Two decades of permafrost region CO2, CH4, and N2O budgets suggest a small net greenhouse gas source to the atmosphere

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Abstract

The long-term net sink of carbon (C), nitrogen (N) and greenhouse gases (GHGs) in the northern permafrost region is projected to weaken or shift under climate change. But large uncertainties remain, even on present-day GHG budgets. We compare bottom-up (data-driven upscaling, process-based models) and top-down budgets (atmospheric inversion models) of the main GHGs (CO2, CH4, and N2O) and lateral fluxes of C and N across the region over 2000-2020. Bottom-up approaches estimate higher land to atmosphere fluxes for all GHGs compared to top-down atmospheric inversions. Both bottom-up and top-down approaches respectively show a net sink of CO2 in natural ecosystems (-31 (-667, 559) and -587 (-862, -312), respectively) but sources of CH4 (38 (23, 53) and 15 (11, 18) Tg CH4-C yr-1) and N2O (0.6 (0.03, 1.2) and 0.09 (-0.19, 0.37) Tg N2O-N yr-1) in natural ecosystems. Assuming equal weight to bottom-up and top-down budgets and including anthropogenic emissions, the combined GHG budget is a source of 147 (-492, 759) Tg CO2-Ceq yr-1 (GWP100). A net CO2 sink in boreal forests and wetlands is offset by CO2 emissions from inland waters and CH4 emissions from wetlands and inland waters, with a smaller additional warming from N2O emissions. Priorities for future research include representation of inland waters in process-based models and compilation of process-model ensembles for CH4 and N2O. Discrepancies between bottom-up and top-down methods call for analyses of how prior flux ensembles impact inversion budgets, more in-situ flux observations and improved resolution in upscaling.

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76 Abstract

The long-term net sink of carbon (C), nitrogen (N) and greenhouse gases (GHGs) in the northern permafrost region is projected to weaken or shift under climate change. But large uncertainties remain, even on present-day GHG budgets. We compare bottom-up (data-driven upscaling, process-based models) and top-down budgets (atmospheric inversion models) of the main GHGs (CO₂, CH₄, and N₂O) and lateral fluxes of C and N across the region over 2000-2020. Bottom-up approaches estimate higher land to atmosphere fluxes for all GHGs compared to top-down atmospheric inversions. Both bottom-up and top-down approaches respectively show a net sink of CO₂ in natural ecosystems (-31 (-667, 559) and -587 (-862, -312), respectively) but sources of CH4 (38 (23, 53) and 15 (11, 18) Tg CH4-C yr-1) and N2O (0.6 (0.03, 1.2) and 0.09 (-0.19, 0.37) Tg N2O-N yr-1) in natural ecosystems. Assuming equal weight to bottom-up and top-down budgets and including anthropogenic emissions, the combined GHG budget is a source of 147 (-492, 759) Tg CO₂-Ceq yr-1 (GWP100). A net CO₂ sink in boreal forests and wetlands is offset by CO₂ emissions from inland waters and CH₄ emissions from wetlands and inland waters, with a smaller additional warming from N₂O emissions. Priorities for future research include representation of inland waters in process-based models and compilation of process-model ensembles for CH4 and N2O. Discrepancies between bottom-up and top-down methods call for analyses of how prior flux ensembles impact inversion budgets, more in-situ flux observations and improved resolution in upscaling.

108 Introduction

109 The northern permafrost region covers 22% of the northern hemisphere land mass, is dominated by 110 taiga and tundra ecosystems, and is an important component in the global cycles of carbon (C) and 111 nitrogen (N) (Obu et al., 2019; Schuur et al., 2022). The permafrost region warms at rates 2-4 times 112 faster than the global average and climate-driven changes in the extent and temperature of permafrost 113 have been observed (Biskaborn et al., 2019; Rantanen et al., 2022). As soils become warmer or thaw, 114 increased microbial decomposition of soil organic matter (SOM) is projected to cause net losses of C 115 and N to the atmosphere or to aquatic ecosystems (Schuur et al., 2022). Climate change and 116 permafrost thaw also affect and interact with other ecosystem properties, including vegetation 117 dynamics, different disturbance regimes, and the distribution and flow of water through the landscape 118 (Treharne et al., 2022). All of these factors affect the seasonal and annual budgets of the important 119 greenhouse gases (GHGs) carbon dioxide (CO₂), methane (CH₄), and nitrous oxide (N₂O) as well as 120 downstream lateral fluxes of C and N.

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122 Permafrost is ground that is at or below 0°C for at least two consecutive years and underlies circa 14 123 million km² of land in the Northern Hemisphere (Obu, 2021). The northern permafrost region by 124 definition also includes areas with spatially discontinuous permafrost coverage, and covers circa 22 125 million km² of land when permafrost-free areas within the region are included (Obu, 2021). Warming 126 of the active layer and permafrost, gradual thaw and abrupt thaw are occurring across the permafrost 127 domain (Nitze et al., 2018; Runge et al., 2022; Smith et al., 2022) and may increase decomposition of 128 SOM, rich in both C and N, which has accumulated over millenia under cold and wet conditions. 129 Permafrost region soils are estimated to store 1000±200 Pg of organic C and 60±20 Pg of N in the 130 upper three metres (Hugelius et al., 2014; Mishra et al., 2021; Palmtag et al., 2022). Of the total C 131 storage, about 330±80 Pg C is stored in peatlands (Hugelius et al., 2014, 2020), and the rest in mineral 132 soil, often enriched in C by repeated deposition or frost heave processes (Tarnocai et al., 2009). 133 Deeper unconsolidated sedimentary deposits store an additional 400-1000 Pg C, making the 134 permafrost region the largest terrestrial C and N pool on Earth (Strauss et al., 2021).

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As soils thaw or become warmer, enhanced microbial processing of soil C and N causes release of GHGs (CO₂, CH₄ and N₂O) into the atmosphere which cause further warming and forming a positive feedback loop known as the "permafrost carbon feedback" (Schuur et al., 2008, 2022). Hereafter we refer to it as the "permafrost GHG feedback", to include non-carbon feedbacks, such as from N₂O. The sixth assessment report of the Intergovernmental Panel on Climate Change (IPCC) (Canadell et al. 2021) estimated that the permafrost GHG feedback from CO₂ per degree of global warming at the end of the century is 18 (3.1 to 41, 5–95% range) PgC °C⁻¹, with an additional permafrost GHG

feedback from CH₄ of 2.8 (0.7 to 7.3) PgCeq $^{\circ}C^{-1}$. However, this IPCC estimate does not fully include 143 144 abrupt permafrost thaw processes which cause strong additional release of CO₂ and CH₄ at decadal to 145 centennial time scales, especially from release of CH₄ from water-logged post-thaw environments 146 (Turetsky et al., 2020). Abrupt thaw, including thaw-lake formation, collapse of permafrost peatlands, 147 and thaw-slump formation, can rapidly affect permafrost at depths of several metres, causing rapid 148 melting of ground ice, land subsidence and a complete restructuring of the landscape. In addition to 149 uncertainties in how climate warming drives increased respiration, there is large uncertainty 150 regarding mediating effects from increased vegetation productivity (and CO₂ uptake) caused by longer 151 growing seasons, increased CO₂ concentrations, and additional nutrient release from thawing 152 permafrost (Abbott et al., 2016; Liu et al., 2022; McGuire et al., 2018). While uncertainties remain 153 large, many studies based on observational GHG flux time series show enhanced net GHG emissions 154 from warming and thawing permafrost soils (Kuhn et al., 2021; Marushchak et al., 2021; Natali et al., 155 2015; Rodenhizer et al., 2022; Voigt et al., 2017, 2019).

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157 A potential shift from a net GHG sink to a source remains uncertain in projections using Earth System 158 models (ESMs). A recent study using the CMIP6 ensemble of ESMs projects a sustained northern CO₂ sink from 2015-2100 across a broad range of human emissions scenarios (Qiu et al., 2023). 159 160 However, the majority of the CMIP6 models do not include an explicit representation of permafrost, 161 or GHG feedbacks from thaw, and are thus likely unable to properly project changes in permafrost 162 GHG balance under future warming. A previous intercomparison of process-models with 163 representation of gradual permafrost thaw found that the northern permafrost region would act as a 164 sustained net C sink under medium emission scenarios (RCP4.5), but would likely act as a C source 165 under higher emissions scenarios, at least over the long term (McGuire et al., 2018). However, there 166 was significant spread between different models, largely reflecting limited representation of processes 167 affecting vegetation productivity, soil respiration and permafrost dynamics. There is also mounting 168 evidence that other localized disturbance processes, still lacking in global models, play an important 169 role in the mobilization of permafrost C and N. This includes disturbances associated with abrupt 170 thaw, coastal erosion, fires, pests or windfalls (Foster et al., 2022; Holloway et al., 2020; Hugelius et 171 al., 2020; Marushchak et al., 2021; Walker et al., 2019). While progress is continuously being made, 172 model simulations of the interactions between permafrost and disturbances are in their infancy 173 (Treharne et al., 2022) and are often only relevant for specific field sites (Aas et al., 2019; Brown et 174 al., 2015; López-Blanco et al., 2022). Similarly, the lateral export of C in the form of dissolved 175 organic C (DOC) is missing in most ESMs. Moreover, dedicated simulations of permafrost region 176 N₂O fluxes are still scarce and process information is insufficiently available (Lacroix et al., 2022; 177 Voigt et al., 2020).

