# Identifying and quantifying the impact of climatic and non-climatic drivers on river discharge in Europe

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#### Abstract

Our water resources have changed over the last century through a combination of water management evolutions and climate change. Understanding and decomposing these drivers of discharge changes is essential to preparing and planning adaptive strategies. We propose a methodology combining a physical-based model to reproduce the natural behavior of river catchments and a parsimonious model to serve as a framework of interpretation, comparing the physical-based model outputs to observations of discharge trends. We show that over Europe, especially in the South, the dominant explanations for discharge trends are non-climatic factors. Still, in some catchments of Northern Europe, climate change seems to be the dominating driver of change. We hypothesize that the dominating non-climatic factors are irrigation development, groundwater pumping and other human water usage, which need to be taken into account in physical-based models to understand the main drivers of discharge and project future changes.

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# Key Points:

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- Attribution of streamflow trends to climate drivers
- Larger effect on streamflow of non-climatic drivers
- Strong impact of human activities, especially over Spain

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#### 11 Abstract

Our water resources have changed over the last century through a combination of wa-12 ter management evolutions and climate change. Understanding and decomposing these 13 drivers of discharge changes is essential to preparing and planning adaptive strategies. 14 We propose a methodology combining a physical-based model to reproduce the natu-15 ral behavior of river catchments and a parsimonious model to serve as a framework of 16 interpretation, comparing the physical-based model outputs to observations of discharge 17 trends. We show that over Europe, especially in the South, the dominant explanations 18 for discharge trends are non-climatic factors. Still, in some catchments of Northern Eu-19 rope, climate change seems to be the dominating driver of change. We hypothesize that 20 the dominating non-climatic factors are irrigation development, groundwater pumping 21 and other human water usage, which need to be taken into account in physical-based mod-22 els to understand the main drivers of discharge and project future changes. 23

## <sup>24</sup> Plain Language Summary

Water is an essential resource. Its access and management are key challenges in the 25 context of climate change. Changes in precipitation distribution and intensity and other 26 climate effects lead to a change in the water availability and in the discharge of rivers. 27 On top of that, humans intervene to uptake water from rivers and change streamflow dy-28 namics. To better assess management practices and prepare for future climate conditions, 29 it is important to understand which part of discharge evolution is due to climate and which 30 part is due to human intervention. In this article, we present an innovative methodol-31 ogy to do so. We show that over Europe, if discharge in the North is mostly impacted 32 by the evolution of climate, in the rest, water management practices are the main cause 33 of discharge changes. This is especially the case for the drying discharge trends in the 34 South. Therefore, the evolution of management practices must be particularly of inter-35 est when constructing adaptation pathways to future climate conditions. 36

## 37 1 Introduction

Water is an essential resource for both ecosystems and human needs. Floods or wa-38 ter scarcity can lead to environmental catastrophes, conflicts and economic hardships. 39 Understanding the evolution of water availability is a key challenge in the context of cli-40 mate change and a highly managed continental water cycle. To study the evolution of 41 water resources, one key variable is streamflow. Being at the surface, it is directly re-42 lated to freshwater available to humans and ecosystems (Dai, 2016). In order to opti-43 mize its availability and reduce the impacts of floods and hydrological droughts, mankind 44 has managed it over the last millennia. Because of its central role in our water resources, 45 it has also been well observed over the last century. 46

From a geophysical perspective, streamflow provides a comprehensive overview of 47 the water dynamics of catchments as it is the result of the catchment-integrated balance 48 between water storage, precipitation and evapotranspiration (Milly et al., 2005; Rottler 49 et al., 2020). These last two fluxes are dominated by climate processes and thus driven 50 by atmospheric variability and trends (Christidis & Stott, 2022; García-Ruiz et al., 2011). 51 On the other hand, it is through the management of water storage (reservoirs or ground-52 water pumping) and evaporation (land use and irrigation) that humans optimize the ben-53 efits they take from surface water and modify streamflows (Schneider et al., 2013; Riedel 54 & Weber, 2020). 55

All of these processes have confounding effects on river discharge, which makes it difficult to detect and attribute trends in water resources (Rottler et al., 2020; Ficklin et al., 2018). With climate change, precipitation distribution, frequency and intensity are evolving, along with an increase in atmospheric water demand due to increased en-

ergy available at the surface and atmospheric water holding capacity and to changes in 60 turbulences (Douville et al., 2021; Christidis & Stott, 2022; García-Ruiz et al., 2011; Ribes 61 et al., 2019; Dezsi et al., 2018). In turn, human activities and management, through ab-62 stractions (for irrigation, domestic uses...) and regulations (dams, reservoirs...), directly 63 impact the partitioning of water between runoff and evapotranspiration along with flow 64 seasonality, due to additional water uptakes and to controlled water releases (Rottler et 65 al., 2020; García-Ruiz et al., 2011; Ficklin et al., 2018). Therefore, streamflow changes 66 are driven by climate change and anthropogenic activities, both influencing catchment 67 dynamics and equilibrium. 68

To project future streamflow changes and adapt water management strategies to climate change, it is essential first to understand the relative weight of these different drivers in streamflow dynamics. Being able to attribute past changes in river discharge to either climatic factors or human intervention on the land surface processes provides invaluable information to water managers in an evolving water cycle.

Physical-based land surface models (LSMs) and global hydrological models (GHMs) 74 have been developed to understand streamflow dynamics, reproduce land surface pro-75 cesses and predict the evolution of the water cycle using different scenarios for the fu-76 ture (W. Zhao & Li, 2015; Nazemi & Wheater, 2015). They have grown more complex 77 over time and represent, at best, the current understanding of surface/atmosphere in-78 teractions, vegetation dynamics and hydrological processes under the control of climate 79 (Tafasca et al., 2020; Quintana-Seguí et al., 2020; Stephens et al., 2023). These models 80 are very useful to study patterns of change and trends and link them to specific processes 81 (Douville et al., 2021; Zanardo et al., 2012; Alkama et al., 2010; Do et al., 2020). How-82 ever, to this day, they fail to effectively include most anthropogenic water usage and man-83 agement, even if progress is being made in that direction (F. Wang et al., 2018; Nazemi 84 & Wheater, 2015). 85

In view of the complexity of land surface processes and the lack of data, another 86 class of models has also been developed: parsimonious or calibrated models. Based on 87 the perceived functioning of the surface hydrology (Beven & Chappell, 2021), relations 88 and parameters are selected and then adjusted over a period to represent, at best, ac-89 tual streamflow characteristics. These models have demonstrated their value for oper-90 ational short-term predictions and to represent and detect current trends in discharge 91 with a simplified interpretation tool (Jiang et al., 2015; Andréassian et al., 2016; Per-92 rin et al., 2003). However, they are limited in their ability to predict changes associated 93 with specific drivers due to the difficulty of physical interpretation of the adjusted pa-94 rameters and the undetermined sensitivity of these parameters to the drivers (Zheng et 95 al., 2018; Andréassian et al., 2016; Coron et al., 2014; Nicolle et al., 2021). Still they have 96 been used to try and separate the effect of anthropogenic activities from climatic drivers, 97 often comparing a reference "untouched" period or area to a post-change period or to 98 a similar but highly anthropized area (Ficklin et al., 2018; W. Wang et al., 2020; Palmer 99 et al., 2008; Ahn & Merwade, 2014; Zheng et al., 2018; Luo et al., 2020; J. Zhao et al., 100 2018). However, these methods all rely on the debatable assumption that the adjusted 101 parameters are independent of climate variability (Coron et al., 2014; Andréassian et al., 102 2016; Reaver et al., 2022). 103

Using both classes of models, we propose a method to analyze observed annual river 104 discharge and decompose observed trends into climate-driven changes and those caused 105 by human intervention on the continental water cycle. The LSM is chosen as the climatic 106 reference as it represents the behavior of catchments and land surface dynamics, respond-107 108 ing to changes in climate variables only. Due to the incomplete representation of the complex land surface processes and the lack of representation of human water management, 109 the direct validation of the predicted river discharge to observation is difficult (Hagemann 110 & Dümenil, 1997). The Budyko space and the one-parameter parsimonious model pro-111 posed by Fu's equation (Zhang et al., 2004) is used as a framework for interpreting both 112

the LSM's simulated and the observed historical discharge. This parsimonious model in-113 troduces a parameter allowing to isolate the partial trends in discharge (Q) due to a change 114 in catchment evaporation efficiency from the partial trend due to changes in the two main 115 average climate variables precipitation (P) and potential evapotranspiration (PET) (Collignan, 116 Polcher, Bastin, & Quintana-Segui, 2023). Projecting the LSM output onto this frame-117 work allows to derive the climate sensitivity of the adjusted parameter of the parsimo-118 nious model. In turn, comparing these results to the interpretation of observed histor-119 ical records by the Budyko model allows to isolate the trends due to changes in evap-120 oration efficiency and land characteristics not represented by the LSM. This separates 121 the observed changes in streamflow into a component that can be attributed to climate 122 variations and another that can be linked to human activities. 123

