

Structure of Urban Landscape and Surface Temperature: a Case Study in Philadelphia, PA

Erik Mitz¹, Peleg Kremer^{2*}, Neele Larondelle³, Justin Stewart^{2,4}

¹ Department of Political Science, Villanova University, Villanova, Pennsylvania, United States of America; E-mail: emitz@villanova.edu

² Department of Geography and the Environment, Villanova University, Villanova, Pennsylvania, United States of America; E-mail: peleg.kremer@villanova.edu

³ Institute of Geography, Humboldt Universität zu Berlin, Berlin, Germany; E-mail: n.larondelle@gmail.com

⁴ Department of Ecological Science, Vrije Universiteit Amsterdam, 1081 HV Amsterdam, Netherlands; E-mail: justin618s@gmail.com

Abstract

Discerning the relationship between urban structure and function is crucial for sustainable city planning and requires examination of how components in urban systems are organized in three-dimensional space. The Structure of Urban Landscape (STURLA) classification accounts for the compositional complexity of urban landcover structures including the built and natural environment. Building on previous research, we develop a STURLA classification for Philadelphia, PA and study the relationship between urban

structure and land surface temperature. Finally, we evaluate the results in Philadelphia as compared to previous case studies in Berlin, Germany and New York City, USA. In Philadelphia, STURLA classes hosted ST that were unique and significantly different as compared to all other classes. We find a similar distribution of STURLA class composition across the three cities, though NYC and Berlin showed strong correlation with each other but not with Philadelphia. Our research highlights the use of STURLA classification to capture a physical property of the urban landscape.

Key Words

Urban Landscape, Urban surface temperature, STURLA, Urban structure, city comparison

Introduction

Urban spatial structure is important to understanding urban social-ecological interactions and provides a bridge to planning sustainable cities (Zhou et al., 2017). Urban structure characteristics including vegetation and other landcover classes influence, and can be used to estimate ecological functions (Bastian et al., 2014; van Oudenhoven et al., 2012). However, defining urban structure and key relationships between structure and ecological processes is challenging in landscapes characterized by variable density and patchy spatial patterns (Pickett & Cadenasso, 2008).

While it is well established that urban areas host ecological communities subject to unique stressors (Jones & Harrison, 2004; Joyner et al., 2019; Reese et al., 2016) absent in natural systems (e.g pollution, high population density), the influence of landscape heterogeneity is currently unknown. Functional classification of urban structure is necessary for understanding the nature of social and ecological relationships in urban areas (Cadenasso

et al., 2007; McPhearson et al., 2016; Zhou et al., 2014). Over the last decade, fine scale landcover classification for selected urban areas have been developed (MacFaden et al., 2012; Pickard et al., 2015) that allows more nuanced understanding of urban landcover. While some functional classification approaches have been suggested (see for example Cadenasso et al., 2007), still major challenges remain in integration of spatial structure and configuration that allows automated and unbiased analysis of fine scale relationships between urban form and process.

A major barrier for understanding the relationship between urban structure and environmental function is the lack of independent measurement of the fine-scale spatial variability of the distribution of environmental and ecological variables. Particularly important is the vertical dimension and variation of the three-dimensional landscape that is rarely addressed (Alavipanah et al., 2017) in ecological studies. Where independent measurements exist, such as data from Environmental Protection Agency (EPA) air pollution monitoring stations or United States Geographical Survey (USGS) water monitoring sites, the spatial distribution is not sufficient to allow intra-urban analysis. Surface temperature is one example of a physical property of the urban environment. Landsat thermal bands have been used in research addressing landcover (Zhou et al., 2011), urban heat island (Rosenzweig et al., 2009; Zhao et al., 2011), and urban ecosystem services (Schwarz et al., 2011). Likewise, ST structures patterns of taxonomic and functional biodiversity (Scherrer & Körner, 2011; Zogg et al., 1997), hydrology (Reyes et al., 2018), air quality (Li et al., 2018; Sillman & Samson, 1995), and social variables relevant for studies of environmental injustice (Huang & Cadenasso, 2016; Zhang et al., 2017). Thus we use ST as a proxy for a wide range of potential variables of interest.

