal of groundwater for irrigation, especially since 1977 (Van Kirk and Naman,
175 2008). Climate change is expected to further reduce summer flows by decreasing snowpack and
176 increasing irrigation demand (Persad et al., 2020). There are ongoing efforts to model
177 interactions between groundwater and surface water (Foglia et al., 2013, 2018; Tolley et al.,
178 2019). Pursuant to California's Sustainable Groundwater Management Act (SGMA), Siskiyou
179 County is developing a groundwater sustainability plan for the valley.

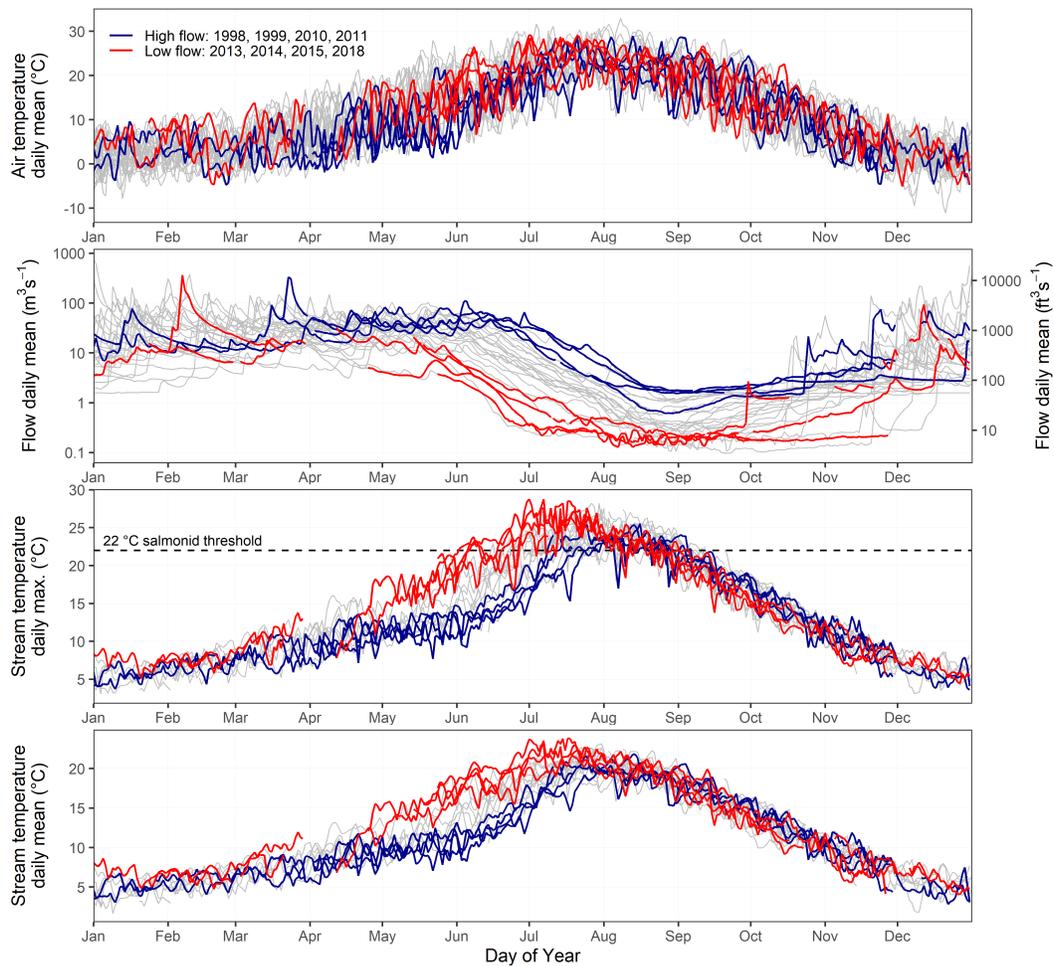
180 The Scott River has the Klamath Basin's largest population of Endangered Species Act-listed
181 coho salmon (*Oncorhynchus kisutch*) population, despite currently impaired habitat (NMFS,
182 2014). High water temperatures are stressful to coho salmon, chinook salmon (*Oncorhynchus*
183 *tshawytscha*) and steelhead (*Oncorhynchus mykiss*) (NCRWQCB, 2005). These fishes'
184 importance to local Native American tribes has led to contention over water management.
185 Government agencies, tribes, and local organizations have studied Scott River stream
186 temperatures for several decades (Asarian et al., 2020; KNF, 2010; Quigley et al., 2001; QVIR,
187 2016). The river is listed as impaired under the Clean Water Act, and California's North Coast
188 Regional Water Quality Control Board developed Total Maximum Daily Loads (TMDLs) for
189 water temperature and sediment (NCRWQCB, 2005).

190 Our study site is located at the outlet of Scott Valley, with a drainage area of 1,714 km² (Figure
191 1). Despite simulated total valley-wide streamflow depletion (i.e., decreased streamflow due to
192 groundwater pumping) of approximately 150,000 m³d⁻¹ (60 ft³/sec) in August (Foglia et al.,
193 2013), the 10 kilometers of river directly upstream of our study site are primarily a gaining reach,
194 receiving groundwater from the alluvial aquifer (Tolley et al., 2019).

195



196
197 **Figure 1.** Maps of study site and weather station within the Scott River watershed, the Klamath
198 Basin, and California. Source map credits: Esri, Earthstar Geographics, NOAA, and USGS.
199



200

201 **Figure 2.** Time series of daily mean air temperature, daily mean flow, daily maximum stream
 202 temperature (DMxST), and daily mean stream temperature (DMST) for the years 1995–2020.
 203 Colored lines are days in four example high-flow (red) and low-flow years (blue). Gray lines are
 204 other years.

205

206 3 Methods

207 3.1 Data sources and data preparation

208 3.1.1 Water temperature and river flow

209 Since 2007, the Quartz Valley Indian Reservation (QVIR) Environmental Department has been
 210 using YSI (Yellow Springs, Ohio) 6600 multi-parameter datasondes to monitor Scott River water
 211 temperatures at the U.S. Geological Survey (USGS) gage 11519500 near the outlet of Scott
 212 Valley (QVIR, 2016; Asarian et al., 2020) (Figure 1). Temperature measurements are recorded
 213 every 30 minutes with a reported accuracy of ± 0.15 °C. We combined QVIR's dataset with
 214 additional temperature data collected at the same site by the U.S. Forest Service (USFS) in the
 215 years 1995–1998, 2003–2005, 2010–2016, and 2019 (KNF, 2010), and U.S. Bureau of
 216 Reclamation (USBR) for the years 1998–2000. Following compilation, we reviewed the data and

217 removed any suspicious values (e.g., when there were calibration issues or probes appear to have
 218 been exposed to air). We then calculated daily mean stream temperature (DMST) and daily
 219 maximum stream temperature (DMxST). For days when data were available from multiple
 220 entities, we averaged values (Text S1).

221 Daily average streamflow for gage 11519500 were downloaded from the USGS National Water
 222 Information System.

223 3.1.2 Air temperature

224 Daily mean air temperature data from USFS' Quartz Hill weather station located approximately
 225 8 km southeast of the flow gage (Figure 1) are available as Global Historical Climatology
 226 Network - Daily station USR0000CQUA (Menne et al., 2012a, 2012b). We excluded all dates
 227 with a quality flag. For days lacking Quartz Hill measurements (0.5% of days with measured
 228 stream temperatures and 3.8% of the all days 1995–2020), we infilled missing values by linear
 229 regression with nearby weather stations or the gridded PRISM dataset (Daly et al., 2008) (Text
 230 S2).

231 Stream temperatures are correlated with air temperatures at multiple time scales. The optimal
 232 number of days to average for regression modeling varies (Webb et al., 2003). In addition to
 233 simple averages across varying numbers of days, other approaches include applying exponential
 234 weights (Koch and Grünwald, 2010; Piotrowski and Napiorkowski, 2019; Soto, 2016) or
 235 including separate terms for air temperatures on the day of interest and preceding days (Siegel
 236 and Volk 2019). We tested five categories of air temperatures covariates in our models, where A_i
 237 is the mean air temperature on the day i , using Equations (1), (2), (3), (4), and (5):

238 Single-day average A_1 :

$$239 \quad A_1 = A_i \quad (1)$$

240

241 Multi-day averages $A_2 \dots A_7$:

$$242 \quad A_2 = \frac{(A_i + A_{i-1})}{2}, \dots, A_7 = \frac{(A_i + A_{i-2} \dots A_{i-6})}{7} \quad (2)$$

243

244 Multi-day weighted averages A_{2w} and A_{3w} , with preceding days discounted by 50% per day:

$$245 \quad A_{2w} = A_i + \frac{(0.5 \cdot A_{i-1})}{1.5} \quad \text{and} \quad A_{3w} = \frac{(A_i + 0.5A_{i-1} + 0.25A_{i-2})}{1.75} \quad (3)$$

246

247 Lagged averages A_{L3} and A_{L5} :

$$248 \quad A_{L3} = \frac{(A_{i-1} + A_{i-2} + A_{i-3})}{3} \quad \text{and} \quad A_{L5} = \frac{(A_{i-1} + A_{i-2} + A_{i-3} + A_{i-4} + A_{i-5})}{5} \quad (4)$$

249

250 Differences between lagged average and day i :

$$251 \quad A_{\Delta 3} = (A_i - A_{L3}) \quad \text{and} \quad A_{\Delta 5} = (A_i - A_{L5}) \quad (5)$$

252 To improve numerical stability, we standardized each air temperature and flow predictor variable
253 by centering (i.e., subtracting the mean) and scaling (i.e., dividing by the standard deviation).

254 3.2 Model development and calibration

255 We developed statistical models to predict DMxST and DMST using river flow and air
256 temperature as predictors (Table 1). We tested three classes of models: non-linear logistic
257 regression, harmonic regression, and generalized additive models (GAM). Models were
258 developed in R version 4.02 (R Core Team 2020).

259

260

261 3.2.1 Generalized additive models (GAMs)

262 We focused our stream temperature modeling on GAMs because they offer powerful flexibility
263 including non-linear smoothers (Pedersen et al., 2019; van Rij et al., 2019). We used the bam
264 function in the mgcv R package version 1.8-31 (Wood, 2017) to develop GAM models, fit using
265 fast restricted maximum likelihood (fREML). GAM model terms can be either linear coefficients
266 or non-linear functions (Wood, 2017; Pedersen et al., 2019). The non-linear functions are smooth
267 curves with the amount of wiggleness automatically determined by a smoothing penalty. We used
268 cyclic cubic regression splines (“cc”) as the smoother for day of the year D and thin plate
269 regression splines (“tp”) as smoothers for other covariates. To avoid overfitting (Siegel and
270 Volk, 2019) we limited smoothers for most variables to a maximum of 3 knots, except D which
271 was allowed up to 5 knots.

272 We compared GAMs that allow the relationships between covariates and the response variable to
273 vary seasonally to GAMs where those relationships are seasonally constant. Our GAM models
274 represented interactions between variables as either partially non-linear or fully non-linear. For a
275 partially non-linear interaction, the linear slope of one variable (e.g., flow) changed as a smooth
276 non-linear function of another variable (i.e., D), an approach used by Jackson et al. (2018) and
277 Siegel and Volk (2019) and specified in mgcv using the “by” option. Fully non-linear
278 relationships between two or more variables were specified as tensor product smooths in mgcv
279 using the syntax “te()” (Wood, 2017). If main effects were included as separate terms, then we
280 used “ti()” to specify a tensor product interaction (Wood, 2017).

281 All our GAM models included a random effect for year and all but one (“GAM11,” Section 4.2)
282 included an AR-1 autocorrelation error structure. The bam function cannot automatically derive
283 the AR-1 coefficient (ρ), so it must be manually assigned. Following Baayen et al. (2018) and
284 van Rij et al. (2019, 2020), we initially fit each model without an autocorrelation term, and then
285 re-ran the model with an autocorrelation term, assigning a ρ value based on the lag 1
286 autocorrelation from the residuals of the initial model. Baayen et al. (2018) and van Rij et al.
287 (2019) advise testing several ρ values using model-comparison procedures, which in our case
288 always confirmed the initial values were optimal.

289

290 3.2.2 Harmonic regression

291 As an alternative to compare to GAMs, we use harmonic regression (also known as
292 trigonometric or periodic regression) (Cox, 2006) with paired sine and cosine interaction terms
293 that allow the slope of covariates to vary as a smooth cycle over the course of the year (Bodeker
294 et al., 1998). For daily periodicity, we multiplied day of the year D by $2\pi/365$ (Helsel et al.,
295 2020). We developed these models using the `lme` function in the `nlme` R package version 3.1-148
296 (Pinheiro et al., 2020) with an AR-1 autocorrelation term and a random intercept for year, fit
297 using maximum likelihood (ML). Harmonic regression of stream temperature is common
298 (Kothandaraman, 1971; Johnson et al., 2020), but we are not aware of previous applications of
299 harmonic interactions between D and other covariates for stream temperatures.