#### 124 2 Method

The Budyko framework is a relatively simple empirical framework which relies on 125 balancing the water and energy fluxes through only a few variables (precipitations P and 126 potential evapotranspiration PET) to express the partitioning of water between evap-127 otranspiration E and runoff. As opposed to other simple empirical models such as lin-128 ear regression models, it accounts for physical boundaries: the water limit and the en-129 ergy limit on the system. For the framework to work, it needs to be applied to a closed 130 system where the boundaries can be defined, such as a watersheds at an equilibrium state 131 (the variations of water storage within the catchment are supposed to be small). It is 132 simple enough to be applicable to a wide variety of observed catchments as only basic 133 variables are needed. In this study we used the parametric equation of Fu-Tixeront (Zhang 134 et al., 2004) (Zhang et al., 2008) (Zheng et al., 2018). It reduces for each catchment its 135 evaporation efficiency to a single specific parameter  $\omega$  (Equ. 1), fitted over hydrologi-136 cal year averages, in a given period. For the same climatic conditions P, PET, a catch-137 ment with a higher  $\omega$  will evaporate more than another one with a smaller  $\omega$ . 138

In the original framework, this parameter is assumed to be constant since the wa tershed is considered to be in a stationary state and only driven by climate.

$$\frac{E}{P} = 1 + \frac{PET}{P} - \left(1 + \left(\frac{PET}{P}\right)^{\omega}\right)^{\frac{1}{\omega}} \tag{1}$$

<sup>141</sup> E is the actual evaporation at the scale of the catchment. With the same assump-<sup>142</sup> tion of a closed system and no water storage change, the water continuity yields for dis-<sup>143</sup> charge Q = P - E.

Here, we consider only that the system is piece-wise stationary and that the parameter can be assumed to the constant over a short period (11 years) (Han et al., 2020). This introduces a time-dependence in the parameter  $\omega$  by successive fits over an 11-year time-moving window. We therefore capture the long-term effects of climate change and anthropogenic activities, both influencing catchments responses.

As a result, both the annual mean of P and PET, regrouped in the variable C later 149 on, and for the evaporation efficiency  $\omega$  are time dependent. This allows to construct a 150 framework of interpretation, with a simple decomposition of discharge trends Q: a par-151 tial trends due to long-term changes in average climate variable and a partial trend due 152 to changes in catchment responses. More details for this methodology are given in (Collignan, 153 Polcher, Bastin, & Quintana-Segui, 2023). In this framework, the anthropogenic water 154 management and water usage will only change the catchment responses and not the cli-155 mate variables. 156

By applying the method to the observed catchments and the representation of these catchments in a land surface model, the relative contribution of climate to trends in discharge can be quantified. The two systems considered are :

160	• A climate-driven system, referred as climatic system: An LSM is used to
161 162	estimate the climate induced changes in the evaporation efficiency $(\omega_c)$ . The LSM stands as our climatic reference. It provides us with the following information:
163	$-\Delta Q_{climat}(C,\omega_c)$ : defined as the climate driven discharge trends.
164 165	- $\Delta Q_c(C, \overline{\omega_c})$ : partial trend due to fluctuations in annual averages of climate variables $(C)$ , where $\overline{\omega_c}$ is the average evaporation efficiency over the entire period.
166 167	$-\Delta Q_c(C_{rand},\omega_c)$ : partial trend due to climatic impact on evaporation efficiency $\omega_c$ , where $C_{rand}$ is a random climate with no trends.
168 · · · · · · · · · · · · · · · · · · ·	• The observation-based system, referred as actual system: the framework is used to decompose the observed discharge changes. We successively fit the frame- work to discharge observations, getting another time series of the evaporation ef- ficiency parameter $\omega_a$ . This provides the following information:
172	$-\Delta Q_{actual}(C,\omega_a)$ : overall trend in Q in the actual system.
173	$-\Delta Q_a(C,\overline{\omega_a})$ : partial trend due to changes in C.
174 175 176 177	- $\Delta Q_a(C_{rand}, \omega_a)$ : By randomizing the climate, the partial trend due to the evo- lution of evaporation efficiency $\omega_a$ can be estimated. In that case, all changes in the watershed characteristics (anthropogenic as well well as its climate in- duced) are considered.
178	We illustrate the differences in both systems over a given catchment, with a fig-

<sup>179</sup> ure of  $Q_{climat}$ ,  $Q_{actual}$ ,  $\omega_c$  and  $\omega_a$ , for the station of Castejon, upstream of the Ebro river <sup>180</sup> in Spain (in supplementary materials, Fig. S3).

We consider that the LSM accurately reproduces dynamic changes in an idealized natural catchment driven by observed climatic conditions, even if it might have biases in the absolute values of discharge. Therefore we only compare trends between both systems. All trends are computed using the Mann-Kendall non-parametric test, associated with the Thiel-Sen slope estimator (Xiong et al., 2020), with a 0.05 p-value threshold for significance.

#### 187 **3 Data**

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#### 3.1 The Land Surface Model ORCHIDEE

The LSM used in this study is the Organizing Carbon and Hydrology In Dynamic 189 Ecosystems (ORCHIDEE) (Krinner et al., 2005) from the Institut Pierre Simon Laplace 190 (IPSL). The current version of the model simulates the global carbon cycle and quan-191 tifies terrestrial water and energy balance through biophysical and natural biogeochem-192 ical processes. It can include some anthropogenic interference such as land cover changes, 193 forest and grassland management or irrigation (Guimberteau et al., 2012). Here the model 194 is used without these options, as only the climatic dependences of hydrology are sought. 195 Used in off-line conditions, the atmospheric conditions are forced by a given data-set. 196

#### 3.2 Forcing datasets

Three different climatic datasets are used to drive the LSM. These datasets are used 198 as input to the off-line LSM and provide the variables needed in Fu's equation 1. The 199 main one is the forcing dataset GSWP3 (Hyungjun, 2017), covering the 1901-2012 pe-200 riod at a 3-hourly resolution with a geographic resolution of 0.5°x 0.5°. It is a dynam-201 ical downscaling of 20th Century Reanalysis using a Global Spectral Model. It is bias-202 corrected using Global Precipitation Climatology Center (GPCC) (Rudolf et al., 2005) 203 and Climate Research Unit (CRU) observational data (Harris et al., 2020). The results 204 presented later on are obtained with this forcing dataset. 205

We also use two other forcings, WFDEI-GPCC (Weedon et al., 2014) and E2OFD (Beck et al., 2017), both covering the 1979 to 2014 period. Testing the methodology with different independent climate datasets allows to verify the robustness of our results comparing the two systems and their sensitivity to the choice of climate forcing used (see supplementary materials).

3.3 Watersheds and discharge observation datasets

The river discharge observations collected by the Global Runoff Data Center (GRDC) 212 from gauging stations all over Europe (GRDC, 2020) are the base of the current study. 213 They were completed over Spain with data obtained from the Geoportal of Spain Min-214 isterio (Ministerio para la Transición Ecológica y el Reto Demográfico, 2020) and over 215 France with data from the database HYDRO (Ministere de l'ecologie, du developpement 216 durable et de l'energie, 2021). In the final analysis, only 814 stations were kept with at 217 least 50 years of observations and for which we were able to satisfyingly reproduce the 218 upstream catchment in the hydrological routing of the LSM (Polcher et al., 2022; Nguyen-219 Quang et al., 2018), based on the dataset HydroSHEDS (Hydrological Data and Maps 220 Based on Shuttle Elevation Derivatives at Multiple Scales) (Lehner et al., 2008). 221

222 4 Results

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#### 4.1 Decomposing the trends in river discharge over the past century

The framework defined by the parsimonious model chosen allows to separate trends in river discharge (Q) into the part explained by the evolution of climate C (x-axis) and a partial trend due to changes in the catchment affecting evaporation efficiency ( $\omega$ ) (yaxis). Figure 1a allows to illustrate the relative importance of both components of the trends as estimated with the methodology presented above, using a 100-year-long simulation with an LSM and the observed discharge at 569 gauging stations.

Positioning the results for one catchment in Fig. 1a allows to illustrate the magnitudes of the partial trends due to each component and whether they are concurrent or opposite and if they tend to increase or decrease discharge. Two different systems are projected on this framework for comparison (see Method): the climatic system ( $Q_{climat}$ ) based on the LSM outputs and the actual system ( $Q_{actual}$ ) based on observed records.