To account for the heterogenous vertical dimension of the built environment in urban landscape in a reproducible and scalable way, we employ STURLA classification

(Hamstead et al., 2016). STURLA has identified patterns of microbial biogeography in the atmosphere of Philadelphia (J. Stewart et al., 2020), and ST in NYC (Hamstead et al., 2016) and Berlin (Kremer et al., 2018). In summary, the urban landscape is characterized as a discrete number composite landclasses that characterize the natural and built environment in Philadelphia, PA, USA. The city is one of the poorest cities in the US, with 26 percent of its population living in poverty (PEW, 2017). It is also one of the most segregated cities in the US, with African American and Asian populations concentrated in neighborhoods in West and North Philadelphia respectively (The Brookings Institution, 2003). The city's population peaked in 1950 with over 2 million people, and was declining until 2010 when it started growing again. Recently, Philadelphia is experiencing strong, yet uneven economic resurgence reflected in job growth and rising housing prices (PEW, 2017).

Philadelphia's urban structure emerged through the evolution of its original plan, laid out by William Penn in 1643. It has a gridded layout with mostly low and mid-rise residential buildings. A long time "gentleman's agreement" kept Penn's statue on top of city hall as the highest building in the city, preventing high-rise development for decades until the 1980s. The most common residential structures in the city are rowhouses. Rowhouses commonly occupy a narrow street frontage and are attached to other homes on both sides (Simmons Schade et al., 2008). Aside from the built environment, green space in the city includes 19% tree cover and 24% grass-shrub cover that are distributed unevenly across the city with some neighborhoods densely vegetated and others with little to no green space (O'Neil-Dunne, 2011). Part of the city's sustainability plan, Greenworks Philadelphia, includes a goal of tree canopy cover of 30% in all city neighborhoods by 2025 (City of Philadelphia, 2015a). However, until recently, the only publicly available data for a comprehensive analysis of the city's green space has been NLCD landuse-landcover datasets that do not have the spatial resolution and functional categories required to identify small and fragmented patches of

green in the city. In 2011, a fine scale dataset of Philadelphia landcover has been released (City of Philadelphia, 2011) that is used here as the basis for the STURLA. Empirical evidence from two cities, Berlin and New York City (NYC), were compared (Larondelle et al., 2014) and more detailed analysis of within class and neighborhood effects were performed in a Berlin case study (Kremer et al., 2018).

The objectives of this short study were to identify if STURLA could explain the variation of urban structure in a new model city (Philadelphia), and quantify this variation using a physical property of the environment (ST). Results suggest STURLA identifies common urban structure units that encompass the majority of the variation in the urban landscape structure. Moreover, when correlated to surface temperature, these common urban structure classifications exhibit distinct temperature signatures for different urban structure units with temperature trends dramatically similar between Berlin and NYC. Here, we contribute to the developing literature on the urban structure-function relationship using STURLA by adding a third case study city of different , Philadelphia, and comparing the results to previous studies.

Materials and methods

Study area

Philadelphia is the sixth largest city in the nation with a city population of 1.6 million inhabitants (U.S. Census Bureau, 2016) and hosts an average population density of 30,297 inhabitants per square kilometer. It is located at the confluence of the Delaware and Schuylkill rivers on the eastern border of Pennsylvania with the Appalachian Mountains to the west and the Atlantic Ocean to the east. The city has a total area of about 370 km² of which 350 km² are land and the rest, water.

127 *Pre-processing urban landscape structure data*

128 To construct the urban structure dataset, we used a 2008 1-meter resolution land cover
129 dataset (City of Philadelphia, 2011), The Property Assessment dataset from the Philadelphia
130 Office of Property Assessment (City of Philadelphia, 2015b) indicating number of floors in
131 buildings for each tax lot in the city in tabular format, and the Philadelphia Department of
132 Water parcels dataset. We joined the property assessment tabular data to the parcels dataset
133 using unique parcel IDs and created a 1-meter resolution raster dataset from the Number of
134 Floors field in the Property Assessment dataset. Number of floors were classified into three
135 categories: lowrise (1–3 stories), midrise (4–9 stories) and highrise (>9 stories) (Larondelle et
136 al., 2014; I. D. Stewart & Oke, 2012). We then combined it with the land cover raster dataset,
137 by replacing all building land cover pixels with a value representing building height category
138 to create our basic urban structure dataset.