300

301 3.2.3 Non-linear logistic regression

302 Since Mohseni et al.'s (1998) non-linear logistic regression of air temperature and stream
303 temperature has been so widely applied, we use it as a benchmark to compare our other model to.
304 Many streams, including the Scott River (Manhard et al., 2018), exhibit hysteresis in which the
305 relationship between stream temperature and air temperature differs between spring and fall
306 (Mohseni et al., 1998). Following Jones et al.'s (2016) code using R's `optim` function, we
307 modeled the ascending (weeks 1–30) and descending (weeks 31–52) limbs separately, fitting
308 models using weekly averages and then apply them to daily data. These models do not include
309 flow, autocorrelation, or random effects. We used 7-day average air temperatures to match the
310 original method.

311

312 3.3 Model selection and validation

313 We compared alternative model configurations (which variables and interactions are included,
314 which are assigned random effects, etc.) to select a final model (Table 1). Initial exploration
315 indicated that A_{2w} (2-day weighted air temperature) provided better model fits than other air
316 temperature variables, so we used A_{2w} for most of our models. After final model selection, we
317 developed a separate set of models to assess the sensitivity of model fits to using different air
318 temperature variables (Figure S1). Rather than slavishly follow a pre-specified procedure such as
319 forward-selection or backward selection, we took a more holistic approach to model selection.
320 We selected a final model after considering multiple models using a variety of methods including
321 Akaike information criterion (AIC), `fREML` (fast restricted maximum likelihood) scores for
322 GAMs, goodness of fit metrics (root mean squared error [RMSE] and coefficient of determination
323 [R^2]), and review of residual plots and auto correlation function plots. Concurvity, the non-linear
324 equivalent of collinearity, is a potential concern for GAMs such as ours that contain smooths for
325 time along with other time-varying covariates (Amodio et al., 2014; Wood, 2017), so we
326 evaluated each GAM's concurvity using `mgcv`'s concurvity function.

327 Prior to modeling, we randomly selected and excluded all data from 4 (17%) of the 24 years.
328 These data were not used in model selection but instead were retained for out-of-sample
329 validation.

330 We validated models using two methods. First, we used leave-one-year-out (LOYO) cross-
331 validation, a version of k -fold variation in which we withhold a year, re-fit the model using the

332 19 remaining years, compare predictions for the withheld year against the measured data using
333 goodness of fit metrics (RMSE and R^2), and then repeat the same process for each year. Second,
334 for out-of-sample validation, we compared model predictions (calibrated with 20 years of data)
335 against data from the four removed years using goodness of fit metrics (RMSE and R^2).

336

337 3.5 Model application to hydroclimatic and flow management scenarios

338 To assess the seasonal response of stream temperatures to variation in flow and air temperatures,
339 we applied our selected GAM models to a group of 15 “quantile air temperature” scenarios
340 representing combinations of 3 air temperature inputs and 5 flow inputs (Table 2, Figure 3). All
341 three air temperature inputs were derived using non-parametric quantile regression (Cade and
342 Noon, 2003; Muggeo et al., 2013) to calculate the air temperature associated with three quantiles
343 (0.05, 0.50, and 0.95, equivalent to 5%, 50%, 95% exceedance probabilities) for each day of the
344 year (Figure 3a), using the `quantregGrowth` R package (Muggeo et al., 2013), with options
345 described in Text S3. For air temperature, the 0.50 quantile represented typical conditions, the
346 0.05 quantile represented hottest conditions, and the 0.95 quantile represented coolest conditions.
347 Three of the five flow inputs were based on quantiles (0.05, 0.50, and 0.95) derived using similar
348 methods as the air temperature inputs, with the 0.50 quantile representing typical conditions, the
349 0.05 quantile representing very low flow conditions, and the 0.95 quantile representing high flow
350 conditions (Figure 3b). The remaining two of the five flow inputs are based on the USFS water
351 right and California Department of Fish and Wildlife (CDFW) Interim Instream Flow Criteria.
352 The USFS first-priority Scheduled D water right varies by month and day, from a high of 200
353 ft^3/sec ($5.67 \text{ m}^3/\text{sec}$) in November through March to a low of 30 ft^3/sec ($0.85 \text{ m}^3/\text{sec}$) in August
354 and September (Superior Court of Siskiyou County, 1980) (Figure 3b). The CDFW criteria vary
355 by month and day, from a minimum of 62 ft^3/sec ($1.75 \text{ m}^3/\text{sec}$) in September to a high of 362
356 ft^3/sec ($10.3 \text{ m}^3/\text{sec}$) in February (CDFW, 2017) (Figure 3b). The CDFW and USFS flows do not
357 follow a particular flow quantile through the entire year, but instead are extreme drought
358 conditions in May (0.05 quantile) and high flows in August and September (0.50 to 0.95
359 quantile).

360 To assess the realistic timing and magnitude of modeled exceedances of stream temperature
361 thresholds, we also applied our selected GAM model to predict stream temperatures in a group of
362 “observed air temperature” scenarios that pair the observed daily air temperature time series for
363 1995-2020 with eight flow conditions: observed USGS flows in addition to the five flows used in
364 the “quantile air temperature” scenarios (low, typical, high, USFS, and CDFW) as well as two
365 additional scenarios in which the CDFW and USFS flows are used as minimums that are
366 supplanted by observed USGS flows on dates when the observed flows are higher (Table 2). We
367 expect that using observed air temperatures instead of quantile air temperatures provides more
368 realistic real-world predictions because air temperatures fluctuate erratically from day to day
369 (Figure 2), instead of staying near the same quantile like flow does during the seasonal flow
370 recession each year from May through September.

371 We summarized the results of each “observed air temperature” scenario by calculating: 1) annual
372 maximum temperature, 2) first and last day each year in which water temperatures exceed $22 \text{ }^\circ\text{C}$,
373 and 3) the annual degree days exceedance of $22 \text{ }^\circ\text{C}$, calculated by subtracting 22 from all
374 DMxST and summing all positive values by year. We chose $22 \text{ }^\circ\text{C}$ as an indicator of biological
375 effects on juvenile salmonids that rear in the mainstem Scott River or outmigrate downstream

376 using the river as a migratory corridor. Given the potential for local genetic adaptation to thermal
377 regimes (Zillig et al., 2021), we prioritized studies near the Scott River in selecting thresholds.
378 When the Klamath River exceeds 22–23 °C, juvenile salmonids move to tributary confluences
379 (Sutton et al., 2007; Sutton and Soto, 2012; Brewitt and Danner, 2014). Similar behavior was
380 observed in the Shasta River (Nichols et al., 2014) and 22 °C was also used by McGrath et al.
381 (2017). The 22 °C threshold is not fully protective for coho salmon (Text S4) but we chose it
382 because our study site is a mainstem river where temperatures are expected to be higher than a
383 cool tributary.

384

385

386 **Table 1.** Comparison of model training statistics.
387

Model Name	Predictor variables	Daily maximum stream temperature (DMxST)						Daily mean stream temperature (DMST)					
		fREML	AIC	AR1	edf	RMSE	R ²	fREML	AIC	AR1	edf	RMSE	R ²
<i>GAM1: tensor Q-A_{2w}-D</i>	te(Q, A _{2w} , D)	5004	9901	0.587	46.5	0.86	0.982	3354	6607	0.747	46.5	0.77	0.979
GAM2: tensors Q-D & A _{2w} -D	s(A _{2w}) + ti(A _{2w} , D) + te(Q, D)	5036	9973	0.603	36.4	0.88	0.981	3364	6639	0.769	35.6	0.80	0.978
GAM3: tensor Q-D & vary A _{2w}	s(D, by = A _{2w}) + te(Q, D)	5039	9978	0.580	39.3	0.86	0.982	3373	6649	0.742	39.1	0.75	0.980
<i>GAM4: tensors Q-D & A_{2w}-Q (final)</i>	s(D, by = A _{2w}) + ti(A _{2w} , Q) + te(Q, D)	5053	10022	0.603	36.3	0.89	0.981	3401	6729	0.763	35.4	0.80	0.978
GAM5: tensors Q-D & A _{2w} -Qv2	s(A _{2w}) + ti(A _{2w} , Q) + te(Q, D)	5095	10116	0.608	34.1	0.90	0.98	3452	6840	0.764	33.2	0.82	0.977
GAM6: tensor Q-D no vary A _{2w}	s(A _{2w}) + te(Q, D)	5105	10139	0.611	30.4	0.90	0.98	3466	6873	0.77	28.6	0.83	0.976
<i>GAM7: varying Q & A_{2w}</i>	s(D, by = A _{2w}) + s(D, by = Q) + s(D)	5160	10254	0.659	28.2	0.96	0.978	3464	6871	0.804	27.5	0.88	0.973
<i>GAM8: A_{2w} no varying</i>	s(A _{2w}) + s(Q) + s(D)	5448	10855	0.773	23.1	1.35	0.956	3626	7208	0.834	23.8	1.08	0.96
<i>GAM9: A_{2w} no Q or varying</i>	s(A _{2w}) + s(D)	5525	11010	0.846	22.3	1.70	0.931	3749	7459	0.875	21.6	1.30	0.941
GAM10: A ₇ only with AR1	s(A ₇)	6606	13186	0.886	21.3	2.75	0.817	5044	10062	0.905	21.2	2.29	0.819
<i>GAM11: A₇ only no AR1</i>	s(A ₇)	10441	20823	N/A	23.2	2.21	0.882	9330	18602	N/A	23.1	1.74	0.895
<i>Harmonic12: varying Q & A_{2w}</i>	A _{2w} + A _{2w} :sin(Dn) + A _{2w} :cos(Dn) + Q + Q:sin(Dn) + Q:cos(Dn) + cos(Dn) + sin(Dn)	N/A	10368	0.73	N/A	1.04	0.969	N/A	6810	0.859	N/A	0.94	0.964
<i>Logistic13: Mohseni</i>	Logistic regression with A ₇	N/A	N/A	N/A	N/A	2.34	0.868	N/A	N/A	N/A	N/A	1.84	0.883

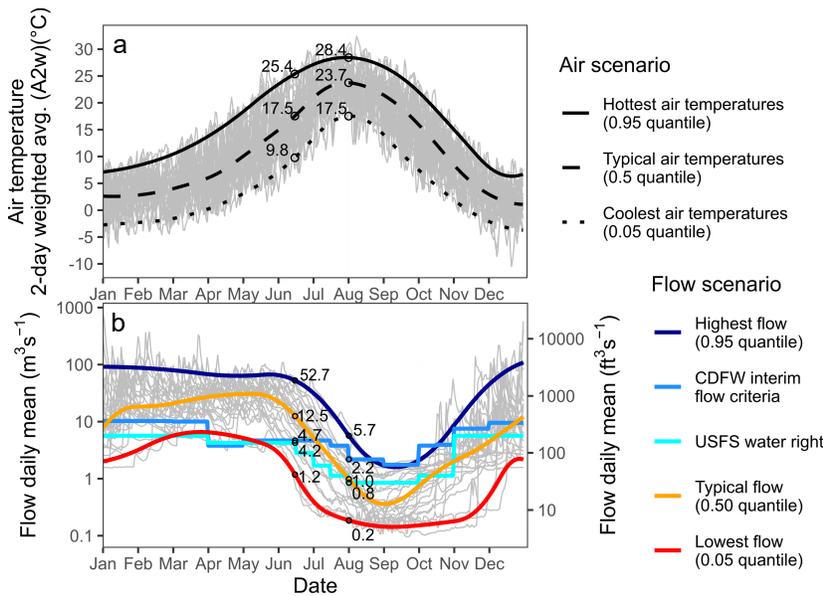
388 *Note:* GAM models are sorted by fREML score for DMxST. Except ‘GAM11 A₇ only no AR1’, all GAM and LMM models also
389 include an AR1 autocorrelation structure and a random effect of year. For models with italicized names, validation statistics are
390 provided in Figures 4 (DMxST) and S5 (DMST). D = day of year from 1 (1 January) to 366 (31 December in leap year), Q = daily
391 mean flow in units of m³/s, see Section 3.1.12 for key to ‘A’ air temperature variables, ‘s()’ is a non-linear function, ‘s(D, by =)’ is a
392 linear interaction that varies smoothly by D, ‘te()’ is a fully non-linear tensor product smooth of two or three variables, ‘ti()’ is a tensor
393 product interaction, ‘:’ is linear interaction, n = 2π/365, fREML = fast restricted maximum likelihood score, AIC = Akaike
394 information criterion, AR1 = autocorrelation coefficient, edf = effective degrees of freedom, RMSE = root mean squared error, and R²
395 = coefficient of determination.

396 **Table 2.** Matrix showing the 23 stream temperature model scenarios representing combinations
 397 of air temperature and flow inputs, and organized into two scenario groups. The 15 scenarios in
 398 Group 1 use “quantile air temperature” inputs and 8 scenarios in Group 2 use “observed air
 399 temperature” inputs.
 400

Air temperature inputs	Flow inputs							Observed
	Lowest (0.05 quantile)	Typical (0.50 quantile)	Highest (0.95 quantile)	USFS exact	CDFW exact	USFS as minimum	CDFW as minimum	
Hottest (0.95 quantile)	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1
Typical (0.50 quantile)	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1
Cooltest (0.05 quantile)	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1	Group 1
Observed	Group 2	Group 2	Group 2	Group 2	Group 2	Group 2	Group 2	Group 2

401 *Note:* USFS = USFS Schedule D first-priority water right (Superior Court of Siskiyou County,
 402 1980), and CDFW = CDFW Interim Instream Flow Criteria (CDFW, 2017). See text for
 403 explanation of quantiles and flow minimums.
 404
 405

406



407

408 **Figure 3.** Inputs to Group 1 scenarios representing 15 combinations of (a) three air temperature
 409 inputs and (b) five flows inputs that vary by day. Observed values for 1995–2020 are shown as
 410 gray lines in both panels. Data values are labeled for 15 June and 1 August.

