

Bedrock controls on water and energy partitioning across the western contiguous United States

Robert Ehlert¹, W. Jesse Hahm¹, David N Dralle², Daniella Rempe³, and Diana M. Allen¹

¹Simon Fraser University

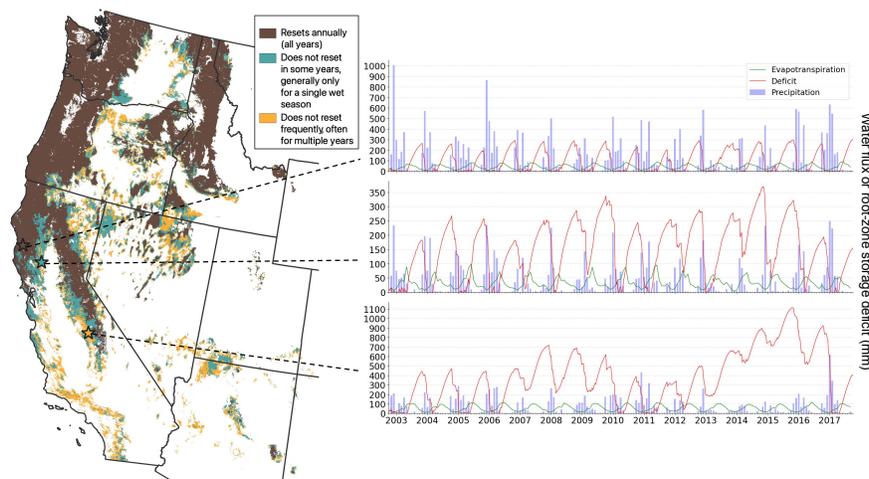
²Pacific Southwest Research Station, United States Forest Service

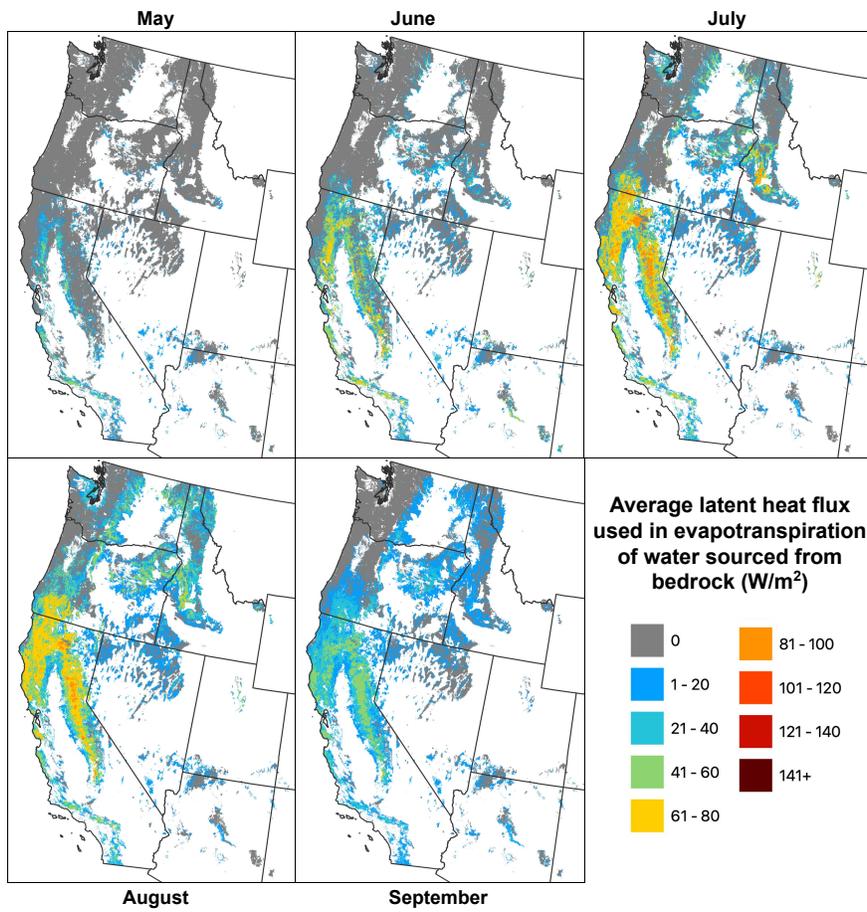
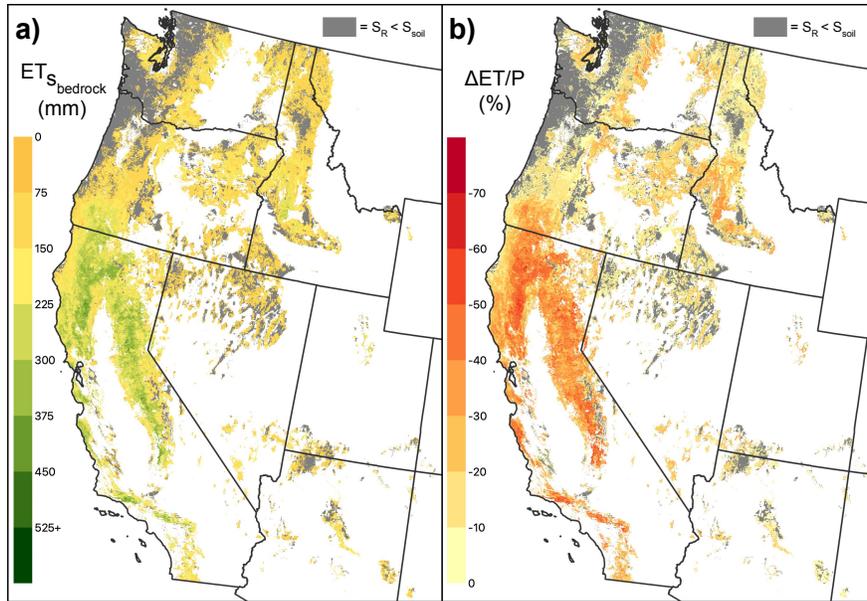
³University of Texas at Austin

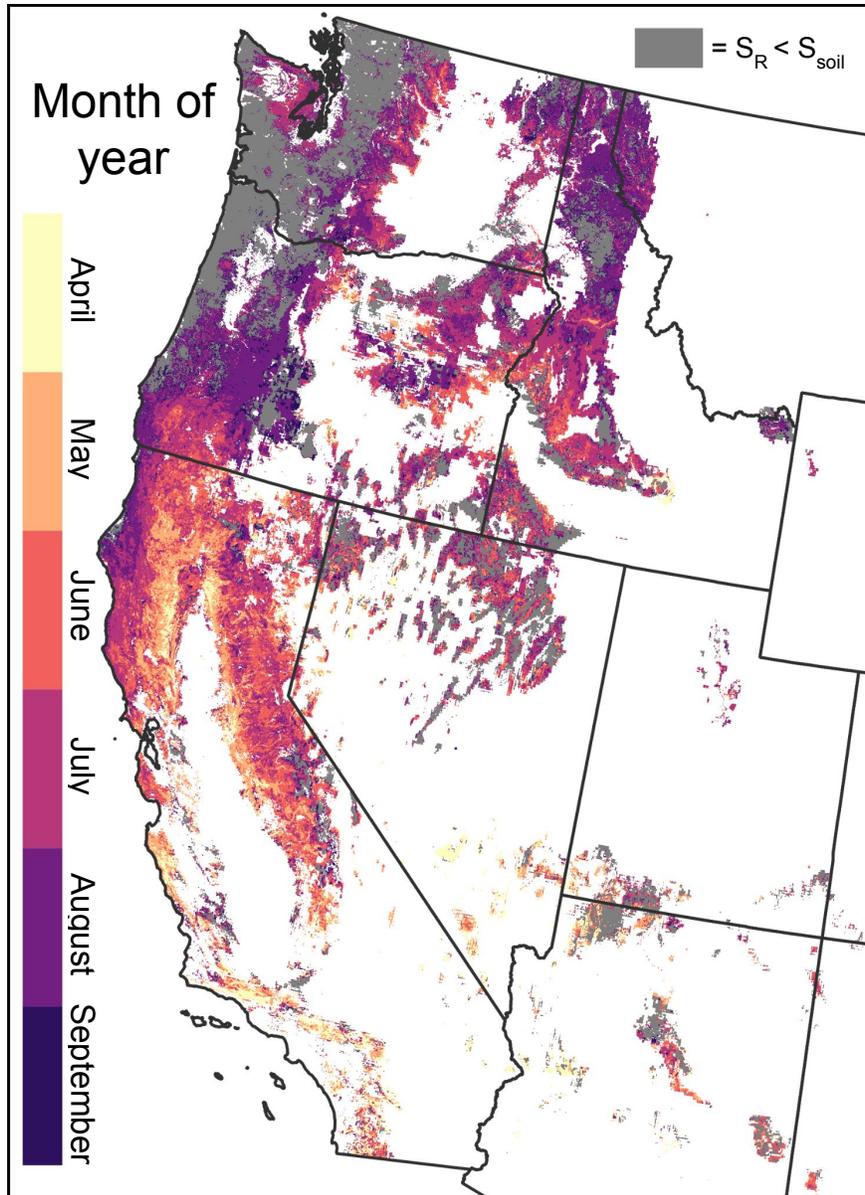
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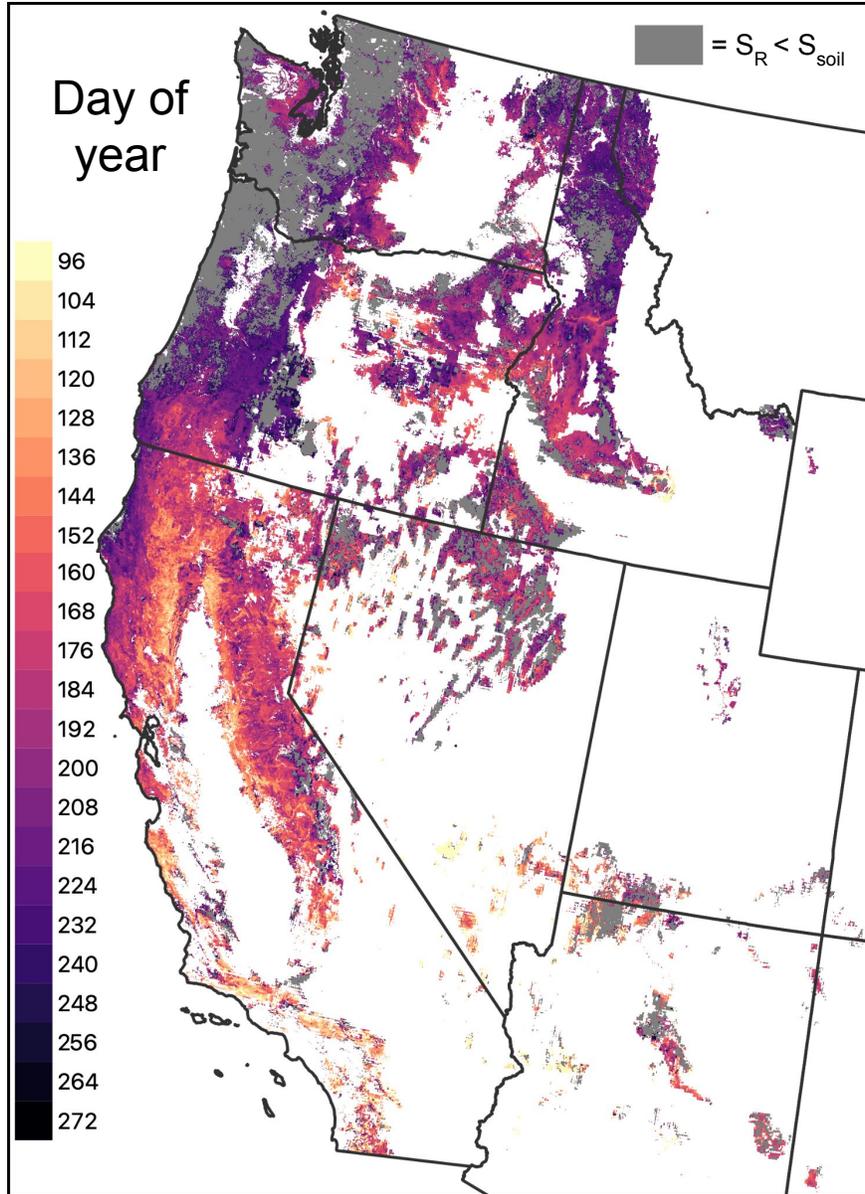
Abstract