Our results show that in the case of the climatic system (Fig. 1b), the changes in 235 annual mean climate variables have about a four times larger impact than the changes 236 in evaporation efficiency on relative annual mean discharge trends. Overall, almost all 237 catchments where  $Q_{climat}$  has significantly changed (catchments with significant trends 238 correspond to colored points) have concurrent trends in both components. A high co-239 variance between these two components allows to better detect the trends. More gen-240 erally, there is a dominance of the trends in annual mean in climate variables P, PET, 241 amplified by the response of evaporation efficiency of the catchment induced by climate 242 change. These cases correspond to catchments where an increase in P and/or a decrease 243 in *PET* tends to increase annual mean discharge or inversely for a decrease. For instance, 244 if an increase in P is inhomogeneous, with an even stronger increase in winter precip-245 itation, the partitioning towards runoff is usually higher, which translates into a decreased 246 evaporation efficiency and thus an even stronger increase of  $Q_{climat}$  (top right quadrant, 247 Fig. 1b). Therefore in this example, the increase in the annual mean  $Q_{climat}$  is not only 248 due to an increase in annual mean P but is amplified by the more contrasted seasonal-249 ity and its impact on evaporation efficiency. More generally, there are fewer catchments 250 where the changes in the evaporation efficiency tend to decrease discharge in the climatic 251 system, except when they concur with a high decrease in relative discharge due to a de-252 crease in P and/or an increase in PET. This is coherent with the increasing intensity 253 and contrasted seasonality of precipitation events observed over Europe (Christidis & 254



(a) Interpretation scheme: comparing significant trends due to climate variables or due to  $\omega$ . The graphs have four quadrants: the top right and the bottom left ones correspond to area of the graph where the climatic trend and the area due to  $\omega$  are complementary/ going into the same direction. The top left one contains the basins for which the trends due to  $\omega$  are positive and the trends due to climate variables are negative and the bottom right quadrant the opposite.



(b) Climatic system  $Q_{climat}$ : relative trends (%/yr over the century) due to changes in  $\omega_c$  versus relative trends due to changes in climate variables C



(c) Actual system  $Q_{actual}$ : relative trends (%/yr over the century) due to changes in  $\omega_a$  versus relative trends due to changes in climate variables C

Figure 1. Comparing the relative trends  $\left(\frac{dQ}{dt}/\overline{Q}\right)$  due to a change in climate variables or due to a change in evaporation efficiency  $\omega$  in the evolution of discharge, for both system  $Q_{climat}$  and  $Q_{actual}$ . One point corresponds to a basin with at least 50 years of river discharge observations over Europe. The scale of trends due to  $\omega_a$  in the actual system is ten times larger than the one for trends due to  $\omega_c$  in the climatic system. The green line is the line y = x. The color scale represents the significance of the trend in Q when all factors are considered. The markers indicate whether the partial trends are significant due to changes in C (x-axis), in  $\omega$  (y-axis), or both. Stott, 2022; Riedel & Weber, 2020; Zveryaev, 2004; Ribes et al., 2019; Douville et al.,
2021), which would logically tend to decrease evaporation efficiency by increasing local
runoff, and therefore lead to a positive partial trend in discharge.

In the case of the observed system (Fig. 1c), the relationship between the two par-258 tial trends looks very different. The one linked to changes in evaporation efficiency is larger 259 than the partial trends linked to the annual mean in P and PET by a factor of 3. More 260 generally, the total trends in  $Q_{actual}$  (catchments with significant trends correspond to 261 colored points) follow the partial trends due to changes in the catchments' evaporation 262 efficiency. Therefore, in the actual system, the trends in discharge are mainly due to changes 263 in catchment behavior due to non-climatic factors. Contrary to when only climate change 264 is considered for natural catchments, land use changes and human water management 265 tend to increase the evaporation efficiency of catchments and, therefore, decrease  $Q_{actual}$ . 266 This is coherent with activities such as irrigation, or agriculture in general, which aim 267 at optimizing the evapotranspiration over catchments. 268

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#### 4.2 Geographical distribution of discharge trend characteristics

The spatial distribution of the significant relative trends (Fig. 2) is spatially coherent, which also attests to the method's robustness. When some specific catchments are referred to, the geographic location of these catchments is illustrated in supplementary material (Fig. S2).

In the climatic system (Fig. 2a), basins in eastern Europe and Spain are getting 274 dryer with trends in discharge between -0.2%/yr and -0.5%/yr over the past century. In 275 central and northern Europe, the climatic discharge is increasing with trends of +0.2%/yr276 to 0.5%/yr over the past century. Similarly to the previous results, we observe that in 277 this system, the trends in discharge  $Q_{climat}$  (Fig. 2a) are mostly driven by changes in 278 average climate variables C (Fig. 2c) and not to changes in evaporation efficiency  $\omega_c$  (Fig. 279 2e). These partial trends due to changes in the evaporation efficiency  $\omega_c$  (Fig. 2e) are 280 small (between -0.2%/yr to +0.2%/yr) and are mostly positive. Their effect is negligi-281 ble when looking at the total trends in discharge (Fig. 2a). It can however amplify the 282 partial trend due to changes in the annual average of climate variables C. It corresponds 283 to the top-right and bottom-left quadrants in Fig. 1b. This effect is illustrated in the 284 Duero basin in north-western Spain, where both partial trends concur to a decrease in 285  $Q_{climat}$ . They can also cancel each other out, for instance, in the Tiber River in Italy, 286 where the decrease in  $Q_{climat}$  due to changes in C is not significant in the overall changes 287 in  $Q_{climat}$ . 288

Again, for the actual system, our results show that the discharge trends (Fig. 2b) 289 are mostly explained by changes in the evaporation efficiencies (Fig. 2f). Here the changes 290 in the evaporation efficiency  $\omega_a$  encompass all changes in the catchment's evaporative 291 behaviors, those induced by climate and those by changing water usage. Similarly to re-292 sults from other studies (Vicente-Serrano et al., 2019) over Western Europe, we find the 293 highest negative trends are over Southern Spain and are mostly driven by non-climatic 294 factors. To facilitate the comparison, scales of the Fig. 2 are fixed thus for Fig. 2d and 295 Fig. 2f they saturate, not showing that the trends are a lot higher in Spain, with up to 296 -4.4%/yr change over the past century in  $Q_{actual}$ . Over the rest of Europe, trends are 297 lower and less significant, with positive trends generally in northern Europe, Great Britain 298 and Sweden and negative trends over central Europe. The south of Spain corresponds 299 to an area where both climate changes and changes related to human activities led to 300 a significant decrease in river discharge over the past century. There, mainly, the changes 301 in evaporation efficiency result in decreasing trends in  $Q_{actual}$  (Fig. 2f). This is coher-302 ent with increasing irrigation water uptakes. However, the Guadiana River stands out 303 in our results. It seems that over that specific catchment, the overall effect of human wa-304 ter management and land use changes tend to increase  $Q_{actual}$ , contrary to the rest of 305



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(a) Relative changes in  $Q_{climat}$ 



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(c) Relative trend in  $Q_c(C, \overline{\omega_c})$ : trends in relative  $Q_{climat}$  changes due to climate variables C.



(e) Relative trends in  $Q_c(C_{rand}, \omega_c(t))$ : trends in relative  $Q_{climat}$  changes due to changes in the evaporation efficiency  $\omega_c$ 



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(b) Relative changes in  $Q_{actual}$ 



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(d) Relative trend in  $Q_a(C, \overline{\omega_a})$ : trends in relative  $Q_{actual}$  changes due to climate variables C.



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(f) Relative trends in  $Q_a(C_{rand}, \omega_a(t))$ : trends in relative  $Q_{actual}$  changes due to changes in the evaporation efficiency  $\omega_a$ .

Figure 2. Significant trends in the relative river discharge  $Q/\overline{Q}$  over the time period 1901-2012 (%/yr over the century). The scales have been forced to be the same for all maps for comparison purposes but the extrema can go higher or lower.

Spain. As discussed later, this could be linked to unsustainable groundwater pumping (Holtz & Pahl-Wostl, 2012), which invalidates the hypothesis of no water storage change and, therefore, results in a lower apparent evaporation efficiency and an artificially underestimated evapotranspiration, increasing the resulting discharge. More generally, there are several basins over Europe where the trends induced by changes in the evaporation efficiency lose their significance when the climate variability is considered in the reconstructed discharge. See, for instance, western France, northern Germany, Serbia.

Interestingly, when we draw similar maps for sub-periods of 10 years, the impact of evaporation efficiency changes on discharge is not dominant anymore. At the decadal scale, the climatic variability is high. This climatic noise covers the effect of changes in the catchment's evaporation efficiency in discharge trends. At the scale of the century, the signal-to-noise ratio is higher, bringing to light the long-term role of changes in the catchment's evaporation efficiency and catchment's behavior on discharge.

#### 319 5 Discussion

Our method uses a parsimonious hydrological model to decompose the observed river discharge trends into climate-driven processes, as estimated with a state-of-the-art LSM, and non-climatic changes that can be attributed to human activities. It can be generalized to the use of any couple of calibrated parsimonious model and physical-based model and be an effective operational tool to estimate and illustrate the effect of nonclimatic drivers and land surface changes not well accounted for in current models and which may have a strong direct impact on water resources.