411

412 4 Results

413 4.1 Measured water temperature, air temperature, and flow

414 From May–July, measured water temperatures were highly variable among years (Figure 2). For
415 those months, the highest-flow years had DMxST averaging 6.8 °C cooler than during lowest-
416 flow years, while DMST averaged 5.3 °C cooler. In contrast, from August through October inter-
417 annual differences in water temperature much less pronounced. Annual maximum water
418 temperatures occurred earlier in the season in low-flow years (i.e., early/mid-July) than in high-
419 flow years (i.e., late July or early August). These observations inspired us to develop seasonally
420 varying models.

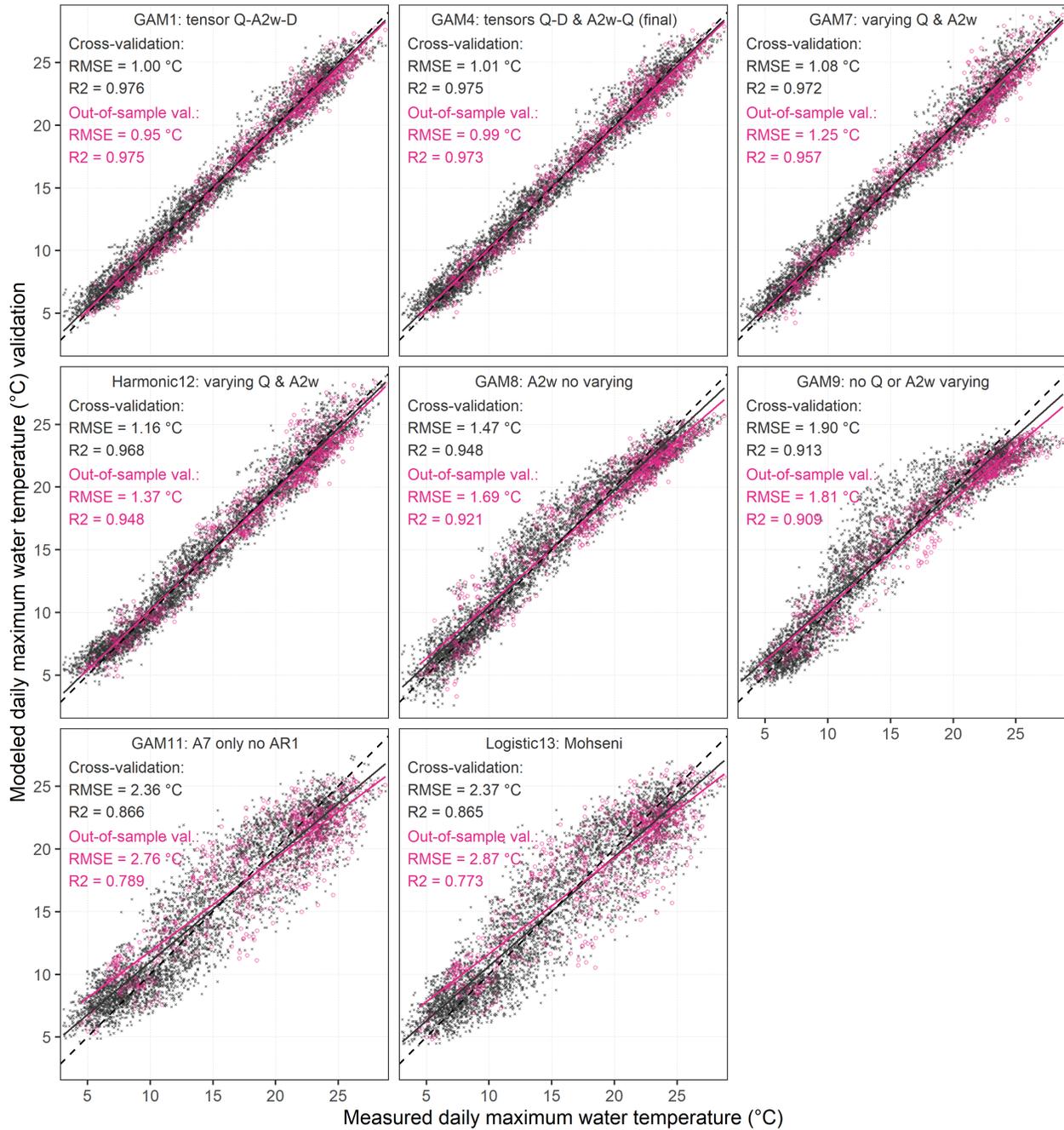
421

422 4.2 Model selection and validation

423 The sensitivity analysis of model training statistics for models using various air temperature
424 metrics indicated similar performance for most of the metrics, except the longest multi-day air
425 temperature averages which had higher RMSE (Figure S1). For DMxST, RMSE ranged from
426 0.88–0.90 °C for all air temperature metrics except the 3-day to 7-day averages which were
427 0.96–1.15 °C (Figure S1). For DMST, RMSE ranged from 0.79–0.82 °C for all air temperature
428 metrics except the 4-day to 7-day averages which were 0.85–0.98 °C and the single-day average
429 (0.87 °C) (Figure S1). Given the excellent performance of the 2-day weighted air temperature
430 (A_{2w}) in predicting both DMxST and DMST (Figure S1), we use A_{2w} for all models except
431 Logistic13 and the two GAM models that mimic it (Table 1).

432 Validation and training statistics indicate a wide range of performance (Table 1, Figure 4), with
433 the tensor models (i.e., GAM1, GAM2, GAM3, GAM4, GAM5, GAM6) performing best while
434 those models that used only air temperature (e.g., Logistic13 and its GAM equivalent GAM11)
435 performed the worst.

436 GAM4, chosen as our selected model for reasons discussed in Section 5.1, had a cross-validated
437 RMSE of 1.01 °C for DMxST (Figure 4) and 0.93 °C for DMST (Figure S5), with similar values
438 for out-of-sample validation. Similar to the measured data (Figure 2), in the May–July period the
439 selected model predicts cool water temperatures during high-flow years and warm water
440 temperatures during low-flow years (Figure S2). The effects plot for the selected models show
441 that stream temperatures are relatively insensitive to flow from 1 December to 1 March, but that
442 flow exerts a strong cooling influence from 1 April to 1 August (Figure 5, Figure S7). The
443 complete time series of measured and modeled water temperature data for all years is available
444 as Figure S3 and S4 for DMxST and DMST, respectively.



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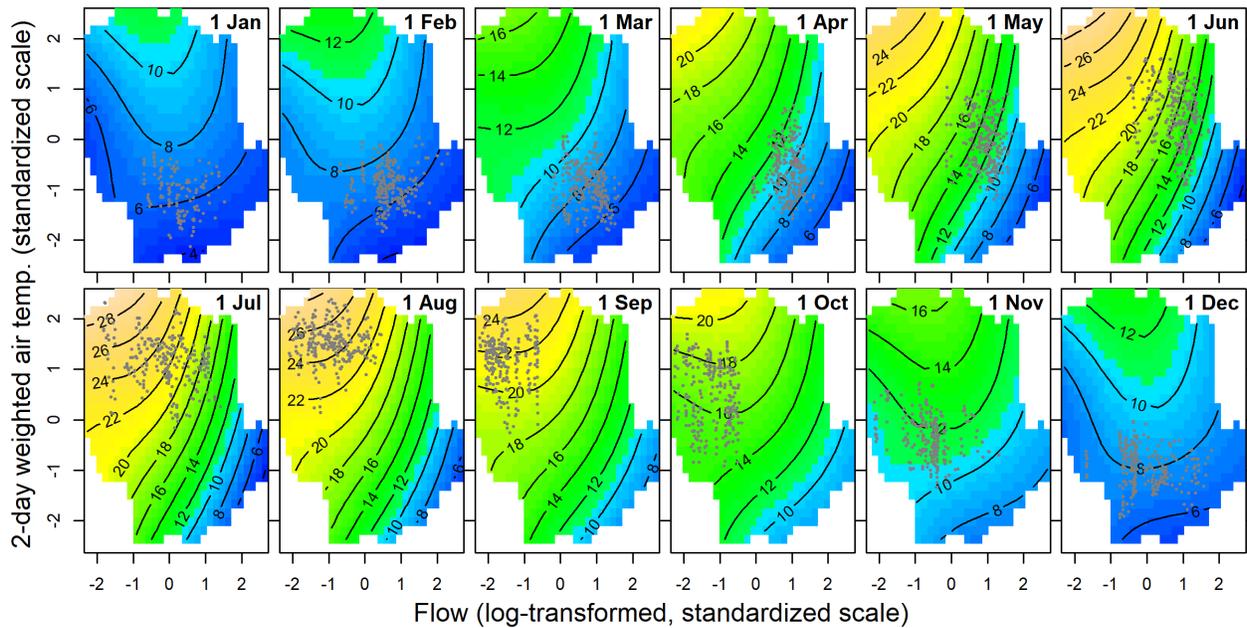
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Figure 4. Comparison of measured DMxST to LOYO cross-validation predictions and out of sample validation predictions for 1995–2020. Solid lines are linear regression and dotted lines are the 1:1 (Y=X) lines.

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453 **Figure 5.** Effects plot showing predictions from selected model “GAM4: tensors Q-D & A2w-
 454 Q” that uses 2-day weighted air temperature (A_{2w}), flow (Q), and day of year (D) as predictors.
 455 Colors show predicted DMxST as function of Q and A_{2w} , with DMxST labeled contour lines
 456 spaced 2 °C apart. Panels represent the first day of each month. Gray dots show position of
 457 calibration points within 5 days of first of each month.

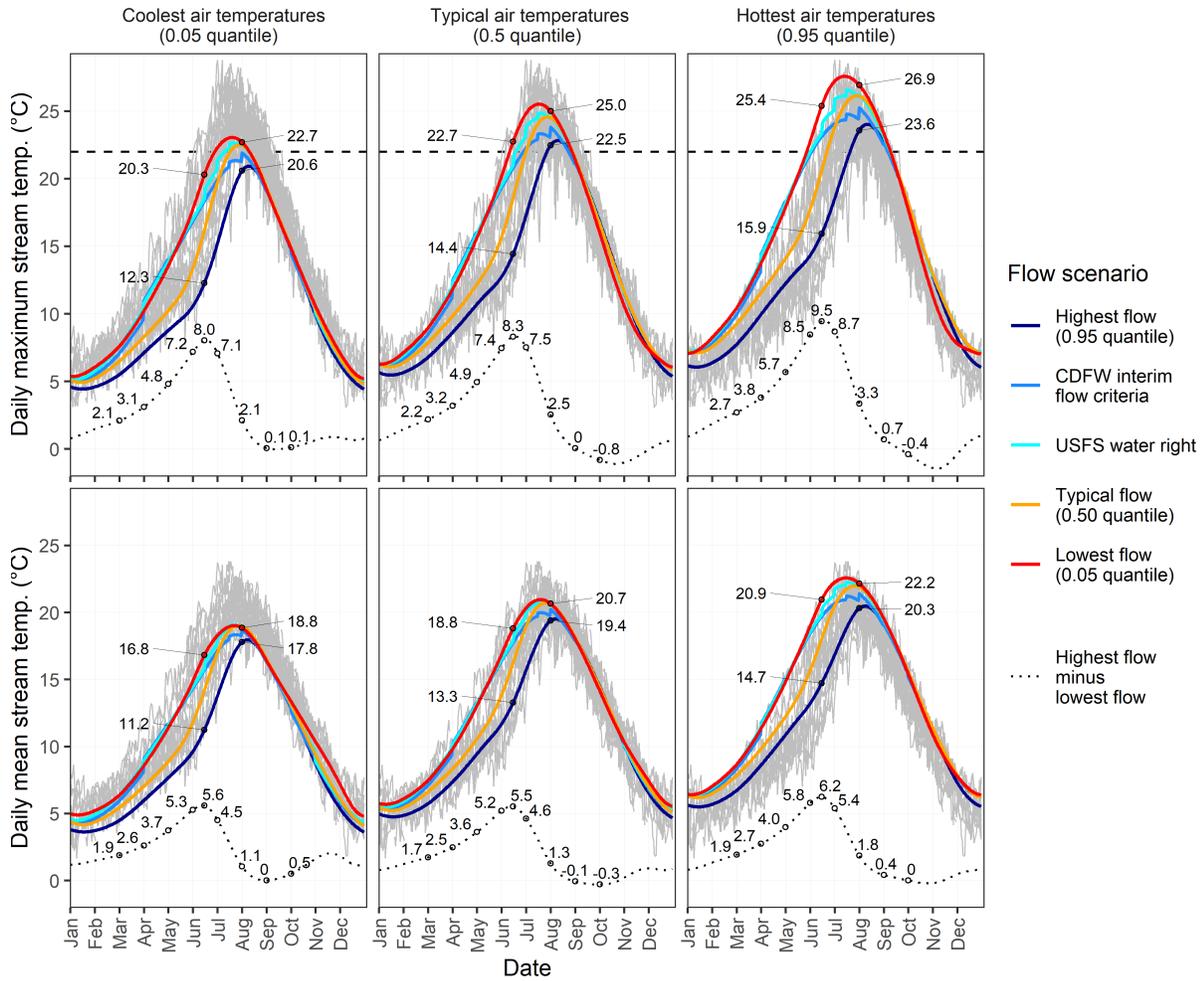
458

459

460 4.3 Model application to hydroclimatic and flow management scenarios

461 The “quantile air temperature” model scenarios show that flow and air temperature both had
462 strong effects on water temperature (Figure 6). The cooling effect of high flow followed a
463 seasonal pattern, rising in March to reach a peak on 15 June (up to 9.5 °C for DMxST and 6.2 °C
464 for DMST), then diminishing to near zero by early September (Figure 6). Cooling effects of high
465 flows were stronger when air temperatures were high than when air temperatures were low (e.g.,
466 15 June difference in DMxST between highest-flow and lowest-flow scenarios is 9.5 °C with the
467 hottest air temperatures and 8.0 °C with the coolest air temperatures). With less solar radiation
468 (due to shorter days and lower solar angle) and lower air temperatures than earlier months,
469 DMxST is almost always less than 22 °C by early September regardless of flow (gray lines in top
470 panels of Figure 6). Consistent with the measured data (Figure 2), modeled annual maximum
471 water temperatures occurred later in the season in high-flow years (i.e., late July or early August)
472 than in low-flow years (i.e., early/mid-July) (Figure 6).