179 Improved understanding of GHG exchange in the permafrost region is therefore crucial for 180 constraining global GHG budget estimates and reducing discrepancies between methods 181 (Friedlingstein et al., 2022; Saunois et al., 2020; Tian et al., 2020). Estimates of GHG budgets are 182 typically done using bottom-up (data-driven ecosystem flux upscaling - hereafter referred to as 183 ecosystem flux upscaling - or process-based models) or top-down (from inversions of atmospheric 184 GHG mole fractions - hereafter referred to as atmospheric inversion models) approaches (Ciais et al., 185 2022). Budgets based on ecosystem flux upscaling combine observations of GHG fluxes with 186 geospatial datasets, while process-based model budgets are based on mathematical representations of 187 ecosystem processes characterizing the functioning of biogeophysical systems. Here, we also compare 188 the other bottom-up approaches to a terrestrial model-data fusion (MDF) approach, where a process-189 model is calibrated at pixel-scale using a Bayesian algorithm and spatially coherent observations 190 interpolated from field data and satellite-based Earth Observation (EO). Atmospheric inversion 191 models use advanced mathematical methods to estimate surface-to-atmosphere net GHG fluxes by 192 combining atmospheric GHG concentration information (in-situ or flask measurements from surface 193 stations, or total column abundances estimated from satellites), gridded prior flux information, and 194 atmospheric transport models. The first comprehensive GHG budget synthesizing bottom-up and top-195 down GHG estimates for the Arctic tundra was published in 2012 as part of the REgional Carbon 196 Cycle Assessment and Processes project (RECCAP, McGuire et al., 2012) and highlighted the high 197 variability between budgeting methods thus calling for more efforts to identify and reduce the sources 198 of discrepancies. Although much progress has been made in the decade that followed, some of these 199 issues remain unresolved and there has been no systematic review of GHG budgets for the full 200 permafrost region (including boreal and tundra biomes) that compares and reconciles different 201 budgeting approaches (bottom-up and top-down).

202 Here we present comprehensive budgets of GHGs (CO₂, CH₄, and N₂O) and lateral fluxes of C and N 203 for the period 2000-2020 across the northern permafrost region. We compare estimated GHG fluxes 204 from the permafrost region using bottom-up and top-down approaches and identify remaining 205 research gaps that must be addressed in order to reconcile the different budget estimates and improve 206 interpretations of GHG budgets. The budgets also include estimated anthropogenic emissions of CO₂ 207 and CH₄. These permafrost regional budgets are part of the REgional Carbon Cycle Assessment and 208 Processes-2 (RECCAP2) project of the Global Carbon Project that aims to collect and integrate 209 regional GHGs budgets covering all global lands and oceans (Ciais et al., 2022) 210 (https://www.globalcarbonproject.org/reccap/).

212 Methods

213 Study area

214 The spatial extent of permafrost defined in this study includes areas within the northern permafrost 215 region as defined in Obu et al. (2021) and restricted to the Boreal Arctic Wetlands and Lakes Dataset 216 area (BAWLD), (Olefeldt et al., 2021) (Fig 1). This restriction was made due to constraints in data 217 availability for ecosystem flux upscaling. As a consequence the targeted BAWLD-RECCAP2 218 permafrost domain does not take into account areas underlain by permafrost in Central Asia and the 219 Tibetan plateau (blue areas outside the black line in figure 1). The BAWLD-RECCAP2 permafrost 220 region considered in this study is 18.42 million km² (excluding ice sheets and glaciers). Because much 221 of the region is underlain by spatially discontinuous permafrost in a mosaic of different land forms 222 and ecosystems, areas and land cover types without permafrost are included to the domain. All flux 223 estimates and models were run or rescaled to 0.5 x 0.5 degree spatial resolution and masked to match 224 the BAWLD-RECCAP2 permafrost region (hereafter permafrost region). We differentiated tundra 225 and boreal forest areas within the permafrost region using a biome delineation (Dinerstein et al., 226 2017). The study area overlaps several other RECCAP2 regions (Ciais et al., 2022) but no specific 227 effort to harmonize the budgets presented here with the RECCAP2 budgets of those regions are made 228 in this paper.



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231 Figure 1: Map of the extent of the study area, defined as were the northern permafrost region (blue shades, data 232 from Obu et al., 2021) overlaps with the spatial extent of the Tundra and Boreal forest biomes (hatched areas) as 233 represented Boreal Arctic Wetlands and Lakes Dataset (BAWLD, Olefeldt et al., 2021). Because the permafrost 234 extent in non-continuous in much of the region, it includes large areas of permafrost-free ecosystems in a 235 mosaic within the broader region. Figure S1 in the supplement shows the additional areas that recorded 236 mean annual air temperature (MAAT) below 0°C between 1990 and 2000 (full extent of ISIMIP3 237 permafrost model intercomparison), but which were excluded from this budget estimate because they 238 are outside the BAWLD extent.

239 Summary of overall budget approach

This paper presents full annual budgets of C and N fluxes in the form of the main GHGs (CO₂, CH₄ and N₂O) and as lateral fluxes in streams and rivers and from coastal erosion for the time period 2000-2020. All budgets are expressed on a C and N mass basis (i.e., for GHGs as CO₂-C, CH₄-C and N₂O-N yr⁻¹). Budgets are reported as Tg C or N and are reported as mean fluxes with 95% confidence intervals (CI). In this paper we aim to present the most complete available budget estimates derived 245 from data-driven ecosystem upscaling (all values from Ramage et al., in prep.), process based models 246 (typically Dynamic Global Vegetation Models, Land Surface Models or ecosystem models) and from 247 atmospheric inversion models. Figure 2 shows a generalized overview of the approach. Consistent 248 with global GHG budgets (Friedlingstein et al., 2022; Saunois et al., 2020; Tian et al., 2020) sinks into 249 the biosphere are reported as negative numbers while sources to the atmosphere are reported as 250 positive numbers. Ecosystem upscaling and process-based model ensembles are considered as 251 bottom-up inventories, while the inverse atmospheric model ensembles are viewed as top-down 252 atmospheric constraints. To estimate the total combined radiative balance of the permafrost region 253 GHG budgets, all budgets are combined to a common unit of CO₂-C equivalents. This is calculated as 254 Global Warming Potential (GWP) for a 100 year time period (GWP-100 from table 7.15 in IPCC, 255 2023).

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Greenhouse gas (GHG) and lateral flux budgets for 2000-2020



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Figure 2: Conceptual figure summarizing the overall approach, including top-down and bottom-up, to compile the RECCAP2 permafrost GHG and lateral flux budgets. Main budget components are presented in blue boxes. The bottom-up process-based models include both ensembles of process-models as well as model-data fusion (MDF) with CARDAMOM. Additional budget components in blue text (white box) include separate inventories of anthropogenic fluxes, lateral fluxes (rivers and coastal erosion) and geological emissions which are used to complete the budgets. Data-driven ecosystem GHG inventories and estimates of lateral fluxes and geological CH4 are taken directly from Ramage et al. (in prep.)

265 CO₂-C budget from process-based models

Estimates of terrestrial ecosystem fluxes of CO_2 were extracted from an 'ensemble of opportunity' consisting of 73 process-based model simulations that have been generated over the past 10 years as 268 part of model intercomparison projects. Supplemental table S1 summarizes all the models included, 269 which model intercomparison project they belonged to, the last year of simulation, and whether they 270 represent processes relevant to fires as well as permafrost carbon. These include one variant of each 271 model taken from the available historical simulations in the following projects: the Coupled Model 272 Intercomparison project phase 5 and 6 (CMIP5 and CMIP6) historical coupled climate simulations 273 (Eyring et al., 2016; Taylor et al., 2012) and the Land Surface, Snow and Soil Moisture Model 274 Intercomparison Project (LS3MIP) land-history simulations driven by observed meteorology (Van 275 Den Hurk et al., 2016). These simulations were performed by various modelling groups and are 276 available from either the CMIP5 archive (https://esgf-node.llnl.gov/search/cmip5) or the CMIP6 277 archive (https://esgf-node.llnl.gov/search/cmip6). We downloaded and extracted carbon stocks and 278 fluxes from both the Permafrost Carbon Network (PCN) and Multi-scale Synthesis and Terrestrial 279 Model Intercomparison Project (MsTMIP) ensembles via the ORNL DAAC (McGuire et al., 2022 280 and (Huntzinger et al., 2018, respectively). Data from the Inter-Sectoral Impact Model 281 Intercomparison Project phase 2a and 2b (ISIMIP2a and ISIMIP2b) were downloaded from 282 https://www.isimip.org. In the case of ISIMIP2a, only the ensemble members driven by Global Soil 283 Wetness Project version 3 (Dirmeyer et al., 2006) data were included. For ISIMIP2b, only the 284 ensemble members driven by bias-corrected climate data from the IPSL-CM5A-LR Earth System 285 model submitted to the CMIP5 archive were considered. The other available ensemble members were 286 very similar, which, if included, would mean ISIMIP2a/2b would contribute the overall majority of 287 the models to the ensemble of opportunity. Supplemental Table 1 shows which model simulations 288 represent permafrost carbon. Initial results from ISIMIP3a/3b simulations are also shown for four 289 models (i.e., JULES, ORCHIDEE-MICT, JSBACH, and ELM-ECA). Results from these model 290 simulations are previously unpublished and described in more detail in the paragraph below.

