

1 **Improvements of Biogenic Emission Estimation in China by Using WRF-CLM4-**
2 **MEGAN Model**

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12 **Key Points:**

- 13 • Regional biogenic emissions were estimated within an ecological framework base on
14 plant-specific physiological parameters.
- 15 • Neglect of leaf-air temperature bias and shaded canopy resulted in a 23% decrease and a
16 102% increase in estimation, respectively.
- 17 • Simulated isoprene flux was generally within a factor of 2 of canopy-scale flux
18 measurements, indicating the good performance of this method.
- 19

20 **Abstract**

21 Biogenic emission models are developed on the foundation of leaf physiological processes and
22 driven by a set of physical and biological factors. To estimate emissions online, many studies
23 used weather forecasting models coupled with simple biogenic emission algorithms, in which the
24 canopy physiological parameters were neglected or oversimplified. In this study, the land surface
25 scheme CLM4 (Community Land Model version 4) coupled in the advanced Weather Research
26 and Forecasting model (WRF) was used to determine canopy physiological parameters. The
27 MEGAN (Model of Emissions of Gases and Aerosols from Nature) algorithms embedded in
28 CLM4 scheme used these parameters to estimate biogenic emissions. The emission estimated by
29 using leaf temperature in our study were about 23% higher than that based on air temperature as
30 in the previous methods. Compared with studies neglecting shaded canopy, the separate
31 treatments of sunlit and shaded leaves in this study lowered the estimations by a factor of 2
32 through decreasing diffuse radiation absorbed by sunlit canopy. Dynamic weather history was
33 used in our study to replace the fixed values in the original MEGAN-CLM4 code. An emission
34 inventory of isoprene and monoterpenes in China was established for the year 2018. The
35 estimates were evaluated against field measurements. Generally, the coupled model produced a
36 reasonable simulation in both emission budgets and spatiotemporal distribution of biogenic
37 emissions.

38 **Plain Language Summary**

39 The gas emission rate of vegetation depends on physiological conditions such as leaf
40 temperature, stomatal opening, and absorbed radiation. We estimated regional vegetation
41 emissions based on plant physiological parameters, which were neglected or oversimplified in
42 previous studies. Emission estimations based on leaf temperature in our study were 23% higher
43 than estimations based on air temperature in previous studies. Neglecting the shaded canopy
44 overestimated emissions by a factor of 2 compared with estimations in our study, which treated
45 sunlit and shaded leaves separately. The effects of weather history on emission rates were
46 considered in our study. Comparisons between simulated and measured emissions showed that
47 this method was able to estimate vegetation emissions reasonably.

48 **1 Introduction**

49 Globally speaking, biogenic volatile organic compounds (BVOCs) emitted by terrestrial
50 vegetation are estimated to be 500~1100 Tg C yr⁻¹, corresponding to about 90% of the emission
51 total (Guenther et al., 1995; Henrot et al., 2017). Many BVOC species are actively involved in
52 the atmospheric chemistry and have a substantial impact on tropospheric oxidation, aerosol
53 concentration, and the global carbon cycle (Fehsenfeld et al., 1992). BVOCs are therefore a
54 crucial component of the earth system and quantitative estimates of their emissions are required
55 for further exploring their impacts on regional and global atmospheric chemistry.

56 Many biogenic emission models are developed with a strong foundation in the physiological
57 processes of a leaf (Guenther et al., 1991; Niinemets et al., 1999). The Model of Emissions of
58 Gases and Aerosols from Nature (MEGAN), a model estimates BVOC emission fluxes as basal
59 emission rates modulated by emission activity factors, has been intensively used for regional and
60 global BVOC emission estimations (Guenther et al., 2006; Guenther et al., 2012). Process-based
61 models link BVOC production rate explicitly to leaf photosynthetic electron transport rate and
62 electron requirement for BVOC synthesis (Niinemets et al., 2002; Niinemets et al., 1999).

63 The MEGAN algorithms are incorporated into the terrestrial component of the earth climate
64 system model, Community Land Model (CLM) for online estimation. In the coupling of
65 MEGAN and CLM, physical and biological variables required by BVOC estimation are
66 determined by comprehensive ecological and physiological processes in CLM at each time step
67 (Lawrence et al., 2011; Levis et al., 2003). Process-based models are typically coupled within
68 dynamic vegetation models that have a mechanistic model for leaf photosynthesis at their core
69 (Arneth et al., 2007). These models require more computational time and resources, so it would
70 be too time-consuming for regional emission estimation.

71 Instead, many studies use weather forecasting models coupled with a simplified version of
72 MEGAN, the parameterized canopy emission activity (PCEEA) algorithm (Guenther et al., 2006;
73 Sakulyanontvittaya et al., 2008). Due to lack of a detailed canopy model which calculates leaf
74 temperature and leaf-level photosynthetic photon flux density (PPFD), the PCEEA algorithm
75 uses air temperature and canopy above solar radiation instead. The leaf temperature is affected
76 by air temperature, as well as other environmental and biological factors. Subin et al. (2011)
77 indicates that the strong advection and boundary layer mixing during the day decouples the air
78 temperature from the vegetation temperature to a great extent, making daytime surface energy
79 budget the primary controlling factors of vegetation temperature changes. Furthermore, due to
80 the different morphological and physiological properties, relationships between air temperature
81 and leaf temperature, and between canopy above PPFD and leaf-level PPFD, vary significantly
82 among tree species. Since the PCEEA algorithm was based on standard MEGAN canopy model
83 simulation for warm broadleaf forests, using the same equations for representations of other plant
84 types leads to unpredictable uncertainties. Therefore, reasonable plant-specific physiological
85 variables are needed to improve the BVOC estimation in weather models.

86 CLM version 4 (CLM4) was coupled and released with the Weather Research and Forecasting
87 model (WRF), a mesoscale numerical model designed to simulate regional weather and climate,
88 as one of the land surface scheme options (Jin et al., 2010). Because MEGAN has been
89 embedded within CLM as mentioned above, the coupling of WRF-CLM4-MEGAN allowed
90 weather forecasting models to estimate regional BVOC emissions within an ecological
91 framework. Besides improvements result from real-time plant physiological variables derived
92 from CLM4, sub-grid vegetation compositions represented in CLM4 are also expected to provide
93 a more reasonable estimation because of the significant variability in basal emission ability
94 among tree species. However, few studies employed the coupled model to estimate regional
95 BVOC emissions (Zhao et al., 2016).

96 Satellite data revealed a significant greening pattern in China from the year 2000 to 2017.
97 Approximately 42% of the greening in China was associated with forest expansion to mitigate
98 land degradation, air pollution, and climate change (Chen et al., 2019) (Chen et al., 2019).
99 Accurate estimations are needed to investigate the trend in BVOC emissions with changes in
100 land cover. In recent decades, many studies estimated national or regional BVOC emissions and
101 reported a wide emission range of 4.1~23.4 Tg C yr⁻¹ for isoprene and 1.8~5.6 Tg C yr⁻¹ for
102 monoterpenes on the national level (Fu and Liao, 2012; Klinger et al., 2002; Li et al., 2013; Liu
103 et al., 2018; Tie et al., 2006; Wu et al., 2020). However, these estimates are either calculated
104 offline or not fully based on plant physiological variables. In this study, we used the WRF-
105 CLM4-MEGAN coupled model to improve BVOC emission estimations in China. Two primary
106 classes of BVOCs, isoprene (C₅H₈) and monoterpenes (C₁₀H₁₆) (including α -pinene, β -pinene, 3-
107 carene, t- β -ocimene, limonene, sabinene, and myrcene), were considered in this study.

108 2 Methods and Data

109 2.1 CLM4 land surface scheme and coupling with MEGAN

110 The CLM4 was coupled and released with WRF since version 3.5 as one of the land surface
 111 scheme options. CLM4 consists of components related to land biogeophysics, hydrological
 112 cycle, biogeochemistry, human dimensions, and ecosystem dynamics. CLM4 includes 5 layers
 113 for snow, 10 layers for soil, and 1 layer for vegetation and is accurate in describing soil and
 114 vegetation processes (Jin and Wen, 2012; Lawrence et al., 2011; Subin et al., 2011).