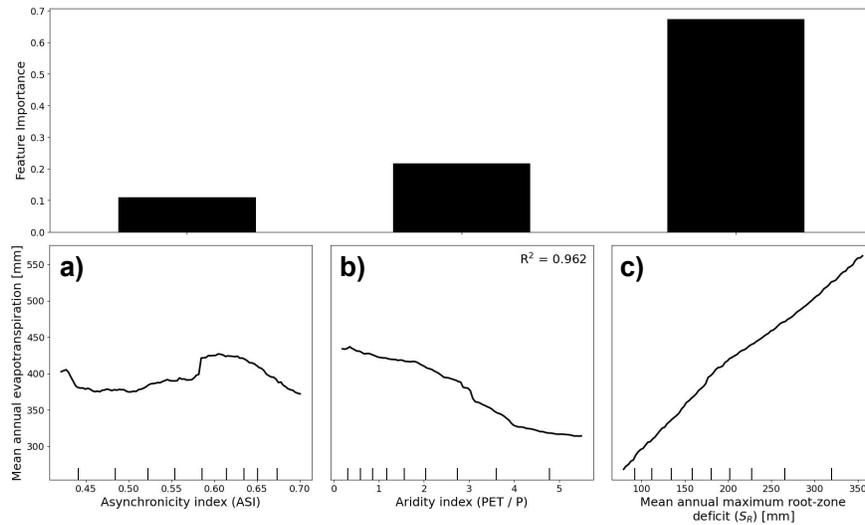
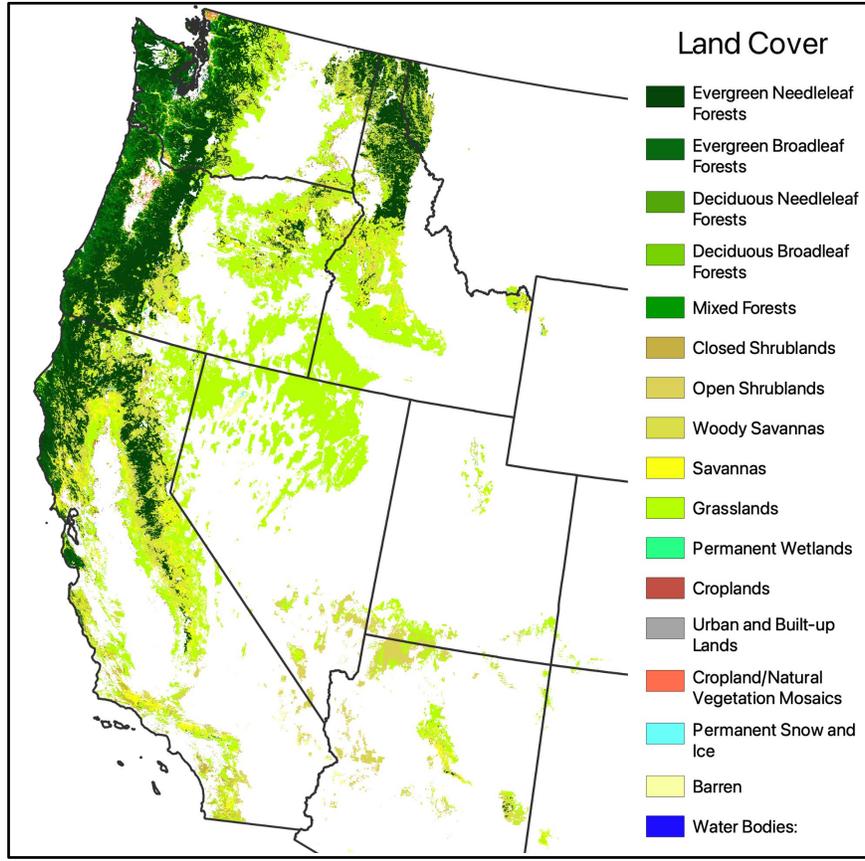
Across diverse biomes and climate types, plants use water stored in bedrock to sustain transpiration. Bedrock water storage (S_{bedrock} , mm), in addition to soil moisture, thus plays an important role in water cycling and should be accounted for in the context of surface energy balances and streamflow generation. Yet, the extent to which bedrock water storage impacts hydrologic partitioning and influences latent heat fluxes has yet to be quantified at large scales. This is particularly important in Mediterranean climates, where the majority of precipitation is offset from energy delivery and plants must rely on water retained from the wet season to support summer growth. Here we present a simple water balance approach and random forest model to quantify the role of S_{bedrock} on controlling hydrologic partitioning and land surface energy budgets. Specifically, we track evapotranspiration in excess of precipitation and mapped soil water storage capacity (S_{soil} , mm) across the western US in the context of Budyko's water partitioning framework. Our findings indicate that S_{bedrock} is necessary to sustain plant growth in forests in the Sierra Nevada — some of the most productive forests on Earth — as early as April every year, which is counter to the current conventional thought that bedrock is exclusively used late in the dry season under extremely dry conditions. We show that the average latent heat flux used in evapotranspiration of S_{bedrock} can exceed $100 \text{ } \$\text{W}/\text{m}^2$ during the dry season and the proportion of water that returns to the atmosphere would decrease dramatically without access to S_{bedrock} .

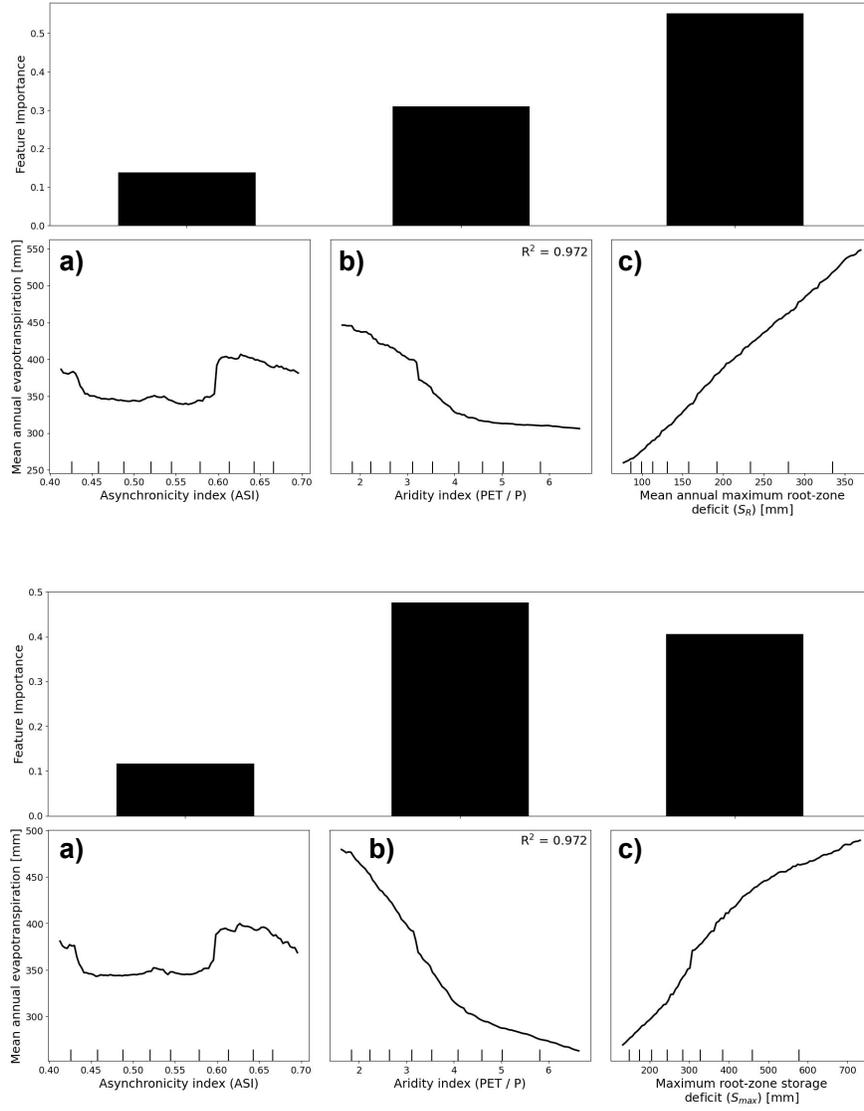


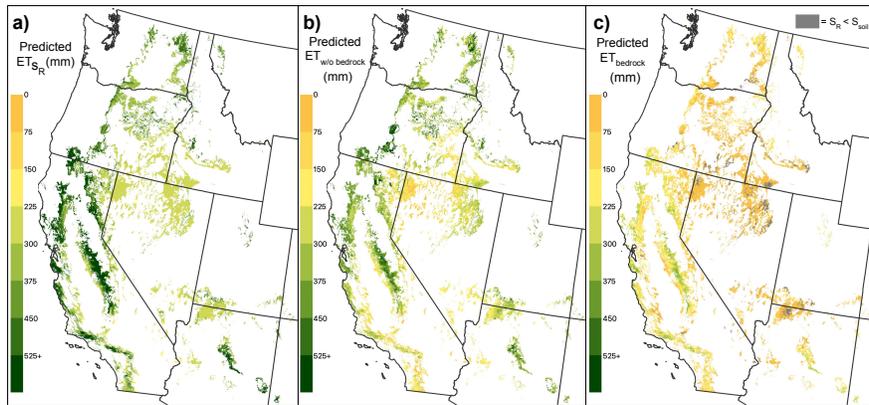
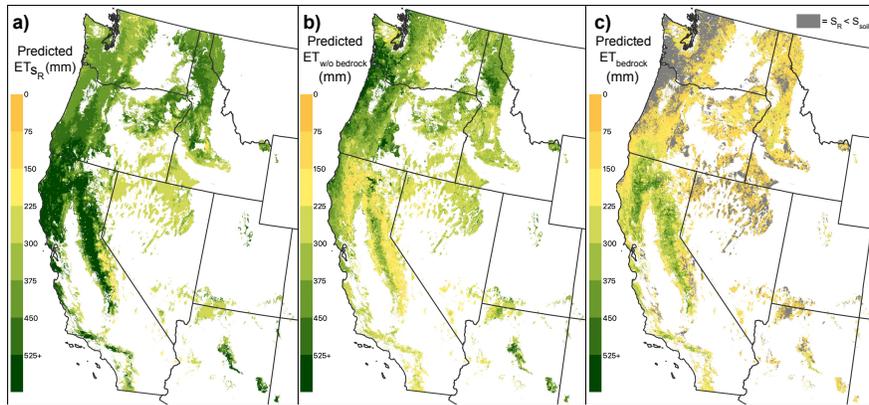
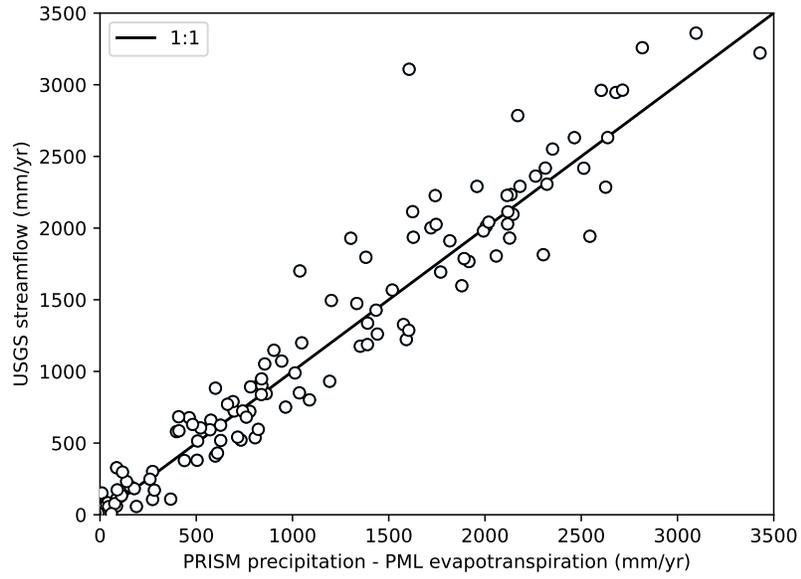