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292 The ISIMIP3 modelling output was from:

(i) The Joint UK Land Environment Simulator (JULES) which was driven by GSWP3 meteorology
bias corrected by the W5E5 data set (Lange, 2019), denoted GSWP3-W5E5 under the ISIMIP3a
protocol. JULES is the land surface component of UKESM (Sellar et al., 2019). The configuration of
JULES presented here includes the representation of C and N cycling (Wiltshire et al., 2021) but not
vertically resolved soil carbon.

(ii) ORCHIDEE-MICT was also driven by GSWP3-W5E5 following the ISIMIP3a protocol.
ORCHIDEE-MICT (Guimberteau et al., 2018) is a version of ORCHIDEE with permafrost C
representation in a multilayered vertically discretized model, interactions between soil C, soil
temperature and hydrology, and a fire module which burns litter and vegetation. The version used in
ISIMIP3a is further improved with representation of grassland management (Chang et al., 2021) and
northern peatlands (C. Qiu et al., 2020).

(iii) JSBACH-wet and JSBACH-dry were driven by GFDL-ESM4 historical forcing, following the
ISIMIP3b protocol. JSBACH is the land surface component of the Max Planck Institute for
Meteorology Earth System Model MPI-ESM version 1.2 (Mauritsen et al., 2019), with methane cycle
according to (Kleinen et al., 2020). The "wet" and "dry" configurations cover plausible ranges of soil
parameters leading to wetter and dryer soil conditions (De Vrese et al., 2023).

309 (iv) ELM-ECA was driven by climate data from GFDL-ESM4 submitted to the CMIP6 archive. 310 GFDL-ESM4 was bias-corrected by the W5E5 data that was also used in ISIMIP3a. This simulation 311 follows the ISIMIP3b protocol. ELM-ECA is a land model from the Energy Exascale Earth System 312 Model (E3SM) (Zhu et al., 2019). It simulates C, N, phosphorus, water and energy cycles for major 313 terrestrial ecosystems (e.g., forest, shrub, grassland, wetland). ELM-ECA considers multiple nutrient 314 competitions among plants, microbial immobilizer, nitrifier, denitrifier, and mineral surfaces to 315 resolve the resource partitioning among different competitors. The version used in ISIMIP3b has 316 improved the parameterization on wetland inundation and upland plant carbon-nutrient interactions.

317

318 All of the models (supplemental table S1) contain their own individual C cycle processes with a range 319 of complexities. Eighteen of the 73 simulations include a representation of permafrost C and were 320 analysed as a separate sub-ensemble (layered and bulk C, respectively). The multi-annual mean 321 estimate of the carbon stocks and fluxes was defined for the period 1980 to the end of the model 322 simulation (supplemental table S1). The mean residence time of dead organic matter (MRT_{SHR}) is 323 defined as the sum of C stored in SOM and litter content divided by heterotrophic respiration. All of 324 these models estimate the net ecosystem productivity of CO_2 (NEP = NPP - SHR), where NPP is the 325 net primary productivity and SHR is the soil heterotrophic respiration and positive values are a land 326 sink. Some of the process-models used here also consider the additional impact of aboveground C 327 emissions from fire but none estimate belowground C loss from fire. In this case the net biosphere 328 productivity is defined which additionally includes the fire emissions (NPP - SHR - fire). This is of 329 particular note given the majority of fire carbon emissions in the circumpolar domain are from 330 belowground sources: roughly 84-90% in arctic-boreal North America and 57-74% in Eurasia (Potter 331 et al., 2023; Veraverbeke et al., 2021; Walker et al., 2020). Other disturbances, such as pest or storm 332 damage to forests, are not considered here. Fluxes from rivers and lakes are also excluded, as well as 333 those from abrupt permafrost thaw since these processes are not represented by the included models. 334 Carbon use efficiency (CUE), defined as the ratio of net primary productivity (NPP = GPP - plant 335 respiration (R_a)) to GPP, is an emergent property of the models which quantifies vegetation efficiency 336 at storing C fixed via photosynthesis on annual timescales. In addition, full ecosystem budgets for 337 CH₄ and N₂O fluxes from process-based models were either not available or not provided for the 338 model intercomparisons (but see separate section for process-based model budgets of wetland CH₄ 339 flux).

341 Observation-informed estimates of terrestrial C stocks, fluxes and ecosystem traits (e.g. MRT_{SHR}) 342 within a consistent mass-balanced framework are essential to support evaluation of the model 343 ensembles described above. To generate these estimates we use the CARbon DAta MOdel fraMework 344 (CARDAMOM) (Anthony Bloom et al., 2016), which has been previously used to inform our 345 understanding of Arctic C-cycling (López-Blanco et al., 2019). CARDAMOM is a model-data fusion 346 framework that uses a Bayesian approach within an adaptive-proposal Markov chain Monte Carlo 347 (AP-MCMC) (Haario et al., 2001) to train a process-model of intermediate complexity, DALEC (A. 348 A. Bloom & Williams, 2015; Smallman et al., 2021; Smallman & Williams, 2019). CARDAMOM 349 estimates ensembles of parameter sets for each location independently from each other as a function 350 of location specific data-constraints. These location specific parameter ensembles result in DALEC 351 simulations consistent with the assimilated datasets, their associated uncertainties and ecological and 352 dynamical constraints (for details see Famiglietti et al., 2021). From these parameter ensembles 353 CARDAMOM can generate pixel-level estimates of terrestrial C-cycling and their associated 354 uncertainty, allowing a more rigorous evaluation of more complicated process-oriented models which 355 have a less direct connection to data (Caen et al., 2022). Specifically in this analysis, CARDAMOM 356 analysed terrestrial C-cycling at a monthly time step and 0.5 x 0.5 degree spatial resolution for 19 357 years (2001-2019). The meteorological drivers were drawn from the GSWP3-W5E5 dataset, while 358 fire was imposed as a function of MODIS burned area (Giglio et al. 2015) and forest loss was 359 constrained using global forest watch (Hansen et al., 2013). Assimilated information was time series 360 estimates of leaf area index (Copernicus Service Information 2021), woody biomass for 2017 and 361 2018 (Santoro et al., 2021), and net biome exchange of CO_2 (Koren 2020). Moreover, the Northern 362 Circumpolar Soil Carbon Database (NCSCD) provided a pixel specific prior for the initial soil C 363 content (G. Hugelius et al., 2013; G Hugelius et al., 2013). Finally a globally applied prior for the 364 ratio of autotrophic respiration and photosynthesis of 0.46 +/- 0.12 (Collalti & Prentice, 2019). For a 365 more detailed description of CARDAMOM, DALEC and its drivers and observational constraints see 366 the supplemental text and figure S2.

367 Estimated wetland CH₄ emissions from process based models

368 Because of limited data availability, no full ecosystem flux budgets of CH₄ from process-based 369 models are available. But there are global-scale estimates of CH₄ budgets for the wetland component 370 of the terrestrial land surface, produced for the Global Methane Budget (Saunois et al., 2020). This 371 study defined wetlands as peatlands (bogs and fens), mineral soil wetlands (swamps and marshes), 372 and seasonal or permanent floodplains. This excludes exposed water surfaces without emergent 373 macrophytes (such as ponds, lakes and rivers) and coastal vegetated ecosystems. In Saunois et al 374 (2020) 13 land surface models which represent CH_4 exchanges were run for the time period 2000-375 2017, using a common climatic forcing (see table 2 in Saunois et al., 2020). For the permafrost region

376 budget, only model runs with wetland extent constrained by the Wetland Area Dynamics for Methane 377 Modeling dataset (WAD2M) (Zhang et al., 2021) were used (called "diagnostic model runs" in 378 Saunois et al., 2020). The annual modelled CH₄ wetlands budgets were extracted for the permafrost 379 region and summarised per decade. Although the spatial extent of wetlands in BAWLD is not exactly 380 the same as in WAD2M, the two datasets are similar. They are based on partly identical source data 381 and the definition of wetlands applied in WAD2M is consistent with definitions in BAWLD. We 382 therefore consider the estimates to be sufficiently similar that it supports comparison of CH₄ budgets 383 from ecosystem upscaling with estimates from process-based models (i.e. differences between the 384 methods themselves are much larger than differences in wetland area). Supplementary table S2 385 summarizes the process-based models used to estimate CH₄ wetland fluxes.