115 The MEGAN model uses mechanistic algorithms to account for the major known process
 116 controlling biogenic emissions. MEGAN estimates emissions (F_i , $\mu\text{g C m}^{-2}$ ground area h^{-1}) of
 117 BVOC species i according to:

$$118 F_i = \gamma_i \times \sum \varepsilon_{i,j} \times \chi_j \quad (1)$$

119 where $\varepsilon_{i,j}$ ($\mu\text{g C m}^{-2} \text{h}^{-1}$) is the emission factor (EF) at standard conditions for PFT j with fraction
 120 coverage χ_j . PFT-specific EFs of isoprene were determined based on observations conducted in
 121 China and EF used in previous studies (as shown in Table S1). Due to lack of detailed
 122 monoterpene EFs reports, the EFs of main monoterpene species were determined by scaling
 123 default MEGAN EFs with the ratio of local isoprene EF to default value presented in Guenther et
 124 al. (2012)). The emission factors of each vegetation type used in this study were shown in Table
 125 S2.

126 The emission activity factor for each compound (γ_i) accounts for emission responses to solar
 127 radiation, leaf temperature, LAI, leaf age, and soil moisture. The effects of variations in CO_2
 128 concentration were neglected in this study. Details of the algorithms could be found in Guenther
 129 et al. (2006) and Guenther et al. (2012).

130 The coupling of CLM4-MEGAN improves the BVOC estimations through reasonable driving
 131 factors and detailed sub-grid representation, as briefly described below. We refer the reader to
 132 the description of Oleson et al. (2010) for the details of computations.

133 1. Leaf temperature

134 Variations in leaf temperature are influenced by net radiation absorbed/emitted by the vegetation
 135 and sensible and latent heat fluxes from vegetation. The two-stream approximation is applied to
 136 vegetation when calculating solar radiation reflected and absorbed by the canopy. Leaf
 137 temperatures are determined by the canopy energy balance equations. Due to the dependence of
 138 heat fluxes on vegetation temperature, the Newton-Raphson iteration is used to solve for folia
 139 temperature and the vegetation fluxes simultaneously.

140 2. Sunlit and shaded fractions of the canopy

141 The canopy in CLM4 is treated as sunlit and shaded leaves. Leaf fractions of different plant
 142 types are determined according to the leaf and stem area index and the solar zenith angle at each
 143 time step. CLM4 assumed that sunlit leaves receive the absorbed direct beam radiation and the
 144 absorbed diffuse radiation apportioned by f_{sun} (the sunlit fraction of the canopy), and that shaded
 145 leaves receive the absorbed diffuse radiation apportioned by f_{sha} (the shaded fraction).

146

147

148 3. The medium-term weather history

149 Current MEGAN algorithms use average leaf temperature, solar radiation, and leaf fractions over
150 the past time to account for the influence of medium-term (days to weeks) weather history.
151 CLM4 contains an accumulation module used to calculate the average of user-specified variables
152 over user-defined time intervals. However, the accumulation of past time leaf temperature and
153 PPF was commented out in the default CLM4 code. Instead, fixed values are assigned to those
154 coefficients based on conditions during previous days. After activating this module, a decrease in
155 average temperature and PPF with increasing simulation time was found. That was because
156 these two variables were not being accumulated but still being averaged over the total running
157 time. We corrected the accumulation code so that the average leaf temperature, PPF, and leaf
158 fraction are calculated at each time step.

159 4. Sub-grid heterogeneity

160 In CLM4, the surface heterogeneity is represented using a sub-grid tile approach in which grid
161 cells are composed of multiple land units (glacier, wetland, lake, urban and vegetated area),
162 snow/soil columns and plant functional types (PFTs). Vegetated surfaces are comprised of up to
163 4 plant functional types (PFTs). An explicit canopy layer represents the PFTs with specific leaf
164 and stem optical properties, root distribution parameters, aerodynamic parameters, and
165 photosynthetic parameters. The detailed representations of sub-grid improve the accuracy of land
166 surface parameterizations and reduce the uncertainty from plant distribution in BVOC estimation
167 (Schultz et al., 2016; Zhao et al., 2016).

168 2.2 Land surface datasets

169 In this study, MODIS datasets of land cover (MCD12Q1) for the year 2016 and water mask
170 (MOD44W) for the year 2015, both with a resolution of 500 m, were used to replace the
171 outdated United States Geological Survey (USGS) data used in default WRF initial static field.
172 The 17 MODIS land-use categories defined by the International Geosphere Biosphere Program
173 (IGBP) were mapped onto the 24 USGS categories. Default CLM4 prescribes the sub-grid PFT
174 composition for each land category in USGS, leading to geographical-invariant plant
175 distribution. We represented the sub-grid surface heterogeneity in this study as the composition
176 of 500 m-resolved land categories contained in a 12 km-resolved model grid. A total of 12
177 vegetation categories in IGBP were converted to 7 PFTs used in the CLM4 scheme. The
178 conversion of IGBP land cover into USGS and PFTs was illustrated in Table S3. The spatial
179 distribution of 7 PFTs was shown in Fig.S1 (the small islands in the South China Sea are not
180 included).

181 LAI data used in the default CLM4 scheme are updated daily by linearly interpolating between
182 prescribed monthly values. In this study, the MODIS LAI data derived from MCD15A2H
183 version 6 with a spatial resolution of 500 m and temporal resolution of 8 days was introduced to
184 CLM4. The seasonal and regional patterns of LAI were shown in Figure S2 (the small islands in
185 the South China Sea are not included). The sub-grid PFT-specific LAI was averaged over the
186 fraction of the land area covered by each PFT within the grid cell. The same LAI data was used
187 as current LAI (LAIc) for 8 days and the past 8-day image was considered as LAI of the previous
188 time step (LAIp). The changes between LAIc and LAIp was used to determine leaf age
189 (Guenther et al., 2006).

190 2.3 Numerical experiments

191 The simulations with the WRF version 4 were performed on a domain at 12 km horizontal
192 resolution covering China and its surrounding areas with 420×380 cells in the horizontal
193 direction and 35 layers in the vertical direction, extending from the surface to 50 hPa. The initial
194 meteorological fields and boundary conditions were from the 6 h NCEP (National Centers for
195 Environmental Prediction) global final analysis with a $1^\circ \times 1^\circ$ spatial resolution. The
196 meteorological fields were initialized at the start of each model run, which covered one month to
197 account for the effects of canopy climate history. We designed four scenarios to evaluate the
198 influence of parameter applications as follows: (1) BASE: standard configuration; (2) C1_T2:
199 replacing leaf temperature with the air temperature at 2 m height; (3) C2_FSUN: neglecting
200 shaded leaves; (4) C3_FIX: using fixed values for variables related to weather history. The
201 simulation time of the BASE case covered the entire year of 2018, while other cases were only
202 performed for July.

203 **3 Results and Discussions**

204 3.1 Evaluation of WRF output

205 Since the temperature and solar radiation exert primary control on BVOC emissions, we
206 evaluated the WRF model performance in simulating the air temperature at 2 m height (T2) and
207 the downward shortwave radiation (SWDOWN). The meteorology observations from 362 sites
208 and daily solar radiation observations from 100 sites in China were used for comparison. The in-
209 situ meteorology observations are provided by the National Climatic Data Center (NCDC,
210 <https://www.ncdc.noaa.gov/>) and radiation observations are derived from the National
211 Meteorological Information Center (<https://data.cma.cn/>). The statistical analyses for four
212 seasons are displayed in Table 1. The mean error (ME), mean bias (MB), correlation coefficient
213 (r), and root-mean-square error (RMSE) of hourly T2 series are 1.70, -0.52 , 0.98, 2.51°C ,
214 respectively. The r value in summer (0.90) is relatively lower than those in spring (0.95), autumn
215 (0.96), winter (0.97), and the simulation shows slight cooling bias in spring, autumn, and winter.
216 The ME, MB, r , RMSE values of the SWDOWN are 71.01, 69.82, 0.86, and 71.11 W m^{-2} . The
217 simulated SWDOWN was $\sim 45\%$ higher than measured data. Overestimated solar radiation is a
218 common issue of WRF model which could be attributed to neglect of radiation effect of aerosols
219 (Lu and Kueppers, 2012). The simulations of WRF-CLM4 successfully reproduced the temporal
220 and spatial patterns of T2 and SWDOWN (Figure S2). Generally, comparisons with observed
221 data indicate that WRF-CLM4 provide a good simulation on meteorological conditions that are
222 desirable for driving the MEGAN algorithm.