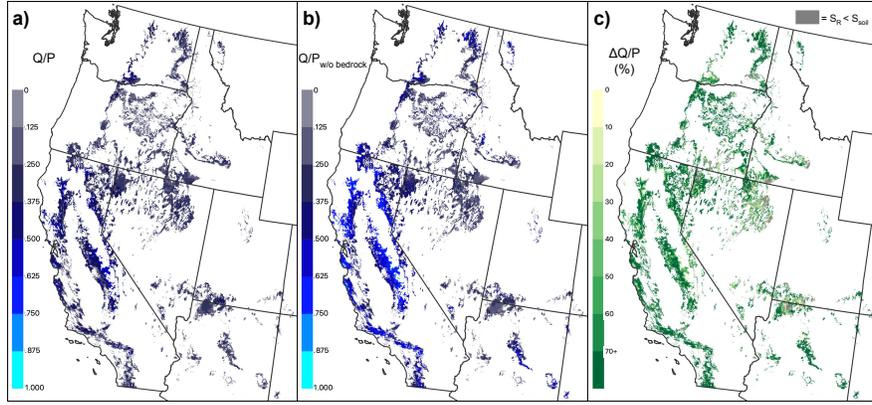


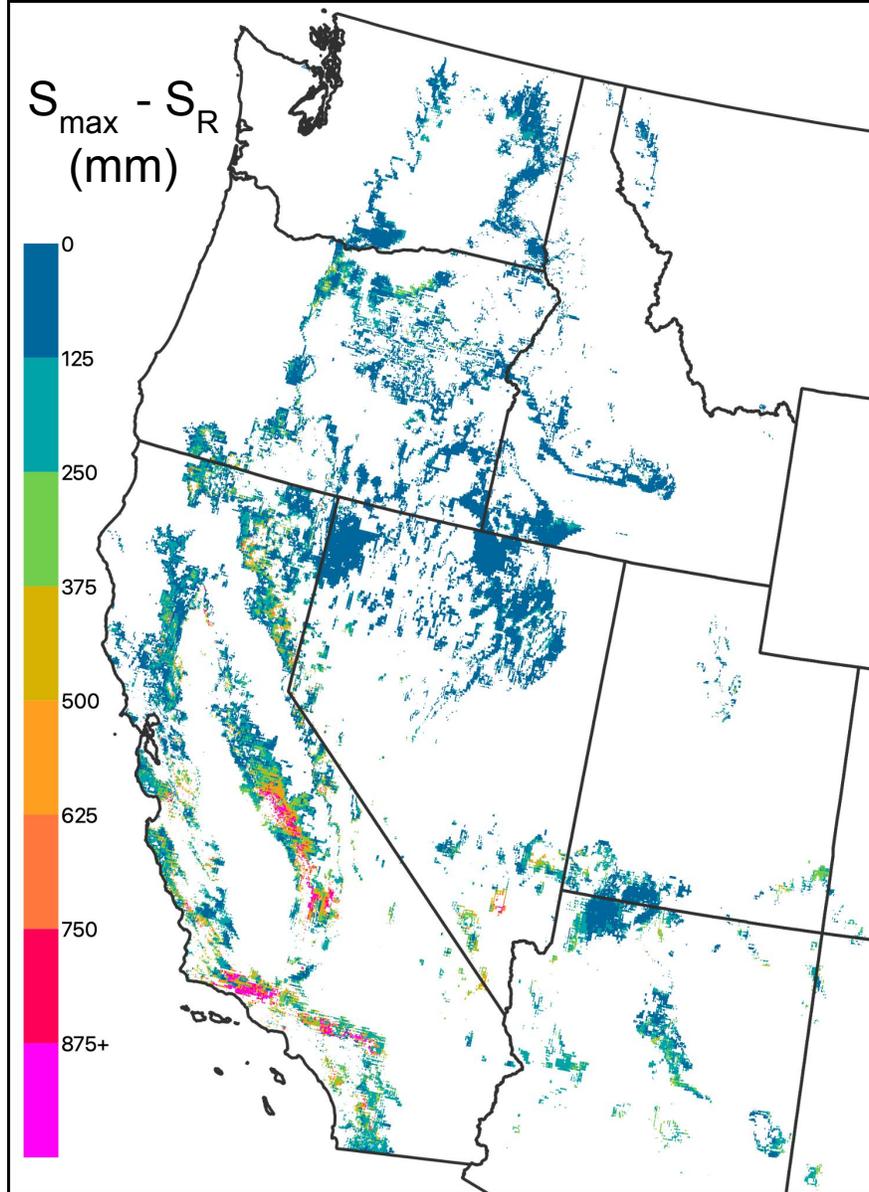


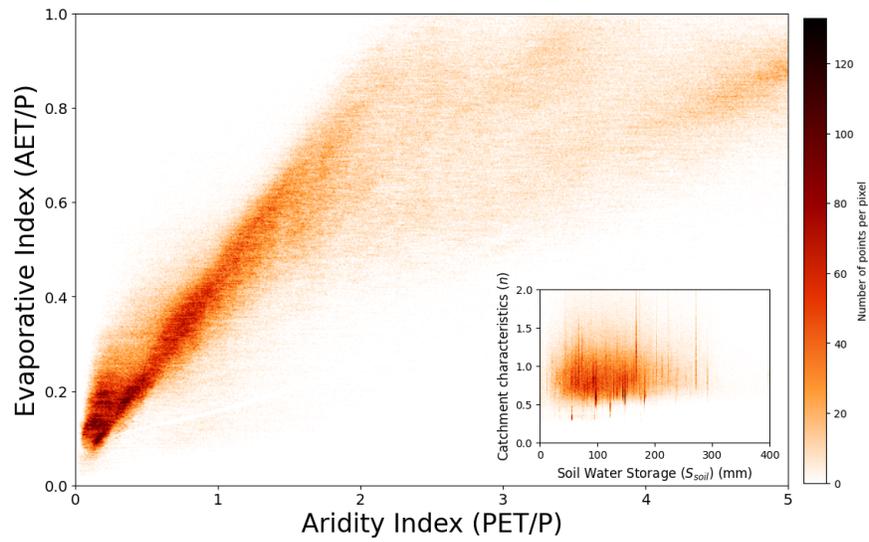
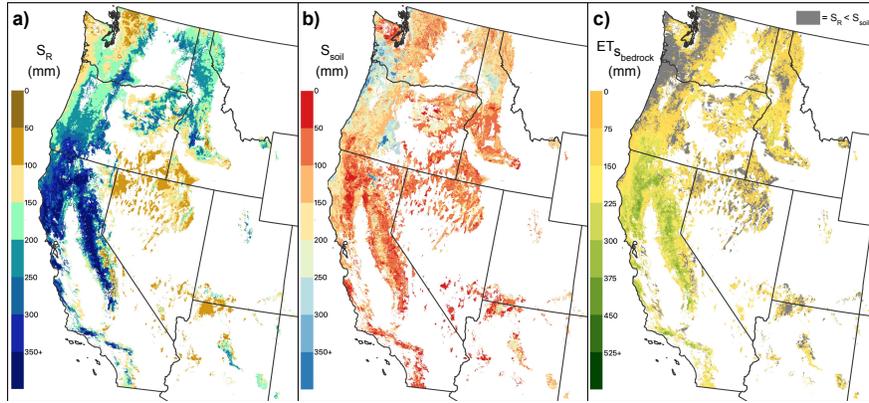
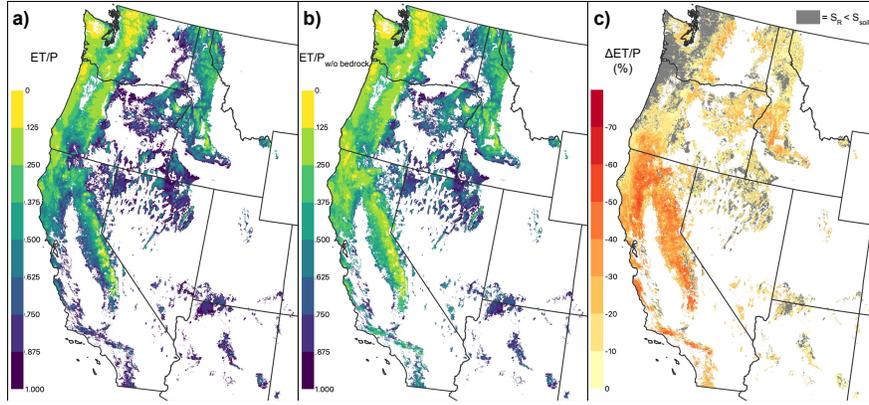


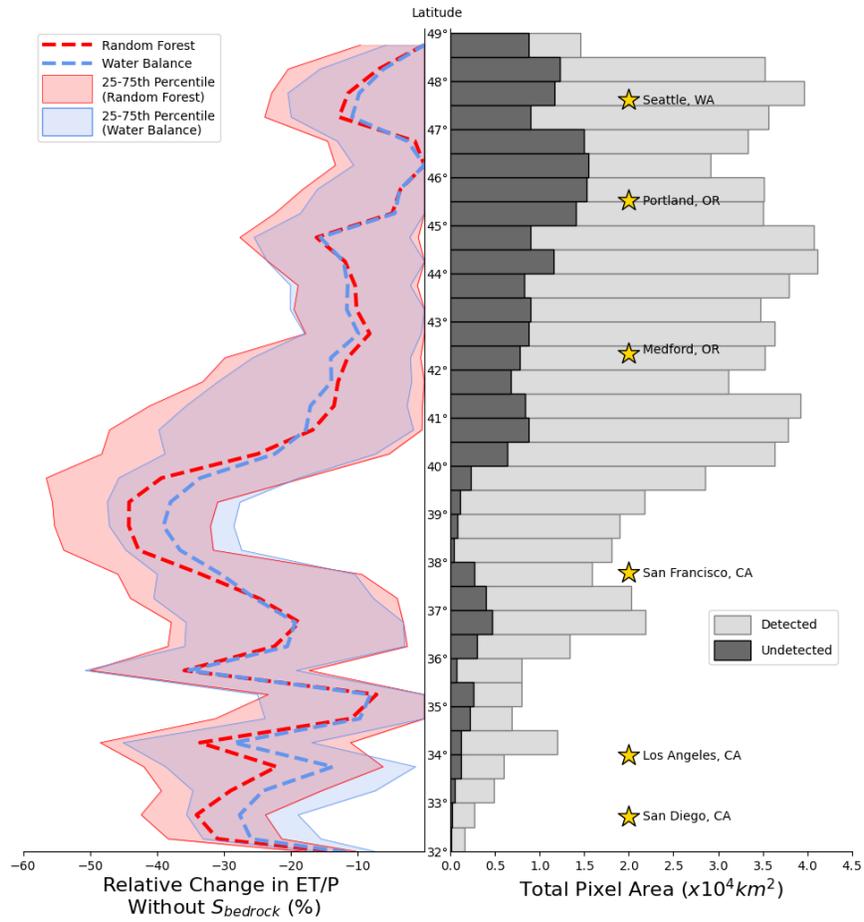
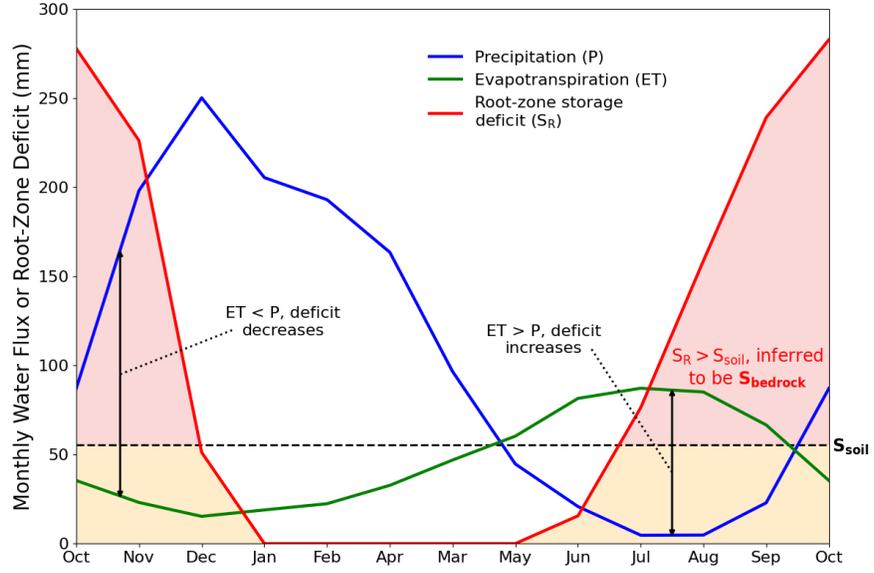


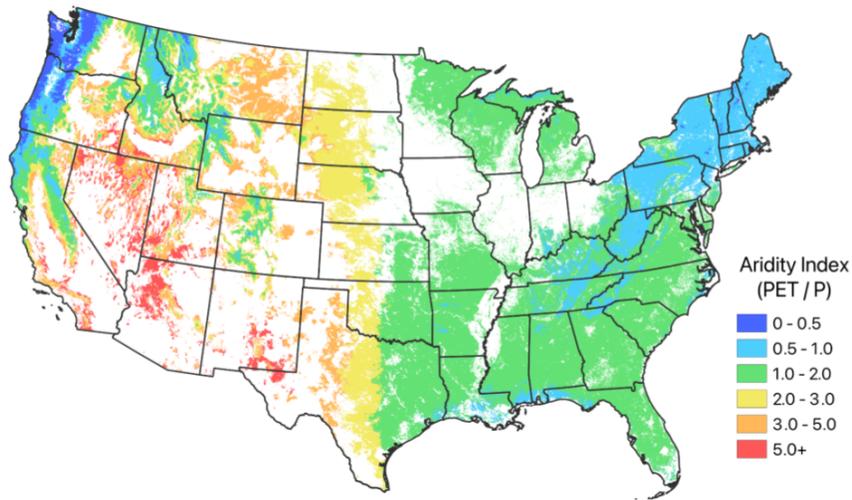
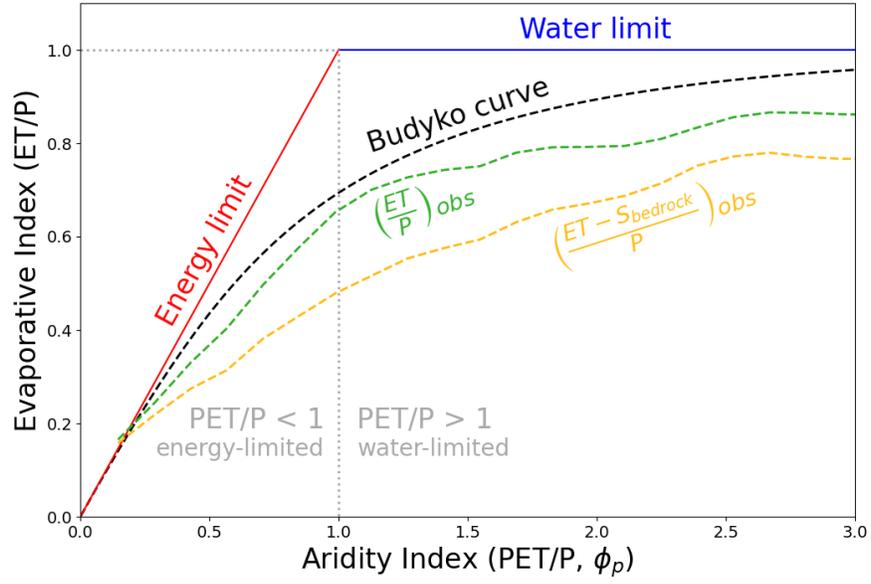