473 In the “observed air temperature” scenarios, we modeled DMxST pairing the observed air
474 temperature time series for 1995–2020 with eight flow scenarios (Table 2, Figures 7 and S8).
475 These scenarios provide an indication of the range (e.g., due to air temperatures) in daily water
476 temperature associated with each flow scenario. Compared to the lowest flow scenario (0.05
477 quantile), the highest flow scenario (0.95 quantile) has annual maximum temperatures that are
478 3.7 °C cooler (Figure 7a) and temperatures first reach 22 °C 51 days later (Figure 7c); in
479 contrast, there is only a 2-day difference in the last day of the year that has temperatures >22 °C
480 (Figure 7d). The scenario with observed flows has the most interannual variation in the annual
481 maximum temperature (Figure 7a) and timing of exceedances of 22 °C (Figure 7c,d), because it
482 includes very low flows as well as very high flows. Water temperatures reach 22 °C 17 days
483 earlier with the exact USFS flows than with observed flows (Figure 7c) because the USFS flows
484 are much lower than average observed flows in May and June. In contrast, in the scenario in
485 which USFS flows are treated as minimums (supplanted by observed flows on days when
486 observed flows are higher), temperatures reach 22 °C on the same day as the observed flow
487 scenario (Figure 7d). Due to high July and August flows in the CDFW scenarios, annual
488 maximum water temperatures are 1.1–1.3 °C cooler in the CDFW scenarios than the observed
489 flow scenario (Figure 7a). Relative to the observed flow scenario, the date that the CDFW as
490 minimum scenario first reaches 22 °C is only 3 days later average (2 July vs. 30 June), but the
491 number of years with exceedances prior to June 20 are reduced from 6 to 2 (Figure 7c) because
492 the CDFW flows are higher in observed flows in drought years. Patterns of inter-scenario
493 differences in annual degree-days exceedance of 22 °C (Figure 7b) are very similar to those of
494 annual maximum temperature (Figure 7a). While the CDFW flows and USFS flows are both
495 predicted to improve (i.e., cool) summer temperatures relative to current conditions, these
496 improvements would be greater with the higher CDFW flows.



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Figure 6. Predicted maximum and mean water temperatures under the 15 “quantile air temperature” scenarios representing combinations of 3 air temperature inputs (arranged in columns) and flow inputs (shown by color). Observed values for 1995–2020 are shown as gray lines. Selected data values are labeled on 15 June and the first day of the months March–October. Horizontal line at 22 °C DMxST is salmonid temperature threshold.

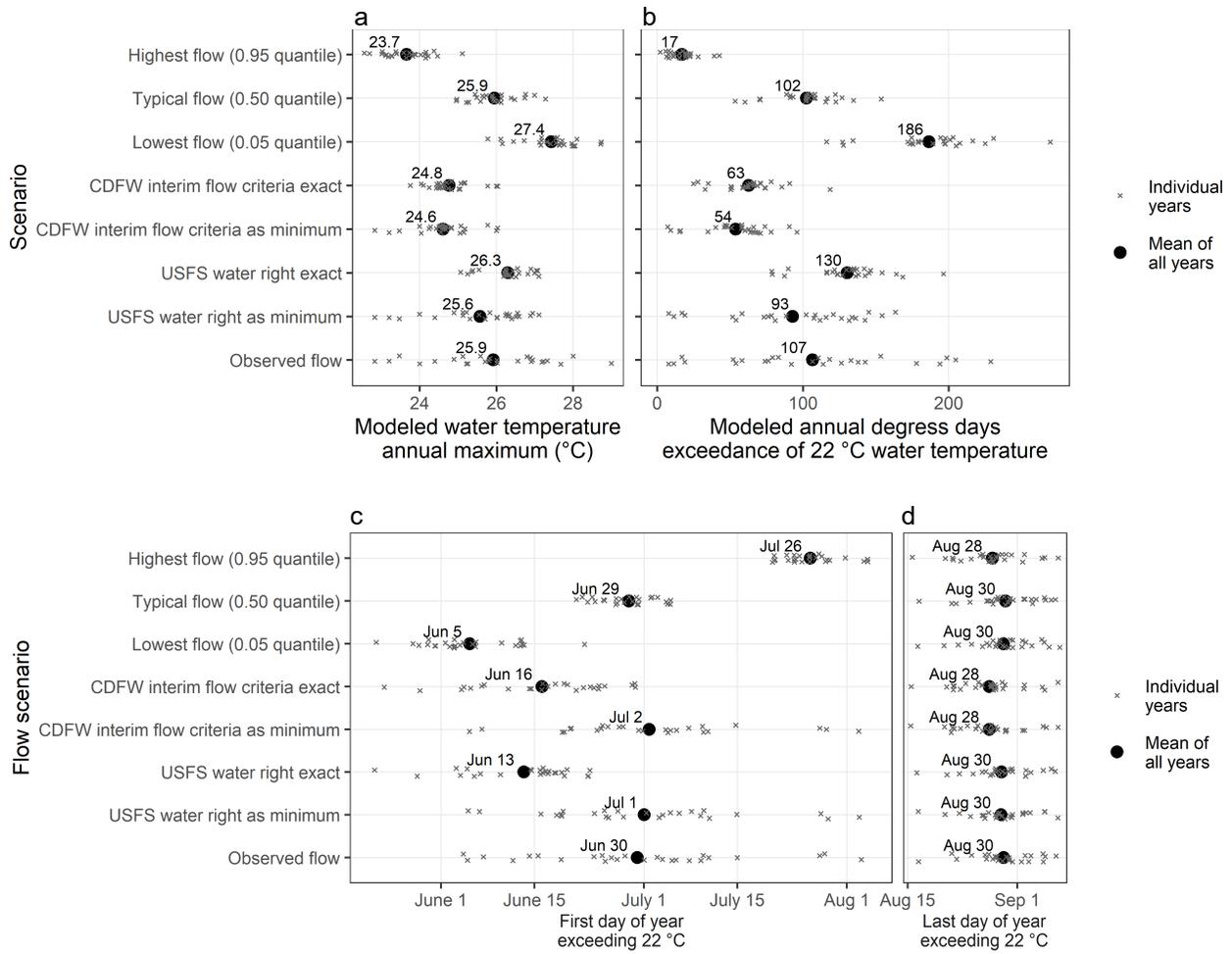


Figure 7. (a) Annual maximum stream temperature, (b) annual degree-days exceeding 22 °C, and (c) first day and (d) last day each year when DMxST exceeds 22 °C predicted using a statistical model pairing observed air temperatures for 1995–2020 with the same eight flow conditions shown in Figure S8. Points for individual years are offset slightly for clarity. Data labels are the mean of all years.

5 Discussion

Consistent with our hypothesis, models with seasonally varying effects of flow outperformed models with a constant relationship between stream temperature and flow. High flows have a strong cooling effect on stream temperatures in April–July, but less influence during other months. The flexibility of GAMs, including non-linear and seasonally varying relationships between stream temperature and flow, produced more accurate predictions than harmonic regression models. Logistic regression of stream temperature with air temperature, based on Mohseni et al.’s (1998) popular method, performed particularly poorly in comparison to the GAMs because it did not include flow as a predictor. Our results confirm previous findings that summer stream temperatures are negatively correlated with flow (Arora et al., 2016; Isaak et al.,

525 2017; Luce et al., 2014; Neumann et al., 2003), and that flow more strongly affects DMxST than
526 DMST (Asarian et al., 2020; Gu et al., 1998; Gu and Li, 2002).

527

528 5.1 Model selection and the importance of seasonally varying and non-linear relationships

529 After considering 13 models, we selected the GAM model with a two-day weighted air
530 temperature (A_{2w}) whose slope varies by day of the year (D), a tensor product smooth of flow
531 and day of the year (Q-D), and an A_{2w} -Q tensor product interaction (Table 1, Figure 4). We
532 chose this model (GAM4 “tensors Q-D & A_{2w} -Q”) based on a combination of model fit (low
533 RMSE, high R^2 , and low fREML score) and fewer effective degrees of freedom (edf) than other
534 models with similar fit. This structure allowed modeled stream temperatures to respond flexibly
535 to varying conditions in all three variables (D, A_{2w} , and Q). Although the three-way tensor
536 GAM1 “tensor Q-D- A_{2w} ” had the lowest fREML score, making it an appealing choice, it also
537 had the highest edf, increasing the risk of being overfit. Indeed, when we experimented with
538 applying GAM1 to model scenarios (not shown), the coolest air temperature scenarios (0.05
539 quantile) had mid-July temperatures that were higher in the typical flow (0.50 quantile) than
540 either the lowest flow (0.05 quantile) or highest flow (0.95 quantile) scenario, which seemed
541 implausible.

542 Comparing the relative performance of models with different smoothers and interactions
543 provides insight into which are most important (Table 1, Figure 4, Figure S5). All models
544 lacking seasonally varying flow effects (i.e., GAM8, GAM9, GAM10, GAM11, and Logistic13)
545 performed worse than any model with seasonally varying flow effects, highlighting the
546 importance of this feature. Modeled temperatures were biased high in April–June in models
547 without seasonally varying flow effects, an issue that is diminished but still present in the
548 Harmonic12 model that represents seasonal effects as perfectly symmetrical sine waves, and
549 completely absent in the models that represents seasonal effects as flexible GAM smoothers
550 (Figure S6). Models with tensors (i.e., GAM1, GAM2, GAM3, GAM4, GAM5, GAM6) had
551 better fit than models with seasonally varying but linear relationships (e.g., GAM7), though the
552 difference was not as great as the difference between seasonally constant models and seasonally
553 varying models. For example, relative to the GAM7 model which is seasonally varying but
554 linear, the GAM8 model with non-linear but seasonally constant relationships had a RMSE 0.4
555 $^{\circ}\text{C}$ lower (0.96 $^{\circ}\text{C}$ vs. 1.35 $^{\circ}\text{C}$) for DMxST and 0.2 $^{\circ}\text{C}$ lower (0.88 $^{\circ}\text{C}$ vs. 1.08 $^{\circ}\text{C}$) for DMST
556 (Table 1). The selected model, GAM4, which has a fully non-linear tensor product smooth of D
557 and Q, and a tensor product interaction of A_{2w} and Q, has improved (relative to GAM8) RMSE
558 of 0.89 $^{\circ}\text{C}$ for DMxST and 0.80 $^{\circ}\text{C}$ for DMST and improved rREML scores (Table 1). In addition,
559 the results for GAM3 (seasonally varying A_{2w} and Q-D tensor product smooth) and GAM5 (A_{2w} -
560 D tensor product interaction and Q-D tensor product smooth), suggests that most of the
561 improvement between GAM7 and GAM4 comes from the Q-D tensor product smooth rather
562 than from the D-varying A_{2w} or Q- A_{2w} tensor product interaction (Table 1).

563 The GAMs work well because they are able to represent the non-linear relationships and
564 interactions between predictor variables present in our dataset. Heeding guidance from previous
565 researchers we prevented overfitting by limiting the number of knots in the tensors (Jackson et
566 al., 2018; Siegel and Volk, 2019). Our flexible approach takes maximal advantage of our multi-
567 decade daily calibration dataset featuring a range of environmental conditions (i.e., hot and cool
568 air temperatures and high and low flows) over the 4696 days. Our validation results suggest that

569 we have enough data to support our rather complex selected model GAM4. Future researchers
570 modeling temperatures at other sites may not have as much data, so should exercise caution and
571 may want to use the simpler GAM7 model.

572

573 5.3 Snow and groundwater mediate the effects of river flow on water temperature

574 Flow magnitude and seasonality at our study site is driven by a mix of valley groundwater
575 dynamics and snowmelt-driven mountain runoff (Foglia et al., 2013; Van Kirk and Naman,
576 2008). Groundwater contributes to the relationship between flow and stream temperature at our
577 study site, as it does in many other rivers (Briggs et al., 2018; Kelleher et al., 2012; Mayer, 2012;
578 Nichols et al., 2014; Isaak et al., 2017). Thermal infrared imagery, field measurements
579 (NCRWQCB, 2005), and a groundwater model (Tolley et al., 2019) all confirm that substantial
580 groundwater is forced into the Scott River where the valley constricts upstream of our site, a
581 common phenomenon at the outlet of alluvial valleys (Stanford and Ward, 1992). Process-based
582 model scenarios predicted a doubling of groundwater-derived flow would cool peak summer
583 Scott River temperatures by 2 °C, and a 50% reduction of groundwater-derived flow would
584 warm temperatures by 2 °C (NCRWQCB, 2005).