139

140 *Constructing the STURLA classification*

141 We constructed a 120 m grid aligned to the Landsat surface temperature dataset and
142 derived STURLA classes as the presence of all land cover and building height types that fell
143 within each grid cell. Following Hamstead et al. (2016) we used a zonal statistics tabulate
144 area operation to compute the area of each land cover or building height category within each
145 cell. Finally, we generated and assigned a STURLA class variable for each grid cell.

146

147 *Comparison of STURLA classification results from current and previous studies*

148 Permutational t-tests with Bonferroni correction were used to test for differences
149 between cities in STURLA classes. The permutational t-test was used because we test data
150 representing the population rather than a sample. The null hypothesis of the permutational t-
151 test is that STURLA class composition does not differ between the cities. Permutational

Pearson correlations were conducted to determine if the cities distribution of STURLA classes were similar between cities. These tests were conducted in R using the package “RVAideMemoire” (Hervé, 2020).

Surface Temperature Processing

Surface temperature was obtained from Landsat 7 thermal band 6(1). We obtained monthly composite data for the month of July 2010 from the Global Web-enabled Landsat Data (WELD) website. Each monthly composite image is normally a composite of two Landsat scenes because LANDSAT returns to any single location every 16 days. Using a composite scene helps address the Landsat 7 scan line corrector error. WELD data is terrain-corrected and radiometrically calibrated Landsat data (Roy et al., 2010). Top-of the - Atmosphere reflectance was converted to surface temperature followed the methodology detailed in Kremer et al. (2018) in processing surface temperature.

Analysis of class surface temperature

We computed the mean, min, max and standard deviation of surface temperature pixels that fell within each cell of the STURLA grid using zonal statistics (Table 1) and joined these results with the STURLA class variable. Averaging was necessary because Landsat 7 thermal bands are resampled to 30 meters for distribution (Roy et al., 2010) while the STURLA grid is 120 m. Thus, we averaged sixteen 30 m pixels that fell within each 120 m cell. Similar to Hamstead et al. (2016) and Larondelle et al. (2014) we focused the class temperature analysis on the most frequently occurring classes, which cumulatively comprise 90% of the city’s land area. As done with comparison of STURLA classes between cities, permutational t-tests with Bonferroni correction were employed to test significance. Likewise, the null hypothesis of the permutational t-test is that ST does not differ between the STURLA classes.

Results

The most prevalent composite class in Philadelphia contains trees, grass, paved surfaces, and low rise buildings ('tgpl') (Table 1). The 'tgpl' class accounts for about 57% of total city area and can be found in all parts of the city and was largely homogenous in spatial distribution (Figure 1A). The second largest class, 'tgplm' at 8.5% of the area, which is similar to 'tgpl' except it includes midrise buildings, is concentrated in the center of the city and along a few main corridors to the North and West. STURLA classes were able to identify the role of urban structure influencing ST (Figure 1B). Classes generally hosted ST that were unique (Figure 1B) and significantly different (Table 2) compared to all other classes with the exception of 'tgbp' with similar ST values to 'tgwp' and 'tgwpl'.

The prevalence and distribution of the STURLA classes in Philadelphia differs from what we found in previous studies of urban structure NYC and Berlin (Figure 2). In Berlin and NYC, ~1/3 of the landscape can be explained by one highly composite STURLA class. Another difference between the results in Philadelphia and previous studies is the number of classes that cumulatively explain 90% of the area of the city. Ten classes covered 90% of the area of Philadelphia while the same number of classes only covered 79% of the area of New York City and 68% of the area in Berlin. Despite these differences, pairwise comparison of each city revealed that STURLA class proportions were not significantly different between the cities (all $p > 0.05$). Still, Berlin and NYC were highly correlated ($r^2 = 0.952$, $p < 0.05$) while Philadelphia remained unassociated to the other cities (both $r^2 > 0.1$, $p > 0.05$).