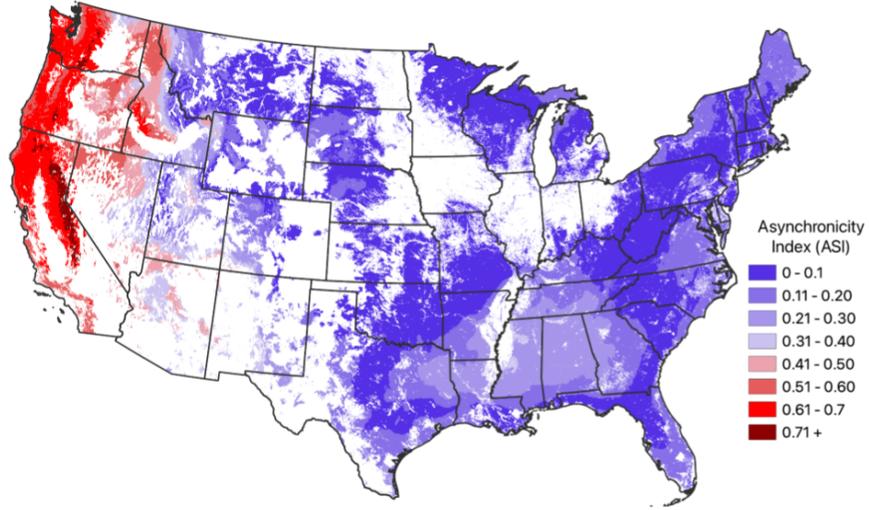












Abstract

Across diverse biomes and climate types, plants use water stored in bedrock to sustain transpiration. Bedrock water storage ($S_{bedrock}$, mm), in addition to soil moisture, thus plays an important role in water cycling and should be accounted for in the context of surface energy balances and streamflow generation. Yet, the extent to which bedrock water storage impacts hydrologic partitioning and influences latent heat fluxes has yet to be quantified at large scales. This is particularly important in Mediterranean climates, where the majority of precipitation is offset from energy delivery and plants must rely on water retained from the wet season to support summer growth. Here we present a simple water balance approach and random forest model to quantify the role of $S_{bedrock}$ on controlling hydrologic partitioning and land surface energy budgets. Specifically, we track evapotranspiration in excess of precipitation and mapped soil water storage capacity (S_{soil} , mm) across the western US in the context of Budyko’s water partitioning framework. Our findings indicate that $S_{bedrock}$ is necessary to sustain plant growth in forests in the Sierra Nevada — some of the most productive forests on Earth — as early as April every year, which is counter to the current conventional thought that bedrock is exclusively used late in the dry season under extremely dry conditions. We show that the average latent heat flux used in evapotranspiration of $S_{bedrock}$ can exceed 100 W/m^2 during the dry season and the proportion of water that returns to the atmosphere would decrease dramatically without access to $S_{bedrock}$.

Plain Language Summary

Plants frequently use water stored in bedrock ($S_{bedrock}$) in order to grow. However, the proportion of precipitation that returns to the atmosphere (evapotranspiration) vs. to streams (runoff) as well as how much latent heat — the energy associated with evaporating water — is used as a result of access to $S_{bedrock}$ has not been measured. In Mediterranean climates, such as the western US, the majority of energy (sunlight) is received during the dry season and plants must rely on water stored belowground during the wet season to sustain summer growth. In this study, we present two methods for calculating how much $S_{bedrock}$ influences the amount of water returning to the atmosphere vs. streams and what that corresponds to in terms of latent heat energy at the surface. We use gridded data to compare the amount of water entering (precipitation) and exiting (evapotranspiration) the area and use a mapped soil water storage capacity product to draw conclusions about the timing and magnitude of plant transpiration that is a result of access to bedrock water. Our findings indicate that some of the Earth’s most productive forests use $S_{bedrock}$ early in the growing season, consuming over 100 W/m^2 of latent heat energy in the summer.

1 Introduction

Globally, a greater proportion of precipitation is returned to the atmosphere via evapotranspiration (ET) compared to oceans via streamflow (Q) (Jasechko et al., 2013; Trenberth et al., 2007). Locally, precipitation partitioning between streamflow and evapotranspiration is mediated by local climate (Budyko, 1974). In asynchronous climates, where the majority of precipitation is offset from energy delivery (Feng et al., 2019; Klos et al., 2018), a substantial proportion of plant transpiration is sourced from bedrock water storage ($S_{bedrock}$) (Hahm et al., 2020; Hubbert, Beyers, & Graham, 2001; McCormick et al., 2021; Rempe & Dietrich, 2018; Rose et al., 2003; Witty et al., 2003). Yet, there have been no attempts to quantify the extent to which bedrock water storage alters annual hydrologic partitioning in asynchronous climates. Moreover, global climate models (GCMs) typically only consider soil moisture dynamics when modelling latent heat flux — the transfer of heat between the terrestrial biosphere and atmosphere — which may work well for humid regions but poorly accounts for climates where plants rely on

67 water stored deep in the subsurface to compensate for a lack of precipitation during the
 68 summer dry season. Soil moisture content has been shown to influence extreme daily tem-
 69 peratures (Durre et al., 2000), regulate the number of large fires (Jensen et al., 2018) and
 70 length of the wildfire season (Rakhmatulina et al., 2021), and was a contributing factor
 71 to the 2003 record-breaking heat wave in Europe (Fischer et al., 2007). It stands to reason
 72 that bedrock water storage, in addition to soil moisture, should be considered when
 73 evaluating land energy budgets and hydrologic partitioning.

74 The relative magnitudes of the water balance components at a location are dictated
 75 by the availability of water supply (precipitation) vs. demand (energy) (Budyko, 1974).
 76 Over long time frames, where change in storage (ΔS) can be considered negligible, the
 77 ratio of evapotranspiration relative to precipitation (i.e. the evaporative index, $\epsilon = ET/P =$
 78 $1 - Q/P$) can be estimated based on the ratio of potential evapotranspiration (PET)
 79 relative to precipitation (the aridity index, $\Phi = PET/P$; see Table 1 for a list of vari-
 80 ables and their definitions). In practice, most catchments fall near a single curve — the
 81 Budyko curve — when plotted in ET/P versus PET/P space, with deviations from this
 82 curve resulting from seasonality (Feng et al., 2012; Hickel & Zhang, 2006; Xing et al.,
 83 2018), vegetation cover (Chen et al., 2013; R. Donohue et al., 2007; M. Liu et al., 2022;
 84 L. Zhang et al., 2001), subsurface storage dynamics (Milly, 1994a), and other catchment-
 85 specific characteristics (e.g. Lhomme & Moussa, 2016; H. Yang et al., 2014). Numerous
 86 parametric extensions have been proposed to the Budyko equation (e.g. Choudhury, 1999;
 87 Fu, 1981, etc.) and a general solution has been mathematically derived that captures the
 88 catchment characteristics in a single parameter (H. Yang et al., 2008). The relationship
 89 described by Budyko also emerges from process-based hydrological models (e.g. R. J. Dono-
 90 hue et al., 2012; Entekhabi & Rodriguez-Iturbe, 1994; Feng et al., 2015; Porporato et al.,
 91 2004, etc.).

92 Early approaches for estimating subsurface storage deficits, calculated by taking
 93 the difference between precipitation and evaporation over time, date back to at least the
 94 1960s (Grindley, 1960, 1968). In the literature, these methods were used mostly to es-
 95 timate groundwater recharge (e.g. Finch, 2001; Rushton & Ward, 1979; Rushton et al.,
 96 2006, etc.) and were limited by spatial and temporal data resolution. More recently, re-
 97 motely sensed water fluxes have been used to estimate root-zone storage capacities (S_R)
 98 at large scales. For example, continental-scale S_R has been estimated using mass bal-
 99 ance approaches (e.g. de Boer-Euser et al., 2016; Gao et al., 2014; Stocker et al., 2023)
 100 and a methodology for estimating S_R at a global scale has been proposed by Wang-Erlandsson
 101 et al. (2016), and extended to account for snow cover by Dralle et al. (2021), which has
 102 been used to investigate ecosystem resilience (Singh et al., 2022), plant water-use sen-
 103 sitivity resulting from interannual rainfall variability (Dralle et al., 2020), and drought
 104 coping mechanisms in rainforest-savanna transects (Singh et al., 2020). Existing field-
 105 scale measurements (e.g. Rempe & Dietrich, 2018), which cannot be extrapolated over
 106 larger scales due to the spatial heterogeneity of plant rooting structures across different
 107 climates soil types and bedrock weathering patterns (Gentine et al., 2012; Sivandran &
 108 Bras, 2013), align well with satellite-derived S_R (McCormick et al., 2021). Root-zone stor-
 109 age capacities calculated via the deficit method influence the proportion of precipitation
 110 that returns to the atmosphere, for a given aridity index, in Australian catchments (Cheng
 111 et al., 2022). When combined with existing soil water storage capacity datasets (e.g. Grid-
 112 ded National Soil Survey Geography Database (gNATSGO); Soil Survey Staff, 2019),
 113 satellite-derived S_R has been used to estimate $S_{bedrock}$ for the contiguous United States
 114 (McCormick et al., 2021).

115 In this study, we examine the extent to which the bedrock root-zone, which extends
 116 beneath the typically thin (< 1 m) soil profile, influences water and energy budgets in
 117 the western US. More specifically, we investigate how plant access to bedrock water con-
 118 trols water partitioning and latent heat fluxes. We use a simple water balance approach
 119 combined with a national soil coverage database (i.e. gNATSGO), gridded water flux data,

120 and a recent dataset of gridded subsurface water storage capacity to provide insights re-
 121 garding the transfer of water found exclusively in bedrock to the atmosphere. We quan-
 122 tify the total amount of annual evapotranspiration accessed from the bedrock root-zone,
 123 and show that plant growth in many parts of the western US relies on bedrock water sur-
 124 prisingly early into the growing season, counter to conventional understandings that bedrock
 125 is used only late in the dry season. Finally, we use a random forest model to corrobo-
 126 rate the mass-balance inferences of yearly evapotranspiration that is attributed to ac-
 127 cess to bedrock water reserves.

128 In providing a simple, reproducible framework for quantifying the impacts of $S_{bedrock}$
 129 on hydrologic and energy partitioning we look to answer three questions: (1) How early
 130 into the growing season do plants in asynchronous climates rely on $S_{bedrock}$ to sustain
 131 summer growth?; (2) How does access to bedrock water impact the partitioning of pre-
 132 cipitation into evapotranspiration versus streamflow?; and (3) What is the latent heat
 133 flux associated with plant use of bedrock water?

134 2 Methods

135 To assess bedrock controls on water and energy partitioning, we apply two approaches:
 136 (1) an annual water balance, which calculates the total inferred yearly evapotranspira-
 137 tion sourced from bedrock by tracking incoming and outgoing water fluxes; and (2) a ran-
 138 dom forest model, which estimates the total yearly evapotranspiration sourced from bedrock
 139 using a selection of input variables considered to be predictors of evapotranspiration. The
 140 water balance method provides conservative, lower-bound constraints on bedrock wa-
 141 ter use based on conservation of mass, while the random forest model represents a 'best
 142 estimate' approach that relies on additional climate predictors beyond evapotranspira-
 143 tion and precipitation fluxes.

144 In both cases, gridded timeseries of water flux data, in combination with an exist-
 145 ing soil water capacity dataset (gNATSGO), are used to estimate the mean annual evap-
 146 otranspiration sourced from bedrock. However, the input variables of the models differ.
 147 The water balance method tracks incoming (precipitation) and outgoing (evapotranspi-
 148 ration) fluxes, at a pixel scale, to determine the amount of evapotranspiration that can
 149 be attributed to bedrock (i.e. ET in excess of soil water storage) in a typical water year.
 150 The random forest approach trains a model that predicts the mean annual evapotran-
 151 spiration based on a set of variables describing climate and total observed subsurface stor-
 152 age capacity, then replaces the total observed subsurface storage capacity with mapped
 153 soil water storage capacity to predict what total mean annual evapotranspiration would
 154 be without access to bedrock water storage; the difference in mean annual ET predic-
 155 tions between the model trained on the total storage vs. soil-storage capacity only is used
 156 to infer the amount of evapotranspiration attributed to bedrock water storage.