223 Table 1 *Verification Statistics of Air Temperature at 2 m Height (T2) and Downward Shortwave*
 224 *Radiation (SWDOWN).*

| Variable | Season | Mean | | ME ^b | MB ^b | r^b | RMSE ^b |
|--------------------------------|--------|-------------------|-------------------|-----------------|-----------------|-------|-------------------|
| | | Obs. ^a | Sim. ^a | | | | |
| T2 (°C) | Spring | 13.99 | 13.95 | 1.54 | -0.04 | 0.95 | 2.26 |
| | Summer | 23.60 | 23.65 | 1.64 | 0.05 | 0.90 | 2.54 |
| | Autumn | 12.70 | 12.10 | 1.47 | -0.60 | 0.96 | 2.16 |
| | Winter | -1.34 | -2.84 | 2.17 | -1.50 | 0.97 | 3.00 |
| | Year | 12.23 | 11.70 | 1.70 | -0.52 | 0.98 | 2.51 |
| SWDOWN (W m ⁻²) | Spring | 176.39 | 261.75 | 85.98 | 85.37 | 0.64 | 90.40 |
| | Summer | 192.25 | 278.29 | 86.50 | 86.05 | 0.59 | 92.00 |
| | Autumn | 129.03 | 184.58 | 57.86 | 55.55 | 0.59 | 61.89 |
| | Winter | 92.04 | 144.36 | 53.69 | 52.32 | 0.71 | 57.60 |
| | Year | 147.43 | 217.25 | 71.01 | 69.82 | 0.86 | 77.11 |

225 ^a Obs.: Mean observed value, Sim.: Mean simulated value;

226 ^b ME: Mean Error, MB: Mean Bias, r : correlation coefficient, RMSE: Root Mean Square Error.

227 3.2 Impacts of physiological variables application on estimations

228 The primary advantage of the coupling of WRF-CLM4-MEGAN is that MEGAN is driven by
 229 real-time physiological variables derived from vegetation physics. Here we discussed how the
 230 physiological parameter application affects estimations.

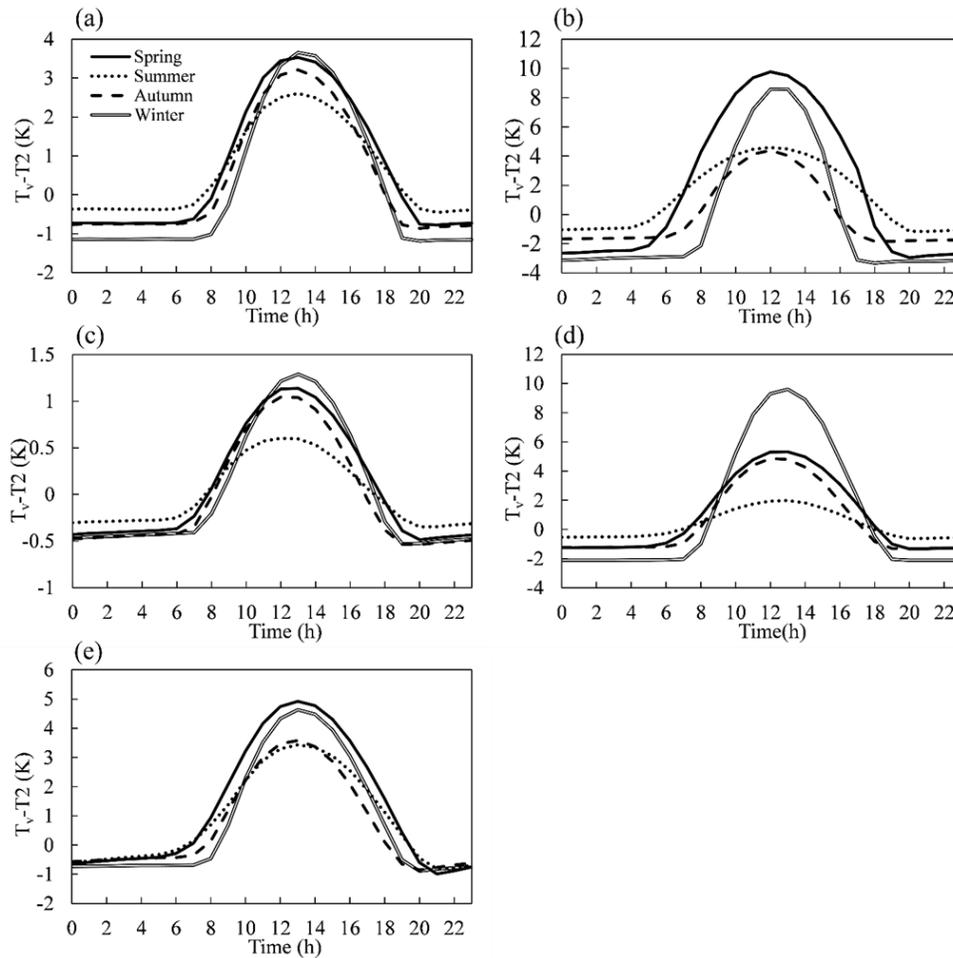
231 3.2.1. Effects of considering the leaf-air temperature bias

232 The difference between vegetation and air temperature is PFT-dependent due to different
 233 physiological conditions. Figure 1 displays the daily profiles of average leaf-air temperature bias.
 234 Due to the weak BVOC emission capacities of grass and crop, as well as the limited biomass in
 235 most time of the year, the temperature of grass and crop was not discussed here. The daily
 236 variety indicated that plants were cooler than the ambient environment in the night and warmer
 237 in the daytime, with positive bias peaked at 12:00~13:00 LT (Local Time). However, the
 238 maximum bias varied significantly among PFTs and seasons. The maximum leaf-temperature
 239 bias of evergreen trees remains relatively low and stable over seasons, ranging from 0.5 K to 1.5
 240 K for evergreen broadleaf trees and 2 K to 4 K for evergreen needleleaf trees. The temperature
 241 bias of deciduous trees and shrubs was similar to that of evergreen trees in summer, but the
 242 maximum bias in winter was higher than that in summer by 3~8 K. The greatest difference
 243 between winter and summer maximum bias was found in simulations of deciduous broadleaf
 244 trees.

245 Varieties in the maximum temperature bias and its seasonal pattern indicated a strong relation
 246 between leaf-air temperature bias and leaf biomass. In summer, strong transpiration of leaves
 247 cool the vegetation effectively and prevent the leaf temperature from rising rapidly under
 248 sunlight. The leaf biomass of evergreen trees did not change significantly over the year, so the
 249 temperature bias remains stable among seasons. The cooling effect of transpiration was
 250 extremely low in winter for deciduous plants. Therefore, the vegetation was significantly

251 warmed by solar radiation absorbed by stems, resulting in a large positive bias between
 252 vegetation and air temperature.

253 Wei et al. (2012) measured the canopy temperature and micrometeorological data of *Quercus*
 254 *variabilis*, a typical broadleaf deciduous tree, during the growing season (May to August). They
 255 reported that the mean canopy temperature was 3.55 K higher than air temperature during the
 256 daytime. The simulated result of our study was 2.2 K. Song et al. (2017) found a positive bias
 257 within 2 K between canopy and air temperature for broadleaf evergreen trees in Xishuangbanna,
 258 southwestern China. The simulated leaf-air temperature bias of broadleaf evergreen trees in our
 259 study was ~1.5 K. These comparisons indicated the good model performance on simulating the
 260 leaf temperature of major BVOC sources.



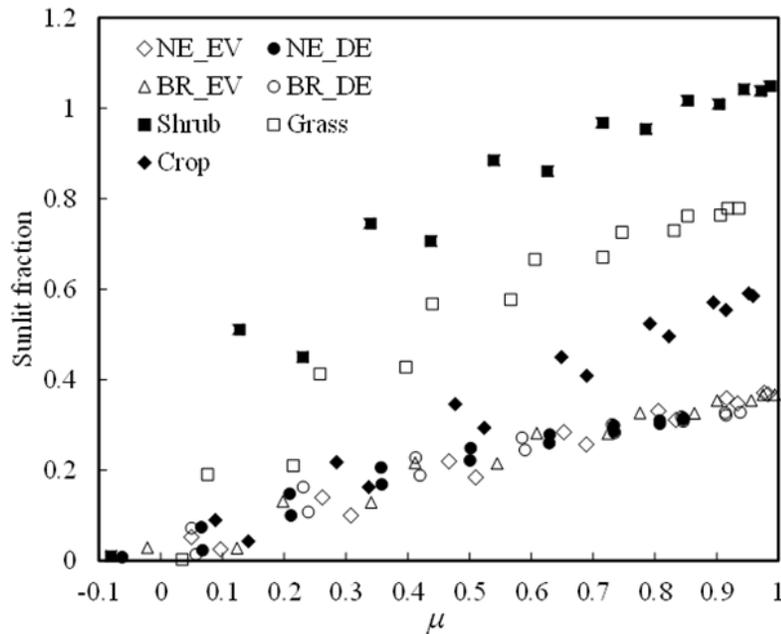
261
 262 Figure 1 Daily Profile of Differences in Average Leaf Temperature (T_v) and Air Temperature at
 263 2 m Height (T_2), (a) evergreen needleleaf trees, (b) deciduous needleleaf trees, (c) evergreen
 264 broadleaf trees, (d) deciduous broadleaf trees, (e) shrub.

