

1 **Modeling the hydrologic influence of subsurface tile drainage using the**
2 **National Water Model**

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11
12 Key Points:

- 13
14 1. A new subsurface tile drainage module is incorporated into the National Water Model
15 (NWM) to predict streamflow over the tile-drained areas
16
17 2. NWM with a tile drainage module can predict high-flows and streamflow peaks better
18 than the original NWM over heavily tile-drained areas
19
20 3. Incorporating tile drainage into the NWM considerably enhanced the streamflow event
21 hit rates and reduced false alarm rates

22 **Abstract**

23 Subsurface tile drainage (TD) is a dominant agriculture water management practice in the United
24 States (US) to enhance crop production in poorly-drained soils. Assessments of field- or
25 watershed-level ($<50 \text{ km}^2$) hydrologic impacts of tile drainage are becoming common; however,
26 a major gap exists in our understanding of regional ($>105 \text{ km}^2$) impacts of tile drainage on
27 hydrology. The National Water Model (NWM) is a distributed 1-km resolution hydrological
28 model designed to provide accurate streamflow forecasts at 2.7 million reaches across the US.
29 The current NWM lacks tile drainage representation which adds considerable uncertainty to
30 streamflow forecasts in tile-drained areas. In this study, we quantify the performance of the
31 NWM with a newly incorporated tile drainage scheme over the heavily tile-drained Midwestern
32 US. Implementing a tile drainage scheme enhanced the uncalibrated model performance by
33 about 20% to 50% of the calibrated NWM (*Calib*). The calibrated NWM with tile drainage
34 (*CalibTD*) showed enhanced accuracy with higher event hit rates and lower false alarm rates than
35 *Calib*. *CalibTD* showed better performance in high-flow estimations as tile drainage increased
36 streamflow peaks (14%), volume (2.3%), and baseflow (11%). Regional water balance analysis
37 indicated that tile drainage significantly reduced surface runoff (-7% to -29%), groundwater
38 recharge (-43% to -50%), evapotranspiration (-7% to -13%), and soil moisture content (-2% to -
39 3%). However, infiltration and soil water storage potential significantly increased with tile
40 drainage. Overall, our findings highlight the importance of incorporating the tile drainage
41 process into the operational configuration of the NWM.

42 **1. Introduction**

43 Agriculture management practices such as irrigation, fertilizer and pesticide application, and
44 tillage are generally employed to enhance crop productivity and are crucial for global food
45 production and food security. Agriculture subsurface drainage, often known as subsurface tile
46 drainage, is a widely-used agriculture water management practice to improve crop growth in
47 regions with shallow water tables or poorly drained soils. According to the United States
48 Department of Agriculture (USDA) National Agricultural Statistics Service (NASS) Census of
49 Agriculture 2017, about 22.48 million hectares (Mha) of croplands in the US are tile-drained,
50 and 83.80% of the total tile-drained croplands of the US are concentrated in six Midwestern
51 states (USDA-NASS, 2017) (Figure 1a), which is one of the world's most productive areas in
52 terms of food and bioenergy, and it is located in the headwater regions of the Mississippi River
53 (Gaunter et al., 2014; Ray et al., 2013).

54 In general, tile drains are buried under the crop root zone to extract saturation water (or free
55 water) from the soil, improve root-zone soil aeration and soil quality, reduce crop root diseases
56 and soil erosion, allow for earlier planting and enhance crop yield (Figure 1b) (Du et al., 2005;
57 Fausey, 2005; Fausey et al., 1987; Kornecki and Fouss, 2001). Furthermore, tile drainage is
58 known to have a significant impact on watershed hydrology (Blann et al., 2009; King et al.,
59 2014; Rahman et al., 2014; Thomas et al., 2016), because it depletes the free water from the root-
60 zone soil layer, resulting in enhanced infiltration and reduced surface runoff, peak flows, and
61 flooding (Golmohammadi et al., 2017; Rahman et al., 2014; Robinson and Rycroft, 1999; Skaggs
62 et al., 1994). Tile drainage may also increase the watershed baseflow, annual runoff volume,
63 instream pollutant concentrations, the timing and shape of the hydrograph, and the local and
64 regional climate by modifying energy and water flux from croplands to the atmosphere (Blann et
65 al., 2009; Eastman et al., 2010; Guo et al., 2018; Khand et al., 2017; King et al., 2014; Magner et
66 al., 2004; Schilling et al., 2012; Schilling and Helmers, 2008; Schilling and Libra, 2003;
67 Schottler et al., 2014; Thomas et al., 2016; Yang et al., 2017). However, the intensity and
68 direction of the tile-drainage impact on hydrology depend on several field-specific factors such
69 as soil properties, antecedent soil moisture storage, climatic conditions, topography, design of the
70 tile drainage system, and tillage practices (Blann et al., 2009; King et al., 2014; Robinson, 1990;
71 Robinson and Rycroft, 1999; Skaggs et al., 1994; Thomas et al., 2016; Wiskow and van der

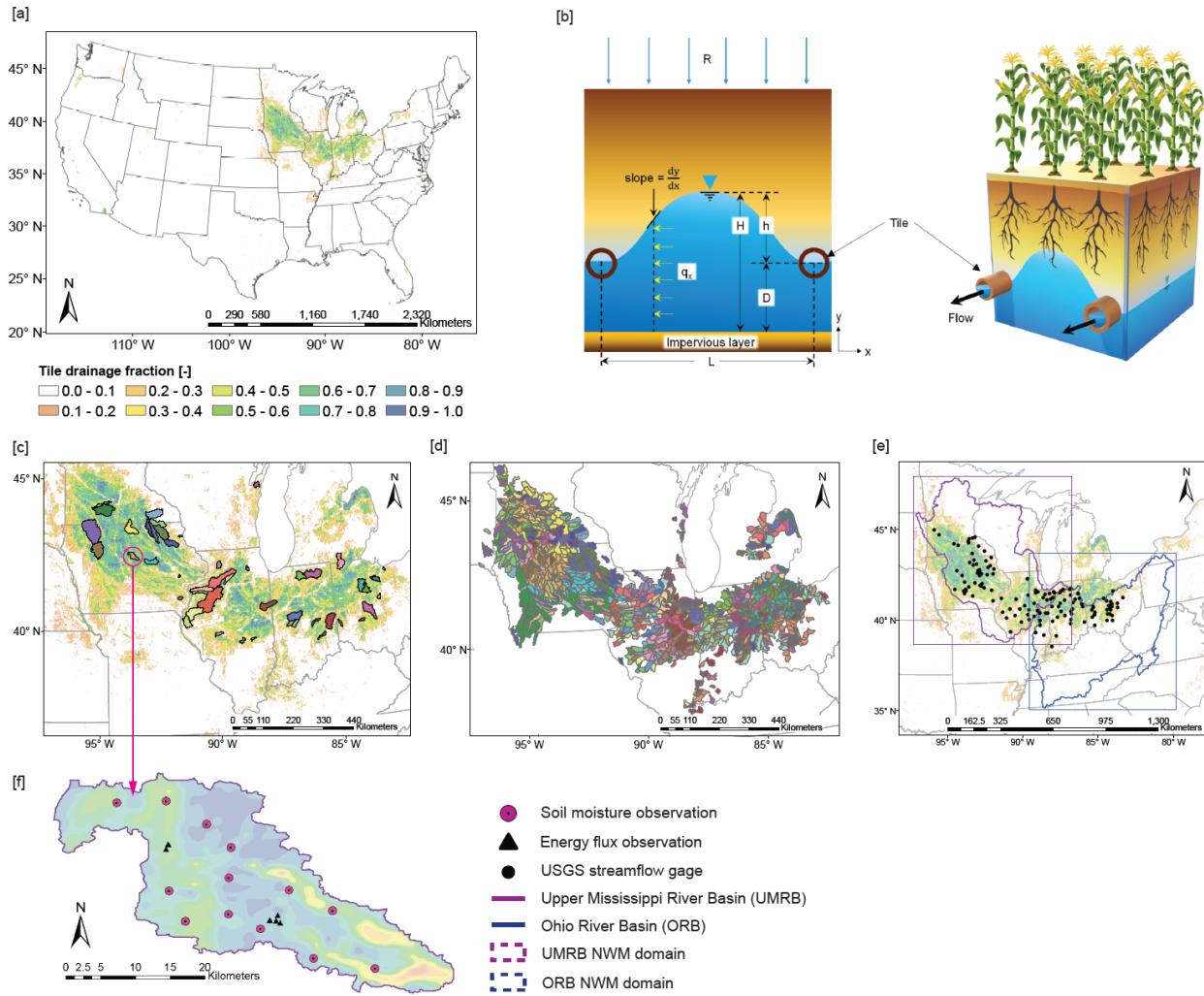
72 Ploeg, 2003). The above findings on the hydrologic impact of tile drainage are based on field-
73 level or small watershed-scale ($<50 \text{ km}^2$) studies. A comprehensive understanding of regional-
74 scale hydrology of tile-drainage is a major knowledge gap (Hansen et al., 2013; King et al.,
75 2014; Thomas et al., 2016). Accurate modeling of tile drainage impacts on the continental or
76 regional water cycle is a daunting challenge due to the lack of continental-scale high-resolution
77 tile drainage data and an efficient, fully distributed, continental-scale hydrology model with a tile
78 drainage scheme.

79 In the recent decade, the flood frequency and intensity have increased over the continental
80 United States (CONUS), especially over the Central US (Mallakpour and Villarini, 2015). To
81 provide flash flood forecasts and other hydrologic guidance with longer lead time and less
82 uncertainties, the National Weather Service (NWS) Office of Water Prediction (OWP) of the
83 National Oceanic and Atmospheric Administration (NOAA) developed a hydrologic modeling
84 framework, the National Water Model (NWM), to simulate observed and forecast streamflow for
85 about 2.7 million stream reaches of the CONUS. However, the NWM has considerable
86 uncertainties in the streamflow prediction over the Midwestern US (Dugger et al., 2017; Karki et
87 al., 2021). One of the reasons for the underperformance of the NWM can be the lack of
88 representation of subsurface tile drainage hydrology in the NWM (Hansen et al., 2013). Field-
89 level studies have already highlighted the importance of defining tile drainage within the
90 hydrologic models to achieve accuracy in simulated water budget components over heavily tile-
91 drained regions (Green et al., 2006; Hansen et al., 2013).

92 To address these shortfalls, in this study, we investigate the regional impact of tile drainage on
93 the NWM performance in simulating streamflow over the upper Midwestern US by developing a
94 new tile drainage scheme and implementing it into the NWM. We evaluate the NWM model
95 performance with tile drainage regarding the streamflow simulation with and without NWM
96 parameter calibration, and explore the influence of tile drainage on regional water budget and
97 regional hydrology. In these simulations, we use the recently developed 30-meter resolution
98 Agriculture Tile drainage data for the US (AgTile-US) (Valayamkunnath et al., 2020) to
99 explicitly define the tile-drained croplands within the NWM.

100 In section 2, we describe the details of the study area, process descriptions of the NWM and the
101 new tile drainage scheme, introduction to the input and evaluation data, calibration and

102 regionalization of model parameters, and details of model simulation experiments. Details of
 103 hydrological and statistical analysis used in this study to evaluate the model performance are
 104 presented in section 2.8. The results on the model performance evaluation, the impact of tile
 105 drainage on energy and water balance components, comparison with parallel works,
 106 perspectives, and limitations of the study are discussed in section 3.