157 Using the water balance approach, we additionally explore which areas in the west-
 158 ern contiguous US are prone to periods when the subsurface deficit is unable to be re-
 159 plenished on an annual basis (Fig. 2). Using these areas, we re-purpose the original ran-
 160 dom forest model, replacing the average annual root-zone storage deficit (S_R) with max-
 161 imum root-zone storage deficit (S_{max}), to make inferences about the water partitioning
 162 properties of regions where the deficit does not always reset annually. Finally, we inves-
 163 tigate the timing of bedrock water use in the growing season and calculate the latent heat
 164 energy used to explore the role of plant use of bedrock water on land surface energy fluxes.

165 2.1 Study Area

166 We restricted our study area to winter-wet, summer-dry climate regions of the west-
 167 ern contiguous US. To identify these climate regions, we use the asynchronicity index
 168 (ASI , (Feng et al., 2019)) calculated from monthly TerraClimate precipitation and po-

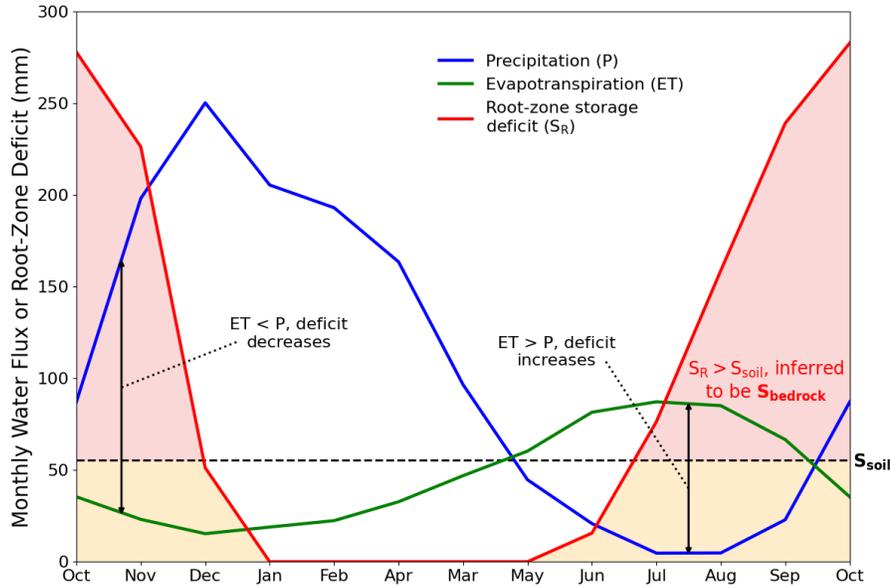


Figure 1. Conceptual diagram describing the root-zone storage characteristics of a typical water year (Oct. 1 - Sep. 30) in regions characterized by asynchronous climates. At the beginning of the wet season, the deficit accrued during the dry season begins to decrease as $P > ET$. Prior to the beginning of the following dry season, the deficit returns to zero and remains at, or near, zero until $ET > P$. When ET remains $> P$ such that the deficit surpasses the soil water storage capacity (S_{soil}), plant transpiration is inferred to be a result of access to water stored below the soil layer, i.e. $S_{bedrock}$. Figure is adapted from Lapides et al. (2022b) Fig. 1d.

169 potential evapotranspiration values (Abatzoglou et al., 2018). We limited the study domain
 170 to pixels with an asynchronicity index greater than or equal to 0.40, which is a slightly
 171 stricter threshold (0.36) than proposed by (Feng et al., 2019) to designate Mediterranean
 172 climates. The masked coverage of the contiguous US, as well as computed asynchronicity
 173 index values, are shown in Fig. S1.

174 We additionally masked out pixels where:

- 175 1. long-term evapotranspiration exceeds precipitation, e.g. due to irrigated agricul-
 176 tural lands or data error;
- 177 2. land cover is classified as urban or water body; or
- 178 3. soil water storage datasets (i.e. gNATSGO) do not have spatial coverage.

179 For this process, we use a gridded climate product from point observations (Parameter-
 180 elevation Regressions on Independent Slopes Model (PRISM); C. Daly et al., 2015), the
 181 Penman-Monteith-Leuning ET product (Y. Zhang et al., 2019), the United States Ge-
 182 ological Survey (USGS) National Land Cover Database (NLCD) land cover classifica-
 183 tion (L. Yang et al., 2018), and the Gridded National Soil Survey Geographic Database
 184 (Soil Survey Staff, 2019). The pixel masking process follows the methodology defined by
 185 McCormick et al. (2021).

186 All gridded timeseries data, including the products described above, are taken for
 187 the 2003 to 2017 water years (Oct. 1 - Sep. 30) and analyzed with the Google Earth En-
 188 gine Python application programming interface (API) (Gorelick et al., 2017).

2.2 Evaluating Storage via Water Balances

Following McCormick et al. (2021), we estimate a lower-bound on the maximum annual root-zone storage deficit (S_R) using the mass balance approached outlined by Wang-Erlandsson et al. (2016) and expanded to account for snow cover by Dralle et al. (2021) (500 m pixel scale). The technique takes the running integrated difference of land-atmosphere water fluxes exiting (F_{out} [L/T] = ET) and entering (F_{in} [L/T] = P) at a pixel. ET is sourced from PML V2 (500 m pixel scale; Y. Zhang et al., 2019) to represent F_{out} and P is extracted from PRISM (4638.3 m pixel scale; C. Daly et al., 2015), to represent F_{in} . Input data was converted from native resolution (shown in parentheses above) to 1000 m and re-projected to the World Geodetic System 1984 (EPGS:4326) for analysis.

First, the accumulated difference between F_{out} and F_{in} is taken for timeframe t_n to t_{n+1} and corrected for the presence of snow based on a snow cover threshold:

$$A_{t_n \rightarrow t_{n+1}} = \int_{t_n}^{t_{n+1}} (1 - [C - C_0]) \cdot F_{out} - F_{in} dt \quad (1)$$

where C_0 is a pre-defined threshold percentage of snow cover, C is snow cover, and $[\cdot]$ is the ceiling operator. When $C > C_0$, F_{out} is unaltered; when $C \leq C_0$, F_{out} is set to 0, thereby ignoring ET when snowmelt may be present. This effectively avoids erroneously accumulating a storage deficit from ET during snowmelt when water may be infiltrating into the root zone (without the need to run a full snowmelt model). We use the Normalized Difference Snow Index (NDSI) snow cover band (Hall et al., 2016) to compute snow cover and set the snow cover threshold to 10%.

Second, the instantaneous root-zone storage deficit can be determined iteratively via the following equation:

$$D_{t_{n+1}} = \max(0, D_{t_n} + A_{t_n \rightarrow t_{n+1}}) \quad (2)$$

where $D_{t_{n+1}}$ is the deficit at time t_{n+1} . If the deficit falls below zero, the cumulative volume resets to zero as the subsurface has been replenished with water.

At each pixel, we compute the mean annual maximum deficit (D_{max} , evaluated Oct. 1 \rightarrow Sep. 30) and infer it to be a lower-bound on annual root-zone storage capacity (S_R). Crucially, this assumes the root-zone storage deficit is replenished on a year-to-year basis which, in many parts of western US, has been shown to not be the case (e.g. Fig. 2; Cui et al., 2022; Goulden & Bales, 2019; Hahm et al., 2022). We then calculate the maximum root-zone storage capacity over the entire study period without the assumption of annual replenishment (S_{max} , evaluated Oct. 1 2002 \rightarrow Sep. 30 2017) to investigate multi-year deficit accrual using the random forest model outlined in Sec. 2.5. We use Eq. 2 to calculate D over the entire study period and take the maximum of those values to represent the lower-bound maximum root-zone storage capacity between Oct. 1 2002 (start) and Sep. 30 2017 (end):

$$S_{max} = \max_{t_{start} \rightarrow t_{end}} (D_{t_{start}}, D_{t_{start+1}}, \dots, D_{t_{end}}) \quad (3)$$

$ET_{bedrock}$, the minimum annual amount of evapotranspiration sourced from bedrock water storage, is inferred to be the difference between the average maximum annual root-zone storage deficit and the soil water storage capacity reported by the Gridded National Soil Survey Geographic Database (Soil Survey Staff, 2019). If the mean annual maximum root-zone storage deficit does not exceed the reported value by gNATGSO, we take this to mean that $S_{bedrock}$ is not needed to explain annual evapotranspiration and set $S_{bedrock} = 0$. This does not necessarily mean that bedrock water storage was not used

226 to support evapotranspiration, but rather that the deficit tracking approach is unable
227 to detect it.

228 Finally, we compute the average first month of year ($MOY_{bedrock}$) when bedrock
229 must be used to explain observed evapotranspiration, by determining the observed month
230 (for the 2003 - 2017 water years) when mean annual root-zone deficit exceeds the total
231 amount of available storage in the soil, implying any evapotranspiration sourced from
232 the subsurface beyond this date must include water sourced from bedrock storage. This
233 does not mean that bedrock storage was not accessed in prior months but rather that
234 it cannot be tracked using the deficit approach. Therefore, this is the latest possible month
235 that bedrock water is used, because it assumes that i) ET is first sourced from S_{soil} un-
236 til it is completely depleted, and ii) that deficits are replenished annually, which may not
237 be the case.

238 2.3 Water Partitioning

239 Within the Budyko (1974) framework, the long-term partitioning of P into ET and
240 Q is a function of the long-term ratio of PET to P . Under these conditions, Q is assumed
241 to include both overland runoff and lateral subsurface flow resulting from infiltration (hence
242 $\Delta S \approx 0$). We took observed evaporative indices ($\epsilon_{obs} = ET/P$) by dividing the mean
243 annual evapotranspiration by precipitation for the 2003 - 2017 water years using data
244 collected from the gridded products described above. We also infer what the evaporative
245 index would be if plants did not have access to bedrock water ($\epsilon_{w/o\ bedrock}$) by re-
246 moving $S_{bedrock} = S_R - S_{soil}$ (the minimum amount of bedrock water used in an av-
247 erage year) from the observed evaporative index. If S_R does not exceed S_{soil} , then our
248 method cannot detect the influence of bedrock on the evaporative index:

$$\epsilon_{w/o\ bedrock} = \begin{cases} ET_{obs} / P & \text{if } S_R \leq S_{soil} \\ [ET_{obs} - (S_R - S_{soil})] / P & \text{if } S_R > S_{soil} \end{cases} \quad (4)$$

Following this, the relative change (expressed as a percentage) in evaporative in-
dex without access to bedrock water is the difference between $\epsilon_{w/o\ bedrock}$ and ϵ_{obs} re-
lative to ϵ_{obs} :

$$\Delta\epsilon = \left(\frac{\epsilon_{w/o\ bedrock} - \epsilon_{obs}}{\epsilon_{obs}} \right) * 100 \quad (5)$$

249 Streamflow data from 128 minimally impacted USGS watershed gauges in the west-
250 ern US are in agreement (Nash-Sutcliffe efficiency of 0.93) with the precipitation (PRISM)
251 and evapotranspiration (PML) data used in our analysis (Fig. S2). Therefore, we find
252 it reasonable to estimate Q from the water balance (i.e. $Q = P - ET$) as, over long time
253 frames, the net groundwater flow out of a catchment is negligible (i.e. $\Delta S \approx 0$). We
254 calculated the runoff ratio (RR) as the difference between one and the observed evap-
255 orative index ($RR = 1 - \epsilon$).

256 2.4 Energy Partitioning

257 We infer the monthly total latent heat flux associated with evapotranspiration sourced
258 from bedrock. The latent heat, i.e. the energy required to change from the liquid to va-
259 por phase, is equal to the the energy required to evaporate the accrued monthly deficit
260 (in mm of water) beyond that provided by soil. We report this value in units of power
261 per unit of area (W/m^2). First, we take the total bedrock water storage extracted for
262 evapotranspiration between two months:

$$ET_{bedrock, month} = \max(0, \min(D_{i+1} - D_i, D_{i+1} - S_{soil})) \quad (6)$$

263 where D_i is the deficit at the beginning of month i . To account for the deficit sourced
 264 from S_{soil} , the difference between months i and $i+1$ is compared against the difference
 265 between month $i+1$ and S_{soil} , returning the lesser of the two values. If this value is be-
 266 low zero, bedrock was not needed to account for evapotranspiration during the month
 267 and $ET_{bedrock, month}$ is set to zero. This calculation assumes that plants first exhaust any
 268 available soil water and subsequently use bedrock water. If plants exhaust soil water and
 269 bedrock water simultaneously throughout the dry season, the method used here to quan-
 270 tify the total latent heat flux associated with bedrock water during the dry season is not
 271 erroneous but rather would shift the bedrock-associated latent heat flux patterns ear-
 272 lier into the dry season.

Secondly, $ET_{bedrock, month}$ (mm) is converted to power per unit area metric (E_e)
 based on the enthalpy of vaporization of a known mass of water:

$$E_e = (ET_{bedrock, month}) * (\rho_w) * (\Delta H_v) * (1/t) \quad (7)$$

273 where ρ_w is the density of water (1000 g/L), ΔH_v is the latent heat of vaporization of
 274 water (2257 J/g), which we do not adjust for local variations in temperature or pressure,
 275 and t is the total seconds between the i th and $i+1$ th month (1 mm of liquid water per
 276 square meter is one liter). The resulting value is an average latent heat flux per second
 277 (i.e. power, W) per m^2 (unit area) for a given time frame.