386 Data driven bottom-up ecosystem GHG budgets and lateral flux budgets

All values reported for data-driven ecosystem flux upscaling of GHG budgets and lateral fluxes presented here are from Ramage et al., (*in prep*). The methods used by Ramage et el., (*in prep*) are briefly described below, but we refer to the original paper for full details.

390 Ramage et al (in prep) calculated C and N budgets (2000-2020) by summing GHG uptake and 391 emissions from terrestrial ecosystems, inland waters, and from disturbances (fire and abrupt thaw), as 392 well as lateral fluxes and geological emissions using several synthesis datasets. The land cover 393 classification used for the analysis was adapted from the BAWLD land cover classification (Olefeldt 394 et al., 2021). The original 19 terrestrial land cover classes in BAWLD were aggregated into five 395 classes: Boreal forest, Non-permafrost wetlands, Dry tundra, Tundra wetlands and Permafrost bogs. 396 The classes Dry tundra, Tundra wetlands, and Permafrost bogs are underlain by surface permafrost 397 and differ largely based on wetness and organic soil depth. Because of spatially discontinuous 398 permafrost coverage, Boreal forests include both permafrost and permafrost-free ecosystems. Mean 399 annual fluxes of CO₂, CH₄, and N₂O were obtained for each of the five terrestrial land cover classes 400 by modifying three comprehensive GHG flux dataset compilations for CO₂ fluxes (A.-M. Virkkala et 401 al., 2022), CH₄ fluxes (Kuhn et al., 2021); and for N₂O fluxes (Voigt et al., 2020) (with addition of 402 N₂O fluxes for Boreal forest).

403 Similarly, inland waters fluxes of CO_2 and CH_4 to the atmosphere were calculated by upscaling mean 404 annual fluxes from lakes and rivers using the estimated surface area of these aquatic classes from the 405 BAWLD classification, adjusted to the permafrost region adjusting for ice-covered duration and 406 fluxes during ice break-up (see Ramage et al., in prep for details). To estimate lake fluxes of N₂O 407 from inland waters, gridded global data of annual flux from (Lauerwald et al., 2019) were used. 408 Estimates of river and stream CO₂ flux were calculated from gridded monthly flux data (Liu et al., 409 2022) (data from https://doi.org/10.5061/dryad.d7wm37pz9), using adjusted surface areas. Riverine 410 CH₄ emissions were determined using the mean CH₄ diffusive flux reported in the MethDB database

- 411 (Stanley et al., 2016). To estimate river fluxes of N_2O , gridded global data of annual full landscape
- 412 flux were used (Maavara et al., 2019).
- 413 Monthly fire emissions of CO₂ and CH₄ were extracted for the study region from the Global Fire

414 Emission Database version 4s (GFED; van der Werf et al., 2017). The GFED spans from 1997-2016

415 and is driven by estimates of burned areas derived from satellite-based remote sensing data at a spatial

- 416 resolution of 0.25 degrees (Van Der Werf et al., 2017).
- 417 Fluxes of CO₂ and CH₄ from landforms caused by abrupt thaw (thermokarst) were extracted from an
- 418 inventory-based abrupt thaw model (Turetsky et al., 2020), in which emissions are estimated for three
- 419 generalized types of abrupt thaw terrains: mineral-rich lowlands, uplands/hillslopes, and organic-rich
- 420 wetlands. The abrupt thaw model was initialized for a historical assessment period (1900-2000) and
- 421 was then run for the period 2000-2020 to assess CO₂ and CH₄ emissions from active and stabilized
- 422 abrupt thaw features. To prevent double counting, fluxes from mineral-rich lowlands and organic-rich
- 423 wetlands were counted as a sub-flux (not added to the total) of terrestrial land cover fluxes.

Lateral C and N fluxes from riverine transport and coastal erosion (i.e., DOC and DON losses from the permafrost region to the ocean) are taken from (Terhaar et al., 2021), representative for all land north of 60° N. Emissions from coastal erosion were calculated by multiplying spatially resolved estimates of coastal erosion rates by estimates of C content in coastal soils (Lantuit et al., 2012).

- 428 Estimates of geological emissions of CH₄ (from subsurface fossil hydrocarbon reservoirs) are taken
- 429 from an upscaled circumpolar permafrost region estimate for gas seeps along permafrost boundaries
- 430 and lake beds (Walter Anthony et al., 2012). No separate estimates of geological emission for CO_2 or
- 431 N_2O are available for the permafrost region.
- 432

433 **Combined best-estimate for bottom-up budgets**

434 To reconcile the differences between the varying bottom-up approaches integrated bottom-up GHG 435 budgets were created by combining results of the ensembles of process-based models and ecosystem 436 upscaling. The integrated bottom-up estimate is calculated as the mean of the process-based models 437 and ecosystem upscaling for upland and wetland ecosystems, respectively, but adding some 438 components which are lacking in the process-based models. The calculator is based on the subset of 439 the process-based model ensembles which include fire flux. Because inland waters are not included in 440 process-based models the numbers from ecosystem upscaling are added to the total. Because models 441 do not account for abrupt thaw wetlands, estimated fluxes from such processes are added to the 442 budgets from process-model wetland flux (with corrections of model fluxes proportional to the 443 respective areas occupied by inland waters and abrupt thaw wetlands).

445 GHG budgets from atmospheric inversions models

supplemental table S3.

446 Independent decadal budgets for CO₂, CH₄ and N₂O are derived from ensembles of inverse flux 447 estimates. Budgets for 2000-2010 are reported for all GHGs. Due to differing data availability during 448 the second decade, the GHG budgets are reported for different time periods; 2010-2020 for CO₂, 449 2010-2017 for CH_4 , and 2010-2019 for N_2O . Inverse systems for estimating GHGs vary significantly 450 in the time scale of the analysis, the spatiotemporal resolution of the inferred fluxes, or the inverse 451 modeling framework used (Gaubert et al., 2019; Peylin et al., 2013), all of which result in differences 452 in the inferred flux estimates. To get the best estimates of GHG budget from atmospheric inversion 453 models, a common method is therefore to derive a mean or median flux estimate and the spread 454 among the estimates from the different atmospheric inversion models (Ciais et al., 2022; Philip et al., 455 2022). The inverse model systems used to derive annual GHG budgets in this study are summarized in 456

The analyses followed the RECCAP2 protocol to estimate CO2, CH4 and N2O fluxes (Ciais et al., 457 458 2022). The specific methodology used to calculate the budgets for CO_2 , CH_4 and N_2O are described in 459 more detail in (Friedlingstein et al., 2022) for CO₂, (Saunois et al., 2020) for CH₄, and (Tian et al., 460 2020) for N2O. These data were retrieved from the GCP/MPI-BGC/RECCAP-2 data portal 461 (https://www.bgc-jena.mpg.de/geodb/) where the total number of inverse modelling estimates 462 available was 6 for CO_2 , 22 for CH_4 and 3 for N_2O . For CO_2 , these 6 estimates have undergone a 463 spatial adjustment for differences in the used fossil fuel, cement emissions and cement carbonation 464 sink. One additional CO_2 flux estimate was added (Chandra et al., 2022) to make a total of 7, with a 465 strong overlap with the older versions of these systems used in (Z. Liu et al., 2022). Inversion 466 estimates for CO₂ and CH₄ were derived from either in-situ (surface) observations of atmospheric 467 GHG mole fractions or satellite derived total column estimates. Note that satellite estimates of CO₂ 468 and CH₄ are available primarily from 2009 onwards (e.g., from JAXA's GOSAT and NASA's OCO-2 469 missions). However, among the 22 available estimates for CH₄, there were multiple submissions from 470 the same group using different configurations of the atmospheric data or errors associated with the 471 data. These submissions were first averaged and then the average estimate was used alongside 472 estimates from the other groups. This resulted in 14 final inverse model estimates that were used to 473 calculate the mean CH₄ budget. There is considerable variability between the analysis systems and a 474 comprehensive assessment of accuracy of individual estimates is lacking for CH_4 and N_2O ; but 475 (Friedlingstein et al., 2022) provide an assessment of the skill of inversions against independent 476 aircraft observations for CO_2 (for 6 out of the 7 used here). We calculate the budget using the mean 477 annual value from these ensembles of estimates. For estimates of N₂O, the estimates from one model 478 system deviated more than an order of magnitude from the other two systems, and these data were not 479 used further. With these changes, the included number of inverse modelling estimates available were 480 7, 14 and 2, for CO₂, CH₄ and N₂O, respectively (Supplemental table S2). There was a large variation

- in the spatial resolution of inverse model systems. In all cases, the permafrost mask, which was available at $\sim 1^{\circ}$, was regridded to the resolution of the estimates from the individual models, then the flux estimates were sampled using the permafrost mask and finally averaged using area-weighting to generate a single value at monthly time steps for the permafrost study domain. This procedure follows the protocol outlined in previous RECCAP2 studies (Ciais et al., 2021). Supplemental tables S5, S6 and S7 (for CO₂, CH₄ and N₂O, respectively) contain mean monthly GHG fluxes across the full study domain, for each specific inverse model used to calculate decadal means.
- 488 Top-down estimated fluxes of CO_2 from fires (not added to the total, reported as sub-flux) are 489 extracted from a separate study combining satellite retrievals and atmospheric inversion of carbon 490 monoxide (CO) converted to CO₂-C emissions using fixed emission factors (Zheng et al., 2023). 491 of this Gridded estimates fire CO_2 -C fluxes from study were retrieved 492 (https://doi.org/10.6084/m9.figshare.21770624, resolution of 3.75°×1.9°), clipped to the extent of our 493 study domain and summarized for the relevant time periods. To report separate sub-fluxes of CH₄ 494 from fires, data was extracted from the Global Methane Budget inventory using the Biomass and 495 biofuel burning component of the top-down ensemble of global inversion estimates (Saunois et al., 496 2020), available for 2000-2017. These are inventory-based, but methodologically consistent with the 497 CH₄ inversions.