We show that the dominant explanation for river discharge trends is non-climatic factors, especially in Southern Europe. In some catchments of Northern Europe, climate change seems to be the dominating driver of change. Still, in accordance with previous studies (Gudmundsson et al., 2017), our results highlight the fact that not accounting for non-climatic trends leads to high under-estimation of discharge changes in the physicalbased model used and therefore to high uncertainties in projections of future water resource trends, especially when looking at long-term trends.

With this methodology, we can only estimate the magnitude of non-climatic trends 334 but not attribute them to specific processes. In some areas where a dominant process 335 can be hypothesized, such as irrigation, correlation with indicators can allow to verify 336 the plausibility of the assumed cause. For instance, over Spain, especially over the Ebro 337 basin, the strong increase in evaporation efficiency and reduced discharge is correlated 338 to the development of dams with a coefficient above 0.7 when correlating  $\omega_a$  to reser-339 voirs levels for 6 sub-basins in this catchment. Dams water storage is an indicator for 340 human management of water resources impacting the evaporation efficiency of water-341 sheds. More generally, we see that the changes in the evaporation efficiency intensified 342 over the second part of the century, where areas equipped for irrigation have been de-343 veloped (Angelak $\iota$ s et al., 2020; Siebert et al., 2015). However, the available data are in-344 sufficient to attest to a correlation with that latter factor or with the effective amount 345 of water used for irrigation. Groundwater pumping and glacier melt can explain posi-346 tive trends in discharge due to additional sources of water not accounted for in the cli-347 matic system, which lead to artificially low evaporation efficiencies in our framework. For 348 the Guadiana River in Spain, the unsustainable groundwater pumping (Holtz & Pahl-349 Wostl, 2012; Llamas et al., 2015; Esteban & Albiac, 2012) can explain the positive trend. 350 In a similar way, for the Po river in Italy, which is highly irrigated (Siebert et al., 2015), 351 we would expect a strong decrease in discharge as in most of Spain, but glacier melt brings 352 additional water to the system (Schaner et al., 2012; Vincent et al., 2017; Huss & Hock, 353 2018), explaining a reduced detected negative trend over the end of the century. Other 354 phenomena, such as soil sealing and river management, would be expected to have sim-355 ilar effects due to a decrease in evapotranspiration or to an artificial enhancement of runoff. 356

Changes in land use as represented in ORCHIDEE (Lawrence et al., 2016) are shown to have little effect on discharge over the studied period and area but could have a more significant effect at finer scale over small catchments.

Quantifying the contribution of climatic and non-climatic factors to changes in river 360 discharge is an important first step. But it should be followed by an attribution. The 361 use of the LSM in its current state allows to attribute the changes due to climate. How-362 ever, the non-climatic factors remain challenging to attribute to specific processes, es-363 pecially since most factors have concurring and competing effects. Detection and attri-364 bution methods have been developed in climate studies to assess anthropogenic climate change. They have allowed to determine the role of different factors by reproducing them 366 first in GCMs (Hegerl & Zwiers, 2011; Douville et al., 2021). Similarly, we would need 367 to simulate water usage such as irrigation, dam management, groundwater pumping and 368 other missing phenomena such as glacier melting in the LSMs so that their impact on 369 the evaporation efficiency can be identified and their contribution to the non-climatic 370 trend quantified. 371

Understanding and quantifying the contribution of various processes contributing 372 to observed discharge changes over the past century is an essential step in developing adap-373 tation strategies to face climate change. In the future, changes in climatic variables are 374 expected to increase even further, with increase in intense precipitation events (Ribes 375 et al., 2019; Douville et al., 2021), an increase in evaporative demand and especially a 376 decrease in average precipitation leading to water scarcity over southern Mediterranean 377 Europe (Gudmundsson et al., 2017; Alkama et al., 2013). Concurrently, in Europe, hu-378 man water management is expected to evolve to adapt to climate change and other con-379 straints, such as changes in water and energy demand and regulations (Arheimer et al., 380 2017). For instance, the extent of irrigated land in Europe peaked at the end of the 20th 381 century and the future irrigation evolution is expected to follow new goals and mostly 382 rely on improved efficiency (Adeveri et al., 2020). Therefore, the balance between the 383 different terms influencing catchment evaporation efficiency and discharge may change. If non-climatic factors dominated over the past century to explain discharge trends, it 385 may not be the same in the future. Attribution needs to be tested over the documented 386 past to improve the representation of non-climatic processes and allow the effective pro-387 jection of future evolutions and eventual changes in this balance. 388

#### <sup>389</sup> Open Research Section

The outputs of the LSM ORCHIDEE for each catchment used in this study with the forcing GSWP3 are gathered in a file freely available on Zenodo.org (Collignan, Polcher, Bastin, & Quintana-Seguí, 2023). This file also contains the description of the stations used in the study: their location, the size of the upstream area used to position the station on the grid and annual averages of streamflow observations.

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# Identifying and quantifying the impact of climatic and non-climatic drivers on river discharge in Europe

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# Key Points:

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- Attribution of streamflow trends to climate drivers
- Larger effect on streamflow of non-climatic drivers
- Strong impact of human activities, especially over Spain

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#### 11 Abstract

Our water resources have changed over the last century through a combination of wa-12 ter management evolutions and climate change. Understanding and decomposing these 13 drivers of discharge changes is essential to preparing and planning adaptive strategies. 14 We propose a methodology combining a physical-based model to reproduce the natu-15 ral behavior of river catchments and a parsimonious model to serve as a framework of 16 interpretation, comparing the physical-based model outputs to observations of discharge 17 trends. We show that over Europe, especially in the South, the dominant explanations 18 for discharge trends are non-climatic factors. Still, in some catchments of Northern Eu-19 rope, climate change seems to be the dominating driver of change. We hypothesize that 20 the dominating non-climatic factors are irrigation development, groundwater pumping 21 and other human water usage, which need to be taken into account in physical-based mod-22 els to understand the main drivers of discharge and project future changes. 23

## <sup>24</sup> Plain Language Summary

Water is an essential resource. Its access and management are key challenges in the 25 context of climate change. Changes in precipitation distribution and intensity and other 26 climate effects lead to a change in the water availability and in the discharge of rivers. 27 On top of that, humans intervene to uptake water from rivers and change streamflow dy-28 namics. To better assess management practices and prepare for future climate conditions, 29 it is important to understand which part of discharge evolution is due to climate and which 30 part is due to human intervention. In this article, we present an innovative methodol-31 ogy to do so. We show that over Europe, if discharge in the North is mostly impacted 32 by the evolution of climate, in the rest, water management practices are the main cause 33 of discharge changes. This is especially the case for the drying discharge trends in the 34 South. Therefore, the evolution of management practices must be particularly of inter-35 est when constructing adaptation pathways to future climate conditions. 36

## 37 1 Introduction

Water is an essential resource for both ecosystems and human needs. Floods or wa-38 ter scarcity can lead to environmental catastrophes, conflicts and economic hardships. 39 Understanding the evolution of water availability is a key challenge in the context of cli-40 mate change and a highly managed continental water cycle. To study the evolution of 41 water resources, one key variable is streamflow. Being at the surface, it is directly re-42 lated to freshwater available to humans and ecosystems (Dai, 2016). In order to opti-43 mize its availability and reduce the impacts of floods and hydrological droughts, mankind 44 has managed it over the last millennia. Because of its central role in our water resources, 45 it has also been well observed over the last century. 46

From a geophysical perspective, streamflow provides a comprehensive overview of 47 the water dynamics of catchments as it is the result of the catchment-integrated balance 48 between water storage, precipitation and evapotranspiration (Milly et al., 2005; Rottler 49 et al., 2020). These last two fluxes are dominated by climate processes and thus driven 50 by atmospheric variability and trends (Christidis & Stott, 2022; García-Ruiz et al., 2011). 51 On the other hand, it is through the management of water storage (reservoirs or ground-52 water pumping) and evaporation (land use and irrigation) that humans optimize the ben-53 efits they take from surface water and modify streamflows (Schneider et al., 2013; Riedel 54 & Weber, 2020). 55

All of these processes have confounding effects on river discharge, which makes it difficult to detect and attribute trends in water resources (Rottler et al., 2020; Ficklin et al., 2018). With climate change, precipitation distribution, frequency and intensity are evolving, along with an increase in atmospheric water demand due to increased en-

ergy available at the surface and atmospheric water holding capacity and to changes in 60 turbulences (Douville et al., 2021; Christidis & Stott, 2022; García-Ruiz et al., 2011; Ribes 61 et al., 2019; Dezsi et al., 2018). In turn, human activities and management, through ab-62 stractions (for irrigation, domestic uses...) and regulations (dams, reservoirs...), directly 63 impact the partitioning of water between runoff and evapotranspiration along with flow 64 seasonality, due to additional water uptakes and to controlled water releases (Rottler et 65 al., 2020; García-Ruiz et al., 2011; Ficklin et al., 2018). Therefore, streamflow changes 66 are driven by climate change and anthropogenic activities, both influencing catchment 67 dynamics and equilibrium. 68

To project future streamflow changes and adapt water management strategies to climate change, it is essential first to understand the relative weight of these different drivers in streamflow dynamics. Being able to attribute past changes in river discharge to either climatic factors or human intervention on the land surface processes provides invaluable information to water managers in an evolving water cycle.