278 2.5 Random Forest Model

279 The random forest regression model represents an alternative approach to calcu-
 280 lating $ET_{bedrock}$, and is employed here as a means of corroborating the lower-bound, wa-
 281 ter mass balance inferences described above. The approach uses climatic (ASI and Φ)
 282 and subsurface storage (S_R) characteristics to train a model to predict observed mean
 283 annual evapotranspiration, and then feeds S_{soil} in place of S_R into the trained model to
 284 determine what mean annual ET would be without access to bedrock water storage.

285 Random forest regression is a predictive machine learning algorithm that consists
 286 of a collection of decision trees, which are randomly populated with samples, where the
 287 final output is the average of the results of each individual tree (Breiman, 2001). Each
 288 individual model (tree) is uncorrelated, producing many unrelated errors which, when
 289 combined into a single collective model, will increase prediction accuracy. We use three
 290 input variables: S_R , ASI , and Φ to predict annual evapotranspiration. All random for-
 291 est regression models were implemented using the Random Forest module provided by
 292 Scikit-learn, an open-source Python machine learning package (Pedregosa et al., 2011).
 293 The random forest model was trained by randomly selecting 70% of the data as a train-
 294 ing set and setting 30% aside for validation purposes. Hyperparameters were set to de-
 295 fault (scikit-learn v1.3.0) with the exception of minimum leaf samples and maximum fea-
 296 tures, which were set to 5 (default is 1) and the square root of the number of features
 297 (1.0), respectively. Hyperparameters were chosen to best optimize computing time as tweak-
 298 ing the hyperparameters did not significantly improve model performance. The model
 299 was run using 20, 40, 80, 120, and 200 trees with improvements in the models perfor-
 300 mance beyond 40 trees being negligible. Therefore, we chose to run the final product us-
 301 ing 40 trees to minimize computing time.

302 ASI values are calculated using the method outlined by Feng et al. (2019). Follow-
 303 ing Eq. 2, S_R values can be quantified by taking the mean of the maximum deficit ob-
 304 served each water year, resetting the deficit annually. Φ is measured by taking annual
 305 cumulative PET (ΣPET) relative to P (ΣP) and averaging across all water years. P and
 306 PET are summed monthly totals taken from TerraClimate (Abatzoglou et al., 2018).
 307 We then replace S_R with S_{soil} and re-run the analysis using the predictive model pro-
 308 duced using S_R . Framed another way, we forced the model to assume there is no longer
 309 access to $S_{bedrock}$ to infer changes in annual evapotranspiration without access to bedrock
 310 reserves.

Table 1. Description of referenced variables

Variable	Dimensions	Description
A	L	Accumulated difference; calculated as $F_{out} - F_{in}$ over a given timeframe
ASI	(-)	Asynchronicity index
C	(-)	Snow cover
C_0	(-)	Snow cover threshold
D_{max}	L	Maximum observed annual root-zone deficit
D_{min}	L	Minimum observed root-zone storage deficit in a year
$D_{t_{n+1}}$	L	Root-zone storage deficit measured at time t_{n+1}
E_e	MT ⁻³	Latent heat flux associated with evapotranspiration sourced from bedrock water storage, expressed as power per unit area
ET	LT ⁻¹	Evapotranspiration
ET_{obs}	LT ⁻¹	Observed evapotranspiration
$ET_{bedrock}$	LT ⁻¹	Minimum annual evapotranspiration sourced from bedrock water storage
$ET_{bedrock, month}$	L	Monthly (dry season) evapotranspiration sourced from bedrock water storage
F_{in}	LT ⁻¹	Inflow
F_{out}	LT ⁻¹	Outflow
$MOY_{bedrock}$	(-)	Average month of year when bedrock is needed to explain evapotranspiration
n	(-)	Variable used to quantify differences in the evaporative index for a particular aridity index, defined by (H. Yang et al., 2008)
P	LT ⁻¹	Precipitation
P_{obs}	LT ⁻¹	Observed precipitation
PET	LT ⁻¹	Potential evapotranspiration
Q	LT ⁻¹	Runoff (streamflow)
RR	(-)	Runoff ratio; calculated as $1 - \epsilon$
$S_{bedrock}$	L	Minimum plant-available water storage capacity in bedrock, inferred from largest deficit in an average water year less mapped S_{soil}
S_{max}	L	The minimum root-zone plant-available storage capacity, inferred from maximum deficit observed over entire time period of analysis
S_R	L	Mean annual root-zone storage capacity inferred from maximum deficit observed over a water year
S_{soil}	L	The maximum amount of plant-available water capable of being stored in the soil profile, from soils mapping
t	T	Time
ΔH_v	ML ² T ⁻²	Enthalpy of vaporization of water
ΔS	LT ⁻¹	Change in storage
$\Delta \epsilon$	(-)	Relative difference between ϵ_{obs} and $\epsilon_{w/o bedrock}$
ϵ	(-)	Evaporative index; calculated as ET/P
ϵ_{obs}	(-)	Observed evaporative index; calculated as ET_{obs}/P_{obs}
$\epsilon_{w/o bedrock}$	(-)	Observed evaporative index without bedrock water storage; calculated as $\epsilon_{obs} (ET_{obs} - S_{bedrock}/P_{obs})$
ρ_w	L ⁻³ M	Density of water
Φ	(-)	Aridity index; calculated as PET/P

311 Finally, following the extended methodology of Sec. 2.2, which removes the assumption
 312 of a yearly resetting deficit, we isolate pixels that have observed multi-year deficit
 313 accruals during the study period to investigate the influence of extended drought conditions
 314 on water partitioning. Between Oct. 1 2002 and Sep. 30 2017, we calculate the
 315 number of years where the deficit was not replenished by taking the minimum root-zone
 316 storage deficit (D_{min}) observed each water year for all pixels. If $D_{min} > 0$ in a given
 317 water year, we take that to mean that the deficit was not replenished in that water year.
 318 The resulting pixels were divided into three classes: 1) Deficit resets annually (all years),
 319 2) Deficit resets most years (deficit resets $>66\%$ of the years), and 3) Deficit resets in-
 320 termittently (deficit resets $<66\%$ of the years) (Fig. 2). For these pixels, we amend the
 321 original random forest model to use S_{max} , as opposed to S_R , in order to better repre-
 322 sent the extent to which $S_{bedrock}$ alters hydrologic partitioning in areas where multi-year
 323 deficits occur. All other model characteristics (i.e. hyperparameters, input variables, etc.)
 324 were retained from the original random forest model.

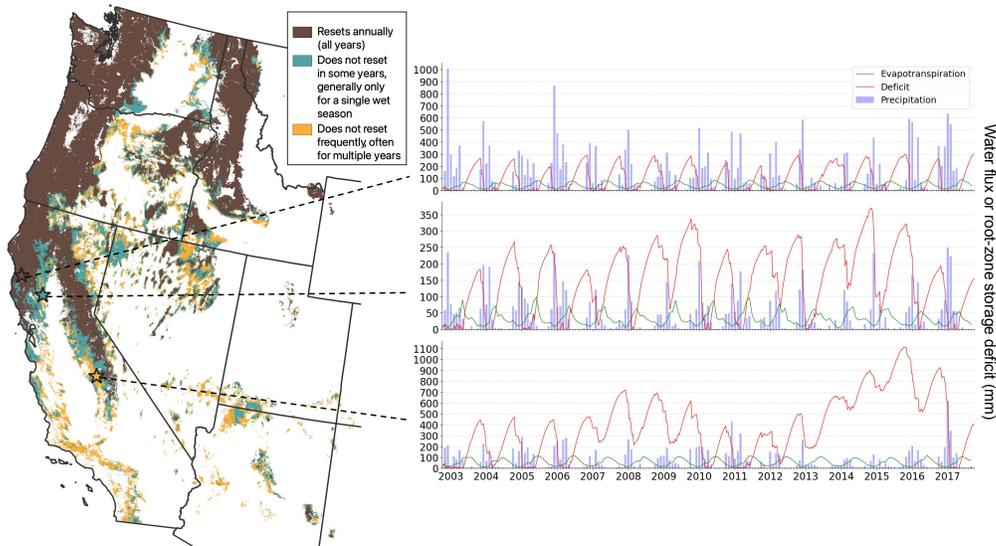


Figure 2. During the study period (Oct. 1 2002 to Sep. 30 2017), the annual deficit returned to zero every year for the regions shown in brown (map on the left). Regions shown in teal and orange, respectively, did not reset in some years ($<33\%$ of the study period) or did not reset frequently, often for multiple years in a row ($>33\%$). For each category, a corresponding example time series of the study period is shown on the right with the relevant fluxes necessary to compute root-zone storage deficit. In forests covering over $26,500 \text{ km}^2$ (land covers 1-5 in Fig. S21), the root-zone storage deficit does not reset annually.

325 3 Results

326 Our primary findings are that i) soil water storage capacity (S_{soil}) does not explain
 327 deviations from the Budyko-curve in asynchronous climates (Fig. 3), ii) the proportion
 328 of terrestrial precipitation returned to the atmosphere (vs. streamflow) is strongly in-
 329 fluenced by plant use of bedrock water reserves (Fig. 4), iii) $S_{bedrock}$ is needed to sus-
 330 tain dry season plant transpiration surprisingly early into the growing season (Fig. 5),
 331 and iv) the summer latent heat flux associated with evapotranspiration of bedrock wa-
 332 ter is substantial (Fig. 7) and warrants further research with respect to land surface en-
 333 ergy interactions. Below, we expand on these findings and highlight particular regions

334 of interest where $S_{bedrock}$ plays an important role in the local water and energy parti-
 335 tioning patterns.

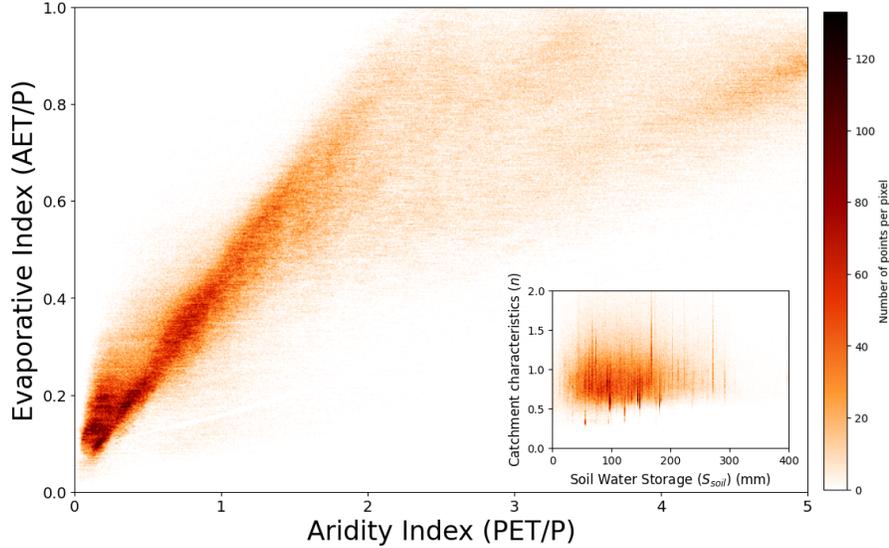


Figure 3. More water is returned to the atmosphere for a given precipitation (higher evaporative index) in locations with more potential evapotranspiration relative to precipitation (aridity index), as shown in this Budyko-space density plot of individual pixels (1000m) with asynchronous climates ($ASI \geq 0.40$) in contiguous United States. The evaporative index for a particular aridity index (expressed in terms of the catchment characteristic, n , where higher n denotes a higher evaporative index for a particular aridity index, see (H. Yang et al., 2008) for derivation) is not well explained by soil water storage (S_{soil}), as shown by the density plot inset.

3.1 Deviations From the Budyko-curve are Poorly Explained by Soil Water Storage Capacity

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338 In our asynchronous climate ($ASI \geq 0.40$) study area, the aridity index explains
 339 the primary trend in the evaporative index for individual pixels, consistent with the Budyko
 340 (1974) findings for catchments (Fig. 3). However, for a given aridity index, there remain
 341 deviations from the curve. It is a commonly hypothesized that, for a particular climate
 342 (held constant here by the use of ASI), subsurface storage capacity may explain devi-
 343 ations from the Budyko-curve (Miller et al., 2012). Using the catchment characteristic
 344 n to quantify differences in the evaporative index for a particular aridity index (H. Yang
 345 et al., 2008), where higher n denotes higher ET/P for a given aridity index, we find that
 346 soil water storage capacity (S_{soil}) alone only explains 11% ($R^2 = 0.11$) of the variance
 347 in n and, therefore, is a poor explanation for deviations from the Budyko-curve across
 348 western US (Fig. 3 inset). Indeed, S_{soil} accounts for only a portion of the below-ground
 349 storage capacity and, in many places, is comparatively small relative to $S_{bedrock}$ (e.g. Mc-
 350 Cormick et al., 2021). Removing ET sourced from bedrock ($ET_{bedrock}$) drastically shifts
 351 the Budyko-curve (Fig. S4). In the following sections, we explore the extent to which
 352 $S_{bedrock}$ may control water and energy partitioning.