498 Anthropogenic emissions

499 Estimates of anthropogenic CO₂ emissions of fossil fuel combustion (coal, gas and oil estimated 500 separately), cement production and cement carbonization were extracted from the Global Carbon 501 Project's gridded dataset for fossil CO₂ emissions and related O₂ combustion (GCP-GridFED) (M. W. 502 Jones et al., 2021), updated for the 2022 edition of the Global Carbon Budget (GCP-503 GridFEDv2022.2; Jones et al., 2022). In GCP-GridFED, the emissions of fossil CO₂ relate to the 504 combustion and use of fossil fuels and the production of cement clinkers. These estimates are 505 consistent at the national and annual level with the emissions inventory compiled by the Global 506 Carbon Project (Friedlingstein et al., 2022; Andrew and Peters, 2022). Emissions are gridded at 1 km 507 resolution based on the Emissions Database for Global Atmospheric Research (EDGAR) dataset, 508 version 4.3.2 (Janssens-Maenhout et al., 2019), and distributed across the months of each year using a 509 relationship with heating and cooling degrees (years 1959-2019) and Carbon Monitor to reflect the 510 impact of COVID-19 (year 2020), as described by Jones et al. (2021, 2022).

511 Estimates of anthropogenic CH₄ emissions were estimated using the *Fossil fuel* plus *Agriculture and*

512 *waste* components of the top-down ensemble of global inversion estimates from Saunois et al. (2020)

- 513 (available for 2000-2017). We refer to Saunois et al., (2020) for more details on how these datasets
- 514 were derived. No separate estimates for anthropogenic N₂O fluxes are included in the study.

515 Results and discussion

516 Bottom-up budgets from models and upscaling

517 Bottom-up ecosystem CO₂ budgets from process-based models

The process-models are divided into two ensembles based on whether they include (*layered C*) or exclude (*bulk C*) an explicit representation of permafrost C (Figure 3). From the 73 available models 55 include permafrost C. The model ensembles are also compared with the observationally-informed CARDAMOM analysis to provide independent quantification of the likelihood of the process-model estimates. Process-model ensemble outputs can be further separated into tundra or boreal, as defined in the BAWLD database (Figures S3 and S4).

524

525 The land surface area in the permafrost region is estimated to be a net sink of CO_2 by both model 526 ensembles (including (layered C) or excluding (bulk C) representation of permafrost C) and the 527 CARDAMOM analysis (Figure 3). When considering net ecosystem productivity (NPP - SHR), the full model ensemble suggests a net sink of -420 (-80 to -1020) Tg CO₂-C y⁻¹ with CARDAMOM 528 showing a net uptake of -960 (sink of -1880 to source of 20) Tg CO₂-C y⁻¹ (table 1). Out of the 73 529 530 process-based model runs, 40 runs did not include fire and as a result are excluded from the net biome 531 productivity (NPP - SHR - F_{fire}) reported (supplemental table S1). The 33 models that include fire, 532 albeit not including combustion of belowground sources, estimate fire emissions of 120 (10 to 460) Tg CO₂-C y⁻¹. This estimate is larger and has a much greater spread than CARDAMOM which has 533 values of 60 (50 to 100) Tg CO₂-C y⁻¹. The inclusion of fire emissions reduces the net land uptake 534 535 (only the models which include fire emissions) to -340 (-90 to -930) Tg CO₂-C y⁻¹, while CARDAMOM estimates a reduced net land uptake of -870 (sink of -1780 to source of 160) Tg CO₂-C 536 v⁻¹ (Figure 3). Although not reflected in the model ensembles, CARDAMOM's ensemble shows the 537 538 source/sink boundary falls between the 75th and 90th quantile and thus the permafrost region could be 539 a small net source of C.



542

543 Figure 3. Carbon fluxes, stocks and relevant ecosystem properties from the process-based models listed in supplemental 544 table S1 over the BAWLD region. The top row shows the following simulated multi-annual mean C fluxes (left to right) -545 heterotrophic respiration (SHR); gross primary productivity (GPP); net primary productivity (NPP); autotrophic respiration 546 (Ra) all in Pg CO2-C y⁻¹. Also shown on the top row are the carbon use efficiency (CUE, dimensionless) and the fire C flux 547 (F_{fire} in Pg CO₂-C y⁻¹). The bottom row shows C stocks (soil and litter carbon and vegetation carbon, both in Pg C), the net 548 ecosystem productivity (NPP - SHR in Pg CO2-C y-1); and the net biosphere productivity (NPP - SHR - Ffire in Pg CO2-C y-1) 549 for the models that include fire emissions. The final plot at the bottom right shows the mean residence time of dead organic 550 matter (MRT_{SHR}; years). The model ensemble is divided into two sub ensembles depending on whether they have a 551 representation of permafrost carbon. In each subplot the left hand box plot ("bulk C", n=55) represents models without 552 permafrost carbon representation and the right hand box plot ("layered C", n=18) represents models which include 553 permafrost carbon. The grey shading represents the likely range estimated by the observationally-informed CARDAMOM 554 analysis. The solid grey line indicates the 50 % quantile, i.e. most likely estimate. The dark grey zone defines the 50 % 555 confidence interval around the 50 % quantile while the light grey zone is the 95 % confidence interval also around the 50 % 556 quantile. In the (NPP - SHR) and (NPP - SHR - Frires) plots the red line is at zero and positive values are a net uptake of 557 carbon.

558

In terms of the plant based C fluxes (i.e. GPP, R_a , NPP and CUE), there is no significant difference (Mann-Whitney, p < 0.01) between the *layered C* models and the *bulk C* models (Figure 3). CARDAMOM's data-informed analysis falls within the spread of the process models. However, the spread of values simulated by the process models is considerably larger than that suggested by CARDAMOM. This is in contrast to the (NPP - SHR) and (NPP - SHR - F_{fires}) fluxes discussed above.

The MRT_{SHR} is significantly longer in models with explicit permafrost C (265 years, *layered C*) than those without (81 year, *bulk C*) (Mann-Whitney p < 0.01; Figure 3). Longer MRT_{SHR} in *layered C* models more closely aligns with CARDAMOM's observationally-informed analysis. Furthermore, the overall majority of the *layered C* models fall within CARDAMOM's 95 % CI. The ones that fall outside have a longer MRT_{SHR} than CARDAMOM. In contrast only a small fraction of the *bulk C* models are consistent with the CARDAMOM MRT_{SHR} with the remainder of the models having a shorter MRT_{SHR} than CARDAMOM. The *layered C* models also have a significantly larger soil C stock which combined with the differences in MRT_{SHR} lead to simulating a similar magnitude of regional heterotrophic respiration as the bulk C models. It is not unexpected that the heterotrophic respiration is similar between the two model ensembles - the additional soil carbon that has been added in the deeper layers is mostly frozen and therefore has very slow heterotrophic respiration. Thus, the *bulk C* models estimate regional heterotrophic respiration which is consistent with both CARDAMOM and *layered C* models but the *layered C* models have the potential for large changes in respiration in the future, whereas the *bulk C* models do not.

The models within this process based model ensemble were run for several different model intercomparison projects so they are not directly comparable in terms of protocol or time period covered. A more constrained ensemble may reduce the uncertainties in the budget estimates. However, the spread of model estimates are still smaller than that from the observationally constrained CARDAMOM assessment.

584 Bottom-up estimate of natural wetland CH₄ fluxes from process-based models

Full model ensemble budgets are only available for CO_2 , but for CH_4 a process-based model ensemble (n=13) of natural wetland CH_4 flux estimates was available from the global CH4 budget (Saunois et al., 2020). The ensemble annual mean is a wetland CH_4 source of 12 (8.6, 16) Tg CH_4 -C yr-1 (Table 1). The interannual variability of the ensemble is low (annual means between 11.2 and 14.1), but there is very large spread within the model ensemble, with annual means over the period varying from 4.9 to 28 Tg CH_4 -C yr-1 for individual models (table S2).