Due to the compositional nature of a STURLA cell where the relative proportions of all elements sum to one Figure 2 shows provides an example compositional variability within the most common class in Philadelphia 'tgpl' using six grid cells taken from a larger city-

wide random sample. The different grid cells and corresponding satellite imagery show the different types of buildings and proportion of each element of the class, trees, grass, paved surfaces, and low-rise buildings, can vary greatly from one another but still fall into the class. Most grid cells from the ‘tgpl’ class show row houses or single-family detached houses since they fall within the size parameters of low-rise buildings (1-3 stories).

Discussion:

One of the main limitations of STURLA classification is the presence/absence nature of class assignment. If the STURLA grid were shifted it would change the relative proportions of the within class elements (e.g. trees decrease). Despite this variation, STURLA classes are a discrete countable number and have a Poisson distribution. Thus, the ranked order abundances of different STURLA classes should not vary in the most frequent classes. For example, since ‘tgpl’ is common in Philadelphia, a reduction in a large number of ‘tgpl’ classes in the city would be relatively less influential than additions/reductions of an uncommon class.

STURLA captured urban structure and characterized the physical property of ST in Philadelphia as previously done in NYC (Hamstead et al., 2016) and Berlin (Kremer et al., 2018), despite variation in size, demography, and historical planning. This suggests that urban areas may be subject to similar processes that result in between city-redundant spatial organizations (Votsis & Haavisto, 2019). Likewise, STURLA may be suited for understanding urban biogeography, environmental justice, and city planning for a sustainable future. Global analyses of cities may also identify clusters of urban areas that would benefit from similar management practices. Likewise, STURLA offers a computationally inexpensive alternative to network analyses of urban structure (Zhong et al., 2014).

227

228 **Conclusion**

229 In this paper we demonstrate the application of STURLA classification to quantify the
 230 relationship between urban structure and surface temperature in Philadelphia. We show it can
 231 be applied to cities with different historical patterns of growth in a reproducible manner.
 232 Furthermore, patterns in class abundance and composition can be used to determine the
 233 surface temperature signature of a composite landscape. Additional research is needed to
 234 compare cities of vastly different urban structure and identify patterns in the relationship
 235 between urban structure with social and ecological properties of the environment.
 236 Understanding general urban structure-environmental function relationships will help build
 237 tools for effective urban planning and management under global change scenarios.

238

239 Table 1: 10 most common STURLA classes in Philadelphia and their ST statistics. STURLA
 240 class codes: t-trees; g-grass; b-bare soil; w-water; p-paved; l-low building; m-medium
 241 building

<i>Class</i>	<i>%</i>	<i>of</i>	<i>%</i>	<i>Mean ST C</i>	<i>Min ST C</i>	<i>Max ST C</i>
	<i>total</i>		<i>cumulative</i>			
tgpl	57.44		57.44	26.95	25.01	28.79
tgpl	8.55		65.99	27.95	25.89	29.93
m						
tgp	7.39		73.37	23.86	22.10	25.75
tgwp	4.36		77.73	22.72	20.77	24.75
w	2.92		80.65	18.34	17.85	19.03
tgwp	2.57		83.22	24.83	22.41	27.29

<i>l</i>					
<i>tgbp</i>	2.46	85.69	26.31	24.16	28.60
<i>l</i>					
<i>tg</i>	1.94	87.63	20.42	19.37	21.62
<i>tgw</i>	1.42	89.05	20.37	19.16	21.69
<i>tgbp</i>	1.29	90.34	24.68	22.81	26.64

242

243

244 Table 2. P-values with Bonferroni correction from pairwise permutational t-tests (n=999) of
 245 ST values for the top ten STURLA classes. Bold values indicate statistical significance
 246 (p<0.05).