585 The timing and magnitude of flow-induced cooling indicated by our models are similar to other
586 snowmelt-dominated rivers. A process-based model of a Sierra Nevada river indicated early
587 summer stream temperatures up to 16 °C cooler in an extreme wet year relative to a dry year
588 (Null et al., 2013). Relative to a statistical model with only air temperature, including flow as a
589 predictor improved stream temperature predictions in April through August in Idaho streams
590 (Sohrabi et al. 2017). Most studies predicting climate change effects do not parse the separate
591 contributions of hydrology and air temperature on stream temperature, but in snow-dominated
592 areas of the western North America, predictions of disproportionate stream temperature warming
593 expected in the summer and/or spring are nearly ubiquitous and attributed to earlier runoff timing
594 from declining snowpack (Caldwell et al., 2013; Crozier et al., 2020; Ficklin et al., 2014; Lee et
595 al., 2020; Leach and Moore, 2019; Luo et al., 2013; Null et al., 2013).

596

597 5.4 Biological implications

598 The prolonged snowmelt-driven flow recession in high-flow years keeps Scott River
599 temperatures cooler longer into the summer than in low-flow years, extending the period when
600 cool water habitat is available for fish (i.e., temperatures less than 22 °C)(Figure 7). These cooler
601 water temperatures give juvenile salmonids additional time to migrate downstream and reduce
602 overall thermal stress for fish that rear in the Scott River through the entire summer. Mean diel
603 range in June–August exceeds 5 °C, providing hours every day when temperatures are less than
604 22 °C even if DMxST exceeds 22 °C. Salmonids can potentially persist by using thermal refugia
605 where cool tributaries, groundwater, or hyporheic flow enters the river during the hotter parts of
606 the day and then moving into the mainstem to feed when temperatures are cooler (Sutton et al.,
607 2007; Sutton and Soto, 2012; Brewitt and Danner, 2014). However, substantial portions of the
608 Scott River and tributaries lack surface flow during summer, especially in dry years, reducing
609 connectivity between thermal refugia and mainstem habitats.

610

611 5.5 Management implications

612 These models can be used not only to identify the seasonally varying influence of flow, but also
613 to predict future stream temperatures based on managed flow recommendations. Instream flow
614 management frameworks are evolving (Mierau et al., 2017; Poff et al., 2017; Yarnell et al.,
615 2020) and accurate stream temperature models provide a valuable tool for use in those processes.
616 Our modeling approach is relatively easy to implement, especially in comparison to a process-
617 based models, which we hope will facilitate water managers' ability to include stream
618 temperature as a management target. For example, Siskiyou County is currently developing a
619 groundwater sustainability plan for Scott Valley (Foglia et al., 2018). The current groundwater
620 model (Tolley et al., 2019) does not simulate water temperatures, so our temperature model
621 could be a used to assess the effects of groundwater management on groundwater-dependent
622 ecosystems. Our results quantify the effect of flow on stream temperatures, including the CDFW
623 and USFS flow thresholds under consideration and could inform state agencies' development of
624 new flow objectives. The CDFW and USFS flows are both predicted to improve (i.e., cool)
625 summer temperatures relative to current conditions, but improvements would be greater with the
626 higher CDFW flows. We caution that while the CDFW and USFS flows are relatively high
627 compared to observed flows in late summer and early fall, for March to early June they represent
628 extreme drought conditions (Figure 2b), which has two implications. First, in dry years
629 temperatures reach 22 °C in early or mid-June in the observed flow scenario, which is only
630 delayed in a small number of years in the scenarios with CDFW and USFS flows as minimums.
631 Second, if river flows were diverted down to the CDFW and USFS flows in May and June, then
632 the 22 °C threshold would be reached an average of approximately two weeks earlier than
633 occurred with the observed flows (Figure 7c). Surface water diversions for in lieu recharge
634 (switching irrigation source from groundwater to surface water) or managed aquifer recharge
635 (Dahlke et al., 2018; Foglia et al., 2013) should not use the CDFW and USFS flows to guide
636 maximum diversion rates, but instead be tailored to reduce deleterious effects on instream habitat
637 including temperatures, such as ceasing by 1 June, the first date when measured (Figure 2) and
638 modeled temperatures (Figure 7) reach 22 °C.

639 As with any regression model, prediction accuracy is likely to degrade when applied to
640 conditions more extreme than those present in the calibration dataset. Our calibration dataset
641 includes a wide range of hydrologic conditions, but no years without surface water diversions or
642 groundwater pumping because those activities occur every year. Groundwater modeling efforts
643 suggest that streamflow depletion from groundwater pumping would be greater in dry years than
644 wet years, because in dry years pumping starts earlier, cumulative amounts pumped are greater,
645 and the aquifer is drawn down lower (Foglia et al., 2013; Tolley et al., 2019). Simulated total
646 valley-wide streamflow depletion peaks around 150,000 m³d⁻¹ (60 ft³/sec) in July and August
647 (Foglia et al., 2013), exceeding streamflow in dry years. Our model should be suitable for
648 modeling stream temperatures in dry years for scenarios with reduced pumping and/or
649 diversions, which would presumably have flows similar to existing wet years (and hence are
650 within the range of calibration flows); however, in wet years such scenarios would likely exceed
651 the range of calibration flows and therefore be subject to considerably higher levels of
652 uncertainty. Any future application of our model to scenarios with flows higher than observed
653 should be done carefully and interpreted with appropriate caveats.

654

655 **6 Conclusions**

656 Statistical models indicate that river flow has a strong cooling effect on river temperatures during
657 April through July in California's Scott River, similar to previous findings from process-based
658 models in many snow-dominated rivers in western North America. A 24-year dataset of daily
659 streams temperatures allowed us to develop a generalized additive model using tensor product
660 smooths and interactions to represent the non-linear and seasonally varying effects of flow and
661 air temperature on stream temperature. Our model also includes the correlation structures
662 inherent in the data, namely daily temporal autocorrelation and random effects for annual
663 variation. Validation indicated excellent model performance, with average errors ≤ 1 °C. This
664 project contributes to an emerging body of work demonstrating the benefits of generalized
665 additive models (GAMs) for modeling daily river temperatures. Given the flexibility of GAMs,
666 there is a risk of overfitting data, but this risk can be minimized by restricting the number of
667 knots in GAM smoothers, confirming that smoother shape matches scientific hypotheses
668 regarding the underlying physical processes, and considering whether sample size is adequate for
669 the complexity of the model.

670 These models identify the specific periods of the year when flow has greatest influence on
671 stream temperatures, and can be used to evaluate the thermal effects of alternative flow
672 management scenarios and prescriptions. The models are implemented in the R software
673 environment with publicly accessible code, and could be applied to model year-round daily
674 temperature in any stream with long-term measurements of flow and water temperature,
675 provided that air temperatures are available from a nearby weather station.

676

677 **CRedit authorship contribution statement**

678 J. Eli Asarian: Conceptualization, Data curation, Methodology, Formal analysis, Visualization,
679 Writing – original draft, Writing – review & editing. Crystal Robinson: Conceptualization,
680 Investigation, Data curation, Funding acquisition, Project administration, Writing - review &
681 editing.

682

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688 Alex Case, Sean Ryan and Marla Bennett (QVIR) assisted with data collection. Edward Jones
689 (USGS) provided code for non-linear regression. Laurel Genzoli reviewed a draft of this
690 manuscript.

691

692 **Data Availability Statement**

693 All data and code are archived in the online repository HydroShare (Asarian and Robinson,
694 2021, <http://www.hydroshare.org/resource/a6653e2919964f9b840ec0340d86e11c>). USBR

695 stream temperature data are also available at: [https://or.water.usgs.gov/cgi-](https://or.water.usgs.gov/cgi-bin/grapher/graph_setup.pl?basin_id=all&site_id=11519500)
 696 [bin/grapher/graph_setup.pl?basin_id=all&site_id=11519500](https://or.water.usgs.gov/cgi-bin/grapher/graph_setup.pl?basin_id=all&site_id=11519500). Gridded PRISM air temperature
 697 data (Daly et al., 2008) are also available at: <https://prism.oregonstate.edu/explorer/>. GHCN-D
 698 air temperature data (Menne et al., 2012a, 2012b) are also available at
 699 <http://doi.org/10.7289/V5D21VHZ>.

700

701 **References**

- 702 Amodio, S., Aria, M., & D'Ambrosio, A. (2014). On concurvity in nonlinear and nonparametric
 703 regression models. *Statistica*, 74(1), 85–98. <https://doi.org/10.6092/issn.1973-2201/4599>
- 704 Arismendi, I., Safeeq, M., Dunham, J. B., & Johnson, S. L. (2014). Can air temperature be used
 705 to project influences of climate change on stream temperature? *Environmental Research Letters*,
 706 9(8), 084015. <https://doi.org/10.1088/1748-9326/9/8/084015>
- 707 Arora, R., Tockner, K., & Venohr, M. (2016). Changing river temperatures in northern Germany:
 708 Trends and drivers of change. *Hydrological Processes*, 30(17), 3084–3096.
 709 <https://doi.org/10.1002/hyp.10849>
- 710 Asarian, J.E., Cressey, L., Bennett, B., Grunbaum, J., Cyr, L., Soto, T., & Robinson, C. (2020).
 711 *Influence of Snowpack, Streamflow, Air Temperature, and Wildfire Smoke on Klamath Basin*
 712 *Stream Temperatures, 1995-2017*. Eureka, CA: Riverbend Sciences.
 713 <https://doi.org/10.13140/RG.2.2.22934.47681>
- 714 Asarian, J. E. & Robinson, C. (2021). Data and Codes for: Modeling Seasonal Effects of River
 715 Flow on Water Temperatures in an Agriculturally Dominated California River. HydroShare,
 716 <http://www.hydroshare.org/resource/a6653e2919964f9b840ec0340d86e11c>
- 717 Asarian, J. E., & Walker, J. D. (2016). Long-Term Trends in Streamflow and Precipitation in
 718 Northwest California and Southwest Oregon, 1953-2012. *Journal of the American Water*
 719 *Resources Association*, 52(1), 241–261. <https://doi.org/10.1111/1752-1688.12381>
- 720 Baayen, R. H., van Rij, J., de Cat, C., & Wood, S. (2018). Autocorrelated Errors in Experimental
 721 Data in the Language Sciences: Some Solutions Offered by Generalized Additive Mixed Models.
 722 In D. Speelman, K. Heylen, & D. Geeraerts (Eds.), *Mixed-Effects Regression Models in*
 723 *Linguistics* (pp. 49–69). Springer International Publishing. [https://doi.org/10.1007/978-3-319-](https://doi.org/10.1007/978-3-319-69830-4_4)
 724 [69830-4_4](https://doi.org/10.1007/978-3-319-69830-4_4)
- 725 Bartholow, J. M. (1991). A modeling assessment of the thermal regime for an urban sport
 726 fishery. *Environmental Management*, 15(6), 833. <https://doi.org/10.1007/BF02394821>
- 727 Benyahya, L., Caissie, D., St-Hilaire, A., Ouarda, T. B. M. J., & Bobée, B. (2007a). A Review of
 728 Statistical Water Temperature Models. *Canadian Water Resources Journal / Revue Canadienne*
 729 *Des Ressources Hydriques*, 32(3), 179–192. <https://doi.org/10.4296/cwrj3203179>
- 730 Benyahya, L., St-Hilaire, A., Quarda, T. B. M. J., Bobée, B., & Ahmadi-Nedushan, B. (2007b).
 731 Modeling of water temperatures based on stochastic approaches: Case study of the Deschutes
 732 River. *Journal of Environmental Engineering and Science*, 6(4), 437–448.
 733 <https://doi.org/10.1139/s06-067>