107

108 **Figure 1.** The Study area. (a) The spatial distribution of tile drainage over the CONUS. The
 109 color grading in (a) indicate the tile drainage area fraction on a 1-km NWM grid. (b) Schematic
 110 representation of tile drainage and parameters of Hooghoudt's tile drainage equation, (c) NWM
 111 tile drainage calibration basins, (d) spatial distribution of regionalization HUC10s. In (d), color
 112 represent corresponding donor basin for the NWM parameters in (c). (e) Represents the two
 113 HUC2 basins identified for the regional NWM simulations. (f) The spatial distribution of soil
 114 moisture and energy flux observations in the South Fork Iowa River watershed, Iowa.

115

116 **2. Study area, modeling approach, and evaluation data**

117 **2.1 Study area description**

118 Our investigation on the influence of tile drainage on the NWM performance and regional
119 hydrology is based on the extensively tile-drained croplands of the upper Midwestern US (Figure
120 1 and S1). Considering computational-resource constraints, we focus on two subdomains with
121 extensive installations of tile drainage: The Upper Mississippi River Basin (UMRB) and the
122 Ohio River Basin (ORB) (Figure 1e). According to the AgTile-US tile drainage data
123 (Valayamkunnath et al., 2020), nearly 50% of total tile-drained croplands of the US are in the
124 UMRB, which accounts for 24.58% of the geographical area of the UMRB and 48% of the total
125 cropland area of the UMRB (Figure S1). Tile-drained croplands of ORB is about 17.2% of the
126 total tile-drained area of the US. Approximately 41.27% of the ORB croplands are tile-drained,
127 which covers 8.79% of the geographical area of the ORB. Together, UMRB and ORB account
128 for nearly 67% of the total tile drainage area of the US. Generally, the croplands of the upper
129 Midwestern region are characterized by moderately to very poorly drained soils and shallow
130 water tables (Barlage et al., 2021; Valayamkunnath et al., 2020). During the 2013-2019 period,
131 the annual average precipitation over UMRB and ORB are 1150 mm and 1370 mm, respectively.
132 Both basins receive the majority of the annual rainfall during the summer (June-August) season.

133 **2.2 The National Water Model (NWM)**

134 The NWM is a joint development between National Center for Atmospheric Research (NCAR)
135 and NOAA NWS to provide water prediction capabilities to advance resilience to water risks.
136 The core of the NWM is the NCAR Weather Research and Forecasting Hydrologic (WRF-
137 Hydro) model (Gochis et al., 2018). WRF-Hydro is a parallelized distributed hydrologic model
138 that is designed to simulate the land surface hydrology and energy states at relatively high spatial
139 resolution (usually 1-km or less). The NWM can either be forced offline (uncoupled) using
140 prescribed atmospheric forcing variables or coupled to the Advanced Research version of the
141 WRF (WRF-ARW) atmospheric model (Skamarock et al., 2008). Atmospheric forcing data
142 required for the model operation include incoming shortwave radiation (W m^{-2}), incoming
143 longwave radiation (W m^{-2}), specific humidity (kg kg^{-1}), air temperature (K), surface pressure
144 (Pa), liquid water precipitation rate (mm s^{-1}), and near-surface wind (both u and v components, m
145 s^{-1}).

146 The NWM uses the Noah-MP land surface model (Niu et al. 2011) to resolve land surface
147 processes and vertical fluxes of energy (sensible and latent heat, net radiation) and water (canopy
148 interception, infiltration, infiltration-excess, deep percolation) within the soil column on a 1-km
149 grid every 60 minutes. Infiltration excess, ponded water depth, and soil moisture are
150 subsequently disaggregated from a 1-km Noah-MP grid to a high-resolution, 250-m, NWM
151 routing grid using a time-step weighted method, and are then used in the subsurface and overland
152 flow terrain-routing modules (Gochis et al., 2018).

153 Prior to the overland flow routing, the NWM subsurface flow module computes the subsurface
154 lateral flow and resulting changes in the water table depth in the 2-m deep soil column using
155 Dupuit–Forcheimer assumptions (Gochis et al. 2018). If subsurface lateral flow fully saturates a
156 model grid, exfiltration is computed and added to the infiltration excess estimated by the Noah-
157 MP and routed as surface runoff. Overland flow is calculated at a 10-seconds time-step using a
158 fully unsteady, spatially explicit, diffusive wave routing formulation based on the steepest
159 gradient around each grid point (Julien et al. 1995). See Gochis et al. (2018) for more details of
160 the surface and subsurface routing schemes of NWM. As the surface flow reaches the grid
161 identified as a channel, it is mapped to the vector channel network and routed downstream using
162 Muskingum-Cunge channel routing formulation. In the NWM, vector channel networks are
163 defined using National Hydrography Dataset (NHD) Plus Version 2 (NHDPlusV2) channel
164 networks. A conceptual exponential bucket model is used to account for the contribution of
165 baseflow to total streamflow in the NWM. Aggregated drainage from the Noah-MP soil column
166 is mapped to a groundwater catchment corresponding to the NHDPlusV2 channel reach or
167 catchment topology. Using an exponential storage-discharge function NWM estimates
168 groundwater discharge for each NHDPlusV2 channel reach/catchment pair at hourly time steps
169 (Gochis et al. 2018).

170 **Table 1.** Calibrated NWM parameters in V2.0. (' \times ' in the values denote that the calibration parameter is a multiplier on the default
 171 value)

Parameter name	Description	Unit	Calibration value ranges (Minimum, Maximum)
BEXP	Pore size distribution index	dimensionless	($\times 0.40$, $\times 1.90$)
SMCMAX	Saturation soil moisture content (i.e., porosity)	volumetric fraction	($\times 0.80$, $\times 1.20$)
DKSAT	Saturated hydraulic conductivity	$m s^{-1}$	($\times 0.20$, $\times 10.00$)
RSURFEXP	Exponent in the resistance equation for soil evaporation	dimensionless	(1.00, 6.00)
REFKDT	Surface runoff parameter. Increasing REFKDT decreases surface runoff	unitless	(0.10, 4.00)
SLOPE	Linear scaling of "openness" of bottom drainage boundary	0-1	(0.00, 1.00)
RETDEPRTFAC	Multiplier on retention depth limit	unitless	(0.10, 20000.00)
LKSATFAC	Multiplier on lateral hydraulic conductivity (controls anisotropy between vertical and lateral conductivity)	unitless	(10.00, 10000.00)
Zmax	Maximum groundwater bucket depth	mm	(10.00, 250.00)
Expon	Exponent controlling rate of bucket drainage as a function of depth	dimensionless	(1.00, 8.00)
CWPVPT	Canopy wind extinction parameter for canopy wind profile formulation	m^{-1}	($\times 0.50$, $\times 2.00$)
VCMX25	Maximum carboxylation at 25°C	$\mu\text{mol m}^{-2} \text{s}^{-1}$	($\times 0.60$, $\times 1.40$)
MP	Slope of Ball-Berry conductance-to-photosynthesis relationship	unitless	($\times 0.60$, $\times 1.40$)
MFSNO	Melt factor for snow depletion curve; larger value yields a smaller snow cover fraction for the same snow height	dimensionless	($\times 0.25$, $\times 2.00$)
TD_SPAC	Tile drain spacing	m	($\times 0.25$, $\times 2.00$)

172

173 In this study, we use NWM version 2.0 (V2.0). The NWM has parameters that can be input into
174 the model as tables and grids and can be tuned or calibrated depending on the research
175 requirements. The list of important NWM V2.0 parameters identified by the NCAR to regionally
176 calibrate NWM (Dugger et al., 2017; Gochis et al., 2019) are listed in Table 1.

177 **2.3 Tile drainage scheme**

178 The current NWM lacks the representation of subsurface tile drainage. To compute tile drainage
179 runoff in the NWM, we implemented a simple analytic solution for subsurface flow to drains
180 based on Hooghoudt's tile-drainage model (Hooghoudt 1940; Ritzema, 1994). Hooghoudt's
181 model computes steady-state flow into the tile by applying Dupuit-Forchheimer assumptions for
182 horizontal flow in an unconfined aquifer and Darcy's Equation. The Hooghoudt's tile-drainage
183 model is computationally simple, and therefore is commonly used to compute the tile drainage
184 runoff in other models, especially in the DRAINMOD model (Skaggs, 1980) and Soil and Water
185 Assessment Tool (SWAT) model (Arnold et al., 1999; Guo et al., 2018; Moriasi et al., 2012).
186 Hooghoudt's steady-state equation that is implemented in the NWM is represented by Equation
187 1.

$$188 \quad q = \frac{8KDh + 4Kh^2}{L^2} \quad (1)$$

189 Where, q is the drainage discharge (m d^{-1}), K is the hydraulic conductivity of the soil (m d^{-1}), L is
190 the distance between tile drains, h is mid-point water table height above the tile drains (m) and D
191 is the height of tile drain from the bottom impervious layer (m) (Figure 1b). If the tile drain is
192 sufficient distance above the impervious layer, the streamlines will converge towards the tile
193 drain and thus no longer be horizontal. This results in longer flowlines and extra head loss. To
194 meet the Dupuit-Forchheimer assumptions of vertical equipotential lines and horizontal flow
195 streamlines and to correct for convergence head loss near the tile drains, D in Equation (1) is
196 replaced with the equivalent depth term (d_e) (Moody, 1967). The equivalent depth (d_e) represents
197 the imaginary thinner soil layer through which the same amount of water will flow per unit time
198 as in the actual situation (Ritzema, 1994). The value of d_e can be obtained using the analytical
199 equations developed from Hooghoudt's solutions as a function of L , D , and radius (r) of tile
200 drain (Moody, 1967) that are provided in Ritzema (1994).

Hooghoudt's model is a suitable option for the NWM framework because it considers most factors determining subsurface flow into tiles: K , L , D , soil profile depth, and water table elevation. Parameter K is already defined in the NWM. Default values of D , r and L are prescribed based on values reported by previous studies (Guo et al., 2018; Huffman et al., 2011; Moriasi et al., 2012; Panuska 2020; Schilling and Helmers 2008; Singh et al. 2006; 2007; Singh and Helmers 2008). The water table depth term, h is diagnosed at each model time-step using the degree of soil saturation simulated by Noah-MP. The tile drainage estimated by the Noah-MP at 1-km is then disaggregated onto a 250-m routing grid. In the NWM channel routing module, the lateral tile drainage runoff is mapped to the nearest vector channel network and routed downstream using Muskingum-Cunge channel routing formulation. We used the 30-meter resolution AgTile-US (Valayamkunnath et al., 2020) tile drainage map re-gridded to a 1-km NWM grid to define the tile-drained area within the model (Figure 1a).

2.4 Data

2.4.1 Observations

The study used hourly streamflow measurements from 188 United States Geological Survey (USGS) streamflow gages spanning across the heavily tile-drained croplands of the Upper Midwestern US (Figure 1c and 1e). These gages are selected from a list of USGS gages over the study area based on two criteria: 1) if the missing data in the streamflow time series is less than 20%, and 2) tile drainage fraction within the catchment is greater than 10%. To further examine the influence of tile drainage on evapotranspiration and soil moisture, we used *in-situ* measurements from the South Fork Iowa River watershed collected by the Agriculture Research Service of the United States Department of Agriculture (Coopersmith et al., 2015; 2021) (Figure 1f), including six sites with hourly flux measurements (latent and sensible heat fluxes) and 12 sites with daily soil moisture measurements. To validate the NWM simulated energy fluxes, we used daytime (9 am - 5 pm local time) hourly flux measurements.