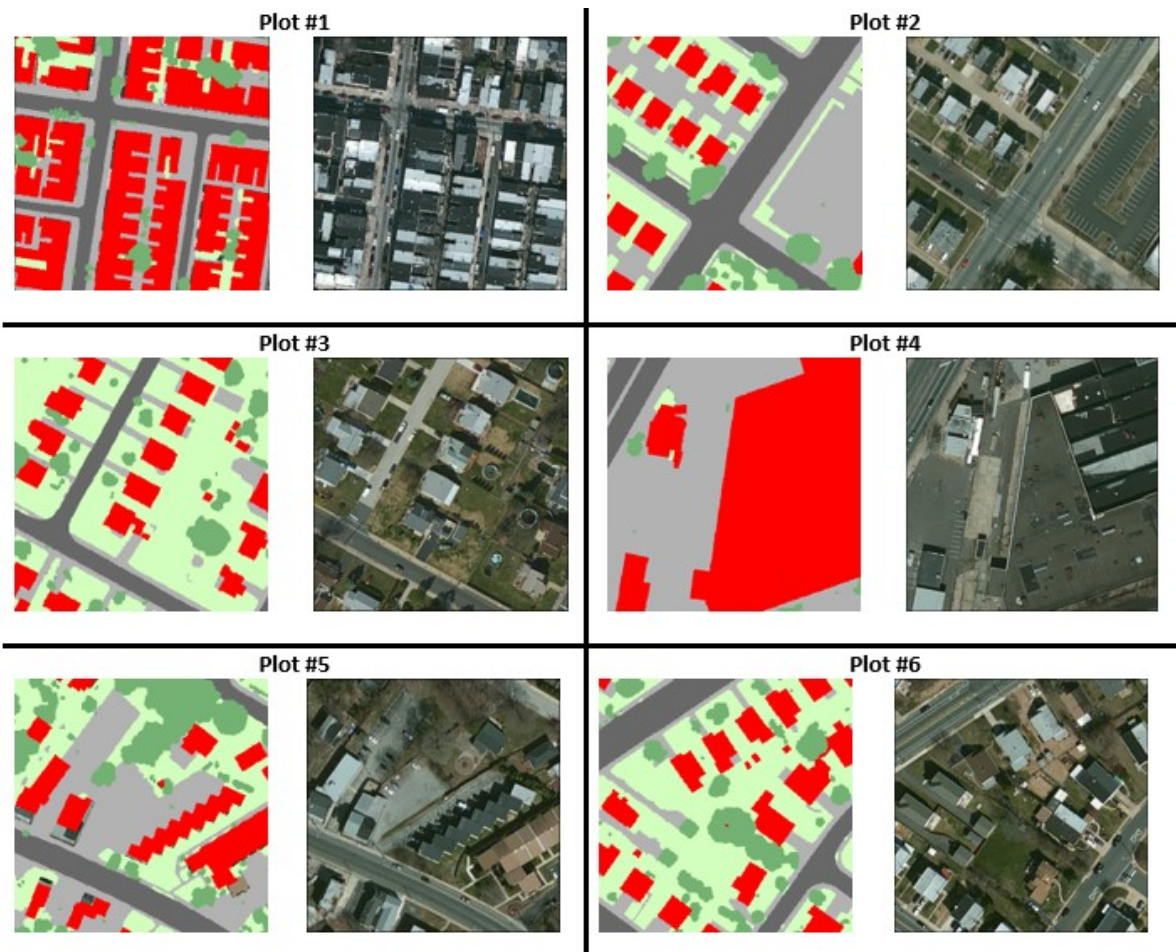
247

<i>Class</i>	<i>tgpl</i>	<i>tgpl</i>	<i>tgpl</i>	<i>tgw</i>	<i>w</i>	<i>tgw</i>	<i>tgbp</i>	<i>tg</i>	<i>tgw</i>	<i>tgw</i>
		<i>m</i>		<i>p</i>		<i>pl</i>	<i>l</i>			<i>p</i>
<i>tgpl</i>	0									
<i>tgpl</i>	0.02	0								
<i>m</i>										
<i>tgpl</i>	0.02	0.02	0							
<i>tgwp</i>	0.02	0.02	0.02	0						
<i>w</i>	0.02	0.02	0.02	0.02	0					
<i>tgwp</i>	0.02	0.02	0.02	0.02	0.02	0				
<i>l</i>										
<i>tgbp</i>	0.02	0.02	0.02	0.02	0.02	0.02	0			

<i>l</i>										
<i>tg</i>	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0	
<i>tgw</i>	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0.02	0	
<i>tgbp</i>	0.02	0.02	0.02	3.74	0.02	4.02	0.02	0.02	0.02	0

248

249



257

258 Figure 2: Example of the composition of STURLA grid cells of the most common STURLA
 259 class in Philadelphia 'tgp1'. STURLA 'tgp1' cells are shown next to corresponding areal
 260 imagery.