- 734 Benyahya, L., St-Hilaire, A., Ouarda, T., Bobee, B., & Dumas, J. (2008). Comparison of non-
735 parametric and parametric water temperature models on the Nivelles River, France. *Hydrological*
736 *Sciences Journal*, 53(3), 640–655.
- 737 Bernhardt, E. S., Heffernan, J. B., Grimm, N. B., Stanley, E. H., Harvey, J. W., Arroita, M.,
738 Appling, A. P., Cohen, M. J., McDowell, W. H., Hall, R. O., Read, J. S., Roberts, B. J., Stets, E.
739 G., & Yackulic, C. B. (2017). The metabolic regimes of flowing waters: Metabolic regimes.
740 *Limnology and Oceanography*. <https://doi.org/10.1002/lno.10726>
- 741 Bodeker, G. E., Boyd, I. S., & Matthews, W. A. (1998). Trends and variability in vertical ozone
742 and temperature profiles measured by ozonesondes at Lauder, New Zealand: 1986–1996. *Journal*
743 *of Geophysical Research: Atmospheres*, 103(D22), 28661–28681.
744 <https://doi.org/10.1029/98JD02581>
- 745 Boudreault, J., Bergeron, N. E., St-Hilaire, A., & Chebana, F. (2019). Stream temperature
746 modeling using functional regression models. *Journal of the American Water Resources*
747 *Association*, 55(6), 1382–1400. <https://doi.org/10.1111/1752-1688.12778>
- 748 Boyd, M., and Kasper, B. (2003). Analytical methods for dynamic open channel heat and mass
749 transfer: Methodology for Heat Source model version 7.0. Portland, OR: Oregon Department of
750 Environmental Quality.
- 751 Brewitt, K. S., & Danner, E. M. (2014). Spatio-temporal temperature variation influences
752 juvenile steelhead (*Oncorhynchus mykiss*) use of thermal refuges. *Ecosphere*, 5(7), art92.
753 <https://doi.org/10.1890/ES14-00036.1>
- 754 Briggs, M. A., Johnson, Z. C., Snyder, C. D., Hitt, N. P., Kurylyk, B. L., Lautz, L., Irvine, D. J.,
755 Hurley, S. T., & Lane, J. W. (2018). Inferring watershed hydraulics and cold-water habitat
756 persistence using multi-year air and stream temperature signals. *Science of The Total*
757 *Environment*, 636, 1117–1127. <https://doi.org/10.1016/j.scitotenv.2018.04.344>
- 758 Brown, G. W. (1969). Predicting Temperatures of Small Streams. *Water Resources Research*,
759 5(1), 68–75. <https://doi.org/10.1029/WR005i001p00068>
- 760 Cade, B. S., & Noon, B. R. (2003). A gentle introduction to quantile regression for ecologists.
761 *Frontiers in Ecology and the Environment*, 1(8), 412–420. [https://doi.org/10.1890/1540-9295\(2003\)001\[0412:AGITQR\]2.0.CO;2](https://doi.org/10.1890/1540-9295(2003)001[0412:AGITQR]2.0.CO;2)
- 762
- 763 Caissie, D. (2006). The thermal regime of rivers: A review. *Freshwater Biology*, 51(8), 1389–
764 1406. <https://doi.org/10.1111/j.1365-2427.2006.01597.x>
- 765 Caldwell, R. J., Gangopadhyay, S., Bountry, J., Lai, Y., & Elsner, M. M. (2013). Statistical
766 modeling of daily and subdaily stream temperatures: Application to the Methow River Basin,
767 Washington. *Water Resources Research*, 49(7), 4346–4361. <https://doi.org/10.1002/wrcr.20353>
- 768 California Department of Fish and Wildlife (CDFW) (2017). Interim Instream Flow Criteria for
769 the Protection of Fishery Resources in the Scott River Watershed, Siskiyou County.
770 [https://web.archive.org/web/20210324204650if_/https://nrm.dfg.ca.gov/FileHandler.ashx?Docu](https://web.archive.org/web/20210324204650if_/https://nrm.dfg.ca.gov/FileHandler.ashx?DocumentID=143476)
771 [mentID=143476.](https://web.archive.org/web/20210324204650if_/https://nrm.dfg.ca.gov/FileHandler.ashx?DocumentID=143476)
- 772 Chandesris, A., Van Looy, K., Diamond, J. S., & Souchon, Y. (2019). Small dams alter thermal
773 regimes of downstream water. *Hydrology and Earth System Sciences*, 23(11), 4509–4525.
774 <https://doi.org/10.5194/hess-23-4509-2019>

- 775 Cox, N. J. (2006). Speaking Stata: In Praise of Trigonometric Predictors. *The Stata Journal: Promoting Communications on Statistics and Stata*, 6(4), 561–579.
776 <https://doi.org/10.1177/1536867X0600600408>
777
- 778 Crozier, L. G., Siegel, J. E., Wiesebron, L. E., Trujillo, E. M., Burke, B. J., Sandford, B. P., &
779 Widener, D. L. (2020). Snake River sockeye and Chinook salmon in a changing climate:
780 Implications for upstream migration survival during recent extreme and future climates. *PLOS ONE*, 15(9), e0238886. <https://doi.org/10.1371/journal.pone.0238886>
781
- 782 Dahlke, H., Brown, A., Orloff, S., Putnam, D., & O'Geen, T. (2018). Managed winter flooding
783 of alfalfa recharges groundwater with minimal crop damage. *California Agriculture*, 72(1), 1–11.
784 <https://doi.org/10.3733/ca.2018a0001>
- 785 Dahlke, F. T., Wohlrab, S., Butzin, M., & Pörtner, H.-O. (2020). Thermal bottlenecks in the life
786 cycle define climate vulnerability of fish. *Science*, 369(6499), 65–70.
787 <https://doi.org/10.1126/science.aaz3658>
- 788 Daly, C., Halbleib, M., Smith, J. I., Gibson, W. P., Doggett, M. K., Taylor, G. H., Curtis, J., &
789 Pasteris, P. P. (2008). Physiographically sensitive mapping of climatological temperature and
790 precipitation across the conterminous United States. *International Journal of Climatology*,
791 28(15), 2031–2064. <https://doi.org/10.1002/joc.1688>
- 792 David, A. T., Asarian, J. E., & Lake, F. K. (2018). Wildfire Smoke Cools Summer River and
793 Stream Water Temperatures. *Water Resources Research*, 54(10), 7273–7290.
794 <https://doi.org/10.1029/2018WR022964>
- 795 Drake, D., Tate, K., & Carlson, H. (2000). Analysis shows climate-caused decreases in Scott
796 River fall flows. *California Agriculture*, 54(6), 46–49. <https://doi.org/10.3733/ca.v054n06p46>
- 797 Dugdale, S. J., Hannah, D. M., & Malcolm, I. A. (2017). River temperature modelling: A review
798 of process-based approaches and future directions. *Earth-Science Reviews*, 175, 97–113.
799 <https://doi.org/10.1016/j.earscirev.2017.10.009>
- 800 Dugdale, S. J., Malcolm, I. A., Kantola, K., & Hannah, D. M. (2018). Stream temperature under
801 contrasting riparian forest cover: Understanding thermal dynamics and heat exchange processes.
802 *Science of The Total Environment*, 610–611, 1375–1389.
803 <https://doi.org/10.1016/j.scitotenv.2017.08.198>
- 804 Dymond, J. R. (1984). Water temperature change caused by abstraction. *Journal of Hydraulic*
805 *Engineering*, 110(7), 987–991. [https://doi.org/10.1061/\(ASCE\)0733-9429\(1984\)110:7\(987\)](https://doi.org/10.1061/(ASCE)0733-9429(1984)110:7(987))
- 806 Erickson, T. R., & Stefan, H. G. (2000). Linear Air/Water Temperature Correlations for Streams
807 during Open Water Periods. *Journal of Hydrologic Engineering*, 5(3), 317–321.
808 [https://doi.org/10.1061/\(ASCE\)1084-0699\(2000\)5:3\(317\)](https://doi.org/10.1061/(ASCE)1084-0699(2000)5:3(317))
- 809 Ficklin, D. L., Barnhart, B. L., Knouft, J. H., Stewart, I. T., Maurer, E. P., Letsinger, S. L., &
810 Whittaker, G. W. (2014). Climate change and stream temperature projections in the Columbia
811 River basin: Habitat implications of spatial variation in hydrologic drivers. *Hydrology and Earth*
812 *System Sciences*, 18(12), 4897–4912. <https://doi.org/10.5194/hess-18-4897-2014>
- 813 FitzGerald, A. M., John, S. N., Apgar, T. M., Mantua, N. J., & Martin, B. T. (2021). Quantifying
814 thermal exposure for migratory riverine species: Phenology of Chinook salmon populations

- 815 predicts thermal stress. *Global Change Biology*, 27(3), 536–549.
816 <https://doi.org/10.1111/gcb.15450>
- 817 Foglia, L., McNally, A., & Harter, T. (2013). Coupling a spatiotemporally distributed soil water
818 budget with stream-depletion functions to inform stakeholder-driven management of
819 groundwater-dependent ecosystems. *Water Resources Research*, 49(11), 7292–7310.
820 <https://doi.org/10.1002/wrcr.20555>
- 821 Foglia, L., Neumann, J., Tolley, D., Orloff, S., Snyder, R., & Harter, T. (2018). Modeling guides
822 groundwater management in a basin with river–aquifer interactions. *California Agriculture*,
823 72(1), 84–95. <https://doi.org/10.3733/ca.2018a0011>
- 824 Folegot, S., Hannah, D. M., Dugdale, S. J., Kurz, M. J., Drummond, J. D., Klaar, M. J., Lee-
825 Cullin, J., Keller, T., Martí, E., Zarnetske, J. P., Ward, A. S., & Krause, S. (2018). Low flow
826 controls on stream thermal dynamics. *Limnologia*, 68, 157–167.
827 <https://doi.org/10.1016/j.limno.2017.08.003>
- 828 Fullerton, A. H., Torgersen, C. E., Lawler, J. J., Faux, R. N., Steel, E. A., Beechie, T. J.,
829 Ebersole, J. L., & Leibowitz, S. G. (2015). Rethinking the longitudinal stream temperature
830 paradigm: Region-wide comparison of thermal infrared imagery reveals unexpected complexity
831 of river temperatures. *Hydrological Processes*, 29(22), 4719–4737.
832 <https://doi.org/10.1002/hyp.10506>
- 833 Gibeau, P., & Palen, W. J. (2020). Predicted effects of flow diversion by Run-of-River
834 hydropower on bypassed stream temperature and bioenergetics of salmonid fishes. *River*
835 *Research and Applications*, 36(9), 1903–1915. <https://doi.org/10.1002/rra.3706>
- 836 Gallice, A., Schaeffli, B., Lehning, M., Parlange, M. B., & Huwald, H. (2015). Stream
837 temperature prediction in ungauged basins: Review of recent approaches and description of a
838 new physics-derived statistical model. *Hydrol. Earth Syst. Sci.*, 19(9), 3727–3753.
839 <https://doi.org/10.5194/hess-19-3727-2015>
- 840 Gu, R. R., & Li, Y. (2002). River temperature sensitivity to hydraulic and meteorological
841 parameters. *Journal of Environmental Management*, 66(1), 43–56.
842 <https://doi.org/10.1006/jema.2002.0565>
- 843 Gu, R., Montgomery, S., & Austin, T. A. (1998). Quantifying the effects of stream discharge on
844 summer river temperature. *Hydrological Sciences Journal*, 43(6), 885–904.
845 <https://doi.org/10.1080/02626669809492185>
- 846 Hahm, W. J., Rempe, D. M., Dralle, D. N., Dawson, T. E., Lovill, S. M., Bryk, A. B., Bish, D.
847 L., Schieber, J., & Dietrich, W. E. (2019). Lithologically Controlled Subsurface Critical Zone
848 Thickness and Water Storage Capacity Determine Regional Plant Community Composition.
849 *Water Resources Research*, 55(4), 3028–3055. <https://doi.org/10.1029/2018WR023760>
- 850 Helsel, D. R., Hirsch, R. M., Ryberg, K. R., Archfield, S. A., & Gilroy, E. J. (2020). Statistical
851 methods in water resources (*U.S. Geological Survey Techniques and Methods*, book 4, chapter
852 A3). <https://doi.org/10.3133/tm4A3>
- 853 Hilderbrand, R. H., Kashiwagi, M. T., & Prochaska, A. P. (2014). Regional and Local Scale
854 Modeling of Stream Temperatures and Spatio-Temporal Variation in Thermal Sensitivities.
855 *Environmental Management*, 54(1), 14–22. <https://doi.org/10.1007/s00267-014-0272-4>