2.4.2 Forcings for NWM

To drive the NWM, we used Analysis of Record for Calibration (AORC) high-resolution (1-km), near-surface, hourly meteorological forcing data (Kitzmiller et al., 2018) is available from 1979 to the present for the CONUS. The AORC delivers hourly accumulated precipitation and other meteorological surface parameters on a 0.0083° grid mesh. It provides superior temperature and

231 precipitation data than the widely-used National Land Data Assimilation System Version 2
232 (NLDAS2) meteorological forcings (Feng et al., 2019; Xia et al., 2012). The AORC is being
233 used as the primary source of forcing data for the calibration of the operational NWM by NCAR
234 and OWP (Feng et al., 2019). To derive high-resolution hourly precipitation, the AORC used
235 different sources of precipitation data such as Livneh (Livneh et al., 2013), NLDAS2 (Xia et al.,
236 2012), Stage IV (Lin and Mitchell, 2005), radar inputs, CMORPH (Joyce et al., 2004), and
237 Climate Forecast System Reanalysis (CFSR) (Saha et al., 2014). For temperature, Livneh,
238 NLDAS2, and Parameter Regression on Independent Slopes Method (PRISM) (Daly et al., 2002)
239 data were used. See Kitzmiller et al. (2018) for more details on the AORC meteorological
240 forcings. Other variables in AORC, including specific humidity, 10-m above ground wind
241 components, terrain-level pressure, surface downward shortwave (solar) radiation flux, and
242 longwave (infrared) radiation flux, were derived from NLDAS2.

243 Additional static data used for the NWM simulations include NLCD land cover (reclassified on
244 to USGS 27-class, 30-arc second), Hybrid STATSGO/FAO Soil Texture (19-class, 30-arc
245 second), and AgTile-US tile drainage map (30-m).

246 **2.5 Calibration of the NWM with a tile drainage scheme**

247 The key elements of an automated calibration workflow are the calibration data, objective
248 function, and the optimization algorithm employed to optimize the objective function in order to
249 minimize the model error (Gupta et al., 1998; Singh and Woolhiser 2002; Tolson and Shoemaker
250 2007). Following the actual NWM calibration procedure (Gochis et al., 2019), we calibrated
251 NWM against the USGS hourly streamflow data. The objective function used for the calibration
252 is provided in Equation 2. The standard Nash–Sutcliffe Efficiency (NSE) emphasizes the high
253 flow performance of the model due to squared error terms. However, combining NSE of log-
254 transformed streamflow with standard NSE provides an additional emphasis on low flows to
255 account for background model bias. During calibration, the objective function will be minimized.

$$256 \quad \text{Objective function} = 1 - \frac{(NSE + NSE_{LOG})}{2} \quad (2)$$

257 Here, NSE is the Nash-Sutcliffe Efficiency and NSE_{LOG} is the Log-transformed NSE (see Table
258 2 for more details).

Table 2. Evaluation metrics used for the performance evaluation of the NWM.

Metrics	Equation	Description
Pearson's Correlation (COR)	$r = \frac{\sum_{i=1}^n (m_i - \bar{m})(o_i - \bar{o})}{\sqrt{\sum_{i=1}^n (m_i - \bar{m})^2} \sqrt{\sum_{i=1}^n (o_i - \bar{o})^2}}$	Here, m_i and \bar{m} are the i^{th} value and mean of NWM simulated streamflow, respectively. o_i and \bar{o} are same as above but for the observation, and n is the length of streamflow series. Values greater than 0.5 are considered acceptable levels of performance. COR is used to capture the flow timing (Benesty et al., 2009; Moriasi et al., 2007). (Optimal value = 1)
Root mean squared error (RMSE)	$RMSE = \sqrt{\frac{\sum_{i=1}^n (m_i - o_i)^2}{n}}$	All terms have same meaning as above. But RMSE is used to capture the flow magnitude. (Optimal value = 0)
Percent bias (Bias)	$Bias = \frac{\sum_{i=1}^n (m_i - o_i) \times 100}{\sum_{i=1}^n o_i}$	All terms have same meaning as above. But Bias is used to capture the flow magnitude. (Optimal value = 0)
Nash-Sutcliffe Efficiency (NSE)	$NSE = 1 - \left[\frac{\sum_{i=1}^n (o_i - m_i)^2}{\sum_{i=1}^n (o_i - \bar{o})^2} \right]$	All terms have same meaning as above. Values between 0.0 and 1.0 are generally viewed as acceptable levels of performance. NSE can capture the flow timing and magnitude errors of the high flows (Moriasi et al., 2007; Nash and Sutcliffe, 1970). (Optimal value = 1)
Log-transformed Nash-Sutcliffe Efficiency (NSE_{LOG})	$NSE_{LOG} = 1 - \left[\frac{\sum_{i=1}^n (\log(o_i) - \log(m_i))^2}{\sum_{i=1}^n (\log(o_i) - \overline{\log(o_i)})^2} \right]$	All terms have same meaning as above. Values between 0.0 and 1.0 are generally viewed as acceptable levels of performance. NSE_{LOG} can capture the flow timing and magnitude errors of the low flows (Moriasi et al., 2007). (Optimal value = 1)
Weighted NSE (NSE_{WT})	$NSE_{WT} = \frac{(NSE + NSE_{LOG})}{2}$	All terms have same meaning as above. Values between 0.0 and 1.0 are generally viewed as acceptable levels of performance. NSE_{WT} is used to capture flow timing and magnitude errors for low flows and high flows. (Moriasi et al., 2007). (Optimal value = 1)
Kling-Gupta Efficiency (KGE)	$KGE = 1 - \sqrt{(r - 1)^2 + \left(\frac{\sigma_m}{\sigma_o} - 1\right)^2 + \left(\frac{\bar{m}}{\bar{o}} - 1\right)^2}$	Here, σ_m and σ_o are standard deviations in simulated and observed streamflow, respectively and other terms have same meaning as above. The range $-0.41 < KGE \leq 1$ could be considered as reasonable levels model performance. KGE is used to capture timing and magnitude errors. (Gupta et al., 2009; Knoben et al., 2019)

261 As in the official calibration strategy of the NWM V2.0, the Dynamically Dimensioned Search
262 (DDS) algorithm (Tolson and Shoemaker, 2007) is used in this study to optimize the objective
263 function. The algorithm is designed to scale the search in model parameter space to the user-
264 defined maximum number of iterations. The algorithm searches globally in its initial iterations
265 and then localizes the searches as the iterations approach the user-defined limit. The transition
266 from global to local search is attained by dynamically and probabilistically reducing the search
267 dimension in the neighborhood. See Tolson and Shoemaker (2007) for more details on DDS. In
268 this study, the maximum number of iterations is set to 300 for the NWM calibration.

269 Since the NWM simulations are data-, time-, and computationally-intensive, calibrating it for the
270 large river basins of the US in a single experiment is a cumbersome task. According to Feng et
271 al. (2019), about 1469 basins across the CONUS are identified from USGS GAGES II reference
272 basins, California Department of Water Resources (CADWR) basins, and NOAA NWS River
273 Forecast Centers (RFC) basins for the CONUS-scale calibration of the NWM V2.0. Calibration
274 basins are selected based on basin size, completeness of the streamflow observation record,
275 distribution within ecoregions level III (Omernik JM. 1995) and hydrograph characteristics in
276 comparison to other basins in the region. A basin is selected if the basin area is between 10 km²
277 and 20,000 km², streamflow data completeness is at least 50% for the calibration period, and the
278 basin has minimal human interventions (i.e., dams, road density, etc.) (Feng et al., 2019). To
279 calibrate NWM for the UMRB and ORB, we used a subset of 49 basins from V2.0 calibration
280 basins that have the tile-drainage area greater than or equal to 10% of the basin area (Figure 1c).

281 Before performing the calibration, we spin-up NWM for the selected 49 basins, separately, from
282 October 1, 2007, through October 1, 2019 period using the default model parameters. Using the
283 model state of October 1, 2019, as the “warm start,” we executed the model calibration from
284 October 1, 2007, through October 1, 2013. A separate 1-year spin-up from October 1, 2007,
285 through September 30, 2008, is considered for each iteration to match the model state to current
286 conditions and suppress most instabilities from parameter changes. The critical parameters of the
287 NWM (V2.0) related to soil, vegetation, runoff, snow, and groundwater and their description are
288 provided in Table 1 along with the most sensitive tile-drainage model parameter, the tile spacing
289 (L) parameter (Moriasi et al., 2012; Sammons et al., 2015; Guo et al., 2018). Using the best
290 parameters determined by the DDS algorithm, we ran the NWM from October 1, 2007, through

291 October 1, 2019. Model outputs for the water years 2007-2013 are discarded as spin-up and
292 calibration periods, and then we evaluated the model for all the 49 basins over the period
293 October 1, 2013, to October 1, 2019.

294 **2.6 Regionalization of calibrated NWM parameters**

295 The total area of the calibrated basins is less than 10% of the area of UMRB and ORB combined.
296 To compare the NWM performance with tile drainage and to quantify impacts of tile drainage on
297 regional hydrology, regional NWM simulation experiments are necessary. To execute the NWM
298 for regional domains presented in Figure (1e), appropriate parameters are required to be assigned
299 for each 1-km model grid cell in the study domain. The purpose of the parameter regionalization
300 is to transfer parameters from the calibration basins (donors) to the uncalibrated basins or 1-km
301 model grids (receiver) (Beck et al., 2016; He et al., 2011; Hrachowitz et al., 2013; Razavi and
302 Coulibaly, 2013). The most critical parts of the parameter regionalization process are identifying
303 donor basins for uncalibrated areas and choosing an optimal regionalization approach. We used
304 the regionalization based on maximum hydrological similarity (or minimum hydrologic distance)
305 to identify donor basins for uncalibrated areas (Beck et al., 2016; Garambois et al., 2015; Sellami
306 et al., 2014; Singh et al., 2014; Wallner et al., 2013). It is reasonable to assume that basins with
307 similar climate, topography, vegetation, geology and soil properties have identical NWM
308 parameters and produce similar hydrological responses. The hydrologic similarity or hydrologic
309 distance is measured by the Gower's distance metric (Gower, 1971).

310 To calculate the Gower's distance between donor and receiver basins, we considered several
311 attributes (see Table 3) based on the Hydrological Landscape Region (HLR) concept (Liu et al.,
312 2008; Winter, 2001; Wolock et al., 2004). Before using the Gower's distance metric, we
313 conducted a principal component analysis (PCA) to remove potential correlation between the
314 basin attributes. Each basin attribute is scaled to [-1, 1] by subtracting the mean and then
315 dividing by the standard deviation before the PCA. We used the following equation to quantify
316 the Gower's distance,

$$S_{ij} = \frac{\sum_{k=1}^n s_{ijk} \delta_{ijk}}{\sum_{k=1}^n \delta_{ijk}} \quad (3)$$

318 **Table 3.** Basin attributes used for characterizing hydrologic similarity in NWM 2.0 with tile drainage scheme

Category	Attribute	Notes
Landform	Percent flatland (total)	Total percent cover of flatland in the basin; flatland refers to areas with a slope of less than 0.01
	Percent flatland (upland)	Upland refers to areas above the middle elevation of the basin
	Percent flatland (lowland)	Lowland refers to areas below the middle elevation of the basin
	Relief	Difference between the highest and lowest elevations
	Circularity index	The ratio of the basin's area over the area of a circle with the same length of perimeter as the basin
Soil and geology	Percent sand	Mean percentage of sand in the soil column (upper 2m)
	Percent clay	Mean percentage of clay in the soil column (upper 2m)
	Depth to bedrock	Average thickness of soil
Land cover	Percent forest	Percent cover of forest (all types) in the basin
	Percent cropland	Percent cover of cropland (all types) in the basin
	Percent urban	Percent cover of urban areas in the basin
	Percent tile drainage	Percent cover of tile drained cropland in the basin
Climate	Feddema moisture index (FMI)	1-(PET/P) (if P>=PET) or (P/PET)-1 (if P<PET), where P & PET are annual mean precipitation and potential evapotranspiration, respectively. See Feddema (2005) and Leibowitz et al. (2016) for more details.