265 Many studies assumed that the leaf temperature was equal to air temperature in BVOC
 266 estimations. A model experiment used air temperature at 2 m height (C1_T2) for BVOC
 267 estimation was performed for July to investigate the impacts of considering leaf-air temperature
 268 bias on emissions. Emissions of isoprene and monoterpene from each PFT estimated in BASE
 269 and C1_T2 scenarios are listed in Table 2. Using T2 underestimated the total isoprene emission

270 and monoterpene emission by 23.9% and 21.9%, respectively. The most significant
 271 underestimation was found in regions covered by shrub (underestimated by 58.6%). As a result
 272 of the small temperature bias in summer, isoprene and monoterpene emissions from broadleaf
 273 deciduous trees were only underestimated by 13.8% and 9.0% in C1_T2, respectively. However,
 274 due to the strong emission capacity, this underestimation contributed to 30% to the total emission
 275 difference. Therefore, due to variations in physiological parameters among tree species, the
 276 impact of unreasonable temperature applications varies greatly among PFTs. To reduce
 277 uncertainty, it is necessary to use PFT-specific leaf temperature for BVOC estimation.

278 3.2.2 Effects of differentiating between sunlit and shaded canopy

279 In CLM4 scheme, shaded leaves affect radiation distribution within the canopy by absorbing part
 280 of diffuse radiation. The sunlit/shaded fraction of the canopy was determined by leaf biomass
 281 and orientation, and the angle of the incident light. Variations in the sunlit fraction of each PFT
 282 (daytime in July) due to changes in solar zenith angle are shown in Fig. 2. For all PFTs, the
 283 fraction of sunlit canopy increased with increasing cosine of solar zenith angle, while the
 284 maximum fraction varied greatly. Due to the dense canopy in summer, trees have a smaller sunlit
 285 fraction than other PFTs under the same solar angle. The maximum fraction of trees was
 286 estimated to be 0.35. Because the leaf biomass in July of evergreen and deciduous trees was
 287 similar to each other, these trees showed a same trend in sunlit fraction with changes in solar
 288 angle. The radiation distribution among the sub-grid PFTs was considered in CLM4. When the
 289 light was coming in vertically, few radiations could be received by shrubs due to the absorption
 290 and reflection of higher canopies the same model grid. CLM4 assumes that the sunlit fraction is
 291 equal to 1 when the leaf area index exposed to light is lower than $0.01 \text{ m}^2 \text{ m}^{-2}$. Therefore, the
 292 sunlit fraction of shrub was set as 1 under vertically incident light in this study. The largest sunlit
 293 fraction of grass and crop was 0.7 and 0.5, respectively.



294

295 Figure 2 Variations in sunlit fraction of each PFT under changing cosine of solar zenith angle
 296 (μ), The meaning of NE_EV, NE_DE, BR_EV and BR_DE refers to Table 2.

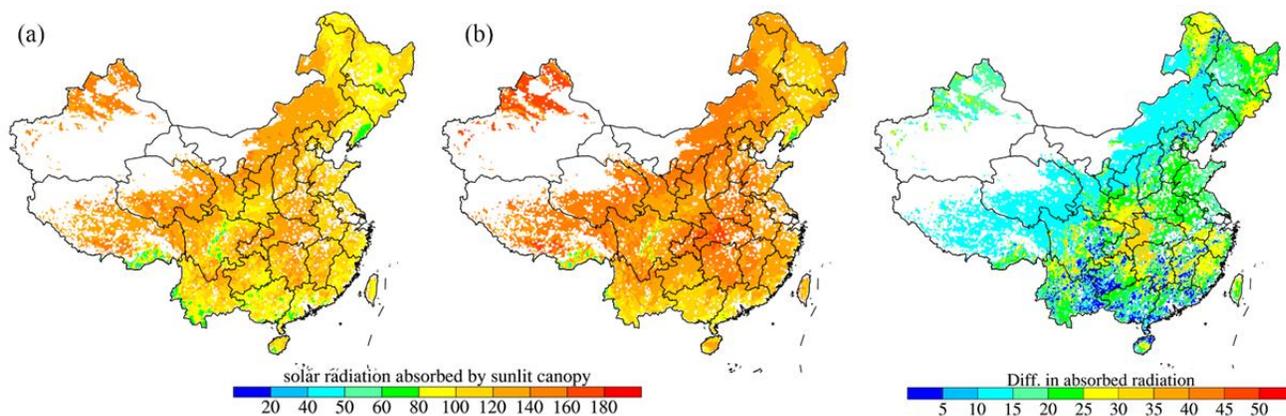
297 Table 2 *Estimations of PFT-specific Isoprene and Monoterpene Emission in Each Scenario (Gg*
 298 *C).*

| PFT | Scenario | Isoprene | Monoterpenes |
|--------------------|----------------------|----------|--------------|
| NE_EV ^a | BASE | 12.3 | 19.1 |
| | C1_T2 ^b | 9.4 | 16.4 |
| | C2_FSUN ^b | 32.5 | 27.0 |
| | C3_FIX ^b | 6.3 | 17.3 |
| NE_DE ^a | BASE | 1.3 | 26.7 |
| | C1_T2 | 0.9 | 20.2 |
| | C2_FSUN | 2.7 | 35.7 |
| | C3_FIX | 0.8 | 21.3 |
| BR_EV ^a | BASE | 249.6 | 56.4 |
| | C1_T2 | 235.8 | 55.4 |
| | C2_FSUN | 663.2 | 83.4 |
| | C3_FIX | 121.3 | 48.1 |
| BR_DE ^a | BASE | 1404.0 | 409.6 |
| | C1_T2 | 1210.0 | 372.6 |
| | C2_FSUN | 3764.0 | 623.2 |
| | C3_FIX | 681.3 | 341.1 |
| Shrub | BASE | 506.2 | 170.6 |
| | C1_T2 | 209.5 | 69.1 |
| | C2_FSUN | 1166.0 | 243.2 |
| | C3_FIX | 180.3 | 134.4 |
| Grass | BASE | 57.2 | 2.6 |
| | C1_T2 | 31.9 | 1.4 |
| | C2_FSUN | 95.3 | 3.4 |
| | C3_FIX | 22.5 | 1.6 |
| Crop | BASE | 35.6 | 12.3 |
| | C1_T2 | 27.1 | 8.9 |
| | C2_FSUN | 79.1 | 18.6 |
| | C3_FIX | 12.6 | 9.1 |
| Total | BASE | 2266.1 | 697.2 |
| | C1_T2 | 1724.5 | 544.2 |
| | C2_FSUN | 5802.8 | 1034.4 |
| | C3_FIX | 1025.5 | 572.9 |

299 ^aNE_EV: Needleleaf Evergreen Tree; NE_DE: Needleleaf Deciduous Tree; BR_EV: Broadleaf evergreen
 300 Tree; BR_DE: Broadleaf Deciduous Tree;

301 ^bC1_T2: replacing leaf temperature with air temperature at 2 m to parameterize temperature response;
 302 C2_FSUN: ignoring the fractions of sunlit and shaded leaves; C3_FIX: using fixed values for variables that
 303 related to weather history.

304 Since the simplified BVOC algorithms were not able to separate the canopy, previous studies
 305 assumed that all the radiation was received by sunlit leaves. Figure 3 displays the difference in
 306 solar radiation absorbed by the sunlit canopy (daytime in July) in differentiating between sunlit
 307 and shaded canopy or not (excluding small islands in the South China Sea). Because the
 308 shadowed canopy absorbed a part of diffuse solar radiation, taking shadowed canopy into
 309 account resulted in a decrease in total radiation absorbed by sunlit leaves. While the radiation
 310 absorbed by shrub and grass showed little difference between the two cases, radiation absorbed
 311 by trees was generally overestimated by over 50% in the scenario which neglected shaded
 312 leaves.