Physical-based land surface models (LSMs) and global hydrological models (GHMs) 74 have been developed to understand streamflow dynamics, reproduce land surface pro-75 cesses and predict the evolution of the water cycle using different scenarios for the fu-76 ture (W. Zhao & Li, 2015; Nazemi & Wheater, 2015). They have grown more complex 77 over time and represent, at best, the current understanding of surface/atmosphere in-78 teractions, vegetation dynamics and hydrological processes under the control of climate 79 (Tafasca et al., 2020; Quintana-Seguí et al., 2020; Stephens et al., 2023). These models 80 are very useful to study patterns of change and trends and link them to specific processes 81 (Douville et al., 2021; Zanardo et al., 2012; Alkama et al., 2010; Do et al., 2020). How-82 ever, to this day, they fail to effectively include most anthropogenic water usage and man-83 agement, even if progress is being made in that direction (F. Wang et al., 2018; Nazemi 84 & Wheater, 2015). 85

In view of the complexity of land surface processes and the lack of data, another 86 class of models has also been developed: parsimonious or calibrated models. Based on 87 the perceived functioning of the surface hydrology (Beven & Chappell, 2021), relations 88 and parameters are selected and then adjusted over a period to represent, at best, ac-89 tual streamflow characteristics. These models have demonstrated their value for oper-90 ational short-term predictions and to represent and detect current trends in discharge 91 with a simplified interpretation tool (Jiang et al., 2015; Andréassian et al., 2016; Per-92 rin et al., 2003). However, they are limited in their ability to predict changes associated 93 with specific drivers due to the difficulty of physical interpretation of the adjusted pa-94 rameters and the undetermined sensitivity of these parameters to the drivers (Zheng et 95 al., 2018; Andréassian et al., 2016; Coron et al., 2014; Nicolle et al., 2021). Still they have 96 been used to try and separate the effect of anthropogenic activities from climatic drivers, 97 often comparing a reference "untouched" period or area to a post-change period or to 98 a similar but highly anthropized area (Ficklin et al., 2018; W. Wang et al., 2020; Palmer 99 et al., 2008; Ahn & Merwade, 2014; Zheng et al., 2018; Luo et al., 2020; J. Zhao et al., 100 2018). However, these methods all rely on the debatable assumption that the adjusted 101 parameters are independent of climate variability (Coron et al., 2014; Andréassian et al., 102 2016; Reaver et al., 2022). 103

Using both classes of models, we propose a method to analyze observed annual river 104 discharge and decompose observed trends into climate-driven changes and those caused 105 by human intervention on the continental water cycle. The LSM is chosen as the climatic 106 reference as it represents the behavior of catchments and land surface dynamics, respond-107 108 ing to changes in climate variables only. Due to the incomplete representation of the complex land surface processes and the lack of representation of human water management, 109 the direct validation of the predicted river discharge to observation is difficult (Hagemann 110 & Dümenil, 1997). The Budyko space and the one-parameter parsimonious model pro-111 posed by Fu's equation (Zhang et al., 2004) is used as a framework for interpreting both 112

the LSM's simulated and the observed historical discharge. This parsimonious model in-113 troduces a parameter allowing to isolate the partial trends in discharge (Q) due to a change 114 in catchment evaporation efficiency from the partial trend due to changes in the two main 115 average climate variables precipitation (P) and potential evapotranspiration (PET) (Collignan, 116 Polcher, Bastin, & Quintana-Segui, 2023). Projecting the LSM output onto this frame-117 work allows to derive the climate sensitivity of the adjusted parameter of the parsimo-118 nious model. In turn, comparing these results to the interpretation of observed histor-119 ical records by the Budyko model allows to isolate the trends due to changes in evap-120 oration efficiency and land characteristics not represented by the LSM. This separates 121 the observed changes in streamflow into a component that can be attributed to climate 122 variations and another that can be linked to human activities. 123

#### 124 2 Method

The Budyko framework is a relatively simple empirical framework which relies on 125 balancing the water and energy fluxes through only a few variables (precipitations P and 126 potential evapotranspiration PET) to express the partitioning of water between evap-127 otranspiration E and runoff. As opposed to other simple empirical models such as lin-128 ear regression models, it accounts for physical boundaries: the water limit and the en-129 ergy limit on the system. For the framework to work, it needs to be applied to a closed 130 system where the boundaries can be defined, such as a watersheds at an equilibrium state 131 (the variations of water storage within the catchment are supposed to be small). It is 132 simple enough to be applicable to a wide variety of observed catchments as only basic 133 variables are needed. In this study we used the parametric equation of Fu-Tixeront (Zhang 134 et al., 2004) (Zhang et al., 2008) (Zheng et al., 2018). It reduces for each catchment its 135 evaporation efficiency to a single specific parameter  $\omega$  (Equ. 1), fitted over hydrologi-136 cal year averages, in a given period. For the same climatic conditions P, PET, a catch-137 ment with a higher  $\omega$  will evaporate more than another one with a smaller  $\omega$ . 138

In the original framework, this parameter is assumed to be constant since the wa tershed is considered to be in a stationary state and only driven by climate.

$$\frac{E}{P} = 1 + \frac{PET}{P} - \left(1 + \left(\frac{PET}{P}\right)^{\omega}\right)^{\frac{1}{\omega}} \tag{1}$$

<sup>141</sup> E is the actual evaporation at the scale of the catchment. With the same assump-<sup>142</sup> tion of a closed system and no water storage change, the water continuity yields for dis-<sup>143</sup> charge Q = P - E.

Here, we consider only that the system is piece-wise stationary and that the parameter can be assumed to the constant over a short period (11 years) (Han et al., 2020). This introduces a time-dependence in the parameter  $\omega$  by successive fits over an 11-year time-moving window. We therefore capture the long-term effects of climate change and anthropogenic activities, both influencing catchments responses.

As a result, both the annual mean of P and PET, regrouped in the variable C later 149 on, and for the evaporation efficiency  $\omega$  are time dependent. This allows to construct a 150 framework of interpretation, with a simple decomposition of discharge trends Q: a par-151 tial trends due to long-term changes in average climate variable and a partial trend due 152 to changes in catchment responses. More details for this methodology are given in (Collignan, 153 Polcher, Bastin, & Quintana-Segui, 2023). In this framework, the anthropogenic water 154 management and water usage will only change the catchment responses and not the cli-155 mate variables. 156

By applying the method to the observed catchments and the representation of these catchments in a land surface model, the relative contribution of climate to trends in discharge can be quantified. The two systems considered are :

160	• A climate-driven system, referred as climatic system: An LSM is used to
161 162	estimate the climate induced changes in the evaporation efficiency $(\omega_c)$ . The LSM stands as our climatic reference. It provides us with the following information:
163	$-\Delta Q_{climat}(C,\omega_c)$ : defined as the climate driven discharge trends.
164 165	- $\Delta Q_c(C, \overline{\omega_c})$ : partial trend due to fluctuations in annual averages of climate variables $(C)$ , where $\overline{\omega_c}$ is the average evaporation efficiency over the entire period.
166 167	$-\Delta Q_c(C_{rand},\omega_c)$ : partial trend due to climatic impact on evaporation efficiency $\omega_c$ , where $C_{rand}$ is a random climate with no trends.
168 · · · · · · · · · · · · · · · · · · ·	• The observation-based system, referred as actual system: the framework is used to decompose the observed discharge changes. We successively fit the frame- work to discharge observations, getting another time series of the evaporation ef- ficiency parameter $\omega_a$ . This provides the following information:
172	$-\Delta Q_{actual}(C,\omega_a)$ : overall trend in Q in the actual system.
173	$-\Delta Q_a(C,\overline{\omega_a})$ : partial trend due to changes in C.
174 175 176 177	- $\Delta Q_a(C_{rand}, \omega_a)$ : By randomizing the climate, the partial trend due to the evo- lution of evaporation efficiency $\omega_a$ can be estimated. In that case, all changes in the watershed characteristics (anthropogenic as well well as its climate in- duced) are considered.
178	We illustrate the differences in both systems over a given catchment, with a fig-

<sup>179</sup> ure of  $Q_{climat}$ ,  $Q_{actual}$ ,  $\omega_c$  and  $\omega_a$ , for the station of Castejon, upstream of the Ebro river <sup>180</sup> in Spain (in supplementary materials, Fig. S3).