319

320

321 Where, S_{ijk} is the distance for variable k between a donor (i) and a receiver (j) and δ_{ijk} is the
322 weight on variable k . For numerical variables, values of S_{ijk} are estimated as the absolute
323 difference in the values of variable k between i and j , normalized by the range of variable k over
324 all observations. For categorical variables, S_{ijk} is assigned to 1 if i and j are equal on variable k
325 and 0 if they are not. The variables used in Equation (3) are the scores of the principal
326 components and weights (δ_{ijk}) are calculated based on the percentages of the total variance
327 explained by individual principal components. The receiver basins depicted in Figure (1d) are
328 extracted from USGS 10-digit Hydrologic Unit Code (HUC10) dataset. We selected 939 HUC10
329 basins over the upper Midwestern US with at least 10% tile drainage (i.e., 10% tile drainage
330 based on the total basin area) to regionalize the calibrated NWM parameters. For each HUC10
331 basin, we calculated Gower's distance from all the 49 calibration basins, identify a donor basin
332 based on minimum Gower's distance (i.e., maximum hydrologic similarity) and spatial distance
333 from the HUC10 basin, and finally transferred all the parameters to the HUC10 basins from their
334 respective donor basin. Using the shapefile of HUC10 basins and the NWM 1-km geogrid, we
335 mapped the parameters to the 1-km model domain. For areas with no tile drainage, we used the
336 parameters from the official NWM V2.0 calibration experiment by NCAR and OWP.

337 **2.7 Simulation experiments**

338 To examine the impact of tile drainage on the NWM performance and land surface hydrology,
339 we conducted the following NWM simulations for the UMRB and ORB regional domains.

- 340 a. *Default*: default NWM V2.0 without parameter calibration
- 341 b. *DefaultTD*: as in *Default*, but including the tile-drainage model
- 342 c. *Calib*: NWM V2.0 with calibrated parameters, mimicking the operational NWM
- 343 d. *CalibTD*: as in *Calib* but using the tile-drainage model with calibrated tile-space
344 parameter.

345 Similar to the calibration experiment, we spin-up all the four regional NWM experiments from
346 October 1, 2012, through October 1, 2019, before performing the analysis run. Using October 1,
347 2019 model state as the initial condition, we re-run the model from October 1, 2012, through
348 October 1, 2019. The first water year (i.e., the water year 2012) model outputs are discarded

349 from the analysis as we use this as an additional model spin-up period. Simulated streamflow
350 from model outputs is extracted for 139 USGS gage locations (Figure 1e). The results presented
351 in this study for the UMRB and ORB regional domains are only for October 1, 2013, through
352 October 1, 2019 period.

353 **2.8 Analysis**

354 The analyses conducted in this study to evaluate the model performance include hydrograph
355 analysis and statistical analysis using various statistical performance metrics provided in Table 2.
356 We evaluated the model simulated high flows, low flows, and streamflow events with
357 observations using hydrograph analysis. We derived high flows and low flows based on observed
358 streamflow quantiles. We split the observed and model estimated streamflow time series into 99
359 segments based on streamflow quantiles ranging from 1 to 100% for every observation. Low
360 flow is defined as streamflow below the median (50th quantile), and high flow is streamflow
361 above the median (see Figure S2 in the supporting information for graphical explanations). For
362 each quantile segment of the streamflow series, we estimated the model performance using
363 metrics listed in Table 3. To identify streamflow events, we use a recently developed R package
364 called “RNWMStat” (<https://github.com/NCAR/RNWMStat>) (Valayamkunnath et al., 2020).
365 RNWMStat can detect and match streamflow events from the observed and simulated
366 streamflow series.

367 The event detection algorithm in the RNWMStat follows a two-step procedure: first, the
368 algorithm smooths the streamflow time series (simulated or observed) using the local weighted
369 regression smoothing (LOESS) technique to remove high-frequency noises in the hydrographs;
370 second, it determines the start, peak, and endpoints of streamflow events from the first derivative
371 (i.e., rate of change) of smoothed streamflow series and remapped on to the original streamflow
372 series. We matched a simulated streamflow event with an observed event if the simulated peak of
373 an event is within the observed event (i.e., between the start and endpoints of an observed event).
374 For the matched events, we estimate peak bias (%), timing error of peak streamflow (hours),
375 event hit rate (%), and false alarm rate (%). Hit rate indicates the percentage of observed events
376 that the model predicts, and false alarm rate is the percentage of model events that are not
377 observed. For the event-based analysis, we used only the events with their peak greater than or
378 equal to the 90th percentile of streamflow. We used the Wilcoxon signed-rank test at 5%

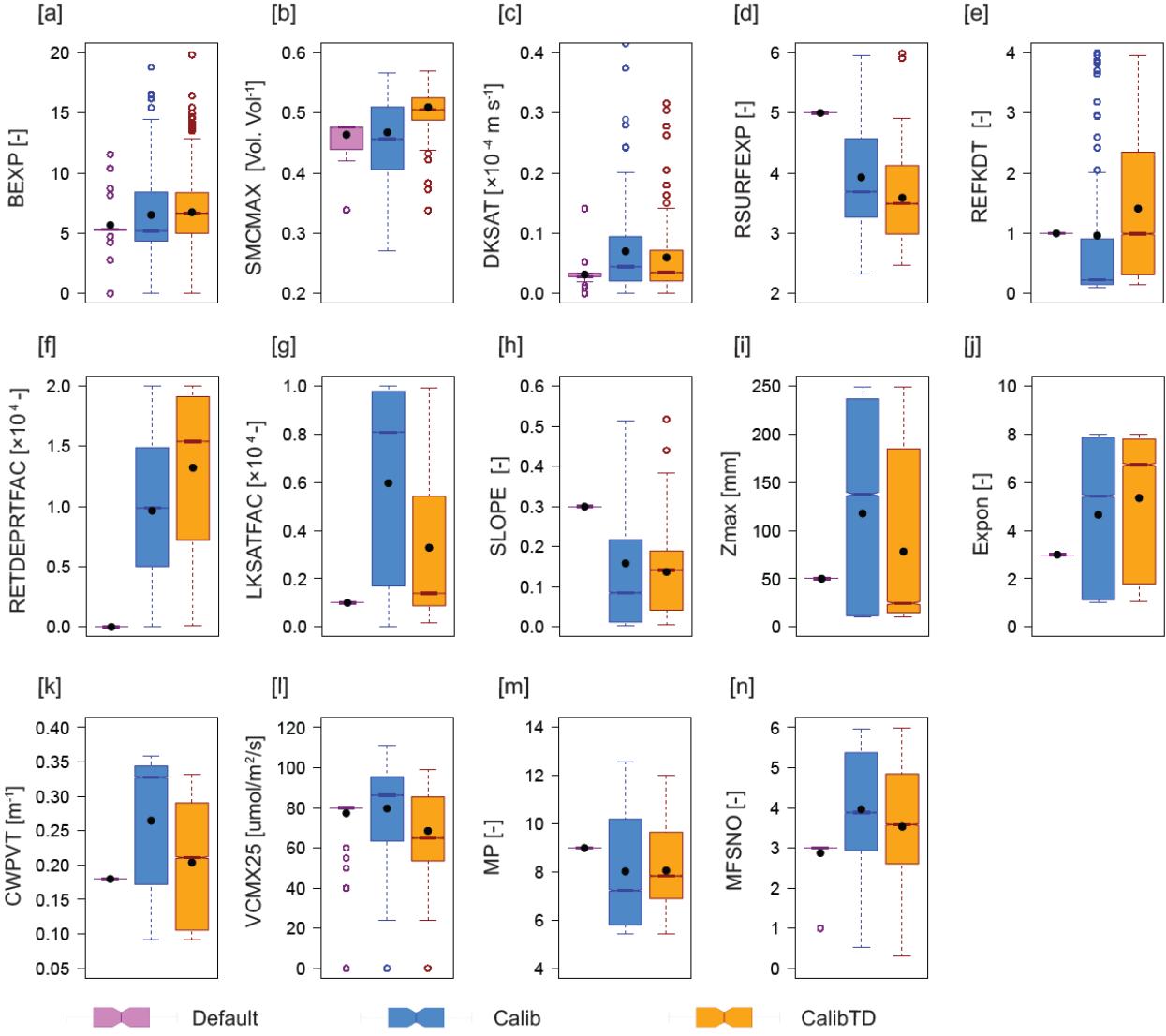
379 significance level to quantify the statistical significance of the median changes in the NWM
380 performance. The estimated p-values are provided in Table S1 to Table S3.

381 **3. Results**

382 **3.1 NWM calibration and parameter estimation**

383 The distributions of 14 sensitive parameters (Dugger et al., 2017; Gochis et al., 2019) from the
384 *Default*, *Calib*, and *CalibTD* are presented in Figure 2. The physical meanings of these
385 parameters are presented in Table 1. The new tile drainage scheme substantially altered the
386 distributions of the NWM parameters. In *CalibTD*, the soil column is relatively water-absorbing
387 or wetter than *Default* and *Calib*, because of its higher median values of pore size distribution
388 index (BEXP) and soil porosity (SMCMAX). We observed a significant reduction in direct soil
389 evaporation (RSURFEXP) and increase in infiltration (REFKDT) and surface water retention
390 depth (RETDEPRTFAC) in *CalibTD* ($p < 0.05$). Additionally, the degree of anisotropy in the
391 soil saturated hydraulic conductivity (LKSATFAC) is greatly reduced ($p < 0.05$) in *CalibTD*
392 compared to *Calib*. However, the estimated LKSATFAC for *CalibTD* is significantly higher
393 compared to *Default* ($p < 0.05$). Furthermore, the degree of openness in the bottom drainage
394 boundary (SLOPE) is slightly higher in *CalibTD* compared to *Calib*.

395 Based on STATSGO2 soil data, the dominant soil types of the study region are loam, silty clay
396 loam, and silt loam (Miller and White, 1998; USDA-NRCS, 2012). Overall, the *CalibTD*
397 parameters ranges are acceptable for the study region with a managed agriculture and above-
398 listed soil types (Clapp and Hornberger 1978; Lipiec et al., 2006; Livneh et al., 2015; Ma et al.,
399 2007; Miller and White, 1998). The distributions of the NWM parameters presented in Figure 2
400 suggest that *CalibTD* creates favorable conditions for low surface runoff rates, high infiltration
401 rates, a saturated soil column, and a shallow water table compared to *Calib* (Kalita et al., 2007).



402

403 **Figure 2.** The distributions of the NWM parameters from *Default*, *Calib* and *CalibTD*
404 experiments.