- 856 Isaak, D. J., Luce, C. H., Horan, D. L., Chandler, G. L., Wollrab, S. P., & Nagel, D. E. (2018).
857 Global Warming of Salmon and Trout Rivers in the Northwestern U.S.: Road to Ruin or Path
858 Through Purgatory? *Transactions of the American Fisheries Society*, 147(3), 566–587.
859 <https://doi.org/10.1002/tafs.10059>
- 860 Isaak, D. J., Wenger, S. J., Peterson, E. E., Hoef, J. M. V., Nagel, D. E., Luce, C. H., Hostetler,
861 S. W., Dunham, J. B., Roper, B. B., Wollrab, S. P., Chandler, G. L., Horan, D. L., & Parkes-
862 Payne, S. (2017). The NorWeST Summer Stream Temperature Model and Scenarios for the
863 Western U.S.: A Crowd-Sourced Database and New Geospatial Tools Foster a User Community
864 and Predict Broad Climate Warming of Rivers and Streams. *Water Resources Research*, 53(11),
865 9181–9205. <https://doi.org/10.1002/2017WR020969>
- 866 Jackson, F. L., Fryer, R. J., Hannah, D. M., Millar, C. P., & Malcolm, I. A. (2018). A spatio-
867 temporal statistical model of maximum daily river temperatures to inform the management of
868 Scotland’s Atlantic salmon rivers under climate change. *Science of The Total Environment*, 612,
869 1543–1558. <https://doi.org/10.1016/j.scitotenv.2017.09.010>
- 870 Johnson, S. L. (2004). Factors influencing stream temperatures in small streams: Substrate
871 effects and a shading experiment. *Canadian Journal of Fisheries and Aquatic Sciences*, 61(6),
872 913–923.
- 873 Johnson, Z. C., Johnson, B. G., Briggs, M. A., Devine, W. D., Snyder, C. D., Hitt, N. P., Hare,
874 D. K., & Minkova, T. V. (2020). Paired air-water annual temperature patterns reveal
875 hydrogeological controls on stream thermal regimes at watershed to continental scales. *Journal*
876 *of Hydrology*, 587, 124929. <https://doi.org/10.1016/j.jhydrol.2020.124929>
- 877 Jones, E. C., Perry, R. W., Risley, J. C., Som, N. A., & Hetrick, N. J. (2016). Construction,
878 calibration, and validation of the RBM10 water temperature model for the Trinity River,
879 Northern California (*U.S. Geological Survey Open-File Report 2016-1056*).
880 <https://doi.org/10.3133/ofr20161056>
- 881 Kelleher, C., Wagener, T., Gooseff, M., McGlynn, B., McGuire, K., & Marshall, L. (2012).
882 Investigating controls on the thermal sensitivity of Pennsylvania streams. *Hydrological*
883 *Processes*, 26(5), 771–785. <https://doi.org/10.1002/hyp.8186>
- 884 Kim, J.-S., & Jain, S. (2010). High-resolution streamflow trend analysis applicable to annual
885 decision calendars: A western United States case study. *Climatic Change*, 102(3–4), 699–707.
886 <https://doi.org/10.1007/s10584-010-9933-3>
- 887 Klamath National Forest (KNF), 2010. Klamath National Forest Sediment and Temperature
888 Monitoring Plan and Quality Assurance Project Plan. Yreka, CA: Klamath National Forest.
889 [https://web.archive.org/web/http://www.waterboards.ca.gov/water_issues/programs/tmdl/records](https://web.archive.org/web/http://www.waterboards.ca.gov/water_issues/programs/tmdl/records/region_1/2013/ref4082.pdf)
890 [/region_1/2013/ref4082.pdf](https://web.archive.org/web/http://www.waterboards.ca.gov/water_issues/programs/tmdl/records/region_1/2013/ref4082.pdf).
- 891 Koch, H., & Grünewald, U. (2010). Regression models for daily stream temperature simulation:
892 Case studies for the river Elbe, Germany. *Hydrological Processes*, 24(26), 3826–3836.
893 <https://doi.org/10.1002/hyp.7814>
- 894 Kothandaraman, V. (1971). Analysis of Water Temperature Variations in Large River. *Journal*
895 *of the Sanitary Engineering Division*, 97(1), 19–31. <https://doi.org/10.1061/JSEDAI.0001242>
- 896 Laanaya, F., St-Hilaire, A., & Gloaguen, E. (2017). Water temperature modelling: Comparison
897 between the generalized additive model, logistic, residuals regression and linear regression

- 898 models. *Hydrological Sciences Journal*, 62(7), 1078–1093.
899 <https://doi.org/10.1080/02626667.2016.1246799>
- 900 Leach, J. A., & Moore, R. D. (2019). Empirical stream thermal sensitivities may underestimate
901 stream temperature response to climate warming. *Water Resources Research*, 55(7), 5453–5467.
902 <https://doi.org/10.1029/2018WR024236>
- 903 Lee, S.-Y., Fullerton, A. H., Sun, N., & Torgersen, C. E. (2020). Projecting spatiotemporally
904 explicit effects of climate change on stream temperature: A model comparison and implications
905 for coldwater fishes. *Journal of Hydrology*, 588, 125066.
906 <https://doi.org/10.1016/j.jhydrol.2020.125066>
- 907 Letcher, B. H., Hocking, D. J., O’Neil, K., Whiteley, A. R., Nislow, K. H., & O’Donnell, M. J.
908 (2016). A hierarchical model of daily stream temperature using air-water temperature
909 synchronization, autocorrelation, and time lags. *PeerJ*, 4, e1727.
910 <https://doi.org/10.7717/peerj.1727>
- 911 Li, H., Deng, X., Kim, D.-Y., & Smith, E. P. (2014). Modeling maximum daily temperature
912 using a varying coefficient regression model. *Water Resources Research*, 50(4), 3073–3087.
913 <https://doi.org/10.1002/2013WR014243>
- 914 Liu, S., Xie, Z., Liu, B., Wang, Y., Gao, J., Zeng, Y., Xie, J., Xie, Z., Jia, B., Qin, P., Li, R.,
915 Wang, L., & Chen, S. (2020). Global river water warming due to climate change and
916 anthropogenic heat emission. *Global and Planetary Change*, 193, 103289.
917 <https://doi.org/10.1016/j.gloplacha.2020.103289>
- 918 Luce, C., Staab, B., Kramer, M., Wenger, S., Isaak, D., & McConnell, C. (2014). Sensitivity of
919 summer stream temperatures to climate variability in the Pacific Northwest. *Water Resources*
920 *Research*, 50(4), 3428–3443. <https://doi.org/10.1002/2013WR014329>
- 921 Luo, Y., Ficklin, D. L., Liu, X., & Zhang, M. (2013). Assessment of climate change impacts on
922 hydrology and water quality with a watershed modeling approach. *Science of The Total*
923 *Environment*, 450–451, 72–82. <https://doi.org/10.1016/j.scitotenv.2013.02.004>
- 924 Lusardi, R. A., Hammock, B. G., Jeffres, C. A., Dahlgren, R. A., & Kiernan, J. D. (2019).
925 Oversummer growth and survival of juvenile coho salmon (*Oncorhynchus kisutch*) across a
926 natural gradient of stream water temperature and prey availability: An in situ enclosure
927 experiment. *Canadian Journal of Fisheries and Aquatic Sciences*, 1–12.
928 <https://doi.org/10.1139/cjfas-2018-0484>
- 929 Manhard, C. V., N. A. Som, E. C. Jones, & R. W. Perry. 2018. Estimation of stream conditions
930 in tributaries of the Klamath River, Northern California (*Arcata Fisheries Technical Report*
931 *Number TR 2018-32*). Arcata, CA: U.S. Fish and Wildlife Service.
932 <https://web.archive.org/web/https://www.fws.gov/arcata/fisheries/reports/technical/2018/EstimationofStreamConditionsinTributariesoftheKlamathRiverNorthernCalifornia.pdf>
- 934 Mayer, T. D. (2012). Controls of summer stream temperature in the Pacific Northwest. *Journal*
935 *of Hydrology*, 475(0), 323–335. <https://doi.org/10.1016/j.jhydrol.2012.10.012>
- 936 McGrath, E. O., Neumann, N. N., & Nichol, C. F. (2017). A Statistical Model for Managing
937 Water Temperature in Streams with Anthropogenic Influences. *River Research and Applications*,
938 33(1), 123–134. <https://doi.org/10.1002/rra.3057>

- 939 Meier, W., Bonjour, C., Wüest, A., & Reichert, P. (2003). Modeling the Effect of Water
940 Diversion on the Temperature of Mountain Streams. *Journal of Environmental Engineering*,
941 129(8), 755–764. [https://doi.org/10.1061/\(ASCE\)0733-9372\(2003\)129:8\(755\)](https://doi.org/10.1061/(ASCE)0733-9372(2003)129:8(755))
- 942 Menne, M.J., Durre, I., Korzeniewski, B., McNeal, S., Thomas, K., Yin, X., Anthony, S., Ray,
943 R., Vose, R.S., Gleason, B.E., & Houston, T.G. (2012a). Global Historical Climatology Network
944 - Daily (GHCN-Daily), Version 3.26. NOAA National Climatic Data Center.
945 <http://doi.org/10.7289/V5D21VHZ>. Accessed 2021-01-11
- 946 Menne, M. J., Durre, I., Vose, R. S., Gleason, B. E., & Houston, T. G. (2012b). An Overview of
947 the Global Historical Climatology Network-Daily Database. *Journal of Atmospheric and*
948 *Oceanic Technology*, 29(7), 897–910. <https://doi.org/10.1175/JTECH-D-11-00103.1>
- 949 Mierau, D. W., Trush, W. J., Rossi, G. J., Carah, J. K., Clifford, M. O., & Howard, J. K. (2017).
950 Managing diversions in unregulated streams using a modified percent-of-flow approach.
951 *Freshwater Biology*. <https://doi.org/10.1111/fwb.12985>
- 952 Mohseni, O., Stefan, H. G., & Erickson, T. R. (1998). A nonlinear regression model for weekly
953 stream temperatures. *Water Resources Research*, 34(10), 2685–2692.
954 <https://doi.org/10.1029/98WR01877>
- 955 Moore, D. R., Spittlehouse, D. L., & Story, A. (2005a). Riparian Microclimate and Stream
956 Temperature Response to Forest Harvesting: A Review. *Journal of the American Water*
957 *Resources Association*, 41(4), 813–834. <https://doi.org/10.1111/j.1752-1688.2005.tb03772.x>
- 958 Moore, R. D., Sutherland, P., Gomi, T., & Dhakal, A. (2005b). Thermal regime of a headwater
959 stream within a clear-cut, coastal British Columbia, Canada. *Hydrological Processes*, 19(13),
960 2591–2608. <https://doi.org/10.1002/hyp.5733>
- 961 Moulton, T. L. (2018). *rMR: Importing Data from Loligo Systems Software, Calculating*
962 *Metabolic Rates and Critical Tensions*. R package version 1.1.0. [https://CRAN.R-](https://CRAN.R-project.org/package=rMR)
963 [project.org/package=rMR](https://CRAN.R-project.org/package=rMR)
- 964 Muggeo, V. M. R., Sciandra, M., Tomasello, A., & Calvo, S. (2013). Estimating growth charts
965 via nonparametric quantile regression: A practical framework with application in ecology.
966 *Environmental and Ecological Statistics*, 20(4), 519–531. [https://doi.org/10.1007/s10651-012-](https://doi.org/10.1007/s10651-012-0232-1)
967 0232-1
- 968 National Marine Fisheries Service (NMFS) (2014). Final Recovery Plan for the Southern
969 Oregon/Northern California Coast Evolutionarily Significant Unit of Coho Salmon
970 (*Oncorhynchus kisutch*). Arcata, CA: National Marine Fisheries Service.
- 971 Neumann David W., Rajagopalan Balaji, & Zagona Edith A. (2003). Regression Model for Daily
972 Maximum Stream Temperature. *Journal of Environmental Engineering*, 129(7), 667–674.
973 [https://doi.org/10.1061/\(ASCE\)0733-9372\(2003\)129:7\(667\)](https://doi.org/10.1061/(ASCE)0733-9372(2003)129:7(667))
- 974 Nichols, A. L., Willis, A. D., Jeffres, C. A., & Deas, M. L. (2014). Water temperature patterns
975 below large groundwater springs: management implications for coho salmon in the Shasta River,
976 California. *River Research and Applications*, 30(4), 442–455. <https://doi.org/10.1002/rra.2655>
- 977 North Coast Regional Water Quality Control Board (NCRWQCB) (2005). *Staff Report for the*
978 *Action Plan for the Scott River Watershed Sediment and Temperature Total Maximum Daily*

- 979 *Loads*. Santa Rosa, CA: North Coast Regional Water Quality Control Board.
 980 https://www.waterboards.ca.gov/northcoast/water_issues/programs/tmdls/scott_river/staff_report
- 981 Null, S. E., Mouzon, N. R., & Elmore, L. R. (2017). Dissolved oxygen, stream temperature, and
 982 fish habitat response to environmental water purchases. *Journal of Environmental Management*,
 983 197, 559–570. <https://doi.org/10.1016/j.jenvman.2017.04.016>
- 984 Null, S. E., Viers, J. H., Deas, M. L., Tanaka, S. K., & Mount, J. F. (2013). Stream temperature
 985 sensitivity to climate warming in California’s Sierra Nevada: Impacts to coldwater habitat.
 986 *Climatic Change*, 116(1), 149–170. <https://doi.org/10.1007/s10584-012-0459-8>
- 987 Ouellet, V., St-Hilaire, A., Dugdale, S. J., Hannah, D. M., Krause, S., & Proulx-Ouellet, S.
 988 (2020). River temperature research and practice: Recent challenges and emerging opportunities
 989 for managing thermal habitat conditions in stream ecosystems. *Science of The Total*
 990 *Environment*, 736, 139679. <https://doi.org/10.1016/j.scitotenv.2020.139679>
- 991 Pedersen, E. J., Miller, D. L., Simpson, G. L., & Ross, N. (2019). Hierarchical generalized
 992 additive models in ecology: An introduction with mgcv. *PeerJ*, 7, e6876.
 993 <https://doi.org/10.7717/peerj.6876>
- 994 Persad, G. G., Swain, D. L., Kouba, C., & Ortiz-Partida, J. P. (2020). Inter-model agreement on
 995 projected shifts in California hydroclimate characteristics critical to water management. *Climatic*
 996 *Change*. <https://doi.org/10.1007/s10584-020-02882-4>
- 997 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team (2020). *nlme: Linear and*
 998 *Nonlinear Mixed Effects Models*. R package version 3.1-148. [https://CRAN.R-](https://CRAN.R-project.org/package=nlme)
 999 [project.org/package=nlme](https://CRAN.R-project.org/package=nlme).
- 1000 Piotrowski, A. P., & Napiorkowski, J. J. (2019). Simple modifications of the nonlinear
 1001 regression stream temperature model for daily data. *Journal of Hydrology*, 572, 308–328.
 1002 <https://doi.org/10.1016/j.jhydrol.2019.02.035>
- 1003 Poff, N. L., Tharme, R. E., & Arthington, A. H. (2017). Chapter 11—Evolution of
 1004 Environmental Flows Assessment Science, Principles, and Methodologies. In A. C. Horne, J. A.
 1005 Webb, M. J. Stewardson, B. Richter, & M. Acreman (Eds.), *Water for the Environment* (pp.
 1006 203–236). Academic Press. <https://doi.org/10.1016/B978-0-12-803907-6.00011-5>
- 1007 Power, M. E., & Dietrich, W. E. (2002). Food webs in river networks. *Ecological Research*,
 1008 17(4), 451–471.
- 1009 Quartz Valley Indian Reservation (QVIR) (2016). *Quality Assurance Project Plan 2016 Revision*
 1010 *Water Quality Sampling and Analysis, CWA 106 grant identification # I-96927206-0*. Fort Jones,
 1011 CA: QVIR Tribal Environmental Protection Department.
- 1012 Quigley, D., Färber, S., Conner, K., Power, J., & Bundy, L. (2001). Water Temperatures in the
 1013 Scott River Watershed in Northern California.
 1014 [https://web.archive.org/web/http://www.fws.gov/yreka/Final-Reports/rmaap/2000-JITW-01-](https://web.archive.org/web/http://www.fws.gov/yreka/Final-Reports/rmaap/2000-JITW-01-SRCD.pdf)
 1015 [SRCD.pdf](https://web.archive.org/web/http://www.fws.gov/yreka/Final-Reports/rmaap/2000-JITW-01-SRCD.pdf)
- 1016 R Core Team (2020). *R: A language and environment for statistical computing*. Vienna, Austria:
 1017 R Foundation for Statistical Computing. <https://www.R-project.org/>.