261

262 References

- 263 Alavipanah, S., Haase, D., Lakes, T., & Qureshi, S. (2017). Integrating the third dimension
 264 into the concept of urban ecosystem services: A review. In *Ecological Indicators* (Vol.
 265 72, pp. 374–398). <https://doi.org/10.1016/j.ecolind.2016.08.010>
- 266 Bastian, O., Grunewald, K., Syrbe, R.-U., Walz, U., & Wende, W. (2014). Landscape
 267 services: the concept and its practical relevance. *Landscape Ecology*, 29(9), 1463–1479.

268 <https://doi.org/10.1007/s10980-014-0064-5>

269 Cadenasso, M. L., Pickett, S. T. A., & Schwarz, K. (2007). Spatial heterogeneity in urban
 270 ecosystems: reconceptualizing land cover and a framework for classification. *Frontiers*
 271 *in Ecology and the Environment*, 5(2), 80–88. [https://doi.org/10.1890/1540-9295\(2007\)5](https://doi.org/10.1890/1540-9295(2007)5)

272 City of Philadelphia. (2011). *Philadelphia Land Cover Raster*. OpenDataPhilly.

273 City of Philadelphia. (2015a). *Greenworks Philadelphia: 2015 Progress Report*.

274 City of Philadelphia. (2015b). *Property Assessment Data*.

275 Hamstead, Z. A., Kremer, P., Larondelle, N., McPhearson, T., & Haase, D. (2016).
 276 Classification of the heterogeneous structure of urban landscapes (STURLA) as an
 277 indicator of landscape function applied to surface temperature in New York City. *Ecol.*
 278 *Indic.*, 70, 574–585. <https://doi.org/10.1016/j.ecolind.2015.10.014>

279 Hervé, M. (2020). *Package ‘RVAideMemoire.’*

280 Huang, G., & Cadenasso, M. L. (2016). People, landscape, and urban heat island: dynamics
 281 among neighborhood social conditions, land cover and surface temperatures. *Landscape*
 282 *Ecology*. <https://doi.org/10.1007/s10980-016-0437-z>

283 Jones, A. M., & Harrison, R. M. (2004). The effects of meteorological factors on atmospheric
 284 bioaerosol concentrations—a review. *Sci. Total Environ.*, 326(1), 151–180.
 285 <https://doi.org/10.1016/j.scitotenv.2003.11.021>

286 Joyner, J. L., Kerwin, J., Deeb, M., Lozefski, G., Prithiviraj, B., Paltseva, A., McLaughlin, J.,
 287 Groffman, P., Cheng, Z., & Muth, T. R. (2019). Green Infrastructure Design Influences
 288 Communities of Urban Soil Bacteria. In *Frontiers in Microbiology* (Vol. 10).
 289 <https://doi.org/10.3389/fmicb.2019.00982>

290 Kremer, P., Larondelle, N., Zhang, Y., Pasles, E., & Haase, D. (2018). Within-Class and
 291 Neighborhood Effects on the Relationship between Composite Urban Classes and
 292 Surface Temperature. *Sustainability*, 10(3), 645. <https://doi.org/10.3390/su10030645>

293 Larondelle, N., Hamstead, Z. A., Kremer, P., Haase, D., & McPhearson, T. (2014). Applying
 294 a novel urban structure classification to compare the relationships of urban structure and
 295 surface temperature in Berlin and New York City. *Appl. Geogr.*, 53, 427–437.
 296 <https://doi.org/10.1016/j.apgeog.2014.07.004>

297 Li, H., Meier, F., Lee, X., Chakraborty, T., Liu, J., Schaap, M., & Sodoudi, S. (2018).
 298 Interaction between urban heat island and urban pollution island during summer in
 299 Berlin. *Science of the Total Environment*.
 300 <https://doi.org/10.1016/j.scitotenv.2018.04.254>