We consider that the LSM accurately reproduces dynamic changes in an idealized natural catchment driven by observed climatic conditions, even if it might have biases in the absolute values of discharge. Therefore we only compare trends between both systems. All trends are computed using the Mann-Kendall non-parametric test, associated with the Thiel-Sen slope estimator (Xiong et al., 2020), with a 0.05 p-value threshold for significance.

#### 187 **3 Data**

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#### 3.1 The Land Surface Model ORCHIDEE

The LSM used in this study is the Organizing Carbon and Hydrology In Dynamic 189 Ecosystems (ORCHIDEE) (Krinner et al., 2005) from the Institut Pierre Simon Laplace 190 (IPSL). The current version of the model simulates the global carbon cycle and quan-191 tifies terrestrial water and energy balance through biophysical and natural biogeochem-192 ical processes. It can include some anthropogenic interference such as land cover changes, 193 forest and grassland management or irrigation (Guimberteau et al., 2012). Here the model 194 is used without these options, as only the climatic dependences of hydrology are sought. 195 Used in off-line conditions, the atmospheric conditions are forced by a given data-set. 196

#### 3.2 Forcing datasets

Three different climatic datasets are used to drive the LSM. These datasets are used 198 as input to the off-line LSM and provide the variables needed in Fu's equation 1. The 199 main one is the forcing dataset GSWP3 (Hyungjun, 2017), covering the 1901-2012 pe-200 riod at a 3-hourly resolution with a geographic resolution of 0.5°x 0.5°. It is a dynam-201 ical downscaling of 20th Century Reanalysis using a Global Spectral Model. It is bias-202 corrected using Global Precipitation Climatology Center (GPCC) (Rudolf et al., 2005) 203 and Climate Research Unit (CRU) observational data (Harris et al., 2020). The results 204 presented later on are obtained with this forcing dataset. 205

We also use two other forcings, WFDEI-GPCC (Weedon et al., 2014) and E2OFD (Beck et al., 2017), both covering the 1979 to 2014 period. Testing the methodology with different independent climate datasets allows to verify the robustness of our results comparing the two systems and their sensitivity to the choice of climate forcing used (see supplementary materials).

3.3 Watersheds and discharge observation datasets

The river discharge observations collected by the Global Runoff Data Center (GRDC) 212 from gauging stations all over Europe (GRDC, 2020) are the base of the current study. 213 They were completed over Spain with data obtained from the Geoportal of Spain Min-214 isterio (Ministerio para la Transición Ecológica y el Reto Demográfico, 2020) and over 215 France with data from the database HYDRO (Ministere de l'ecologie, du developpement 216 durable et de l'energie, 2021). In the final analysis, only 814 stations were kept with at 217 least 50 years of observations and for which we were able to satisfyingly reproduce the 218 upstream catchment in the hydrological routing of the LSM (Polcher et al., 2022; Nguyen-219 Quang et al., 2018), based on the dataset HydroSHEDS (Hydrological Data and Maps 220 Based on Shuttle Elevation Derivatives at Multiple Scales) (Lehner et al., 2008). 221

222 4 Results

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#### 4.1 Decomposing the trends in river discharge over the past century

The framework defined by the parsimonious model chosen allows to separate trends in river discharge (Q) into the part explained by the evolution of climate C (x-axis) and a partial trend due to changes in the catchment affecting evaporation efficiency ( $\omega$ ) (yaxis). Figure 1a allows to illustrate the relative importance of both components of the trends as estimated with the methodology presented above, using a 100-year-long simulation with an LSM and the observed discharge at 569 gauging stations.

Positioning the results for one catchment in Fig. 1a allows to illustrate the magnitudes of the partial trends due to each component and whether they are concurrent or opposite and if they tend to increase or decrease discharge. Two different systems are projected on this framework for comparison (see Method): the climatic system ( $Q_{climat}$ ) based on the LSM outputs and the actual system ( $Q_{actual}$ ) based on observed records.

Our results show that in the case of the climatic system (Fig. 1b), the changes in 235 annual mean climate variables have about a four times larger impact than the changes 236 in evaporation efficiency on relative annual mean discharge trends. Overall, almost all 237 catchments where  $Q_{climat}$  has significantly changed (catchments with significant trends 238 correspond to colored points) have concurrent trends in both components. A high co-239 variance between these two components allows to better detect the trends. More gen-240 erally, there is a dominance of the trends in annual mean in climate variables P, PET, 241 amplified by the response of evaporation efficiency of the catchment induced by climate 242 change. These cases correspond to catchments where an increase in P and/or a decrease 243 in *PET* tends to increase annual mean discharge or inversely for a decrease. For instance, 244 if an increase in P is inhomogeneous, with an even stronger increase in winter precip-245 itation, the partitioning towards runoff is usually higher, which translates into a decreased 246 evaporation efficiency and thus an even stronger increase of  $Q_{climat}$  (top right quadrant, 247 Fig. 1b). Therefore in this example, the increase in the annual mean  $Q_{climat}$  is not only 248 due to an increase in annual mean P but is amplified by the more contrasted seasonal-249 ity and its impact on evaporation efficiency. More generally, there are fewer catchments 250 where the changes in the evaporation efficiency tend to decrease discharge in the climatic 251 system, except when they concur with a high decrease in relative discharge due to a de-252 crease in P and/or an increase in PET. This is coherent with the increasing intensity 253 and contrasted seasonality of precipitation events observed over Europe (Christidis & 254



(a) Interpretation scheme: comparing significant trends due to climate variables or due to  $\omega$ . The graphs have four quadrants: the top right and the bottom left ones correspond to area of the graph where the climatic trend and the area due to  $\omega$  are complementary/ going into the same direction. The top left one contains the basins for which the trends due to  $\omega$  are positive and the trends due to climate variables are negative and the bottom right quadrant the opposite.



(b) Climatic system  $Q_{climat}$ : relative trends (%/yr over the century) due to changes in  $\omega_c$  versus relative trends due to changes in climate variables C



(c) Actual system  $Q_{actual}$ : relative trends (%/yr over the century) due to changes in  $\omega_a$  versus relative trends due to changes in climate variables C

Figure 1. Comparing the relative trends  $\left(\frac{dQ}{dt}/\overline{Q}\right)$  due to a change in climate variables or due to a change in evaporation efficiency  $\omega$  in the evolution of discharge, for both system  $Q_{climat}$  and  $Q_{actual}$ . One point corresponds to a basin with at least 50 years of river discharge observations over Europe. The scale of trends due to  $\omega_a$  in the actual system is ten times larger than the one for trends due to  $\omega_c$  in the climatic system. The green line is the line y = x. The color scale represents the significance of the trend in Q when all factors are considered. The markers indicate whether the partial trends are significant due to changes in C (x-axis), in  $\omega$  (y-axis), or both. Stott, 2022; Riedel & Weber, 2020; Zveryaev, 2004; Ribes et al., 2019; Douville et al.,
2021), which would logically tend to decrease evaporation efficiency by increasing local
runoff, and therefore lead to a positive partial trend in discharge.

In the case of the observed system (Fig. 1c), the relationship between the two par-258 tial trends looks very different. The one linked to changes in evaporation efficiency is larger 259 than the partial trends linked to the annual mean in P and PET by a factor of 3. More 260 generally, the total trends in  $Q_{actual}$  (catchments with significant trends correspond to 261 colored points) follow the partial trends due to changes in the catchments' evaporation 262 efficiency. Therefore, in the actual system, the trends in discharge are mainly due to changes 263 in catchment behavior due to non-climatic factors. Contrary to when only climate change 264 is considered for natural catchments, land use changes and human water management 265 tend to increase the evaporation efficiency of catchments and, therefore, decrease  $Q_{actual}$ . 266 This is coherent with activities such as irrigation, or agriculture in general, which aim 267 at optimizing the evapotranspiration over catchments. 268

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#### 4.2 Geographical distribution of discharge trend characteristics

The spatial distribution of the significant relative trends (Fig. 2) is spatially coherent, which also attests to the method's robustness. When some specific catchments are referred to, the geographic location of these catchments is illustrated in supplementary material (Fig. S2).