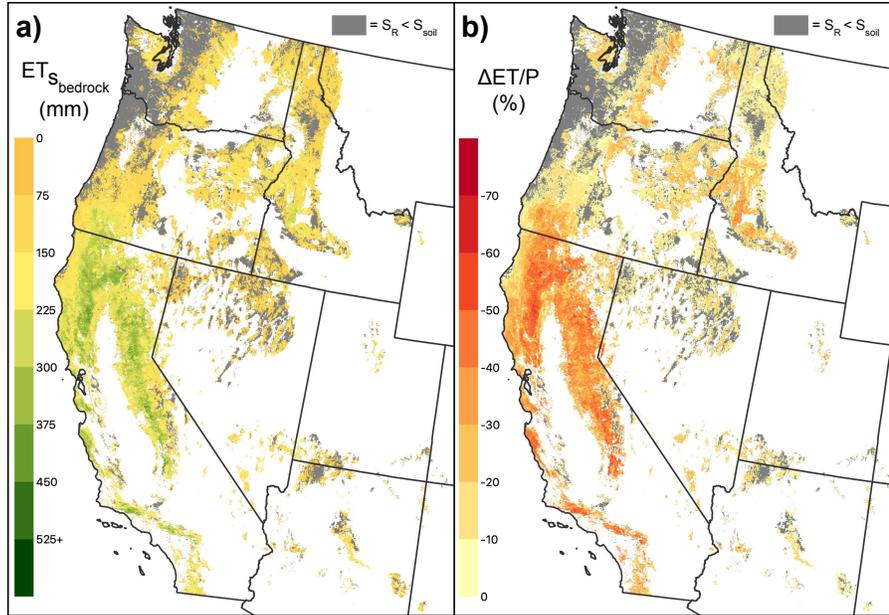


Figure 4. (a) Average value of the largest annually (water year) observed root-zone storage deficit (S_R) in excess of S_{soil} . (b) The relative change in evaporative index when evapotranspiration inferred to be sourced from bedrock storage ($ET_{bedrock}$) is removed, i.e. $(ET - ET_{S_{bedrock}})/P$. (a) inspired by Fig. 3 of McCormick et al. (2021). Across large areas of the western US, annual evapotranspiration would be hundreds of millimeters less and the proportion of precipitation returned to the atmosphere would decrease without access to bedrock water.

3.2 Large Proportions of the Precipitation Returned to the Atmosphere is Sourced from $S_{bedrock}$

In the following section, the spatial patterns of $ET_{bedrock}$, ET/P , and Q/P in the western US, derived using the water balance and random forest methods, are presented. Fig. 4 shows the spatial patterns of evapotranspiration inferred to be sourced from bedrock ($ET_{bedrock}$) and relative change in evaporative index without access to bedrock ($\Delta ET/P$) using the water balance method (see Fig. S5 and S6 for derivation). The corresponding figures using the random forest model can be found in the supplementary information Figs. S9 (ET), S10 (ET/P), and S11 (Q/P). In each case, areas shown in gray represent pixels where bedrock-derived ET was unable to be identified by the proposed methods.

Across the western US, the evaporative index is up to 91% higher (favoring ET) as a result of plant access to bedrock water reserves as opposed to using soil water storage alone. Broadly, evapotranspiration inferred to be sourced from bedrock ($ET_{bedrock}$) increases and relative evaporative index decreases moving south from the USA-Canada border (Fig. 4). In particular, the Northern California Coast Ranges, the southern Cascades, the Transverse Ranges and the Sierra Nevada are most reliant on $S_{bedrock}$ for dry season transpiration. The mean and median changes in relative evaporative index of all pixels in the western US that detected $S_{bedrock}$ use were -16.6 and -13.2%, respectively, using the water balance method. Up to 782 mm of evapotranspiration is inferred to be sourced from bedrock water with mean and median values of 75.8 and 47.9 mm across all pixels, respectively. Areas highlighted in gray did not detect evapotranspiration sourced from bedrock using the deficit approach (i.e. $S_R < S_{soil}$). These areas are mostly lim-

376 ited to the coastal Pacific Northwest, where the aridity index tends to be lower than the
377 rest of the region (Fig. S3), and account for roughly one quarter of all pixels in the study.

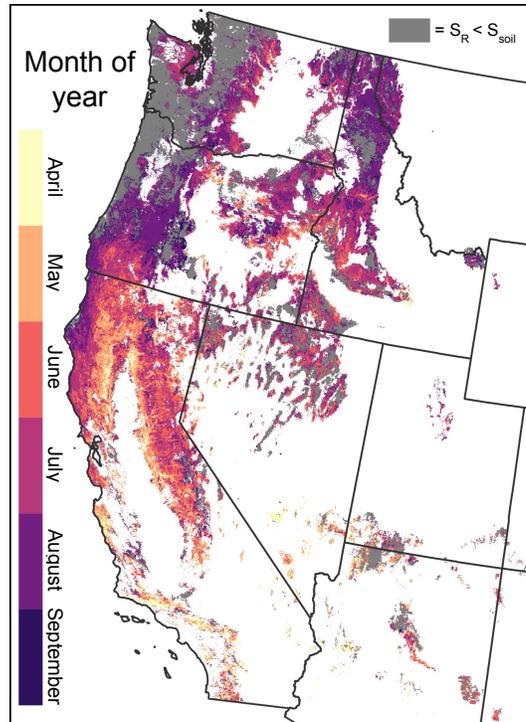


Figure 5. Typical month in which the annual root-zone storage deficit (S_R) exceeds the soil water storage capacity (S_{soil}), implying plant transpiration beyond this point must be using $S_{bedrock}$ to sustain growth. Patterns suggest bedrock water is needed to sustain plant growth very early into the growing season for many parts of the western US.

378 The random forest model driven by mean annual maximum observed $S_{bedrock}$ makes
 379 qualitatively similar predictions ($R^2 = 0.965$; Fig. S8) to the water balance approach based
 380 on mean yearly values (Fig. 6). Areas with a non-resetting deficit are more reliant on
 381 $S_{bedrock}$ to sustain mean annual evapotranspiration when S_{max} is substituted for S_R . In
 382 these areas, S_{max} exceeds S_R by a median value of 82.4 mm (Fig. S20) and S_{soil} val-
 383 ues are low (Fig. S5). For transparency the original model informed by S_R was re-run
 384 in non-resetting pixels as well (Figs. S12-S15). Using only non-resetting pixels (i.e. teal
 385 and orange in Fig. 2), predicted mean and median $ET_{bedrock}$ increased from 87.1 and
 386 54.4 mm to 100.1 and 70.4 mm, respectively, when S_R was substituted with S_{max} and
 387 a new random forest model was run (Fig. S17). Similarly, mean (median) relative evap-
 388 orative index ($\epsilon_{w/o\ bedrock}$) decreased from -19.0% (-14.8) to -23.5% (-22.6) when account-
 389 ing for a non-resetting deficit (Fig. S18-S19). Interestingly, when S_{max} is used as a pre-
 390 dictor instead of S_R , the relative importance of aridity index as a predictor increases sub-
 391 stantially (Fig. S16).

392 **3.3 $S_{bedrock}$ is Needed to Sustain Plant Growth Early into the Grow-** 393 **ing Season and Contributes Substantial Latent Heat Flux as Sum-** 394 **mer Progresses**

395 Regions of high $ET_{bedrock}$ (Fig. 4, S9) also correspond to areas that require $S_{bedrock}$
 396 to sustain plant growth surprisingly early into the growing season (Fig. 5) and involve
 397 large bedrock-water associated latent heat fluxes in the hot summer months (Fig. 7). The
 398 average first day of the year when $S_{bedrock}$ is needed to account for evapotranspiration
 399 (in other words when $S_R > S_{soil}$) is 190 (July 9) and over 21% of the study area must
 400 use bedrock water to account for ET prior to the beginning of summer (June 21) (Fig.

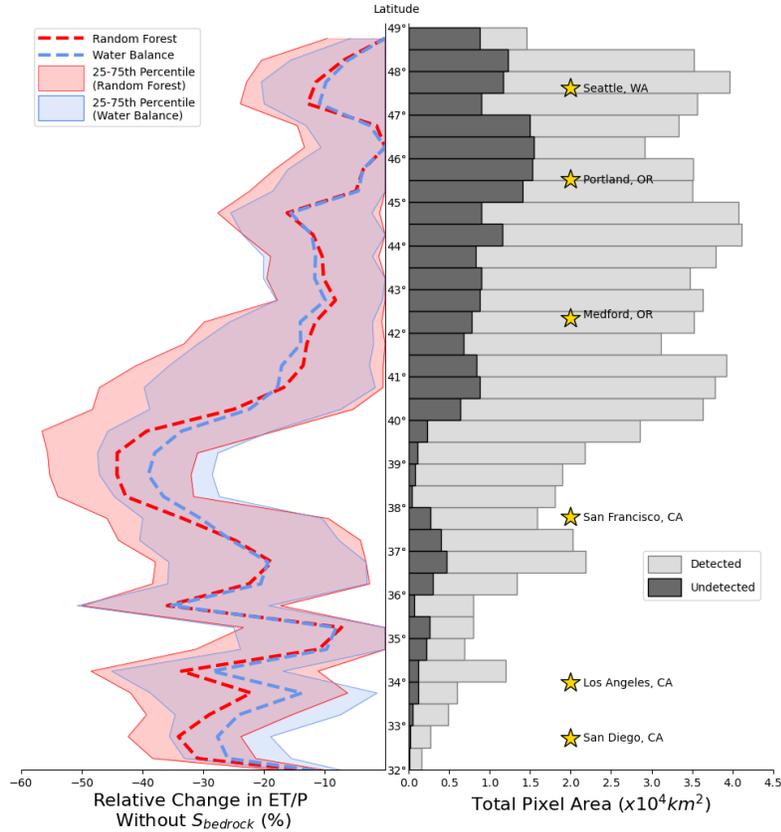


Figure 6. On the left, binned (0.5°) relative change in evaporative index (ET/P) without access to $S_{bedrock}$ using the water balance method (blue) and random forest model (red). On the right, the total area of pixels within each latitudinal band (gray) and the total area of pixels where ET sourced from bedrock water ($ET_{bedrock}$) was not detected using the water balance method (black). Stars show to the latitudinal locations of relevant major cities in the western United States. Across all latitudes, the random forest model predictions align closely with the results of the annual water balance model.

401 S7). In June, the majority of the study area in California has a noticeable latent heat
 402 flux associated with ET sourced from bedrock. By August, there is widespread latent
 403 heat flux across the western US, with July and August having the highest average val-
 404 ues.

405 **4 Discussion**

406 The findings presented in this study highlight the importance of $S_{bedrock}$ on wa-
 407 ter and energy partitioning in the western US. Below we discuss the possible implica-
 408 tions of these findings on land-atmosphere interactions. We begin by situating our study
 409 within the context of the Budyko framework and discuss how this influences hydrologic
 410 partitioning. We then discuss the role of factors like geology on controlling the amount
 411 of $S_{bedrock}$ and, consequently, hydrologic and energy partitioning. Finally, we address
 412 limitations to our study and offer potential future opportunities to advance the topic.

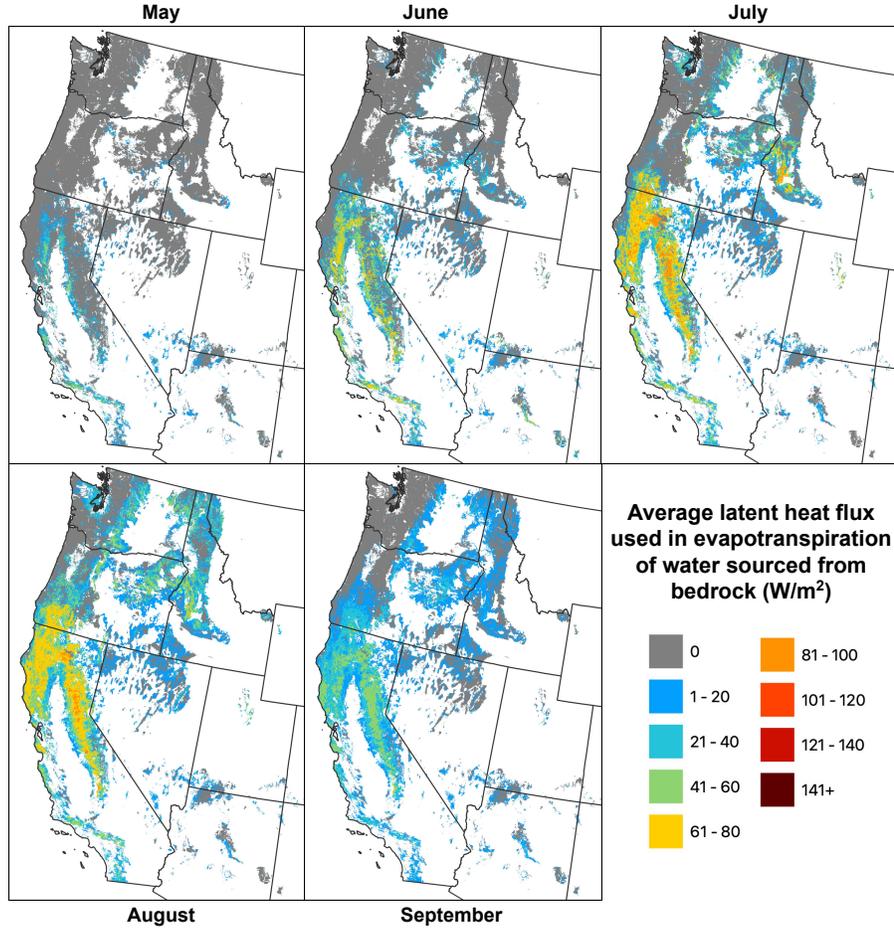


Figure 7. Average latent heat flux (equivalent units to solar irradiance, W/m^2) used in evapotranspiration that is sourced from $S_{bedrock}$ during the growing season. In large parts of the western US, in particular northern California and the Sierra Nevada mountain ranges, a large latent heat flux is associated with evapotranspiration of bedrock water every summer.