405

406 3.2 NWM performance evaluation: calibration and validation periods

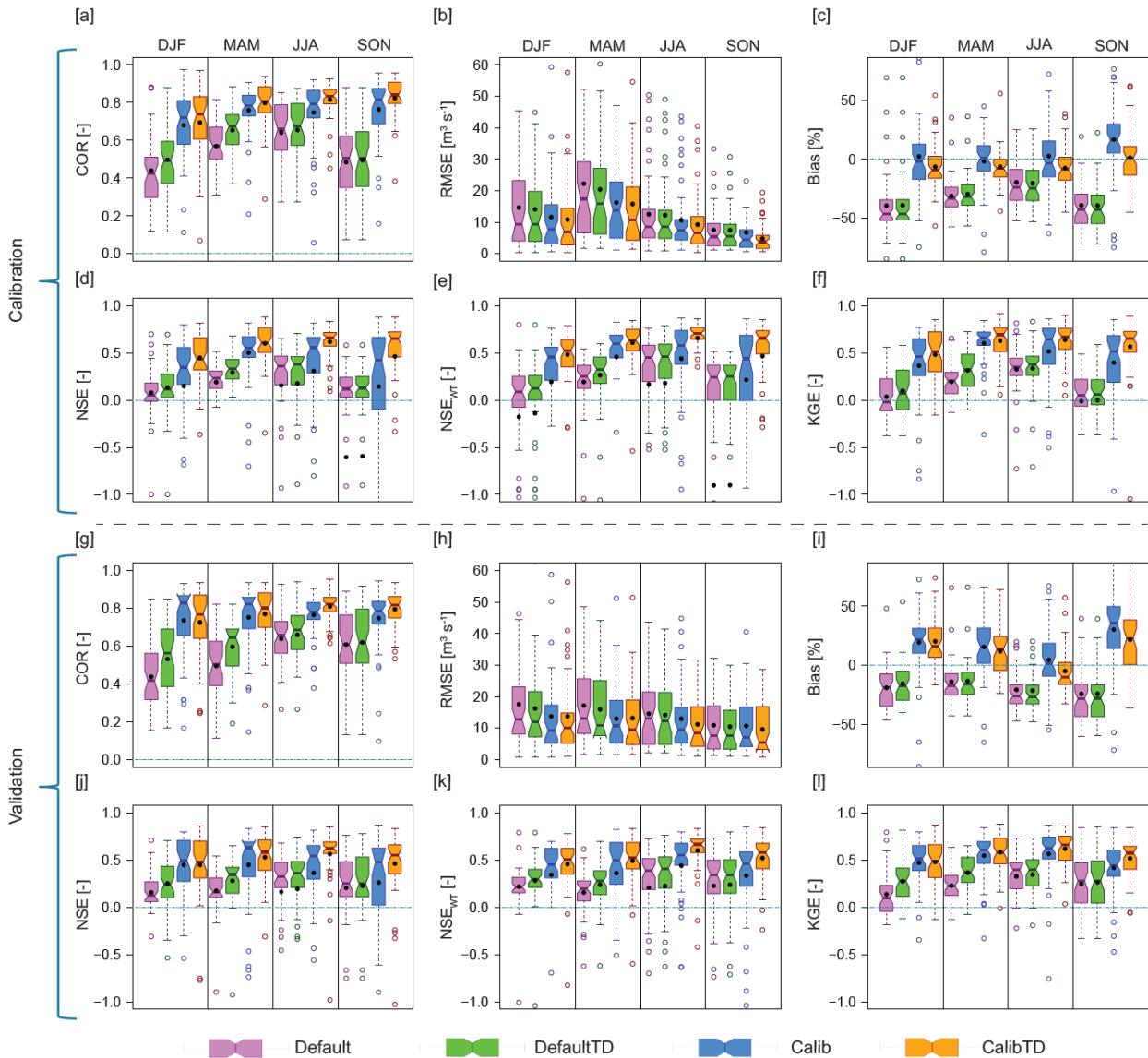
407 Seasonal distributions of NWM performance evaluation metrics for calibration and validation
408 periods are depicted in Figure 3. Representing the tile drainage process in the NWM improves
409 the model performance during the calibration period (Figure 3a-f). Examining the *DefaultTD*
410 model evaluation metrics indicated significant improvements in COR, NSE, NSEWT, and KGE
411 during all seasons than *Default* ($p < 0.05$). Furthermore, the median and spread of RMSE are
412 considerably reduced in *DefaultTD* during all seasons than *Default*. There are no considerable

413 differences in the estimated Bias between *Default* and *DefaultTD*. Overall, *DefaultTD*
414 performance is halfway between *Default* and *Calib*. That is, incorporating tile-drainage modeling
415 into NWM using default parameters (i.e., *DefaultTD*) enhanced the NWM performance by 20%
416 to 50% of the improvements attained by the fully-calibrated NWM (or *Calib*) from *Default* (e.g.,
417 for spring, the median NSE improved from 0.22 (*Default*) to 0.55 (*Calib*) in the non-tiled model,
418 and from 0.22 to 0.33 in the *Default* versus *DefaultTD*). The improvement seen in the *DefaultTD*
419 emphasizes the benefit of incorporating more physical process representation into hydrologic
420 models, rather than relying on calibration to compensate for model deficiencies, which ultimately
421 leads to uncertainty in model reliability across time (Andréassian, 2012; Gharari et al., 2014;
422 Ljung, 1999).

423 Compared to *Default*, the biggest improvement was brought by the *Calib* based on all the metrics
424 we considered (Figure 3a-f and Table S1). However, examining NSE, NSE_{WT}, and KGE
425 indicated that *Calib* has considerable discrepancies in the simulated streamflow over many
426 calibration basins. Based on the valid ranges of evaluation metrics presented in Table 2, the
427 performance of *Calib* is unacceptable in about 18%, 6%, 20%, and 30% of the calibration basins
428 during winter, spring, summer, and fall, respectively (Figure 3d-f). In *CalibTD*, these
429 underperforming basin percentage is reduced to 4%, 2%, 0%, and 6%, respectively for winter,
430 spring, summer, and fall. Additionally, we observed higher metrics medians with lower
431 variabilities for the *CalibTD*. Seasonal analysis indicated that the NWM performance is best
432 during summer and fall. It is due to the high amount of precipitation and streamflow during these
433 seasons. Overall, calibration of the NWM with a tile drainage scheme (i.e., *CalibTD*)
434 significantly improved the model performance than other model experiments ($p < 0.05$) (Figure
435 3a-f and Table S1). Despite the improvements seen in the *DefaultTD*, it was necessary to
436 calibrate to attain improved model performance.

437 Using the best parameters identified by the optimization algorithm, we executed the model for
438 the validation period. As shown in Figure 3g-i, the *DefaultTD* outperformed *Default*. The
439 improvements in NSE, NSE_{WT}, KGE, COR for the *DefaultTD* are significant ($p < 0.05$) during
440 winter and spring compared to *Default*. Similarly, *CalibTD* performed better than *Calib* during
441 the validation period (Figure 3g-I and Table S2), especially during summer and fall. Examining,
442 COR, NSE, and KGE indicated that *CalibTD* performed slightly worse during winter and spring

because it failed to reproduce the flow timings and peaks accurately. Biases in the timing and intensity of snowmelt can be another reason (Suzuki and Zupanski, 2018). Overall, incorporating the tile drainage process into the NWM substantially enhanced the accuracy of the NWM over heavily tile-drained basins in the upper Midwest.



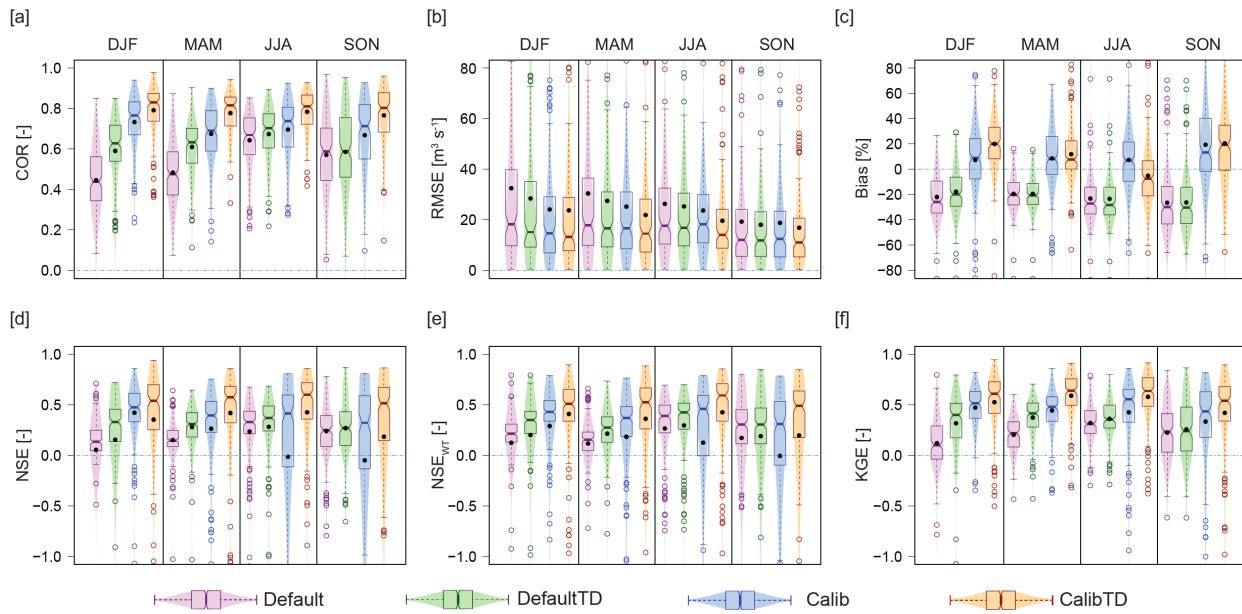
447

Figure 3. The NWM performance evaluation over 49 calibration basins for the calibration and validation periods. Comparison of the distribution of six evaluation metrics estimated based on the four NWM parameter experiments for the calibration (a-f) and validation (g-l) periods. Here, DJF=winter, MAM=spring, JJA=summer and SON=fall. Detailed descriptions of these metrics are provided in Table 2.

453

454 **3.3 NWM performance evaluation: Regional Simulation experiments**

455 By employing the regionalized parameters, we conducted the same set of four NWM simulations
 456 (see section 2.7) to quantify the influence of tile drainage on the NWM performance over the
 457 heavily tile-drained UMRB and ORB. The distributions of model evaluation metrics estimated
 458 using 139 USGS streamflow observations are provided in Figure 4. As mentioned earlier,
 459 *DefaultTD* is able to attain more than 50% of the improvement brought by the fully calibrated
 460 NWM from *Default* over the regional domain. It substantially enhanced the ability of NWM to
 461 capture the timing, peaks, and quantity of observed streamflow. The estimated RMSE for the
 462 *DefaultTD* is 3% to 17% less than that of the *Default*. The improvements we observed in NSE,
 463 NSE_{WT}, and KGE for the *DefaultTD* are significant ($p < 0.05$) compared to *Default* in all seasons
 464 except fall (Figure 4 and Table S3). Except for RMSE in all seasons, NSE_{WT} during summer and
 465 fall, and NSE during fall, all the model evaluation metrics for the *Calib* showed significant
 466 improvements from *Default* ($p < 0.05$) (Figure 4 and Table S3).