591 Bottom-up ecosystem GHG budgets from data driven upscaling

Bottom-up estimates of GHG budgets from data driven upscaling used for the GHG budgets presented
here are all based on Ramage et al. (in prep). More in depth results and discussion can be found in that
paper. Table 2 summarizes the main findings for the larger budgets posts of the three bottom-up data

- 595 driven ecosystem GHG budgets from that paper. All numbers are annual means estimated over the
- 596 full reporting period of 2000-2020.
- 597 The total budget of CO₂ is near neutral, but with a large uncertainty range 0.4 (-620, 652) Tg-CO2-C
- 598 yr-1 (table 2). Sinks of CO₂, mainly in Boreal forest (-270 (-540, -1) Tg-CO₂-C yr-1) and Permafrost-
- free wetland land cover types (-69 (-125, -14) Tg-CO₂-C yr-1) are offset by sources of CO_2 from fires
- 600 (109 (84, 135) Tg-CO₂-C yr-1) and inland waters (streams, rivers, lakes and ponds combined; 231
- 601 (132, 360) Tg-CO₂-C yr-1). The land cover types Dry tundra ecosystems, Permafrost bogs and Tundra
- 602 wetlands have CO_2 budgets within ± 3 Tg of neutral.
- 603 The total bottom-up data driven budget for CH₄ shows a net source of 38 (21, 53) Tg CH₄-C yr-1
- 604 (table 2). The strongest CH4 sources are Permafrost-free wetlands and inland waters (21 (14, 27) and
- 605 9.4 (4.5, 13) Tg CH₄-C yr-1, respectively). In addition, all other land cover types types, as well as

- fires and geological emissions, represent weak sources of CH₄ to the atmosphere (from 1.2 to 3.3 Tg
- 607 CH₄-C yr-1), with the exception of Boreal forests which are a weak sink (-1.1 (-2.2, 0) CH₄-C yr-1).
- 608 The total bottom-up data driven budget for N₂O shows a net source of 0.62 (0.03, 1.2) Tg N₂O-N yr-1
- 609 (table 2). All land cover types, and fires represent net sources of N₂O to the atmosphere. The strongest
- 610 sources are Boreal forest and Dry tundra ecosystems (0.14 (-0.01, 0.3) and 0.23 (0.04, 0.42)Tg N₂O-N
- 611 yr-1 respectively).
- 612
- Table 2. Summary of all main budget posts for the three GHGs from bottom-up ecosystem upscaling as presented byRamage et al., (in prep).
- 615 Due to the large formats, tables are submitted in a separate Excel file which is hopefully more 616 convenient
- 617 Integrated bottom-up budget combining process models and ecosystem upscaling

Integrated bottom-up GHG budgets were created by combining results of the ensembles of processbased models and ecosystem upscaling, but adding fluxes from components known to be missing in process-based models (abrupt wetland thaw and inland water fluxes, Table 3). These integrated bottom-up budgets are viewed as a best-estimate of bottom-up methods to be contrasted against the top-down atmospheric inversions constraints.

- 623 For CO₂, the combined sinks estimated using data-driven upscaling for upland ecosystems and
- 624 wetlands is somewhat lower than ecosystem sink estimated by the adapted process-based model
- 625 ensemble estimate (-230 and -293 Tg-CO₂-C yr⁻¹, respectively). When adding inland water fluxes, the
- 626 integrated bottom up estimate for CO2 is a weak sink with a wide uncertainty range on either side of a
- 627 neutral budget (-31 (-667, 559) Tg-CO₂-C yr⁻¹). For CH₄ the picture is similar, with the combined sum
- of data-driven ecosystem types and fires being similar to the combined wetland process-based models
- 629 plus wetland abrupt thaw fluxes (27 and 31 Tg CH₄-C yr-1, respectively). When adding inland water
- fluxes the integrated bottom up estimate for CH_4 is a source of 38 (23, 53) Tg- CH_4 -C yr⁻¹. For N₂O,
- only the data-driven upscaling estimate is available, so this is used as a best estimate for the bottom-
- 632 up budget (a sink of 0.62 (0.03, 1.2) Tg N₂O-N yr-1).
- 633
- Table 3. Summary of the different budget components used to generate an integrated bottom-up budget for all three GHGs.
- Due to the large formats, tables are submitted in a separate Excel file which is hopefully moreconvenient

637 Top-down ecosystem GHG budget from atmospheric inversion models

- Top-down ecosystem GHG budgets are derived from ensembles of atmospheric inversion model systems. To facilitate comparison to bottom-up estimates, the top-down inversion budgets are ecosystem budgets, excluding anthropogenic emissions (these are reported below).
- The ensemble of atmospheric inversion models for CO_2 (n=7) indicates that the ecosystems of the permafrost region is a total net CO_2 sink with a multi-annual mean of -587 (-862, -312) Tg-CO₂-C vr⁻¹
- permafrost region is a total net CO_2 sink with a multi-annual mean of -587 (-862, -312) Tg-CO₂-C yr⁻¹
- 643 (2000-2020, Table 4). Included within this net sink are CO_2 sources from fires, estimated by one 644 inverse model system as 78 (51, 104) Tg-CO₂-C yr⁻¹. The inversion systems shows a stronger mean
- 645 annual sink in 2010-2020 compared to 2000-2009 (-643 (-917, -369) and -526 (-802, -250) Tg-CO₂-C
- 646 yr^{-1} , respectively).
- The ensemble of inversion models analyzing CH_4 (n=14) shows a multi-annual mean source from natural ecosystems of 15 (11, 18) Tg CH_4 -C yr-1 (2000-2017; table 4). This estimated source includes small fluxes from fires (1.4 (1.2,1.6) Tg CH_4 -C yr-1).
- 650 For N₂O, only two separate inverse model estimates are available and they show a neutral balance or
- weak source of N_2O , with a multi-annual mean of 0.09 (-0.19, 0.37) Tg N_2O -N yr-1 (table 4). There is
- no notable difference between the first and second decade for CH₄, but the N₂O source was weaker in
- 653 2010-2019 compared to 2000-2009 (Table 4).
- 654

Table 4: Summary table of GHG emissions (annual mean and 95% CI) from the RECCAP2 permafrost domain from
atmospheric inversion models. The reported inverse model fluxes are the ecosystem fluxes, not including anthropogenic
emissions. Sub-fluxes from fires (already included in the total) are also shown for CO₂ and CH₄.

- 658 Due to the large formats, tables are submitted in a separate Excel file which is hopefully more 659 convenient
- 660

661 There are no previous studies synthesizing atmospheric inversion model estimates of these three 662 GHGs for permafrost regions, but the results are in line with studies of similar scope. A tundra biome 663 synthesis from the first generation of RECCAP (McGuire et al., 2012) estimated a net CO₂ sink (-120 Tg C yr⁻¹) between 2000-2006 but with very large differences between individual flux estimates in the 664 665 ensemble (range -440, +210). Since then, more global top-down inverse GHG flux estimates have 666 become available (Friedlingstein et al., 2022), and some recent studies provided GHG budgets for 667 northern regions. (Bruhwiler et al., 2021) analyzed inverse model ensembles across Boreal (50-60 N) and Arctic (60-90 N) domains and describe sinks of CO₂ (-290 and -130 Tg CO₂-C yr⁻¹, respectively 668 for 1980-2017) and sources of CH₄ (16 and 9 Tg CH₄-C yr⁻¹, respectively for 2000-2017). Using a 669 670 similar time series of estimates from atmospheric inversion models, Liu et al. (2022) found that the 671 permafrost region changed from being CO₂ neutral (1980-2000) to a CO₂ sink in 2000-2017 (ca. -200±100 Tg CO₂-C yr⁻¹). There are no previous atmospheric inverse modelling estimates of N₂O for 672 673 the permafrost region with which we can compare our results.

674 Comparison of bottom-up and top-down ecosystem GHG budgets

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676 Synthesis and comparison of the different methods show both convergence and divergence in the

- 677 different GHG budgets (Figure 4). Below the budgets for individual GHGs as well as total C and N
- 678 stock change budgets are discussed.
- 679



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 Figure 4. Summary of main budget items for all three GHGs over the time period 2000-2020 calculated using different methods. The error bars represent the 95% confidence interval.