In the climatic system (Fig. 2a), basins in eastern Europe and Spain are getting 274 dryer with trends in discharge between -0.2%/yr and -0.5%/yr over the past century. In 275 central and northern Europe, the climatic discharge is increasing with trends of +0.2%/yr276 to 0.5%/yr over the past century. Similarly to the previous results, we observe that in 277 this system, the trends in discharge  $Q_{climat}$  (Fig. 2a) are mostly driven by changes in 278 average climate variables C (Fig. 2c) and not to changes in evaporation efficiency  $\omega_c$  (Fig. 279 2e). These partial trends due to changes in the evaporation efficiency  $\omega_c$  (Fig. 2e) are 280 small (between -0.2%/yr to +0.2%/yr) and are mostly positive. Their effect is negligi-281 ble when looking at the total trends in discharge (Fig. 2a). It can however amplify the 282 partial trend due to changes in the annual average of climate variables C. It corresponds 283 to the top-right and bottom-left quadrants in Fig. 1b. This effect is illustrated in the 284 Duero basin in north-western Spain, where both partial trends concur to a decrease in 285  $Q_{climat}$ . They can also cancel each other out, for instance, in the Tiber River in Italy, 286 where the decrease in  $Q_{climat}$  due to changes in C is not significant in the overall changes 287 in  $Q_{climat}$ . 288

Again, for the actual system, our results show that the discharge trends (Fig. 2b) 289 are mostly explained by changes in the evaporation efficiencies (Fig. 2f). Here the changes 290 in the evaporation efficiency  $\omega_a$  encompass all changes in the catchment's evaporative 291 behaviors, those induced by climate and those by changing water usage. Similarly to re-292 sults from other studies (Vicente-Serrano et al., 2019) over Western Europe, we find the 293 highest negative trends are over Southern Spain and are mostly driven by non-climatic 294 factors. To facilitate the comparison, scales of the Fig. 2 are fixed thus for Fig. 2d and 295 Fig. 2f they saturate, not showing that the trends are a lot higher in Spain, with up to 296 -4.4%/yr change over the past century in  $Q_{actual}$ . Over the rest of Europe, trends are 297 lower and less significant, with positive trends generally in northern Europe, Great Britain 298 and Sweden and negative trends over central Europe. The south of Spain corresponds 299 to an area where both climate changes and changes related to human activities led to 300 a significant decrease in river discharge over the past century. There, mainly, the changes 301 in evaporation efficiency result in decreasing trends in  $Q_{actual}$  (Fig. 2f). This is coher-302 ent with increasing irrigation water uptakes. However, the Guadiana River stands out 303 in our results. It seems that over that specific catchment, the overall effect of human wa-304 ter management and land use changes tend to increase  $Q_{actual}$ , contrary to the rest of 305



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(a) Relative changes in  $Q_{climat}$ 



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(c) Relative trend in  $Q_c(C, \overline{\omega_c})$ : trends in relative  $Q_{climat}$  changes due to climate variables C.



(e) Relative trends in  $Q_c(C_{rand}, \omega_c(t))$ : trends in relative  $Q_{climat}$  changes due to changes in the evaporation efficiency  $\omega_c$ 



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(b) Relative changes in  $Q_{actual}$ 



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(d) Relative trend in  $Q_a(C, \overline{\omega_a})$ : trends in relative  $Q_{actual}$  changes due to climate variables C.



-0.4 -0.3 -0.2 -0.1 0.0 0.1 0.2 0.3 0.4

(f) Relative trends in  $Q_a(C_{rand}, \omega_a(t))$ : trends in relative  $Q_{actual}$  changes due to changes in the evaporation efficiency  $\omega_a$ .

Figure 2. Significant trends in the relative river discharge  $Q/\overline{Q}$  over the time period 1901-2012 (%/yr over the century). The scales have been forced to be the same for all maps for comparison purposes but the extrema can go higher or lower.

Spain. As discussed later, this could be linked to unsustainable groundwater pumping (Holtz & Pahl-Wostl, 2012), which invalidates the hypothesis of no water storage change and, therefore, results in a lower apparent evaporation efficiency and an artificially underestimated evapotranspiration, increasing the resulting discharge. More generally, there are several basins over Europe where the trends induced by changes in the evaporation efficiency lose their significance when the climate variability is considered in the reconstructed discharge. See, for instance, western France, northern Germany, Serbia.

Interestingly, when we draw similar maps for sub-periods of 10 years, the impact of evaporation efficiency changes on discharge is not dominant anymore. At the decadal scale, the climatic variability is high. This climatic noise covers the effect of changes in the catchment's evaporation efficiency in discharge trends. At the scale of the century, the signal-to-noise ratio is higher, bringing to light the long-term role of changes in the catchment's evaporation efficiency and catchment's behavior on discharge.

#### 319 5 Discussion

Our method uses a parsimonious hydrological model to decompose the observed river discharge trends into climate-driven processes, as estimated with a state-of-the-art LSM, and non-climatic changes that can be attributed to human activities. It can be generalized to the use of any couple of calibrated parsimonious model and physical-based model and be an effective operational tool to estimate and illustrate the effect of nonclimatic drivers and land surface changes not well accounted for in current models and which may have a strong direct impact on water resources.

We show that the dominant explanation for river discharge trends is non-climatic factors, especially in Southern Europe. In some catchments of Northern Europe, climate change seems to be the dominating driver of change. Still, in accordance with previous studies (Gudmundsson et al., 2017), our results highlight the fact that not accounting for non-climatic trends leads to high under-estimation of discharge changes in the physicalbased model used and therefore to high uncertainties in projections of future water resource trends, especially when looking at long-term trends.

With this methodology, we can only estimate the magnitude of non-climatic trends 334 but not attribute them to specific processes. In some areas where a dominant process 335 can be hypothesized, such as irrigation, correlation with indicators can allow to verify 336 the plausibility of the assumed cause. For instance, over Spain, especially over the Ebro 337 basin, the strong increase in evaporation efficiency and reduced discharge is correlated 338 to the development of dams with a coefficient above 0.7 when correlating  $\omega_a$  to reser-339 voirs levels for 6 sub-basins in this catchment. Dams water storage is an indicator for 340 human management of water resources impacting the evaporation efficiency of water-341 sheds. More generally, we see that the changes in the evaporation efficiency intensified 342 over the second part of the century, where areas equipped for irrigation have been de-343 veloped (Angelak $\iota$ s et al., 2020; Siebert et al., 2015). However, the available data are in-344 sufficient to attest to a correlation with that latter factor or with the effective amount 345 of water used for irrigation. Groundwater pumping and glacier melt can explain posi-346 tive trends in discharge due to additional sources of water not accounted for in the cli-347 matic system, which lead to artificially low evaporation efficiencies in our framework. For 348 the Guadiana River in Spain, the unsustainable groundwater pumping (Holtz & Pahl-349 Wostl, 2012; Llamas et al., 2015; Esteban & Albiac, 2012) can explain the positive trend. 350 In a similar way, for the Po river in Italy, which is highly irrigated (Siebert et al., 2015), 351 we would expect a strong decrease in discharge as in most of Spain, but glacier melt brings 352 additional water to the system (Schaner et al., 2012; Vincent et al., 2017; Huss & Hock, 353 2018), explaining a reduced detected negative trend over the end of the century. Other 354 phenomena, such as soil sealing and river management, would be expected to have sim-355 ilar effects due to a decrease in evapotranspiration or to an artificial enhancement of runoff. 356

Changes in land use as represented in ORCHIDEE (Lawrence et al., 2016) are shown to have little effect on discharge over the studied period and area but could have a more significant effect at finer scale over small catchments.

Quantifying the contribution of climatic and non-climatic factors to changes in river 360 discharge is an important first step. But it should be followed by an attribution. The 361 use of the LSM in its current state allows to attribute the changes due to climate. How-362 ever, the non-climatic factors remain challenging to attribute to specific processes, es-363 pecially since most factors have concurring and competing effects. Detection and attri-364 bution methods have been developed in climate studies to assess anthropogenic climate change. They have allowed to determine the role of different factors by reproducing them 366 first in GCMs (Hegerl & Zwiers, 2011; Douville et al., 2021). Similarly, we would need 367 to simulate water usage such as irrigation, dam management, groundwater pumping and 368 other missing phenomena such as glacier melting in the LSMs so that their impact on 369 the evaporation efficiency can be identified and their contribution to the non-climatic 370 trend quantified. 371

Understanding and quantifying the contribution of various processes contributing 372 to observed discharge changes over the past century is an essential step in developing adap-373 tation strategies to face climate change. In the future, changes in climatic variables are 374 expected to increase even further, with increase in intense precipitation events (Ribes 375 et al., 2019; Douville et al., 2021), an increase in evaporative demand and especially a 376 decrease in average precipitation leading to water scarcity over southern Mediterranean 377 Europe (Gudmundsson et al., 2017; Alkama et al., 2013). Concurrently, in Europe, hu-378 man water management is expected to evolve to adapt to climate change and other con-379 straints, such as changes in water and energy demand and regulations (Arheimer et al., 380 2017). For instance, the extent of irrigated land in Europe peaked at the end of the 20th 381 century and the future irrigation evolution is expected to follow new goals and mostly 382 rely on improved efficiency (Adeveri et al., 2020). Therefore, the balance between the 383 different terms influencing catchment evaporation efficiency and discharge may change. If non-climatic factors dominated over the past century to explain discharge trends, it 385 may not be the same in the future. Attribution needs to be tested over the documented 386 past to improve the representation of non-climatic processes and allow the effective pro-387 jection of future evolutions and eventual changes in this balance. 388

#### <sup>389</sup> Open Research Section

The outputs of the LSM ORCHIDEE for each catchment used in this study with the forcing GSWP3 are gathered in a file freely available on Zenodo.org (Collignan, Polcher, Bastin, & Quintana-Seguí, 2023). This file also contains the description of the stations used in the study: their location, the size of the upstream area used to position the station on the grid and annual averages of streamflow observations.