- 1018 Rahmani, F., Lawson, K., Ouyang, W., Appling, A., Oliver, S., & Shen, C. (2020). Exploring the
1019 exceptional performance of a deep learning stream temperature model and the value of
1020 streamflow data. *Environmental Research Letters*. <https://doi.org/10.1088/1748-9326/abd501>
- 1021 Santiago, J. M., Muñoz-Mas, R., Solana-Gutiérrez, J., García de Jalón, D., Alonso, C., Martínez-
1022 Capel, F., Pórtoles, J., Monjo, R., & Ribalaygua, J. (2017). Waning habitats due to climate
1023 change: The effects of changes in streamflow and temperature at the rear edge of the distribution
1024 of a cold-water fish. *Hydrology and Earth System Sciences*, 21(8), 4073–4101.
1025 <https://doi.org/10.5194/hess-21-4073-2017>
- 1026 Segura, C., Caldwell, P., Sun, G., McNulty, S., & Zhang, Y. (2015). A model to predict stream
1027 water temperature across the conterminous USA. *Hydrological Processes*, 29(9), 2178–2195.
1028 <https://doi.org/10.1002/hyp.10357>
- 1029 Siegel, J. E., & Volk, C. J. (2019). Accurate spatiotemporal predictions of daily stream
1030 temperature from statistical models accounting for interactions between climate and landscape.
1031 *PeerJ*, 7, e7892. <https://doi.org/10.7717/peerj.7892>
- 1032 Sinokrot, B. A., & Gulliver, J. S. (2000). In-stream flow impact on river water temperatures.
1033 *Journal of Hydraulic Research*, 38(5), 339–349. <https://doi.org/10.1080/00221680009498315>
- 1034 Sohrabi, M. M., Benjankar, R., Tonina, D., Wenger, S. J., & Isaak, D. J. (2017). Estimation of
1035 daily stream water temperatures with a Bayesian regression approach. *Hydrological Processes*,
1036 31(9), 1719–1733. <https://doi.org/10.1002/hyp.11139>
- 1037 Soto, B. (2016). Assessment of Trends in Stream Temperatures in the North of the Iberian
1038 Peninsula Using a Nonlinear Regression Model for the Period 1950–2013. *River Research and*
1039 *Applications*, 32(6), 1355–1364. <https://doi.org/10.1002/rra.2971>
- 1040 Stanford, J.A. & Ward, J.V. (1992). Management of Aquatic Resources in Large Catchments:
1041 Recognizing Interactions Between Ecosystem Connectivity and Environmental Disturbance. In
1042 R. J. Naiman (Editor). *Watershed Management: Balancing Sustainability and Environmental*
1043 *Change* (pp. 91–124). New York, NY: Springer.
- 1044 St-Hilaire, A., Boyer, C., Bergeron, N., & Daigle, A. (2018). Water temperature monitoring in
1045 Eastern Canada: A case study for network optimization. *WIT Transactions on Ecology and the*
1046 *Environment*, 228, 269–275. <https://doi.org/10.2495/WP180251>
- 1047 Steel, E. A., Kennedy, M. C., Cunningham, P. G., & Stanovick, J. S. (2013). Applied statistics in
1048 ecology: Common pitfalls and simple solutions. *Ecosphere*, 4(9), art115.
1049 <https://doi.org/10.1890/ES13-00160.1>
- 1050 Steel, E. A., Tillotson, A., Larsen, D. A., Fullerton, A. H., Denton, K. P., & Beckman, B. R.
1051 (2012). Beyond the mean: The role of variability in predicting ecological effects of stream
1052 temperature on salmon. *Ecosphere*, 3(11), 1–11. <https://doi.org/10.1890/ES12-00255.1>
- 1053 Stenhouse, S., Pisano, M., Bean, C., & Chesney, W. (2012). Water temperature thresholds for
1054 coho salmon in a spring fed river, Siskiyou County, California. *California Fish and Game*, 98(1),
1055 19–37.
- 1056 Superior Court of Siskiyou County (1980). Scott River Adjudication, Decree No. 30662. Scott
1057 River Stream System, Siskiyou County. Sacramento, CA: State Water Resources Control Board.

- 1058 Sutton, R. J., Deas, M. L., Tanaka, S. K., Soto, T., & Corum, R. A. (2007). Salmonid
1059 observations at a Klamath River thermal refuge under various hydrological and meteorological
1060 conditions. *River Research and Applications*, 23(7), 775–785.
- 1061 Sutton, R., & Soto, T. (2012). Juvenile coho salmon behavioural characteristics in Klamath river
1062 summer thermal refugia. *River Research and Applications*. <https://doi.org/10.1002/rra.1459>
- 1063 Tan, J., & Cherkauer, K. A. (2013). Assessing stream temperature variation in the Pacific
1064 Northwest using airborne thermal infrared remote sensing. *Journal of Environmental*
1065 *Management*, 115, 206–216. <https://doi.org/10.1016/j.jenvman.2012.10.012>
- 1066 Toffolon, M., & Piccolroaz, S. (2015). A hybrid model for river water temperature as a function
1067 of air temperature and discharge. *Environmental Research Letters*, 10(11), 114011.
1068 <https://doi.org/10.1088/1748-9326/10/11/114011>
- 1069 Tolley, D., Foglia, L., & Harter, T. (2019). Sensitivity Analysis and Calibration of an Integrated
1070 Hydrologic Model in an Irrigated Agricultural Basin With a Groundwater-Dependent Ecosystem.
1071 *Water Resources Research*, 55(9), 7876–7901. <https://doi.org/10.1029/2018WR024209>
- 1072 Van Kirk, R. W., & Naman, S. W. (2008). Relative Effects of Climate and Water Use on Base-
1073 Flow Trends in the Lower Klamath Basin. *Journal of the American Water Resources*
1074 *Association*, 44(4), 1035–1052. <https://doi.org/10.1111/j.1752-1688.2008.00212.x>
- 1075 van Rij, J., Hendriks, P., van Rijn, H., Baayen, R. H., & Wood, S. N. (2019). Analyzing the Time
1076 Course of Pupillometric Data. *Trends in Hearing*, 23, 233121651983248.
1077 <https://doi.org/10.1177/2331216519832483>
- 1078 van Rij, J., Wieling, M., Baayen, R., van Rijn, H. (2020). *itsadug: Interpreting Time Series and*
1079 *Autocorrelated Data Using GAMMs*. R package version 2.4. [https://cran.r-](https://cran.r-project.org/package=itsadug)
1080 [project.org/package=itsadug](https://cran.r-project.org/package=itsadug)
- 1081 van Vliet, M. T. H., Ludwig, F., Zwolsman, J. J. G., Weedon, G. P., & Kabat, P. (2011). Global
1082 river temperatures and sensitivity to atmospheric warming and changes in river flow. *Water*
1083 *Resources Research*, 47(2). <https://doi.org/10.1029/2010WR009198>
- 1084 Wanders, N., Vliet, M. T. H. van, Wada, Y., Bierkens, M. F. P., & Beek, L. P. H. (Rens) van.
1085 (2019). High-Resolution Global Water Temperature Modeling. *Water Resources Research*,
1086 55(4), 2760–2778. <https://doi.org/10.1029/2018WR023250>
- 1087 Webb, B. W., Clack, P. D., & Walling, D. E. (2003). Water–air temperature relationships in a
1088 Devon river system and the role of flow. *Hydrological Processes*, 17(15), 3069–3084.
1089 <https://doi.org/10.1002/hyp.1280>
- 1090 Webb, B. W., Hannah, D. M., Moore, R. D., Brown, L. E., & Nobilis, F. (2008). Recent
1091 advances in stream and river temperature research. *Hydrological Processes*, 22(7), 902–918.
1092 <https://doi.org/10.1002/hyp.6994>
- 1093 Webb, B. W., & Walling, D. E. (1993). Temporal variability in the impact of river regulation on
1094 thermal regime and some biological implications. *Freshwater Biology*, 29(1), 167–182.
1095 <https://doi.org/10.1111/j.1365-2427.1993.tb00752.x>
- 1096 Welsh, H. H., Hodgson, G. R., Harvey, B. C., & Roche, M. F. (2001). Distribution of Juvenile
1097 Coho Salmon in Relation to Water Temperatures in Tributaries of the Mattole River, California.

- 1098 *North American Journal of Fisheries Management*, 21(3), 464–470.
1099 [https://doi.org/10.1577/1548-8675\(2001\)021<0464:DOJCSI>2.0.CO;2](https://doi.org/10.1577/1548-8675(2001)021<0464:DOJCSI>2.0.CO;2)
- 1100 Wenger, S. J., Isaak, D. J., Luce, C. H., Neville, H. M., Fausch, K. D., Dunham, J. B., Dauwalter,
1101 D. C., Young, M. K., Elsner, M. M., Rieman, B. E., Hamlet, A. F., & Williams, J. E. (2011).
1102 Flow regime, temperature, and biotic interactions drive differential declines of trout species
1103 under climate change. *Proceedings of the National Academy of Sciences*, 108(34), 14175–14180.
1104 <https://doi.org/10.1073/pnas.11030971108>
- 1105 Wondzell, S. M., Diabat, M., & Haggerty, R. (2019). What Matters Most: Are Future Stream
1106 Temperatures More Sensitive to Changing Air Temperatures, Discharge, or Riparian Vegetation?
1107 *Journal of the American Water Resources Association*, 55(1), 116–132.
1108 <https://doi.org/10.1111/1752-1688.12707>
- 1109 Wood, S.N. (2017). *Generalized Additive Models: An Introduction with R (2nd edition)*.
1110 Chapman and Hall/CRC.
- 1111 Yang, G., & Moyer, D. L. (2020). Estimation of nonlinear water-quality trends in high-frequency
1112 monitoring data. *Science of The Total Environment*, 715, 136686.
1113 <https://doi.org/10.1016/j.scitotenv.2020.136686>
- 1114 Yard, M. D., Bennett, G. E., Mietz, S. N., Coggins, L. G., Stevens, L. E., Hueftle, S., & Blinn, D.
1115 W. (2005). Influence of topographic complexity on solar insolation estimates for the Colorado
1116 River, Grand Canyon, AZ. *Ecological Modelling*, 183(2), 157–172.
1117 <https://doi.org/10.1016/j.ecolmodel.2004.07.027>
- 1118 Yarnell, S. M., Stein, E. D., Webb, J. A., Grantham, T., Lusardi, R. A., Zimmerman, J., Peek, R.
1119 A., Lane, B. A., Howard, J., & Sandoval-Solis, S. (2020). A functional flows approach to
1120 selecting ecologically relevant flow metrics for environmental flow applications. *River Research*
1121 *and Applications*, rra.3575. <https://doi.org/10.1002/rra.3575>
- 1122 Zhu, S., Heddam, S., Nyarko, E. K., Hadzima-Nyarko, M., Piccolroaz, S., & Wu, S. (2018).
1123 Modeling daily water temperature for rivers: Comparison between adaptive neuro-fuzzy
1124 inference systems and artificial neural networks models. *Environmental Science and Pollution*
1125 *Research*. <https://doi.org/10.1007/s11356-018-3650-2>
- 1126 Zhu, S., & Piotrowski, A. P. (2020). River/stream water temperature forecasting using artificial
1127 intelligence models: A systematic review. *Acta Geophysica*, 68(5), 1433–1442.
1128 <https://doi.org/10.1007/s11600-020-00480-7>
- 1129 Zillig, K. W., Lusardi, R. A., Moyle, P. B., & Fanguie, N. A. (2021). One size does not fit all:
1130 Variation in thermal eco-physiology among Pacific salmonids. *Reviews in Fish Biology and*
1131 *Fisheries*. <https://doi.org/10.1007/s11160-020-09632-w>