683 Budgets of CO₂

684 The CO_2 budget for natural ecosystems (excluding anthropogenic fluxes) using bottom-up approaches is a weak sink of -31 (-667, 559) Tg-CO₂-C yr⁻¹ while atmospheric inversion models show a stronger 685 sink of -587 (-862,-312) Tg-CO₂-C yr⁻¹. The 95% CI range of the bottom-up budget is wide and spans 686 687 across the top-down budget, while the range of the top-down budget is narrower and remains a clear 688 sink. In comparison to top-down and bottom-up budgets, the observationally constrained 689 CARDAMOM system (figure 3, table 1) estimates a stronger sink of CO2 -870 (-1780, 160) Tg-CO₂-C yr⁻¹, but CARDAMOM's ensemble estimates crosses the source/sink boundary between the 75th 690 691 and 90th quantile and may be consistent with the other bottom-up sources. Especially for the Tundra 692 biome, process-based models (and CARDAMOM) have CO₂ budgets close to neutral, in line with 693 data-driven upscaling for terrestrial land cover types in that region. 694 Altogether, bottom-up and top-down approaches both show a sink of CO_2 in the region, but of

- 695 different magnitude, and with some approaches not excluding a shift to a weak source.
- 696

697 Because we present a simplified top-down view in this paper (i.e. no analyses of spatial patterns in 698 individual or ensemble models), it is difficult to pin-point the sources of discrepancies between top-699 down and bottom-up approaches. However, for the two bottom-up methods of CO₂ budget estimation, 700 the components can be contrasted. Boreal forest and wetland ecosystems are consistent sinks across 701 the bottom-up approaches, but in the integrated bottom-up CO₂ budget, this sink is offset by fluxes 702 from inland waters and fires. Only data-driven upscaling approaches are available to estimate inland 703 water fluxes, which are typically not represented in process-models. If inland waters are excluded 704 from ecosystem upscaling, the net budget for terrestrial land cover types, including fires, is very 705 similar to the (NPP - SHR - Ffire) flux of process models with layered soil C pools (table 3 and figure 706 3). This suggests that the CO_2 budget for these ecosystems is in agreement between the observational 707 datasets and the generation of models which explicitly represents soil layers and permafrost. It further 708 suggests that process-based models would likely project a weaker CO₂ sink with better representation 709 of inland waters. The three separate methods for calculating fire fluxes spread with estimates of 120, 60, 77 and 110 Tg CO2-C yr⁻¹, from process-based models, CARDAMOM, one atmospheric 710 711 inversion, and the GFED inventories respectively. Comparison of different inventories and process-712 based models in this paper shows that fluxes not represented in the process-based models are 713 potentially large, and should be targeted for inclusion within these complex models. In addition to 714 natural ecosystem fluxes, there may be geological sources of CO_2 which we do not account for, as no 715 separate estimates are available from the permafrost region. The full global geological CO₂ emissions 716 are estimated to 160 Tg CO₂-C yr-1 (Mörner & Etiope, 2002), and it is likely that a small fraction of 717 those fluxes occurs within our study region, but unaccounted for in this budget.

718 Budgets of CH₄

For the natural balance of CH_4 , both the integrated bottom-up budget and the top-down atmospheric inversion models show consistent sources, albeit of different magnitudes at 38 (23, 53) and 15 (11, 18) Tg CH_4 -C yr-1, respectively. Even though both the bottom-up and top-down methods show a source of CH_4 , their uncertainty ranges do not overlap, suggesting that there may be a systematic bias between the methods.

724 In the data-driven upscaling, fluxes of CH₄ are characterized by high fluxes per areal unit from the 725 different wetland land cover types, while other classes with more extensive areal coverage such as 726 Boreal forests and Dry tundra are neutral or even weak sinks due to CH₄ oxidation occurring in dry 727 soils. Land cover types with high CH₄ fluxes are often spatially heterogeneous (with large 728 uncertainties in total area of classes) and sometimes fluxes can be especially large along the margins 729 of these land cover patches. These conditions make CH₄ challenging to upscale, and it also means that 730 the spatial landscape heterogeneity, and the spatial resolution of upscaling or modelling becomes very 731 important for determining accurate budgets (Treat et al., 2018a). For inverse models, this scale issue

should not have large effects on how the systems adjust between prior and posterior fluxes. But theresolution and magnitude of the prior flux ensembles may affect the budget.

734 There are no full CH₄ budgets available from process-based models, but the bottom-up budget 735 includes a model ensemble (n=13) of wetland CH4 flux estimates a source of 12 (8.6, 16) Tg CH₄-C 736 yr-1. This is circa half of the data-driven ecosystem upscaling estimates of combined wetland flux of 737 25 (17, 32) Tg CH₄-C yr-1. Much of this difference may be explained by the lack of abrupt thaw 738 wetlands in the models, which are included among data-driven land cover types and to the integrated 739 bottom-up budget. In addition, the discrepancy may be partly explained by the poor representation of 740 cold-season methane emissions in process-based models, which tend to be underestimated relative to 741 field-based observations (Treat et al., 2018b). The large differences between estimates suggest that 742 future development of process-based model estimates should target inclusion of inland waters, abrupt 743 thaw, but also upland ecosystems and the potential CH4 oxidation occurring there.

744 Budgets of N2O

745 For N₂O, both the bottom-up ecosystem flux upscaling and the top-down atmospheric inversion 746 models show sources but with large differences between estimates (0.62, (0.03, 1.23)) and (0.09, (-0.19, 1.23))747 0.37) Tg N₂O-N yr-1, respectively). Both methods have relatively wide uncertainty ranges that 748 overlap each other, and the inverse model estimate cannot with confidence be distinguished from a 749 neutral budget. The bottom-up estimates are seven times higher than the top-down estimates, showing 750 a clear need to further refine the methods and to gather more observational data. The high bottom-up 751 budget is mainly driven by fluxes from large areas of upland Dry tundra and Boreal forest (despite 752 small per unit area fluxes), but uncertainty ranges are wide for all land cover classes and the mean 753 values for classes may be driven up by preferential reporting from measurement sites with high 754 fluxes ..

755 Tian et al., (2020) presented a global quantification of N2O sources and sinks, where top-down and 756 bottom-up estimates were very similar (ca. 17 Tg N₂O-N yr-1) Although they do not present numbers 757 specifically by biome, the estimates for northern regions in Tian et al. (2020) are generally low, and 758 more consistent with our top-down estimates. We provide no model ensemble estimates for N₂O, as 759 few process-based models simulate cycling of N₂O in permafrost ecosystems. One exception is a 760 recent study using the QUINCY model to estimate an average mean annual flux of 4 mg N₂O-N m-2 761 year-1 across several tundra ecosystem sites (Lacroix et al., 2022). If upscaled to the full tundra domain (5.58 M km2 including classes Wet tundra and Dry tundra), this would yield an annual flux of 762 763 0.022 Tg N₂O-N yr-1, an order of magnitude lower than our bottom-up estimates for these same 764 classes. Conversely, another recent model study estimates much higher N₂O fluxes. An ongoing study 765 (in review) uses the TEM model to estimates a pan-Arctic N₂O budget between 1.1 - 1.2 Tg N₂O-N 766 yr-1 (Yuan et al., 2023), which surpasses any previous estimates presented in this study. We note that the cited study is currently in the form of an open discussion paper and we interpret this estimate

cautiously. It underscores, however, that the uncertainty of the N_2O budget for the permafrost region

is still very large and that modelling work is in its infancy.

Anthropogenic CO₂ and CH₄ emissions, lateral export of C and N and total budgets of C

- 772 Separate estimates of anthropogenic emissions are available from global gridded data for CO₂ and 773 CH_4 (table 5, see table S8 for more details). For both these gases, the emissions are dominated by 774 combustion of fossil fuels occurring within the permafrost region. Anthropogenic emissions of CO2 775 are estimated to be 73 (56, 89) Tg-CO₂-C yr⁻¹, mainly from gas and oil (26 (19, 34) and 32 (27, 36) Tg-CO₂-C yr⁻¹, respectively) (table S8). Anthropogenic emissions of CH4 are estimated to be 5.3 (3.8, 776 6.7) Tg-CH₄-C yr⁻¹, mainly from Fossil fuels but with a small contribution from Agriculture and waste 777 (4.7 (3.4, 6.0) and 0.54 (0.43, 064) Tg-CH₄-C yr⁻¹, respectively) (table S8). These anthropogenic 778 779 fluxes have not been included to bottom-up and top-down budgets reported above, but are included 780 for total budgets of C as well as calculation of the full combined Global Warming potential of all 781 three GHGs.
- 782 In addition to the natural and anthropogenic GHG exchange with the atmosphere, C and N is laterally 783 exported from the permafrost region to the Arctic Ocean via riverine transport and coastal erosion (Table 5). The lateral fluxes of organic C were estimated at 94 (79, 110) Tg C y⁻¹, with riverine 784 transport of dissolved organic carbon contributing 78 (70, 87) Tg C y⁻¹ and coastal erosion of soil and 785 sediment organic C contributing 15 (9, 24) Tg C y⁻¹. Lateral export of N from the permafrost region 786 is estimated to be 2.6 (1.9, 3.6) Tg N y^{-1} with a smaller component of riverine transport (1.0 (0.9, 1.1) 787 Tg N y⁻¹) compared to coastal erosion (1.6 (1.0, 2.5) Tg N y⁻¹). The ratio of C:N lost via lateral 788 789 transport is very high (ca. 80) while the C:N ratio in material lost via coastal erosion (ca. 10) is
- consistent with mature mineral soil organic matter in the permafrost region (Harden et al., 2012).
- 791 Combining the net CO₂ and CH₄ budgets with lateral fluxes yields annual estimated organic C stock 792 change budgets. The sign and magnitude of combined organic C stock change budgets depend on 793 whether bottom-up or top-down approaches to GHG budgeting are used (101 (-565, 723) and -479 (-772, -183) Tg C y⁻¹, respectively). Using a mean of the bottom-up and top-down budgets for GHGs, 794 we estimate a net sink of C from the atmosphere into the terrestrial permafrost region of -189 (-669, 795 270) Tg C y⁻¹. Because we do not account for all N fluxes, we cannot close the full N budget. Our 796 797 combined N2O and lateral flux data spreads less between bottom-up and top-down approaches than 798 for C. The mean estimate shows a net loss of nitrogen to the atmosphere and ocean of 2.9 (1.8, 4.4) Tg N y^{-1} (Table 5). As there is a net sink of C in the region, and the overall ecosystem C:N (in soil and 799 800 vegetation) is unlikely to grow increasingly wider, we expect that if a full N budget was available, a 801 net source of N is more likely. This suggests an unquantified source of N into the system not