313
 314 Figure 3 *Difference in solar radiation absorbed by sunlit canopy during daytime in July between*
 315 *cases considering or not considering sunlit/shaded leaves separately. (a) separate treatments of*
 316 *canopy; (b) no separate treatments.*

317 Leaf-level radiation controls estimation of radiation response of BVOC emission rates. To
 318 investigate the effects of differentiating between sunlit and shaded canopy, the fraction of shaded
 319 leaves was neglected in the C2_FSUN scenario. The results of C2_FSUN are listed in Table 2.
 320 C2_FSUN overestimated isoprene and monoterpene emissions in July by a factor of 2.6 and 1.5,
 321 respectively. Emissions from broadleaf and needleleaf trees were overestimated by a factor of
 322 2.7 due to the large fractions of shadowed leaves in summer. The least discrepancy was found in
 323 the estimation of grass emissions, which were within a factor of 2 of the BASE estimations. In
 324 conclusion, ignoring the shaded part of the canopy could significantly overestimate BVOC
 325 emissions.

326 3.2.3 Effects of improved parameterization of medium-term weather history

327 MEGAN algorithms require average leaf temperature and solar radiation over past days to
 328 simulate the emission response to medium-term environmental changes. In the original WRF-
 329 CLM4 scheme, the accumulated module used to calculate running mean leaf temperature and
 330 leaf fraction was inactive. Variables based on weather history were assigned with fixed values. In
 331 this study, we modified this module to provide dynamic past-day's physiological parameters for
 332 the MEGAN algorithm. We conducted C3_FIX scenario to investigate the impacts (Table 2).
 333 Emission of isoprene and monoterpene in C3_FIX was lower than that estimated in BASE by
 334 54.7% and 17.8%, respectively. The relatively slight influence on monoterpene emissions could
 335 be attributed to the partial dependence of monoterpene emissions on solar radiation. Isoprene
 336 emission is highly sensitive to solar radiation so that variations in past-time's leaf fraction and
 337 PAR greatly affected isoprene emission estimation. Emissions from needleleaf trees were

338 underestimated by about 40% and those from other PFTs were underestimated by about 60%.
 339 We further calculated emissions in January using fixed values (not shown). The isoprene and
 340 monoterpene emissions were estimated as 64.7 (61.3 in BASE) and 36.4 Gg C (34.0 in BASE),
 341 respectively. The discrepancy between using fixed and dynamic values was within 10%.
 342 Although the fixed values could make estimations similar to dynamic variables in low
 343 temperature and radiation conditions, using fixed values for all seasons leads to significant
 344 uncertainty.

345 3.3 BVOC emission budgets and spatiotemporal distribution

346 Hourly emissions of 8 chemical species were calculated by the WRF-CLM4-MEGAN model on
 347 a 12 km×12 km grid for the year 2018. In the following section, all the results are measured as
 348 carbon weights of the constituent compounds, unless stated otherwise.

349 The annual amount emitted for all listed BVOCs reaches 14.7 Tg C with isoprene accounting for
 350 78.3% (11.5 Tg) and the sum of monoterpenes for 21.7% (3.2 Tg). Emissions from each PFT are
 351 shown in Table 3. Province-level emissions are listed in Table S4. For both isoprene and
 352 monoterpene, the predominant source was broadleaf deciduous forests with a contribution of
 353 64.5% and 60.6% to isoprene and monoterpene, respectively. Shrubs ranked second of the
 354 emission contribution, accounting for 19.9% of total BVOC emissions, followed by broadleaf
 355 evergreen trees (12.1%). Grass and Crop were responsible for only 1.3% of total isoprene
 356 emission and 1.1% of monoterpene emission.

357 Table 3 *BVOC Emission Budgets of Each Plant Functional Type (PFT) (Gg C)*.

| PFTs | ISO ^b | MT ^a | | | | | | | T_MT ^c | T_ALL ^c |
|--------------------|------------------|------------------|------------------|--------------------|------------------|------------------|------------------|------------------|-------------------|--------------------|
| | | API ^b | BPI ^b | 3-CAR ^b | OCI ^b | LIM ^b | SAB ^b | MYR ^b | | |
| NE_EV ^a | 59.3 | 35.6 | 38.6 | 20.6 | 3.4 | 12.9 | 5.0 | 5.0 | 121.0 | 180.3 |
| NE_DE ^a | 4.2 | 33.7 | 20.0 | 8.0 | 3.0 | 13.0 | 2.6 | 4.0 | 84.4 | 88.6 |
| BR_EV ^a | 1421.0 | 169.0 | 61.6 | 20.5 | 28.8 | 41.0 | 22.5 | 22.5 | 365.8 | 1786.8 |
| BR_DE ^a | 7442.0 | 806.6 | 426.0 | 97.8 | 179.7 | 262.4 | 100.8 | 60.2 | 1934.0 | 9376.0 |
| Shrub | 2295.0 | 183.3 | 132.7 | 89.0 | 70.9 | 89.0 | 42.9 | 30.6 | 638.3 | 2933.3 |
| Grass | 188.1 | 2.2 | 2.2 | 0.3 | 1.8 | 1.0 | 0.7 | 0.2 | 8.4 | 196.5 |
| Crop | 119.4 | 9.2 | 9.9 | 2.0 | 7.4 | 4.6 | 3.2 | 1.4 | 37.8 | 157.2 |
| Nation | 11528 | 1240 | 691 | 242 | 295 | 424 | 178 | 124 | 3193 | 14721 |

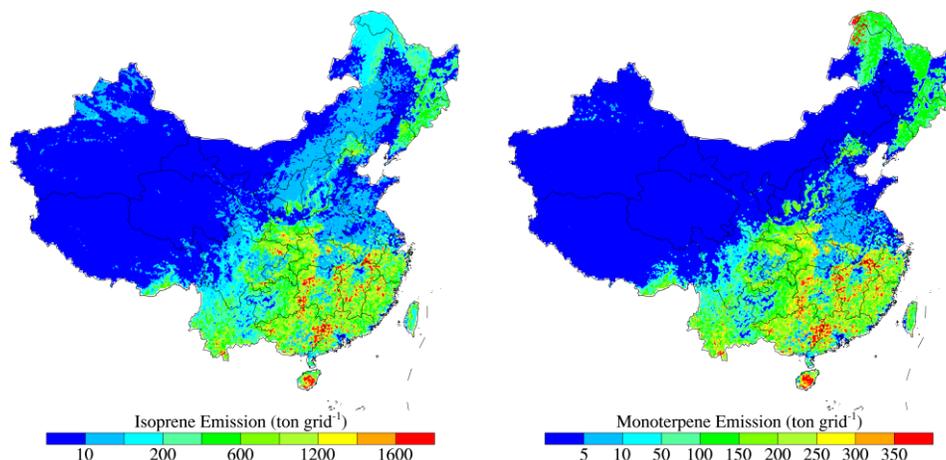
358 ^aRefer to Table 2;

359 ^bISO: isoprene, API: α -pinene, BPI: β -pinene, 3-CAR: 3-carene, OCI: ocimene, LIM: limonene, SAB:
 360 sabinene, MYR: myrcene;

361 ^cT_MT: Total monoterpenes emissions of each PFT, T_ALL: Total BVOC (includes isoprene and
 362 monoterpenes) emissions of each PFT.

363 The spatial distributions of annual emissions of isoprene and monoterpenes are displayed in
 364 Fig.4 (excluding small islands in the South China Sea). The south and northeast of China, as well
 365 as the Qinling Mountains in central China, were estimated with high emission budgets,
 366 accounting for 91.3% of the national isoprene emission and 91.8% of the monoterpene emission.
 367 According to Fig.S1 and the survey results from the Plant Research Institute, these areas are
 368 covered by vegetation species with a high emission capacity of isoprene (broadleaf forests,
 369 shrub) or monoterpenes (coniferous forest). Northeast China is primarily covered by deciduous