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4.1 $S_{bedrock}$ Controls on Water and Energy Partitioning

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The catchment water balance in asynchronous climates often deviates substantially from expectations set by the Budyko curve (Berghuijs et al., 2020; De Lavenne & Andréassian, 2018; Potter et al., 2005; Viola et al., 2017). We found that in the western US, asynchronicity and root-zone storage capacity are two of the strongest predictors for mean annual evapotranspiration (Figs S8, S12, S16), consistent with previous studies focusing on soil water storage that showed that ET is favored with increasing soil water storage capacity (Feng et al., 2012; Milly, 1994a, 1994b; Padrón et al., 2017; Porporato et al., 2004), as well as studies highlighting the importance of seasonality (Feng et al., 2012; Gerrits et al., 2009; Hickel & Zhang, 2006; Xing et al., 2018; Yokoo et al., 2008) and water storage capacity (Chen et al., 2013; Cheng et al., 2022; E. Daly et al., 2019; R. J. Donohue et al., 2012; Gentine et al., 2012; Hickel & Zhang, 2006; Milly, 1994a, 1994b; Potter et al., 2005; Rodriguez-Iturbe et al., 1999; Williams et al., 2012; Woods, 2003). We take this analysis a step further, by differentiating soil from bedrock, to elucidate basic features of how root-zone water is divided between hydrogeologically distinct subsurface layers. Our results suggest S_{soil} alone poorly explains deviations from the Budyko-curve

(Fig. 3), and indicate that $S_{bedrock}$ plays a comparatively larger role on controlling hydrologic partitioning in the western US. This is confirmed by our findings in Figs. 4, S4-S6, S9-11, and is in agreement with similar findings by McCormick et al. (2021), and hillslope-scale observational studies (Dralle et al., 2018; Hahm et al., 2019b, 2022; Lapidés et al., 2022a; Rempe & Dietrich, 2018). Moreover, if $S_{bedrock}$ influences near-surface climate properties in a similar manner to soil moisture (e.g. Brabson et al., 2005; Koster et al., 2004; Haarsma et al., 2009), current GCMs may under-estimate the influence of subsurface storage on extreme temperatures and heat waves (e.g. Seneviratne et al., 2006; Diffenbaugh et al., 2007), precipitation formation (e.g. Alfieri et al., 2008; Ek & Holtslag, 2004; Taylor, 2015), and changes in planetary boundary layer (PBL) circulation patterns (e.g. Sousa et al., 2020; Ookouchi et al., 1984).

4.2 $S_{bedrock}$ Influences on Runoff Generation

Some forms of runoff generation require unsaturated storage deficits to be replenished prior to significant runoff production (McDonnell et al., 2021; Sayama et al., 2011). Recently, Lapidés et al. (2022b) showed that the 'missing' snowmelt runoff during the 2021 spring melt period in California (California Department of Water Resources, 2021) could be attributed to deep root-zone storage deficits caused by drought conditions. These areas, and many other parts of the western US, have among the largest observed S_R in the contiguous US and the fraction of S_R attributed to bedrock is substantial (McCormick et al., 2021). Our results agree with these findings and highlight that $S_{bedrock}$ has major implications for runoff generation in the mountainous West. Deficit-based approaches represent a potential method for scaling up hillslope (e.g. S. P. Anderson et al., 1997; Salve et al., 2012; Tromp-van Meerveld et al., 2007), catchment (e.g. Ajami et al., 2011), and watershed-scale (e.g. Sayama et al., 2011) studies to explain and predict runoff production—the "Holy Grail" of hydrology (Beven, 2006)—at large scales. While our findings suggest bedrock storage heavily influences runoff patterns, especially in southwest (Fig. S11), there is a need for more studies investigating these dynamics and, in particular, field-scale studies to confirm the trends presented here.

4.3 Geological Influences on $S_{bedrock}$ as a Controlling Factor in Vegetation Structure

Evidence supporting the notion that forest ecosystems rely on moisture stored in weathered bedrock to sustain dry season growth goes back several decades (e.g. Arkley, 1981; Jones & Graham, 1993; Rose et al., 2003; Witty et al., 2003). In many cases, bedrock water constitutes a majority of the total subsurface water available to sustain transpiration (e.g. M. Anderson et al., 1995; Hubbert, Graham, & Anderson, 2001; Rose et al., 2003; McCormick et al., 2021). Here, we demonstrate that bedrock storage dynamics influence water and energy partitioning at large scales and throughout many parts of the western US. The extent of bedrock weathering impacts its pore size distribution with depth, and therefore plant-available water storage properties (Klos et al., 2018; Dawson et al., 2020). These properties in turn depend on climate, tectonics, and geology. The mechanisms responsible for the transformation of fresh to weathered bedrock, which in turn increases subsurface moisture storage potential, are well established (see for overview, e.g. S. L. Brantley, 2010; Graham et al., 2010) but remain difficult to investigate due to limitations in accessing deep bedrock samples (see for overview, e.g. Zanner & Graham, 2005). Recently, the Critical Zone (CZ) sciences community has proposed methods for predicting weathered bedrock patterns (Riebe et al., 2017) based on advancements in geophysics (e.g. Slim et al., 2015; St. Clair et al., 2015), geochemistry (e.g. S. Brantley et al., 2013; Lebedeva et al., 2007; Lebedeva & Brantley, 2013), and geomorphology (e.g. R. S. Anderson et al., 2013; Rempe & Dietrich, 2014). A reliable and testable method for predicting weathered bedrock patterns would serve as an important stepping stone in understanding the complex interactions between subsurface properties and aboveground

480 processes. For example, root-zone storage capacities and plant community composition
 481 have been shown to differ drastically in two adjacent, climatically similar watersheds in
 482 California due to contrasting geological substrates (Hahm et al., 2019b). More recently,
 483 Hahm et al. (2023) highlighted areas where geologic substrates overlapped with lower
 484 than 'climatically expected' S_R and argued that plant growth in these areas is inhibited
 485 directly by porosity and/or permeability (e.g. Hahm et al., 2019b; Jiang et al., 2020; H. Liu
 486 et al., 2021) or indirectly via nutrient limitation (e.g. Hahm et al., 2014) and toxicity
 487 (e.g. Kruckeberg, 1992). However, extending these findings to include the influence of
 488 bedrock structure and geology on hydrologic partitioning has not been investigated. The
 489 present study underscores the necessity to further investigate bedrock weathering mech-
 490 anisms as we move towards a holistic approach in CZ sciences.

491 5 Limitations

492 Limiting our study to distributed, remotely sensed, or spatially interpolated datasets
 493 may introduce substantial uncertainty in the results. Although the prevalence of system-
 494 atic errors (e.g. cloud filtering, sensors, etc.) is a known limitation to using remotely sensed
 495 data, we found that precipitation (PRISM) in excess of evapotranspiration (PML) aligned
 496 well with USGS streamflow data in 128 minimally impacted catchments in our study area
 497 (Fig. S2 and Rempe et al. (2022)). There are limited field data to validate our inferences;
 498 however, McCormick et al. (2021) synthesised existing datasets and found the observa-
 499 tions that were consistent with deficit-based methods. The accuracy of satellite-based
 500 data has improved dramatically in recent decades (Dubovik et al., 2021) and, when cou-
 501 pled with finer-scale field studies (i.e. watershed to hillslope), allows for macro-scale as-
 502 simation of topics that underpin important hydrologic problems. While we are confi-
 503 dent in the data presented here we emphasize the need to further implement field-based
 504 studies.

505 To calculate the annual water balance, we first explored the scenario in which the
 506 subsurface storage deficit returned to zero annually. This is not always the case. There
 507 is ample evidence suggesting that many western forests have prolonged, multi-year deficits
 508 (e.g. Cui et al., 2022; Goulden & Bales, 2019; Hahm et al., 2022; P.-W. Liu et al., 2022).
 509 During our analysis we calculated the number of instances per pixel where the subsur-
 510 face storage deficit did not return to zero in a given year and concluded that, in many
 511 cases, the deficit either resets intermittently or very infrequently. When isolating for ar-
 512 eas where the deficit has been shown to not reset, our findings suggest that $S_{bedrock}$ plays
 513 an even bigger role in hydrologic and energy partitioning than previously suggested by
 514 our annual water balance and corroborative random forest model. Despite being limited
 515 by some of the lowest soil water storage capacities in the contiguous US (see McCormick
 516 et al. (2021) Extended Data Fig. 2b), these areas boast many of the largest maximum
 517 root-zone storage (S_{max}) values computed between 2003 - 2017 and, consequently, the
 518 largest $S_{bedrock}$. The importance of $S_{bedrock}$ to dry season plant transpiration in asyn-
 519 chronous climates is not a new idea (e.g. McCormick et al., 2021; Milly, 1994a); how-
 520 ever, research underpinned by these ideas rarely accounts for the possibility of multi-year
 521 deficits. We posit that $S_{bedrock}$ is likely underestimated in areas with non-resetting deficits
 522 and that, in regions that are currently transitioning towards Mediterranean climates as
 523 a result of warming trends (e.g. British Columbia), the magnitude of available $S_{bedrock}$
 524 may be a limiting factor of future plant growth. In the results section, we reported the
 525 typical day (and month) of year when evapotranspiration begins using $S_{bedrock}$ based on
 526 the proposed water balance model and argued that $S_{bedrock}$ is necessary to sustain growth
 527 early into the dry season for many parts of the western US. We did not recalculate this
 528 value using a multi-year deficit for regions where the deficit does not return to zero an-
 529 nually. However, assuming wet season precipitation fully percolates into $S_{bedrock}$ prior
 530 to the dry season, we expect many areas are permanently using $S_{bedrock}$ to sustain sum-
 531 mer growth.

6 Conclusion

In this study, we introduce a simple and reproducible annual water balance framework for assessing the role of $S_{bedrock}$ on water partitioning within the context of the Budyko framework. We employ this framework to investigate the timing of evapotranspiration inferred to be sourced from $S_{bedrock}$ and the magnitude of summer latent heat flux produced as a result of access to $S_{bedrock}$. Finally, we use a random forest regression algorithm to corroborate our findings and then re-purpose the random forest model to explore further areas where the root-zone storage deficit does not reset annually. Our findings suggest that, in the western contiguous US: 1) $S_{bedrock}$ is necessary to explain plant transpiration very early into the growing season; 2) the proportion of precipitation returning to atmosphere would drastically decrease without access to $S_{bedrock}$; 3) the amount of latent heat flux produced as a result of evapotranspiration sourced from bedrock is substantial during the summer; and 4) in regions where the root-zone storage deficit frequently does not reset, the magnitude of evapotranspiration sourced from $S_{bedrock}$ is greater, thereby further influencing the water and energy partitioning properties. These results confirm that $S_{bedrock}$ plays a key role in the local hydrologic cycle and potentially influences the severity and frequency of wildfire and mass die-off events. Further research contributing to the role of $S_{bedrock}$ on the land surface energy balance — e.g. extreme temperatures, heat waves, wind patterns, etc. — would prove beneficial in understanding the factors governing tree death and wildfire, an issue that is prevalent across the western US.

7 Open Research

Flux data (ET, P, PET, and Q) were obtained from Penman-Monteith-Leuning Evapotranspiration (L. Zhang et al., 2001), Parameter-elevation Regressions on Independent Slopes Model (<https://prism.oregonstate.edu>), TerraClimate (<https://www.climatologylab.org/terraclimate.html>), and Catchment Attributes and Meteorology for Large-sample Studies (<https://ral.ucar.edu/solutions/products/camels>), respectively. Land cover, soil water storage, and snow cover were obtained from USGS National Land Cover Database (<https://www.mrlc.gov/data>), Gridded National Soil Survey Geographic Database (<https://www.nrcs.usda.gov/resources/data-and-reports/gridded-national-soil-survey-geographic-database-gnatsgo#download>), and the National Snow and Ice Data Center (<https://nsidc.org/data/mod10a1/versions/61>). All data products were analyzed using the Google Earth Engine Python API (Gorelick et al., 2017). Data, figures, and code associated with this manuscript are available publicly at the following repository on Hydroshare: (Ehlert et al., 2023; Bedrock controls on water and energy partitioning across the western contiguous United States, HydroShare, <https://doi.org/10.4211/hs.191353753cc44de891ee392b95aae22b>).

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