467

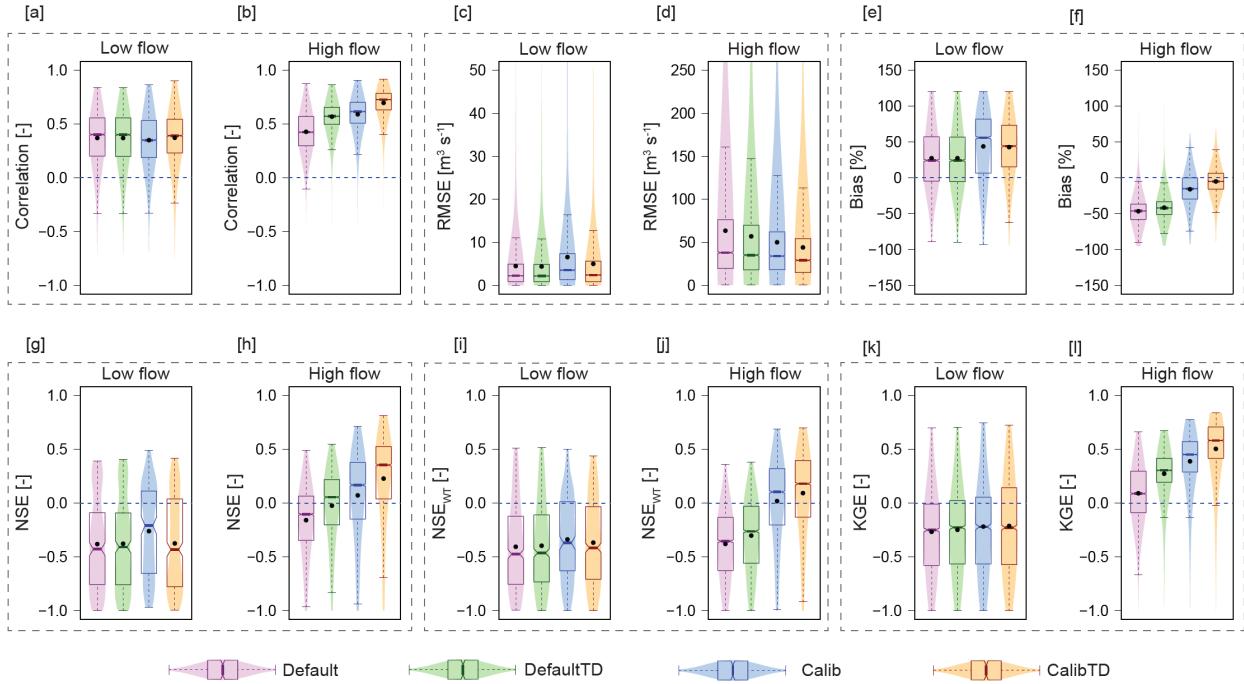
468 **Figure 4.** Seasonal NWM performance evaluation over the two HUC2 regional domains based
 469 on 139 USGS streamflow observations. Comparison of the distribution of six evaluation metrics
 470 estimated based on the four NWM parameter experiments for the regional simulation period (a-
 471 f). In (a-f), the color shading indicate the data distribution density.

472

473 One of the main focuses of this study is to quantify the impact of the tile drainage scheme on
474 calibrated NWM performance over the regional domain, and Figure 4a-f clearly shows a better
475 performance of the *CalibTD* than *Calib*. Seasonal distributions of the model evaluation metrics
476 showed significant ($p < 0.05$) improvements in the *CalibTD* performance in reproducing the flow
477 time, quantity, variance, and dynamics in the observed streamflow than in other model
478 experiments. RMSE in *CalibTD* is considerably reduced by 9% to 23% compared to *Calib*
479 (Figure 4b). However, *CalibTD* slightly overestimated (underestimated) streamflow during
480 winter (summer) compared to observation and *Calib*, but there are no significant differences
481 between them for spring and fall (Figure 4c and Table S3).

482 **3.3.1 Hydrograph analysis**

483 To understand the causes of discrepancies in the NWM simulated streamflow (mainly Bias and
484 RMSE), we conducted hydrograph analysis using the NWM simulated streamflow from four
485 experiments and observations. Results of the high-flow and low-flow hydrograph analysis are
486 presented in Figure 5. The median values of performance metrics estimated for the low-flows are
487 almost the same for *Default* and *DefaultTD* (Figure 5a, c, e, g, i, and k). The median low-flow
488 Bias estimated for *Calib* is twice that of *Default* (Figure 5e). Even though *CalibTD* reduced low-
489 flow biases compared to *Calib*, it still overestimated low-flows by 50%. Analyzing the
490 distributions of NSE (Figure 5g), NSEWT (Figure 5i), and KGE (Figure 5k) indicated that the
491 NWM, in general, failed to reproduce observed low-flow accurately, consistent with previous
492 studies assessing the NWM performance in estimating low-flows have reported similar findings
493 (Hansen et al., 2019; Jachens et al., 2021; Karki et al., 2021). One of the reasons for the
494 overestimation of low-flows can be the high groundwater recharge (deep percolation loss) rate in
495 the NWM (Karki et al., 2021). The existing groundwater scheme in the NWM represents surface
496 water–groundwater connectivity using a one-way connection from the underlying aquifer to the
497 stream channel and omitted the influences of the stream on groundwater, and ignoring the two-
498 way stream–aquifer fluxes in the NWM lead to overprediction of low flows (Jachens et al.,
499 2021). Our results indicate significant reductions in the low-flow Bias and RMSE in *CalibTD*
500 compared to *Calib*. Because tile drainage substantially reduced the groundwater recharge and
501 rerouted the saturated soil water into the stream directly (see section 3.4 for more detailed
502 discussion).



503

504 **Figure 5.** Evaluation of the NWM simulated high-flows and low-flows based on regional
 505 simulation. The model performance metrics are calculated by comparing the NWM estimates
 506 with 139 USGS streamflow observations. In (a-l), the color shading behind the boxplot indicate
 507 the data distribution density.

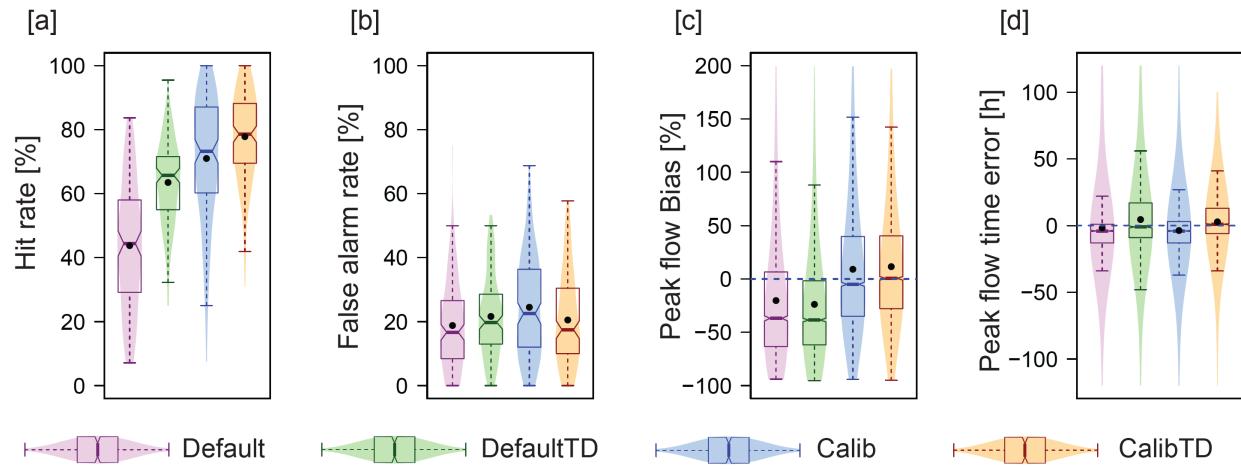
508

509 Results on high-flows revealed considerable improvements in the *DefaultTD* and *CalibTD*
 510 performance over the regional domain (Figure 5b, d, f, h, j, and l). As we highlighted before,
 511 *DefaultTD* significantly ($p < 0.05$) improved the high-flow performance of the NWM compared to
 512 *Default* by increasing COR by 0.15, NSE by 0.16, and KGE by 0.22. Furthermore, *DefaultTD* is
 513 able to reduce RMSE by $-2.84 \text{ m}^3 \text{s}^{-1}$ and improve the Bias by 4.2%. The variability in the model
 514 performance metrics is considerably lower in *DefaultTD* compared to *Default*. *Calib*
 515 substantially enhanced performance in reproducing the observed high-flow characteristics than
 516 *Default*. Analyzing the evaluation metrics of *Calib* indicated a significant ($p < 0.05$) increase in
 517 COR by 0.19, NSE by 0.27, NSE_{WT} by 0.46, and KGE by 0.36 than in *Default*. *Calib* can better
 518 capture the timing and magnitude of observed high-flows with reduced mean error compared to
 519 *Default*. *CalibTD* further enhanced the accuracy in estimating the observed high-flow
 520 characteristics by significantly increasing COR by 0.11, NSE by 0.19, and KGE by 0.13 in
 521 *CalibTD* compared to *Calib* (Figure 5b, h, and l). Furthermore, *CalibTD* reduced the mean error

522 by $4.88 \text{ m}^3\text{s}^{-1}$ and Bias by 10% (Figure 5d and f). Overall, the NWM with *CalibTD* is able to
 523 better capture the timing, magnitude, and dynamics of observed high-flows very well compared
 524 to other experiments.

525 **3.3.2 Event-based evaluation**

526 One important goal of the NWM is to provide flash flood forecasts with longer lead times and
 527 reduced uncertainties. Thus, we analyzed the performance of NWM to capture the different
 528 characteristics of observed streamflow events using 139 USGS gage measurements. Event-based
 529 metrics estimated for different NWM experiments are presented in Figure 6. *Default* is able to
 530 reproduce about 44% of the observed streamflow events (Figure 6a). The *DefaultTD*
 531 significantly increased the event hit rate by 47% ($p<0.001$) than *Default*, and also reduced the
 532 variability in the hit rate. *Calib* significantly enhanced the hit rate of NWM by 67% ($p<0.001$)
 533 compared to *Default*. Among the four NWM experiments considered, *CalibTD* showed the
 534 highest streamflow event hit rate. The estimated hit rate in *CalibTD* is 78%, which is 7% higher
 535 than *Calib*. Moreover, the spread in the hit rate estimated for *CalibTD* is considerably lower than
 536 that of *Calib* (Figure 6a). The median false alarm rate in *Calib* is 22.5%. But in *CalibTD*, the
 537 false alarm rate is substantially reduced to 17.5% (Figure 6b).



539 **Figure 6.** Event-based evaluation of the NWM based on regional simulation. The event-based
 540 statistics are calculated by comparing the NWM estimates with 139 USGS streamflow
 541 observations. In (a-d), the color shading behind the boxplot indicate the data distribution density.