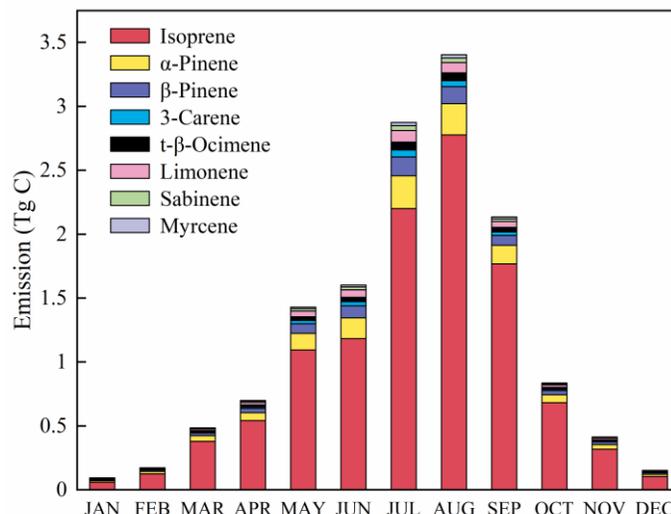
370 coniferous forests (mainly *Larix gmelini*) and deciduous broadleaf forests (mainly *Quercus*
 371 *mongolica*, *Tilia Mongolia*, and *Betula platyphylla*). Large areas of evergreen coniferous forests
 372 (mainly *Pinus massoniana* and *Cunninghamia lanceolata*) and shrubs are found in Southeast
 373 China, and the main plant genera in Southwest China are evergreen tree species, including
 374 evergreen broadleaf forests (e.g. *Quercus aquifolioides*), evergreen coniferous forests (*Picea*
 375 *likiangensis var. balfouriana* and *Pinus yunnanensis*) and shrubs. The Qinling Mountains are
 376 covered by large areas of deciduous broadleaf forests (mainly *Quercus variabilis* and *Quercus*
 377 *liaotungensis*). Regions covered by a large area of crop or grassland, such as the North China
 378 Plain and Inner Mongolia, played a very small role in BVOC budgets due to the low emitting
 379 capacities.



380

381 Figure 4 *Spatial distribution of isoprene and monoterpenes emissions in the year 2018.*

382 BVOC emissions showed strong seasonal variation. The temporal profile of national monthly
 383 emissions of individual species is presented in Fig.5. Because of the highest light intensities,
 384 temperatures, and plant biomass density, the emissions peaked in summer (from June to August)
 385 and around 55.1% (8.1 Tg C) of the total annual budgets were released during this period.
 386 Previous studies estimated the highest emissions in July; however, our results showed a higher
 387 emission in August. This could be attributed to the consideration of the effects of canopy climate
 388 history on estimation. This results were consistent with a whole-year measurement reported by
 389 Chen et al. (2020).

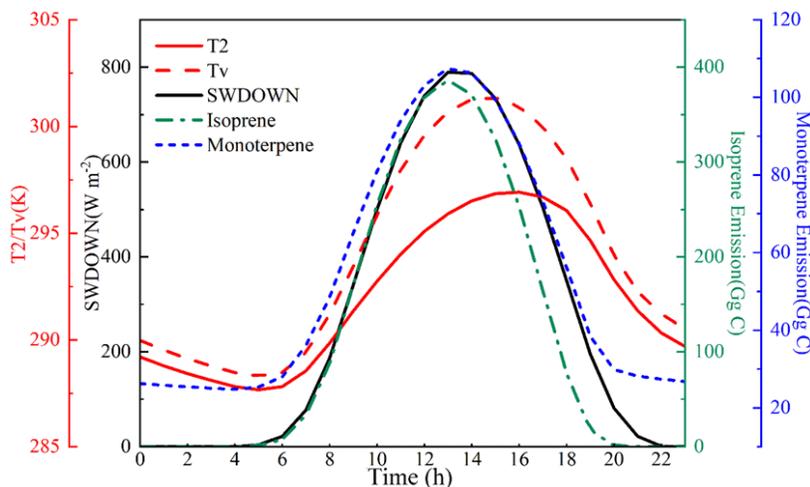


390

391

Figure 5 Monthly variations in BVOC emission budget (Tg C).

392 The diurnal variations of temperature, solar radiation, and emission of isoprene and
 393 monoterpenes in summer are shown in Fig.6. Simulated emissions increased rapidly in the
 394 morning and peaked in the afternoon. Time for peak emissions was closed to time for the highest
 395 leaf temperature and solar radiation, while T2 reached the highest value about an hour later.
 396 Emissions of monoterpenes are not highly dependent on the light while isoprene emissions are
 397 strong light-dependent. As a result, monoterpene emissions maintain a relatively high level
 398 during the night while isoprene emissions cease at nighttime.



399

400 Figure 6 Diurnal variations in summer of air temperature at 2 m height (T_2), leaf temperature
 401 (T_v), downward solar radiation (SWDOWN) and emissions of isoprene and monoterpenes (Gg
 402 C).

403 3.4 Comparisons with previous studies

404 We evaluated the inventory against canopy-scale measurements conducted at different sites in
 405 China. Given that years of interest in the present study and observations are different, the
 406 average data or monthly total emissions were used for comparison.

407 Using the eddy covariance technique, Baker et al. (2005) measured isoprene fluxes in
408 Xishuangbanna, Yunnan Province (21.92°N, 101.27°E). Daytime isoprene fluxes during the wet
409 season (July 2002) was approximately $1 \text{ mg C m}^{-2} \text{ h}^{-1}$. Our model predicted the daytime average
410 isoprene flux over July as $1.5 \text{ mg C m}^{-2} \text{ h}^{-1}$, similar to the observed data. Based on Relaxed
411 Eddy Accumulation (REA) technique, emissions of isoprene and monoterpenes of a temperate
412 forest in Changbai Mountain (42.4°N, 128.1°E) were measured during the summer seasons in
413 2010 and 2011 (Bai et al., 2015). Average isoprene fluxes were measured as 1.3 and 1.5 mg m^{-2}
414 h^{-1} , and the simulated fluxes were 2.1 and $2.0 \text{ mg m}^{-2} \text{ h}^{-1}$, respectively, about 50% higher than
415 observations. The average PAR and temperature during experimental periods were $837.5 \text{ } \mu\text{mol}$
416 $\text{m}^{-2} \text{ s}^{-1}$ and $22.6 \text{ } ^\circ\text{C}$, respectively. The simulation resulted in an average PAR of $1160.1 \text{ } \mu\text{mol}$
417 $\text{m}^{-2} \text{ s}^{-1}$ and a temperature of $22.23 \text{ } ^\circ\text{C}$. The average leaf temperature was $23.37 \text{ } ^\circ\text{C}$. The slight
418 overestimation could be attributed to higher PAR in the model.

419 Using REA method, Bai et al. (2016) measured emissions from a bamboo (*Phyllostachys*
420 *violascenes*) plantation in Zhejiang Province (30.3°N, 119.57°E) and the average isoprene
421 emission fluxes were 2.81, 1.07, 0.186, $0.068 \text{ mg m}^{-2} \text{ h}^{-1}$ for the experimental periods in July,
422 August, September, and October. The predicted monthly average fluxes were 2.04, 1.95, 0.39,
423 $0.14 \text{ mg m}^{-2} \text{ h}^{-1}$, respectively. Estimations were within a factor of 2 of observed values.

424 As illustrated in Table 4, the annual emission budgets estimated by this study fall in the range of
425 past studies. Both the improvement in driving variables and the representation of sub-grid plant
426 composition contribute to the difference in estimates. Formaldehyde (HCHO), as a major
427 intermediate product in the degradation of isoprene in the atmosphere, has been widely used as a
428 proxy for estimates isoprene emissions. Fu et al. (2007) used a continuous 6-year record (1996–
429 2001) of Global Ozone Monitoring Experiment (GOME) HCHO columns to estimate isoprene
430 emission as 12.7 Tg yr^{-1} in China, which is comparable to our model outputs. Stavrou et al.
431 (2014) found that isoprene emissions in China decrease from 8.6 Tg in the year 2007 to 6.5 Tg in
432 the year 2012 based on GOME-2 HCHO columns, lower than emissions in this study by
433 33.7%~76.9%. The isoprene emission from China in 2010 was estimated to be 6.5 Tg based on
434 OMI (Ozone Monitoring Instrument) HCHO observations (Stavrou et al., 2015). Aside from
435 the influence of different meteorological conditions and land cover changes during the past
436 years, the reliability of satellite-based constraints also needs to be improved (Fu et al., 2019).

437 Table 4 Comparison of BVOC Budgets ($Tg\ C\ yr^{-1}$) estimated in This Study and in Previous
 438 Studies.

| Base year | Emission Budget | | | Reference |
|----------------------------|------------------|-----------------|-------|-------------------------------------|
| | ISO ^a | MT ^a | Total | |
| 2018 | 11.5 | 3.2 | 14.7 | This study |
| - | 15.0 | 4.3 | 19.3 | Guenther et al. (1995) ^a |
| - | 4.1 | 3.5 | 7.6 | Klinger et al. (2002) ^a |
| 2004 | 6.8 | 2.8 | 9.6 | Tie et al. (2006) ^b |
| 2000 | 10.0 | 2.5 | 12.5 | Guenther et al. (2006) ^c |
| Averaged over 2001-2006 | 9.6 | 2.8 | 12.4 | Fu and Liao (2012) ^c |
| - | 12.7 | | | Fu et al. (2007) ^d |
| 2010 | 5.9 | | | Stavrou et al. (2014) ^d |
| 2010 | 6.5 | | | Stavrou et al. (2015) ^e |

439 ^aBased on G95 algorithms (Guenther et al., 1995);

440 ^bBased on G20 algorithms (Guenther et al., 2000);

441 ^cBased on MEGAN2.0 algorithms (Guenther et al., 2006);

442 ^dTop-down annual emission estimates which were inferred by inversion of GOME-2 formaldehyde columns.