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# Supporting Information for "Identifying and quantifying the impact of climatic and non-climatic drivers on river discharge in Europe"

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# Contents of this file

- 1. Complements on validation and robustness of the method
- 2. Map of the specific catchments referred to in the study
- 3. Illustration of the analysis at the catchment level

# Introduction

This supplementary information file contains a complementary explanation of the sensitivity and robustness tests done for the study. It also includes a map of the catchments specifically referred to in the article. Finally, it includes a detailed illustration of the methodology at the catchment scale for a given catchment in Northern Spain.

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# Complements on validation and robustness of the method

# a- Adequacy of the framework

The method aims to compare the trends in river discharge  $Q_{LSM}$  from the model, which represents the climatic conditions only, and  $Q_{obs}$  deduced from observations, which represent all conditions at once, the actual conditions. They are not compared directly but through their substitutes  $Q_{climat}$  and  $Q_{actual}$  determined with the Budyko framework, which facilitates the interpretation of partial trends. We, therefore, need to attest to the quality of the Budyko framework to reproduce  $Q_{LSM}$  and  $Q_{obs}$  through their parametric representation.

We use the Nash-Sutcliffe coefficient (NSC) and the Percent bias (PBIAS) (Fig. S1). The Budyko fraemwork is able to reproduce correctly the annual mean of observed river discharge over all European basins with a very good PBIAS (<10 % for all river basins) (Fig. S1d) and a good NSC > 0.5 for 569 stations out of 849, except for north-eastern Europe, where we locate the majority of stations which fail this test (Fig. S1c). This second test is more demanding and attests to the quality of Budyko framework to reproduce the inter-annual variations of discharge. It is also efficient to reproduce the climatic river discharge from the model (Fig. S1a and S1b) with NSC > 0.5 and  $PBIAS \leq 15\%$  except for a few basins and still an under-performance for NSC over Eastern Europe.

Therefore, the Budyko framework is an adequate parametric representation of annual mean discharge in both systems and we can use  $Q_{climat}$  and  $Q_{actual}$  derived from this framework to compare the climatic behavior of the watershed and its actual behavior.

In this study, we filter out the stations for which NSC < 0.5. We only keep the 569 stations for which the Budyko framework is efficient for both reproducing  $Q_{LSM}$  and  $Q_{obs}$ . Therefore, the analysis when comparing  $Q_{climat}$  and  $Q_{total}$  will not be tinted by the ability of Budyko framework to effectively reproduce  $Q_{LSM}$  and  $Q_{obs}$  respectively.

#### b- Robustness to climate data

We also tested the method's robustness and its sensitivity to data driving the LSM by comparing its application with different forcing datasets. Three independent atmospheric datasets are available over the 1979-2010 period.

Over such a short period, trends are mostly non-significant and can't be appropriately statistically compared. However, for all forcings considered, the patterns are very similar. Here we focus on the efficiency parameters  $\omega_c$  and  $\omega_a$  correlation and variance for each forcing, to analyze the impact of the forcing choice on how our method attributes variations of  $\omega$  to climatic behavior with our LSM.

Comparing  $\omega_c$  and  $\omega_a$  obtained for the three forcings over the common period and for each system, we obtain very similar results when looking at the average variance over all basins for each evapotranspiration efficiencies time-series and the two-by-two correlations (Tab. S1).

The variances have a similar order of magnitude no matter the forcing used to calculate  $\omega_c$  and  $\omega_a$ , consistently producing  $\omega_a$  larger than  $\omega_c$  by a factor of ten with all forcings. E2OFD has a finer resolution, increasing the results' variability relative to the other two coarser climate datasets. The forcing datasets are not fully independent given the limited

number of observations. For instance, GSWP3 and WFDEI use the same precipitation product to bias correct the re-analyses on which they are based. E2OFD and WFDEI use the same re-analysis but interpolated to different resolutions and corrected with two distinct observational precipitation estimates. Given that the results are closer for GSWP3 and WFDEI, we can hypothesize that the method is more sensitive to the precipitation data used than the other variables.

These results show globally that the method is robust, since it is not very sensitive to the forcing used. The differences in variance between forcings are smaller than those between the variance of  $\omega_a$  and  $\omega_c$  for all tested forcings. The poorest correlation is between E2OFD and GSWP3 (the forcings most different from each other) and mostly for  $\omega_c$ , which has the smallest average variance. Therefore, it will impact our results less when comparing trends. However, the absolute values of  $\omega$  are significantly different depending on the forcing used, comforting the idea that this method can only be used to assess and compare trends.

Results presented in the article are obtained with the forcing dataset GSWP3, which covers the longest time period 1901-2012 and is thus most relevant for evaluating the driver of river discharge trends.



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Figure S1. Using Nash-Sutcliffe coefficient (NSC) and absolute Percent bias (PBIAS) to compare river discharge modelled  $Q_{LSM}$  or observed  $Q_{obs}$  to river discharge  $Q_{climat}$  and  $Q_{actual}$  calculated with Fu's equation, to attest the quality of the Budyko framework. Colors from yellow to pink are considered as satisfactory.

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**Table S1.** Comparison of the evaporation efficiencies time-series calculated with the different forcings for each system over the period 1979-2010:  $\omega_c$  for the climatic system and  $\omega_a$  for the actual system.

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Average variance over all catchments			
	$\omega_c$	$\omega_a$	
GSWP3	0.0023	0.039	
WFDEI	0.0033	0.036	
E2OFD	0.0110	0.031	

# Correlations:

% of stations with average correlation > 0.6 and median correlation between all catchments

		$\omega_c$		$\omega_a$	
E2OFD/GSWP3	38%	0.50	53%	0.65	
WFDEI/GSWP3	73%	0.75	77%	0.99	
E2OFD/WFDEI	64%	0.70	59%	0.73	



Map of the specific catchments referred to in the study

Figure S2. Catchments of specific rivers in Spain (Ebro, Duero, Guadiana) and Italy (Tiber, Po), referred to in the article as specific examples of some results and hypotheses.

# Illustration of the analysis at the catchment level

The discharge (Fig. S3b) at the station level has continuous observations from the 1950's  $(Q_{obs})$ . We see that if the variability of  $Q_{obs}$  and  $Q_{mod}$  are very similar (Fig. S3b), we see that over the observation period covered by the observation, at the beginning of the period (1950-1970),  $Q_{obs} > Q_{mod}$  while at the end of the period (1990-2010),  $Q_{obs} < Q_{mod}$ . Both tend to decrease but  $Q_{obs}$  has a steeper decrease. Looking at the variations of  $\omega$  (Fig. S3c) in both systems helps to explain that difference.  $\omega_c$  is not constant over time but its variability is smaller than that of  $\omega_a$ . There are other non-climatic factors inducing higher trends. For the particular case of Castejon, there are two time periods at the end of the 1960's end in the 1985-1995 period

where there are trends in  $\omega_a$  with a slope which is higher than 90% of all of  $\omega_c$  slopes over the entire century (Fig. S3d). Therefore, there is a high probability that these slopes can not be explained only by climatic phenomena. They are positive trends: non-climatic factors tend to increase evaporation efficiency (associated with a decrease in discharge, not significant, however, at the decadal scale).

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6226300: Ebro, Rio : Castejon, Lon: -1.69° Lat: 42.18°







(b) Discharge at the station outlet: Observed discharge  $Q_{obs}$  (orange), modeled discharge from the LSM ORCHIDEE  $Q_{mod}$  (blue) and from the Budyko framework fitted on the model  $Q_{climat}$  (dotted blue) and on the observations  $Q_{actual}$  (dashed orange)



(c)  $\omega$  fitted on the model outputs ( $\omega_c$  (blue) corresponding to the "climatic"  $\omega$ , compared to  $\omega_a$  (orange) fitted on the observations).



(d) Slopes of  $\omega$  calculated with an 11-year time moving window (slope calculated over 11 years, 5 years prior and after the referenced year), for  $\omega_c$  (blue) and for  $\omega_a$  (orange). The red points corresponds to years for which the absolute slope of  $\omega_a$  is different from 90% of all  $\omega_c$  slopes (grey area).

February 1, 2024, 2:43pm

Figure S3. Example of the results at the station level for the gauging station Castejon on the

Ebro river in Spain