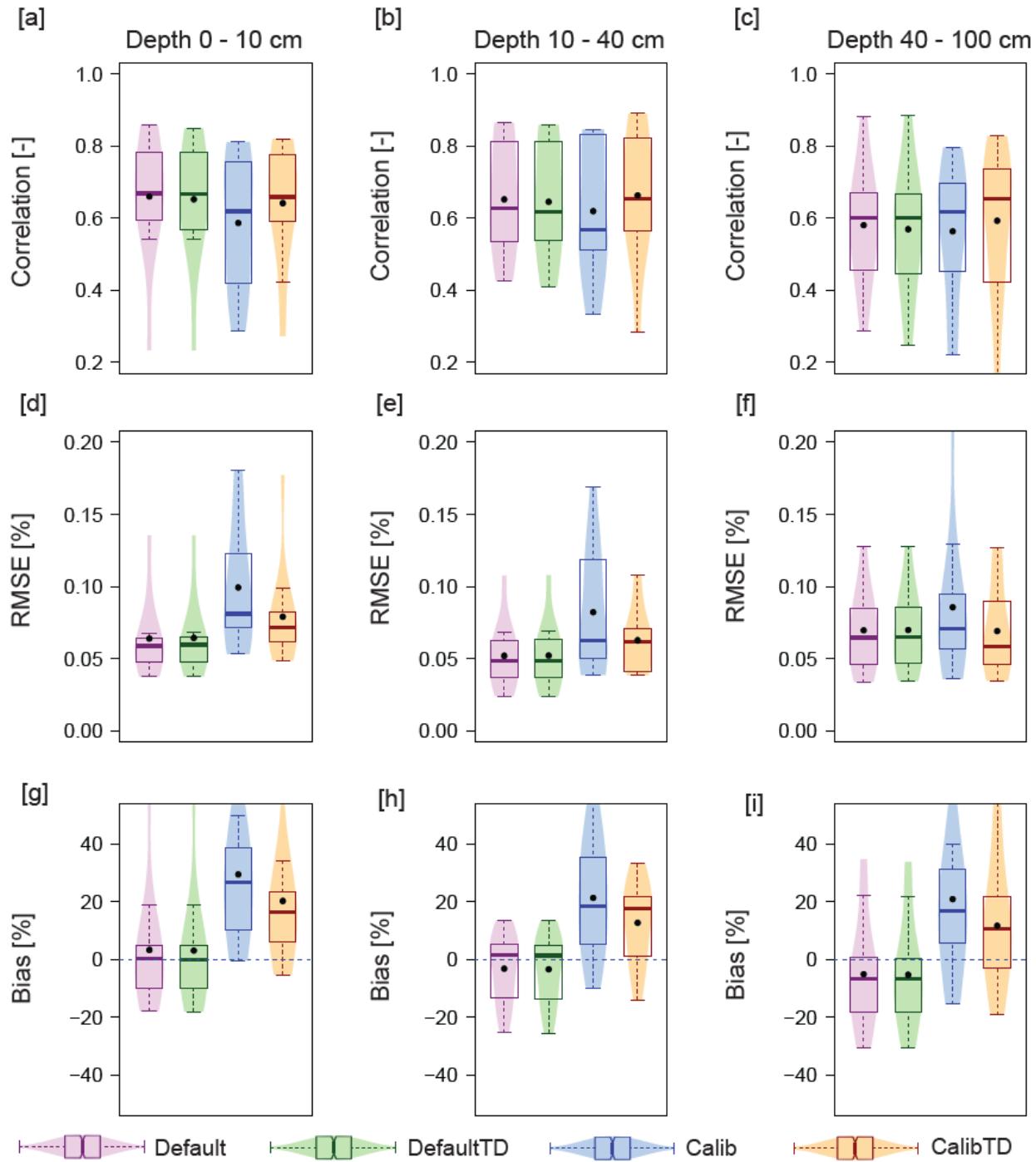
542

543 Tile drainage can significantly impact the peaks and timings of streamflow events, with an
544 earlier peak of greater magnitude (Rahman et al., 2014; Robinson et al., 1985), so we also
545 quantified the NWM's ability to capture the peak flows, and timing of peak flows for each
546 streamflow event. The estimated peak flow bias (%) and peak flow timing error (h) from
547 different NWM experiments are presented in Figures 6c and 6d, respectively. There is no
548 considerable difference between *Default* and *DefaultTD* in the estimated peak flow bias.
549 However, *CalibTD* outperformed *Calib* and produced a lower peak flow bias of 0.57% compared
550 to 5% in *Calib*. The median values of the estimated peak flow timing error are -3h, 0h, 4h, and
551 2h for *Default*, *DefaultTD*, *Calib*, and *CalibTD*, respectively. Overall, the event-based
552 streamflow analysis indicated that NWM with *CalibTD* outperformed other NWM experiments
553 over the heavily tile-drained UMRB and ORB. Our findings are consistent with previous studies
554 in that the model performance to simulate streamflow over a heavily tile-drained watershed was
555 considerably improved when they incorporated tile drainage into the model (Green et al., 2006;
556 Hansen et al., 2013; Robinson et al., 1985; Wiskow and van der Ploeg, 2003).

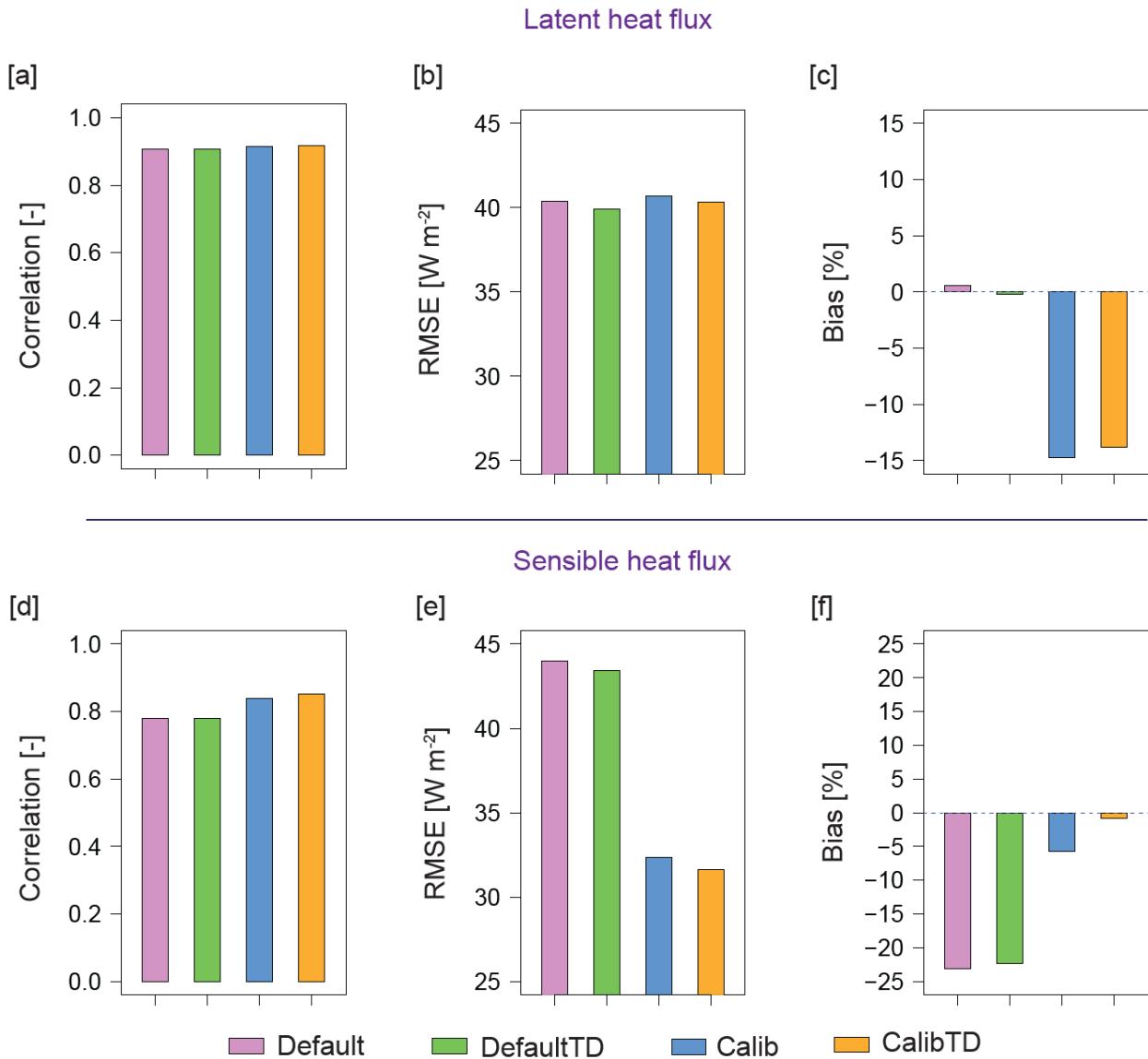
557 3.3.3 Soil moisture evaluation

558 In addition to streamflow, tile drainage modifies the soil water storage. We evaluated the NWM
559 performance using soil moisture measurements (volumetric) from 12 sites in the South Fork
560 Iowa River watershed (Figure 1f). Using the soil moisture measurements from three different
561 depths and NWM estimates at three model levels, we estimated COR, RMSE, and Bias in the
562 model estimated soil moisture (Figure 7). The NWM performance in estimating the soil
563 moisture using *Default* and *DefaultTD* is nearly identical regarding the medians of COR, RMSE.
564 Both *Default* and *DefaultTD* showed higher median COR (0.68) and zero median Bias for the
565 first soil layer (0-10 cm) of the NWM. A lower COR (0.60) and Bias (8%) and higher RMSE
566 (0.062%) are estimated for the third soil layer of the NWM. Calibration substantially impacted
567 the performance of the NMW to estimate soil moisture. For instance, *Calib* significantly reduced
568 the NWM performance compared to *Default* by degrading COR, increasing RMSE, Bias, and
569 their variance. This is not surprising, because the model was calibrated to optimize streamflow
570 prediction. Although *CalibTD* underperformed compared to *Default* and *DefaultTD*, it produced
571 better estimates of soil moisture compared to *Calib*. Also, the medians of COR, RMSE, and Bias

572 are significantly improved, and their variances are reduced when NWM employed *CalibTD*
 573 instead of *Calib*.



574
 575 **Figure 7.** Evaluation of the NWM simulated soil moisture with field measurements. In (a-i), the
 576 color shading behind the boxplot indicate the data distribution density.



577

578 **Figure 8.** Accuracy assessment of NWM simulated energy balance components. (a-c) Represent
 579 the evaluation of NWM simulated latent heat fluxes (evapotranspiration), (d-f) same as (a-c), but
 580 for sensible heat fluxes.

581

582 3.3.4 Energy flux evaluation

583 Using the eddy covariance flux measurements from seven sites in the South Fork Iowa River
 584 watershed (Figure 1f), we evaluated the NWM simulated hourly sensible heat (SH) fluxes and
 585 latent heat (LH) fluxes (equivalent to evapotranspiration). Results of the energy flux analysis are
 586 presented in Figure 8. The results shown in Figure 8 are the averaged values of evaluation