443 ^eTop-down annual emission estimates which were inferred by inversion of OMI formaldehyde columns

444 4 Uncertainty

445 The basal emission factors are identified as the most important uncertainty source in BVOC
 446 emission estimations (Guenther et al., 2006). Local emission factors for isoprene reported by
 447 previous observations conducted in China were used in this study. Since measurements of the
 448 monoterpene emission factors are scarce, we calculated local emission factors based on the ratio
 449 of local isoprene emission factor to default emission factor in MEGAN literature. There are large
 450 uncertainties associated with the conversion approach. More in-situ observations on emission
 451 rates of different PFTs in China are required.

452 CLM4 parameterizes one layer of the canopy, however, solar radiation is attenuated by foliage
 453 and leaf temperature varies among layers. A relatively simple representation of canopy is also a
 454 source of uncertainty. Guenther et al. (1995) found a less than 5% difference in global annual
 455 isoprene emission estimated with one or five layers and no change in the estimations of other
 456 BVOC emissions, suggesting that BVOC emissions are relatively insensitive to the number
 457 layers. However, many studies indicated that the treatment of microclimatic factors such as light
 458 and leaf temperature within the canopy resulted in a substantial difference in estimated emissions
 459 (Keenan et al., 2011).

460 5 Conclusions

461 This study estimated the emission budgets and spatial-temporal patterns of BVOC in China in
 462 the year 2018 using the WRF-CLM4-MEGAN modeling system. This framework improved
 463 biogenic emission estimations by using PFT-specific physiological parameters derived from soil
 464 and vegetation physics in the CLM4 scheme. The simulated vegetation temperature was typically
 465 higher than air temperature by 1~12 K in the daytime and lower than ambient value by

466 approximately 2 K during the night. Using air temperature instead of leaf temperature
467 underestimated isoprene and monoterpene emissions in July by 23.9% and 21.9%, respectively.
468 Because the shaded fraction of broadleaf trees could be higher than 60 % in July, ignoring the
469 influence of shaded canopy on radiation distribution overestimated emissions by a factor of 2.6.
470 Assigning fixed values to variables that related to weather history made a similar estimation to
471 that based on dynamic variables in January, while underestimated emissions in July by
472 approximately 50%. Due to the significant discrepancy caused by these physiological variables,
473 more reasonable parameter applications are important for accurately estimating biogenic
474 emissions. Using the CLM4-MEGAN framework, the annual emissions of BVOC in China was
475 estimated to be 14.7 Tg C, with isoprene and monoterpenes accounting for 78.3% and 21.7% of
476 the totals, respectively. The coupled model successfully reproduced the spatial and temporal
477 patterns of BVOC emissions. The predicted values were within a factor of 2 of most observed
478 values. Comparisons indicated that this coupled model are able to estimate BVOC emissions
479 reasonably in China.

480 **Acknowledgments and Data**

481 The MODIS land cover data (MCD12Q1), water mask data (MOD44W) and leaf are index
482 product (MCD15A2H) were provided by Land Process Distributed Active Archive Center
483 (LPDAAC), USA.

484 The MCD12Q1 product can be freely accessed at 10.5067/MODIS/MCD12Q1.006. MOD44W
485 data can be download from 10.5067/MODIS/MOD44W.006. MCD15A2H data can be found at
486 10.5067/MODIS/MCD15A2H.006.

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490 and ML. LY and YS led the writing of the papers and prepared the figures with contributions
491 from all co-authors.

492 The authors declare that they have no conflict of interest.

493 **References**

- 494 Arneth, A., P. A. Miller, M. Scholze, T. Hickler, G. Schurgers, B. Smith, and I. C. Prentice
495 (2007), CO₂ inhibition of global terrestrial isoprene emissions: Potential implications for
496 atmospheric chemistry, *Geophysical Research Letters*, 34(18), doi:10.1029/2007gl030615.
- 497 Bai, J., A. Guenther, A. Turnipseed, and T. Duhl (2015), Seasonal and interannual variations in
498 whole-ecosystem isoprene and monoterpene emissions from a temperate mixed forest in
499 Northern China, *Atmospheric Pollution Research*, 6(4), 696-707,
500 doi:[10.5094/APR.2015.078](https://doi.org/10.5094/APR.2015.078).
- 501 Bai, J., A. Guenther, A. Turnipseed, T. Duhl, S. Yu, and B. Wang (2016), Seasonal variations in
502 whole-ecosystem BVOC emissions from a subtropical bamboo plantation in China,
503 *Atmospheric Environment*, 124, 12-21, doi:[10.1016/j.atmosenv.2015.11.008](https://doi.org/10.1016/j.atmosenv.2015.11.008).
- 504 Baker, B., et al. (2005), Wet and dry season ecosystem level fluxes of isoprene and
505 monoterpenes from a southeast Asian secondary forest and rubber tree plantation,
506 *Atmospheric Environment*, 39(2), 381-390, doi:[10.1016/j.atmosenv.2004.07.033](https://doi.org/10.1016/j.atmosenv.2004.07.033).

- 507 Chen, C., et al. (2019), China and India lead in greening of the world through land-use
508 management, *Nature Sustainability*, 2(2), 122-129, doi:10.1038/s41893-019-0220-7.
- 509 Chen, J., J. Tang, and X. Yu (2020), Environmental and physiological controls on diurnal and
510 seasonal patterns of biogenic volatile organic compound emissions from five dominant
511 woody species under field conditions, *Environmental Pollution*, 259, 113955,
512 doi:[10.1016/j.envpol.2020.113955](https://doi.org/10.1016/j.envpol.2020.113955).
- 513 Fehsenfeld, F., et al. (1992), Emissions of volatile organic compounds from vegetation and the
514 implications for atmospheric chemistry, *Global Biogeochemical Cycles*, 6(4), 389-430,
515 doi:10.1029/92gb02125.
- 516 Fu, D., D. B. Millet, K. C. Wells, V. H. Payne, S. Yu, A. Guenther, and A. Eldering (2019),
517 Direct retrieval of isoprene from satellite-based infrared measurements, *Nature*
518 *Communications*, 10(1), 3811, doi:10.1038/s41467-019-11835-0.
- 519 Fu, T.-M., D. J. Jacob, P. I. Palmer, K. Chance, Y. X. Wang, B. Barletta, D. R. Blake, J. C.
520 Stanton, and M. J. Pilling (2007), Space-based formaldehyde measurements as constraints
521 on volatile organic compound emissions in east and south Asia and implications for ozone,
522 *Journal of Geophysical Research: Atmospheres*, 112(D6), doi:10.1029/2006jd007853.
- 523 Fu, Y., and H. Liao (2012), Simulation of the interannual variations of biogenic emissions of
524 volatile organic compounds in China: Impacts on tropospheric ozone and secondary organic
525 aerosol, *Atmospheric Environment*, 59, 170-185, doi:[10.1016/j.atmosenv.2012.05.053](https://doi.org/10.1016/j.atmosenv.2012.05.053).
- 526 Guenther, A., C. N. Hewitt, D. Erickson, R. Fall, C. Geron, T. Graedel, P. Harley, L. Klinger, M.
527 Ler dau, and W. McKay (1995), A global model of natural volatile organic compound
528 emissions, *Journal of Geophysical Research: Atmospheres*, 100(D5), 8873-8892.
- 529 Guenther, A., T. Karl, P. Harley, C. Wiedinmyer, P. I. Palmer, and C. Geron (2006), Estimates of
530 global terrestrial isoprene emissions using MEGAN (Model of Emissions of Gases and
531 Aerosols from Nature), *Atmos. Chem. Phys.*, 6(11), 3181-3210, doi:10.5194/acp-6-3181-
532 2006.
- 533 Guenther, A. B., X. Jiang, C. L. Heald, T. Sakulyanontvittaya, T. Duhl, L. K. Emmons, and X.
534 Wang (2012), The Model of Emissions of Gases and Aerosols from Nature version 2.1
535 (MEGAN2.1): an extended and updated framework for modeling biogenic emissions,
536 *Geosci. Model Dev.*, 5(6), 1471-1492, doi:10.5194/gmd-5-1471-2012.
- 537 Guenther, A. B., R. K. Monson, and R. Fall (1991), Isoprene and monoterpene emission rate
538 variability: Observations with eucalyptus and emission rate algorithm development, *Journal*
539 *of Geophysical Research: Atmospheres*, 96(D6), 10799-10808, doi:10.1029/91jd00960.
- 540 Henrot, A. J., T. Stanelle, S. Schröder, C. Siegenthaler, D. Taraborrelli, and M. G. Schultz
541 (2017), Implementation of the MEGAN (v2.1) biogenic emission model in the ECHAM6-
542 HAMMOZ chemistry climate model, *Geosci. Model Dev.*, 10(2), 903-926,
543 doi:10.5194/gmd-10-903-2017.
- 544 Jin, J., N. L. Miller, and N. Schlegel (2010), Sensitivity Study of Four Land Surface Schemes in
545 the WRF Model, *Advances in Meteorology*, 2010, 11, doi:10.1155/2010/167436.