301 MacFaden, S. W., O’Neil-Dunne, J. P. M., Royar, A. R., Lu, J. W. T., & Rundle, A. G.
 302 (2012). High-resolution tree canopy mapping for New York City using LIDAR and
 303 object-based image analysis. *Journal of Applied Remote Sensing*, 6(1), 063567.
 304 <https://doi.org/10.1117/1.JRS.6.063567>

305 McPhearson, T., Pickett, S. T. A. A., Grimm, N. B., Niemelä, J., Alberti, M., Elmqvist, T.,
 306 Weber, C., Haase, D., Breuste, J., & Qureshi, S. (2016). Advancing Urban Ecology
 307 toward a Science of Cities. *BioScience*, biw002. <https://doi.org/10.1093/biosci/biw002>

308 O’Neil-Dunne, J. (2011). *A Report on the City of Philadelphia’s Existing and Possible Tree*
 309 *Canopy*.

310 PEW. (2017). *Philadelphia 2017: The State of the City*.

311 Pickard, B. R., Daniel, J., Mehaffey, M., Jackson, L. E., & Neale, A. (2015). EnviroAtlas: A
 312 new geospatial tool to foster ecosystem services science and resource management.
 313 *Ecosystem Services*, 14, 45–55. <https://doi.org/10.1016/j.ecoser.2015.04.005>

314 Pickett, S. T. A., & Cadenasso, M. L. (2008). Linking ecological and built components of
 315 urban mosaics: an open cycle of ecological design. *Journal of Ecology*, 96, 8–12. <https://doi.org/10.1111/j.1365-2745.2007.01310.x>

317 Reese, A. T., Savage, A., Youngsteadt, E., McGuire, K. L., Koling, A., Watkins, O., Frank,

318 S. D., & Dunn, R. R. (2016). Urban stress is associated with variation in microbial
 319 species composition—but not richness—in Manhattan. *ISME J.*, *10*(3), 751–760. <https://doi.org/10.1038/ismej.2015.152>
 320

321 Reyes, B., Hogue, T., & Maxwell, R. (2018). Urban irrigation suppresses land surface
 322 temperature and changes the hydrologic regime in semi-arid regions. *Water*
 323 *(Switzerland)*. <https://doi.org/10.3390/w10111563>
 324

325 Rosenzweig, C., Solecki, W. D., Cox, J., Hodges, S., Parshall, L., Lynn, B., Goldberg, R.,
 326 Gaffin, S., Slosberg, R. B., Savio, P., Watson, M., & Dunstan, F. (2009). Mitigating
 327 New York City's Heat Island: Integrating Stakeholder Perspectives and Scientific
 328 Evaluation. *Bulletin of the American Meteorological Society*, *90*(9), 1297–1312. <https://doi.org/doi:10.1175/2009BAMS2308.1>
 329

330 Roy, D. P., Ju, J., Kline, K., Scaramuzza, P. L., Kovalsky, V., Hansen, M., Loveland, T. R.,
 331 Vermote, E., & Zhang, C. (2010). Web-enabled Landsat Data (WELD): Landsat ETM+
 332 composited mosaics of the conterminous United States. *Remote Sensing of Environment*,
 333 *114*(1), 35–49. <https://doi.org/10.1016/j.rse.2009.08.011>
 334

335 Scherrer, D., & Körner, C. (2011). Topographically controlled thermal-habitat differentiation
 336 buffers alpine plant diversity against climate warming. *Journal of Biogeography*. <https://doi.org/10.1111/j.1365-2699.2010.02407.x>
 337

338 Schwarz, N., Bauer, A., & Haase, D. (2011). Assessing climate impacts of planning policies
 339 —An estimation for the urban region of Leipzig (Germany). *Environmental Impact*
 340 *Assessment Review*, *31*(2), 97–111. <https://doi.org/10.1016/j.eiar.2010.02.002>
 341

342 Sillman, S., & Samson, P. J. (1995). Impact of temperature on oxidant photochemistry in
 343 urban polluted rural and remote environments. *Journal of Geophysical Research*. <https://doi.org/10.1029/94jd02146>
 344