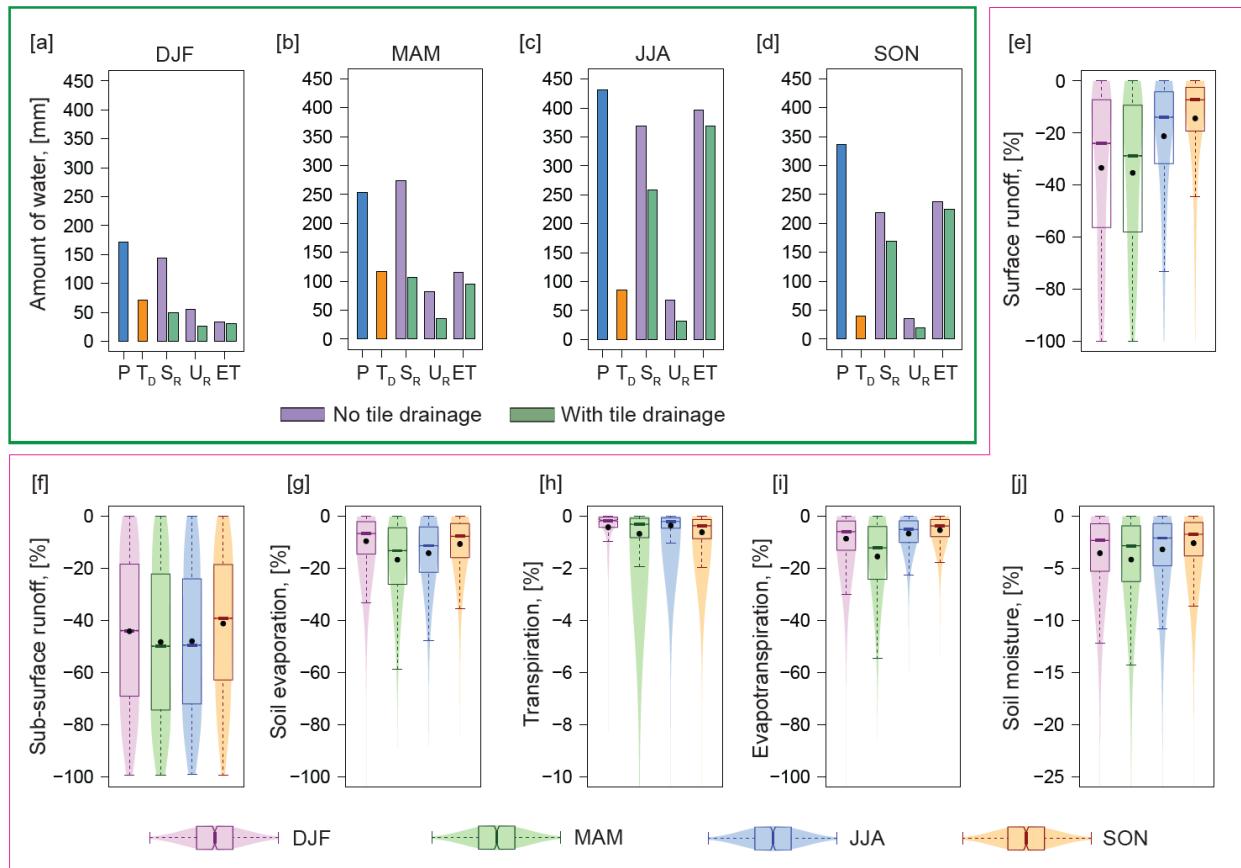
metrics estimated for the observation sites. The estimated COR and RMSE of LH for all the four NWM experiments are almost identical. Despite high correlation, the NWM estimated LH incurred a high mean error ($\sim 40 \text{ W m}^{-2}$) (Figure 8b). NWM with *Default* and *DefaultTD* produced better estimates of LH with Bias equal to $\pm 1\%$. However, *Calib* and *CalibTD* noticeably underestimated LH by -15% and -14%, respectively. In the case of SH, *CalibTD* outperforms other NWM experiments with higher COR (0.83) and lower RMSE (32 W m^{-2}) and Bias (1%). *Calib* considerably enhanced the NWM performance in SH estimation compared to *Default* and *DefaultTD*. However, *Calib* slightly underperformed compared to *CalibTD*. Even though there are discrepancies in the NWM estimated SH and LH, our results of LH and SH indicate that the performance of the NWM is acceptable (see Table 2 for metrics ranges).

3.4. Effect of tile drainage on regional hydrology

To quantify the effects of tile drainage on regional hydrology, we analyzed land surface water balance. For this purpose, we conducted one additional NWM simulation with *CalibTD* parameters and deactivated the tile drainage scheme. This simulation with a deactivated tile drainage scheme is designated as “No tile drainage,” (which is not equal to *Calib* as it uses *CalibTD* parameter set) and the NWM with *CalibTD* is defined as “With tile drainage” in this section. The results of the seasonal water balance analysis are presented in Figure 9. The results shown in Figure 9a-d are the averaged values of water balance components estimated for the tile-drained grids of the NWM within UMRB and ORB. The maximum amount of tile drainage over UMRB and ORB occurred during spring ($117 \pm 50 \text{ mm}$) followed by summer ($85 \pm 32 \text{ mm}$), winter ($71 \pm 40 \text{ mm}$), and fall ($40 \pm 20 \text{ mm}$) (Figure 9a-d). Values in the parenthesis indicate mean and one spatial standard deviation. The ratio of tile-drained water (T_D) to precipitation (P) is highest during spring (0.46), followed by winter (0.41), summer (0.20), and fall (0.12).

The results shown in Figure 9e-j are the distributions of percentage changes in the average values of water balance components that are calculated for each tile-drained grid of the NWM within UMRB and ORB. Analyzing seasonal distributions of surface runoff (S_R) changes indicated a significant decrease in S_R due to tile drainage (Figure 9e), which is consistent with previous studies (Natho-Jina et al., 1987; Robinson et al., 1985; Robinson and Rycroft, 1999; Skaggs et al., 1994). Following the seasonal tile drainage pattern, the highest decline in S_R is estimated for spring (-29%), followed by winter (-24%), summer (-14%), and fall (-7%). Tile drainage

617 significantly decreased subsurface runoff or groundwater recharge (U_R) for all the seasons we
 618 considered (Figure 9f). This is similar to the findings of Golmohammadi et al. (2017). However,
 619 a maximum decrease is identified during spring (-50%) and summer (-50%). During winter and
 620 fall, U_R decreased by -43% and -39%, respectively. The impact of tile drainage on S_R is higher
 621 than U_R because tile drainage increases infiltration. However, all the saturation water from the
 622 infiltration are not removed by the tile drainage and a considerable amount of saturation water
 623 (5% to 10%) is still available to U_R .



624
 625 **Figure 9.** Impact of tile drainage on the NWM water balance components. (a-d) The seasonal
 626 totals of precipitation (P), tile drainage (T_D), surface runoff (S_R), underground runoff or
 627 groundwater recharge (U_R), and evapotranspiration (ET). The values represented in (a-d) are the
 628 averages of all the NWM tile-drained grids in the UMRB and ORB. (e-j) The changes in water
 629 balance components due to tile drainage. The results presented in (e-j), are estimated as “with tile
 630 drainage” minus “no tile drainage”. In (e-j), the color shading behind the boxplot indicate the
 631 distribution density.

632 The main components of evapotranspiration (ET) are direct soil evaporation, transpiration, and
633 canopy evaporation. Our analysis indicated that tile drainage significantly impacted soil
634 evaporation (Figure 9g). The seasonal distributions of soil evaporation changes showed a more
635 significant decrease in spring (-13%) and summer (-11%) ($p < 0.05$). The reduction in soil
636 evaporation estimated for winter and fall are -7% and -8%, respectively. Since the results on
637 transpiration indicated minimal changes (<1%) due to tile drainage, the estimated seasonal
638 changes in ET are almost equal to soil evaporation (Figure 9i). Studies of Khand et al. (2017),
639 Kjaersgaard et al. (2014), and Yang et al. (2017) based on remote sensing and eddy covariance
640 ET measurements from tile-drained croplands of the US reported similar findings on ET
641 changes. Furthermore, we also evaluated the impact of tile drainage on root-zone soil moisture.
642 Our results indicate that the soil moisture considerably decreased by 2% to 3% due to tile
643 drainage. Similar findings were previously reported by many studies (Fausey 2005; Fraser and
644 Flemming, 2001; King et al., 2014).

645 Additionally, we quantified the impact of tile drainage on streamflow by comparing “No tile
646 drainage” with “With tile drainage”. Tile drainage substantially altered the streamflow events by
647 increasing peaks by 14%, increasing volume by 2.3%, delaying event start time by 2 hours, and
648 reducing the end time by 7 hours. As indicated by previous studies, tile drainage is responsible
649 for more short-term flashy streamflow events (De Schepper, 2017; Miller and Lyon, 2021;
650 Rahman et al., 2014; Robinson et al., 1985). Our results indicated a considerable increase in
651 seasonal streamflow volume due to tile drainage. The highest increase is estimated for winter
652 (17%), followed by spring (13%), fall (13%), and summer (2.8%). Moreover, our analysis found
653 that tile drainage enhanced the baseflow volume by 11.52%, which consistent with findings from
654 previous studies (King et al., 2014; Moore and Larson, 1980; Schilling and Libra, 2003).
655 However, the baseflow index is estimated as the ratio of total baseflow to the total streamflow is
656 decreased by -9.10%. In other words, the impact of tile drainage on direct runoff (or quick flow)
657 is more substantial compared to baseflow (Miller and Lyon, 2021). Overall, tile drainage has
658 significant effects on most of the water balance components in the study domain.

659 4. Conclusion

660 The purpose of the study is to quantify the impacts of representing subsurface tile drainage on
661 the National Water Model’s simulated regional hydrology. We implemented Hooghoudt’s tile

662 drainage scheme into the NWM V2.0 and used 30-m resolution AgTile-US to identify tile-
663 drained grids within the model domain. We followed the operational NWM calibration approach
664 and calibrated 14 sensitive NWM parameters (Dugger et al., 2017; Gochis et al., 2019) along
665 with tile spacing. Overall, the changes in these parameters suggested a water-absorbing soil
666 column with higher infiltration rates and moisture storage potential. The calibration results also
667 indicated reduced surface runoff and evapotranspiration over the tile-drained croplands.

668 Representing the tile drainage process in the NWM significantly improved its performance in
669 estimating streamflow over the UMRB and ORB. More interestingly, the NWM with
670 uncalibrated parameters but including a tile drainage scheme (i.e., *DefaultTD*) attained 20% to
671 50% of the improvements brought by the calibrated NWM (*Calib*) from *Default*. The *CalibTD*
672 outperformed other experiments with reduced RMSE, Bias, and increased NSE, COR, and KGE.
673 Furthermore, *CalibTD* accurately captured the dynamics in magnitude, timing, and variability of
674 observed streamflow, especially the high-flows and low-flows. Tile drainage substantially
675 increased peak flows, baseflow, and event volume. This significantly enhanced accuracy of the
676 NWM to simulate high-flows in *CalibTD*. Even though *CalibTD* produced better estimates of
677 low-flows than *Calib*, there is considerable uncertainty in the estimated low-flow timings and
678 magnitudes. The overestimation of low-flows by the NWM can be caused by high groundwater
679 recharge rates or lack of realism in the groundwater scheme in the NWM. Despite these
680 discrepancies, NWM with a tile drainage scheme better estimates soil moisture, latent heat fluxes
681 (or evapotranspiration), and sensible heat fluxes for the tile-drained croplands.

682 We quantified the impact of tile drainage on different water balance components, and our results
683 indicated a significant decrease in the surface runoff, underground runoff or groundwater
684 recharge, and evapotranspiration over UMRB and ORB. The impact of tile drainage on direct
685 runoff (or quick flow) is more profound than on baseflow. The drainage of saturated water from
686 the soil column by the subsurface tiles reduced the deep percolation of free water into the
687 groundwater reservoir (Golmohammadi et al., 2017). Tile drainage removed saturated water
688 from the soil column above the tiles and increased soil storage potential (Rahman et al., 2014).
689 The decrease in ET over the tile drained croplands is mainly due to reduced direct soil
690 evaporation resulting from low soil water content (Moriasi et al., 2012; Rahman et al., 2014).

691 Overall, tile drainage has a significant impact on regional hydrology. The representation of tile
692 drainage process in the NWM can enhance the model's accuracy to estimate the dynamics of
693 streamflow mainly, the timing, peaks, and volume of streamflow over a heavily tile-drained
694 basin. Thus, our findings demonstrate the importance of incorporating tile drainage into the
695 operational NWM for accurate flood forecasts.

696 **Data and Code Availability Statement**

697 All data used to generate the major figures are publicly available. The AORC data are accessed
698 from <https://hydrology.nws.noaa.gov/pub/aorc-historic/>. The USGS streamflow data are
699 available at: <https://waterdata.usgs.gov/nwis/inventory/>. The NLCD land cover data are available
700 at: <https://www.mrlc.gov/data/>. The AgTile-US 30-m tile drainage map is available at:
701 <https://figshare.com/articles/dataset/AgTile-US/11825742/>. NHDPlusV2 data can be accessed
702 from https://nhdplus.com/NHDPlus/NHDPlusV2_data.php. The South Fork Iowa River
703 watershed soil moisture and flux data are obtained from Coopersmith et al. (2015; 2021)
704 (<https://hrsl.ba.ars.usda.gov/southfork/index.html>). The NWM source code used in this study is
705 publicly available at: https://github.com/NCAR/wrf_hydro_nwm_public/. The RNWMStat R-
706 Package is available at: <https://github.com/NCAR/RNWMStat/>.

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