- 546 Jin, J., and L. Wen (2012), Evaluation of snowmelt simulation in the Weather Research and
 547 Forecasting model, *Journal of Geophysical Research: Atmospheres*, 117(D10),
 548 doi:10.1029/2011jd016980.
- 549 Keenan, T. F., R. Grote, and S. Sabaté (2011), Overlooking the canopy: The importance of
 550 canopy structure in scaling isoprenoid emissions from the leaf to the landscape, *Ecological*
 551 *Modelling*, 222(3), 737-747, doi:[10.1016/j.ecolmodel.2010.11.004](https://doi.org/10.1016/j.ecolmodel.2010.11.004).
- 552 Klinger, L. F., Q.-J. Li, A. B. Guenther, J. P. Greenberg, B. Baker, and J.-H. Bai (2002),
 553 Assessment of volatile organic compound emissions from ecosystems of China, *Journal of*
 554 *Geophysical Research: Atmospheres*, 107(D21), ACH 16-11-ACH 16-21,
 555 doi:10.1029/2001jd001076.
- 556 Lawrence, D. M., et al. (2011), Parameterization Improvements and Functional and Structural
 557 Advances in Version 4 of the Community Land Model, *Journal of Advances in Modeling*
 558 *Earth Systems*, 3, doi:10.1029/2011ms000045.
- 559 Levis, S., C. Wiedinmyer, G. B. Bonan, and A. Guenther (2003), Simulating biogenic volatile
 560 organic compound emissions in the Community Climate System Model, *Journal of*
 561 *Geophysical Research: Atmospheres*, 108(D21), doi:10.1029/2002jd003203.
- 562 Li, L. Y., Y. Chen, and S. D. Xie (2013), Spatio-temporal variation of biogenic volatile organic
 563 compounds emissions in China, *Environmental Pollution*, 182, 157-168,
 564 doi:[10.1016/j.envpol.2013.06.042](https://doi.org/10.1016/j.envpol.2013.06.042).
- 565 Li, M., X. Huang, J. Li, and Y. Song (2012), Estimation of biogenic volatile organic compound
 566 (BVOC) emissions from the terrestrial ecosystem in China using real-time remote sensing
 567 data, *Atmos. Chem. Phys. Discuss.*, 2012, 6551-6592, doi:10.5194/acpd-12-6551-2012.
- 568 Liu, Y., L. Li, J. Y. An, W. Zhang, R. S. Yan, L. Huang, C. Huang, H. L. Wang, Q. Wang, and
 569 M. Wang (2018), Emissions, Chemical Composition, and Spatial and Temporal Allocation
 570 of the BVOCs in the Yangtze River Delta Region in 2014.
- 571 Lu, Y., and L. M. Kueppers (2012), Surface energy partitioning over four dominant vegetation
 572 types across the United States in a coupled regional climate model (Weather Research and
 573 Forecasting Model 3–Community Land Model 3.5), *Journal of Geophysical Research:*
 574 *Atmospheres*, 117(D6), doi:10.1029/2011jd016991.
- 575 Niinemets, Ü., G. Seufert, R. Steinbrecher, and J. D. Tenhunen (2002), A model coupling foliar
 576 monoterpene emissions to leaf photosynthetic characteristics in Mediterranean evergreen
 577 *Quercus* species, *New Phytologist*, 153(2), 257-275, doi:10.1046/j.0028-
 578 646X.2001.00324.x.
- 579 Niinemets, Ü., J. D. Tenhunen, P. C. Harley, and R. Steinbrecher (1999), A model of isoprene
 580 emission based on energetic requirements for isoprene synthesis and leaf photosynthetic
 581 properties for Liquidambar and *Quercus*, *Plant, Cell & Environment*, 22(11), 1319-1335,
 582 doi:10.1046/j.1365-3040.1999.00505.x.
- 583 Oleson, K., et al. (2010), *Technical Description of version 4.0 of the Community Land Model*
 584 *(CLM)*.
- 585 Sakulyanontvittaya, T., T. Duhl, C. Wiedinmyer, D. Helmig, S. Matsunaga, M. Potosnak, J.
 586 Milford, and A. Guenther (2008), Monoterpene and Sesquiterpene Emission Estimates for

- 587 the United States, *Environmental Science & Technology*, 42(5), 1623-1629,
588 doi:10.1021/es702274e.
- 589 Schultz, N. M., X. Lee, P. J. Lawrence, D. M. Lawrence, and L. Zhao (2016), Assessing the use
590 of subgrid land model output to study impacts of land cover change, *Journal of Geophysical*
591 *Research: Atmospheres*, 121(11), 6133-6147, doi:10.1002/2016jd025094.
- 592 Song, Q.-H., et al. (2017), Canopy temperature variability in a tropical rainforest, subtropical
593 evergreen forest, and savanna forest in Southwest China, *iForest - Biogeosciences and*
594 *Forestry*, 10(3), 611-617, doi:10.3832/ifor2223-010.
- 595 Stavrakou, T., et al. (2015), How consistent are top-down hydrocarbon emissions based on
596 formaldehyde observations from GOME-2 and OMI?, *Atmos. Chem. Phys.*, 15(20), 11861-
597 11884, doi:10.5194/acp-15-11861-2015.
- 598 Stavrakou, T., J. F. Müller, M. Bauwens, I. De Smedt, M. Van Roozendaal, A. Guenther, M.
599 Wild, and X. Xia (2014), Isoprene emissions over Asia 1979–2012: impact of
600 climate and land-use changes, *Atmos. Chem. Phys.*, 14(9), 4587-4605, doi:10.5194/acp-14-
601 4587-2014.
- 602 Subin, Z. M., W. J. Riley, J. Jin, D. S. Christianson, M. S. Torn, and L. M. Kueppers (2011),
603 Ecosystem Feedbacks to Climate Change in California: Development, Testing, and Analysis
604 Using a Coupled Regional Atmosphere and Land Surface Model (WRF3-CLM3.5), *Earth*
605 *Interactions*, 15, doi:10.1175/2010ei331.1.
- 606 Tie, X., G. Li, Z. Ying, A. Guenther, and S. Madronich (2006), Biogenic emissions of
607 isoprenoids and NO in China and comparison to anthropogenic emissions, *Science of The*
608 *Total Environment*, 371(1), 238-251, doi:[10.1016/j.scitotenv.2006.06.025](https://doi.org/10.1016/j.scitotenv.2006.06.025).
- 609 Wei, D. D., J. S. Zhang, M. Ping, Z. Ning, and Y. F. Ren (2012), Variations of canopy
610 temperature in *Quercus variabilis* plantation and their relations with micrometeorological
611 factors, *Chinese Journal of Applied Ecology*, 23(7), 1767-1773.
- 612 Wu, K., et al. (2020), Estimation of biogenic VOC emissions and their corresponding impact on
613 ozone and secondary organic aerosol formation in China, *Atmospheric Research*, 231,
614 104656, doi:[10.1016/j.atmosres.2019.104656](https://doi.org/10.1016/j.atmosres.2019.104656).
- 615 Zhao, C., et al. (2016), Sensitivity of biogenic volatile organic compounds to land surface
616 parameterizations and vegetation distributions in California, *Geosci. Model Dev.*, 9(5),
617 1959-1976, doi:10.5194/gmd-9-1959-2016.