345 Simmons Schade, R., Architects, B., Spina, L. M., Farnham, J., Fine, A., Gallery, J. A.,

343 Gladstein, E., Gregorski, M., Hauck, P., Jastrzab, G., Koksuz, B., Michel, L., Nikolic,
 344 T., Placke, C., Sauer, R., Schaaf, D., Urek, A., Wilds, S., & Wilson, D. (2008).
 345 *Philadelphia Rowhouse Manual A PRACTICAL GUIDE FOR HOMEOWNERS*
 346 *SPECIAL THANKS TO Steering Committee.*
 347 Stewart, I. D., & Oke, T. R. (2012). Local Climate Zones for Urban Temperature Studies.
 348 *Bulletin of the American Meteorological Society*, 93(12), 1879–1900.
 349 <https://doi.org/doi:10.1175/BAMS-D-11-00019.1>
 350 Stewart, J., Kremer, P., Shakya, K. M., Conway, M., & Saad, A. (2020). Outdoor
 351 Atmospheric Microbial Diversity is Associated with Three-Dimensional Urban
 352 Landscape Structure and Differs from Indoor-Transit Systems. *BioRxiv*.
 353 <https://doi.org/10.1101/2020.06.17.157651>
 354 The Brookings Institution. (2003). *Philadelphia in Focus: A Profile from Census 2000*. The
 355 Brookings Institution.
 356 U.S. Census Bureau. (2016). *State and County Quick Facts*.
 357 van Oudenhoven, A. P. E., Petz, K., Alkemade, R., Hein, L., & de Groot, R. S. (2012).
 358 Framework for systematic indicator selection to assess effects of land management on
 359 ecosystem services. *Ecological Indicators*, 21, 110–122.
 360 <https://doi.org/10.1016/j.ecolind.2012.01.012>
 361 Votsis, A., & Haavisto, R. (2019). Urban DNA and Sustainable Cities: A Multi-City
 362 Comparison. *Frontiers in Environmental Science*, 7(JAN), 4.
 363 <https://doi.org/10.3389/fenvs.2019.00004>
 364 Zhang, X., Estoque, R. C., & Murayama, Y. (2017). An urban heat island study in Nanchang
 365 City, China based on land surface temperature and social-ecological variables.
 366 *Sustainable Cities and Society*. <https://doi.org/10.1016/j.scs.2017.05.005>
 367 Zhao, C., Fu, G., Liu, X., & Fu, F. (2011). Urban planning indicators, morphology and

climate indicators: A case study for a north-south transect of Beijing, China. *Building and Environment*, 46(5), 1174–1183. <https://doi.org/10.1016/j.buildenv.2010.12.009>

Zhong, C., Arisona, S. M., Huang, X., Batty, M., & Schmitt, G. (2014). Detecting the dynamics of urban structure through spatial network analysis. *International Journal of Geographical Information Science*. <https://doi.org/10.1080/13658816.2014.914521>

Zhou, W., Cadenasso, M., Schwarz, K., & Pickett, S. (2014). Quantifying Spatial Heterogeneity in Urban Landscapes: Integrating Visual Interpretation and Object-Based Classification. *Remote Sensing*, 6(4), 3369–3386. <https://doi.org/10.3390/rs6043369>

Zhou, W., Huang, G., & Cadenasso, M. L. (2011). Does spatial configuration matter? Understanding the effects of land cover pattern on land surface temperature in urban landscapes. *Landscape and Urban Planning*, 102(1), 54–63. <https://doi.org/10.1016/j.landurbplan.2011.03.009>

Zhou, W., Pickett, S. T. A., & Cadenasso, M. L. (2017). Shifting concepts of urban spatial heterogeneity and their implications for sustainability. *Landscape Ecology*, 32(1), 15–30. <https://doi.org/10.1007/s10980-016-0432-4>

Zogg, G. P., Zak, D. R., Ringelberg, D. B., White, D. C., MacDonald, N. W., & Pregitzer, K. S. (1997). Compositional and Functional Shifts in Microbial Communities Due to Soil Warming. *Soil Sci. Soc. Am. J.*, 61, 475–481. <https://doi.org/10.2136/sssaj1997.03615995006100020015x>