802 quantified in our budgets. If we assume that the total ecosystem C:N ratio (for vegetation and 0-1 m 803 soils) is stable over time at a value around 15-20 (Palmtag et al., 2022), a total N source of 11-14 Tg 804 N y⁻¹ would be needed to balance the N budget in relation to the C budget. Likely missing N sources to balance the budget may be atmospheric N deposition or biological N2 fixation, both important 805 806 sources of available N for subarctic and arctic ecosystems (Rousk et al., 2018; Yuan et al., 2023). 807 Yuan et al. (2023) report an estimated N deposition between 10 and 15 Tg N yr-1 for the pan-Arctic, in 808 close agreement with the fluxes needed to close the budget. Mean N2 fixation rates of boreal forest (0.12 809 g N m-2 y-1) and tundra ecosystems (0.33 g N m-2 y-1) (Yu & Zhuang, 2020) would yield an 810 additional N sinks of 1.2 and 2.1 Tg N yr-1, respectively, if upscaled to the full spatial extent of the 811 Boreal forests and tundra biomes within the region. Although these estimates of additional N sinks 812 into ecosystems are uncertain, they suggest that our estimated net loss of N via N₂O flux and lateral 813 losses from the domain are a small component of the N cycle, and that a balanced full N-budget 814 would be in agreement with an organic C stock change budget based on a GHG C-flux calculated as 815 the mean between bottom-up and top-down approaches.

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Table 5. Summary of all main budget posts for the three GHGs, including anthropogenic fluxes as well as lateral fluxes andtotal sum changes of C and N.

819 Due to the large formats, tables are submitted in a separate Excel file which is hopefully more 820 convenient

821

822 Estimated combined Global Warming Potential of the permafrost region

823 Using a common unit of CO2-C equivalents over 100 years (GWP100), and including both 824 ecosystems and anthropogenic emissions, the net balance of the three GHGs from bottom-up approaches shows a net source of 579 (-317, 1432) Tg CO₂-C eq yr⁻¹ while top-down atmospheric 825 inversions show a sink of -242 (-576, 83) Tg CO₂-C eq yr⁻¹ (Figure 5, Table 6). Using the mean of the 826 827 two approaches gives an estimated combined GHG source of 147 (-492, 759) Tg CO₂-C eq yr⁻¹). In this mean GWP100 estimate, a CO₂ sink (-237 (-708, 212) Tg CO₂-C eq yr⁻¹) is offset by sources of 828 CH₄ (343 (225, 455) Tg CO₂-C eq yr⁻¹) and N₂O (40 (-9.3, 92) Tg CO₂-C eq yr⁻¹)). Our results 829 830 highlight the importance of monitoring non-CO2 trace gases, since they are responsible for the 831 regions crossing over to a net source of CO₂-C equivalents. Because of their different properties, life 832 times and concentrations in the atmosphere, the total radiative balance of the three GHGs together 833 varies depending on the timescale. Therefore, the estimates of the combined GHG sink or source 834 strengths based on GWP calculations should be interpreted with care. Supplementary table S9 shows a 835 summary of annual GHG budgets converted to CO2-Ceq using 20- and 500-year Global Warming 836 Potential (GWP20 and GWP500). Using CO₂-C equivalents over shorter time-scales (GWP20) yields

net sources from both bottom-up and top-down approaches (1,361 (168, 2512) and 81 (-399, 533), Tg CO₂-C eq yr⁻¹ respectively) with a mean source of 721 (-115, 1522) Tg CO₂-C eq yr⁻¹. Over multiple centuries (GWP500), the bottom-up estimate remains a source (124 (-310,)), but both the top-down and mean estimates are of net GHG sinks (-451 (-789, -116) and -132 (-664, 377) Tg CO₂-C eq yr⁻¹, respectively)

842 The uncertainty ranges of the combined GHG budgets (using GWP100), in both bottom-up and top-843 down approaches, span across a neutral budget. Because the budget is so close to neutral, recent and 844 future shifts in disturbance regimes may shift the sign of the net GHG budget. The combined effect of fire (from bottom-up scaling) is a source of ca 130 Tg CO₂-Ceq yr⁻¹. Without fires, the mean net GHG 845 846 balance would be close to neutral, but unusually strong fire years can significantly increase the net 847 GHG source. For instance, in the summer of 2021, global boreal fire emissions were nearly three 848 times larger than the 2000-2020 mean (Zheng et al., 2023). Emissions of CO₂,CH₄, and N₂O from 849 abrupt thaw landforms, both at present and in the future, is a large but highly uncertain source of 850 GHGs. Abrupt thaw (thermokarst) may expand or shift rapidly over time and can be triggered by fires 851 or by unusually warm summers (Turetsky et al., 2020). The estimated fluxes from abrupt thaw lakes 852 and wetlands are conservative in the bottom-up data driven upscaling used for this assessment, but they may be as large as 300 CO₂-Ceq yr-1 for the 2000-2020 period (see Ramage et al. (in prep.)) 853

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Figure 5. Annual GHGs budgets for bottom-up, top-down approaches, as well as the mean between these two approaches.
All numbers converted to CO₂-equivalents (Tg CO₂-Ceq, with 95% CI) using a 100-year Global Warming Potential (GWP-100). Table 6 complements this figure by showing more detailed sub-fluxes for the different categories.

Table 6. Summary of annual GHGs budgets for all main budget posts, converted to CO₂-equivalents (Tg CO₂-Ceq, with 95% CI) using a 100-year Global Warming Potential (GWP-100) from bottom-up approaches, top-down atmospheric

- inversion models and from anthropogenic fluxes. A GWP-100 of 29.8 for CH_4 and 273 or N_2O relative to CO2 was used (IPCC, 2023; Table 7.15).
- Due to the large formats, tables are submitted in a separate Excel file which is hopefully more convenient

867 Main sources of uncertainty within and between GHG budgets

868 The spread between methods in this study is smaller than in some earlier studies (McGuire et al., 869 2012; Saunois et al., 2020), the differences between bottom-up and top-down estimates are 870 substantial. Combined for all three GHGs over natural ecosystems, the bottom-up ecosystem scaling 871 predicts a source of 449 (-415, 1272) Tg CO₂-Ceq yr-1 while the top-down inversions estimate a sink 872 of -414 (-765, -74) Tg CO₂-Ceq yr-1 (table 6, excluding anthropogenic fluxes). Both the systematic 873 discrepancies between bottom-up and top-down methods, and the wide uncertainty ranges of data and 874 model estimates point to a need for further refinement of methods, process representation and 875 additional observational data.

The flux estimates from the global GHG inversions for the permafrost region show a relatively large spread between the different systems (tables S5-S7). This large range may partly be explained by a relatively limited set of atmospheric concentration observations being available for the region (from the surface networks or satellite observing systems). Furthermore, the resolution at which fluxes are estimated is relatively coarse in comparison to the large landscape heterogeneity of the studied region - this might increase variability between different inversion systems and contribute to the systematic differences in relation to bottom-up approaches.

883 For all three GHGs, the ecosystem flux upscaling estimates suggest stronger sources than the 884 atmospheric inverse models. Because the budgets presented here integrate the full permafrost domain, 885 it is challenging to assess where and why differences occur between bottom-up and top-down 886 estimates. These consistent differences between bottom-up and top-down methods may indicate a 887 systematic bias in either or both of the methods, possibly related to the representativeness of the 888 observations used, as atmospheric mole fraction measurements have a larger footprint compared to 889 EC observations. Strong sources and sinks of all GHGs are associated with certain land cover classes, 890 land forms or processes (e.g. wetlands, inland waters, thermokarst lakes, and fires). It would be of 891 interest for further studies to analyze the spatial patterns of inversion budgets relative to the 892 distribution of specific land cover types or processes in the landscape.

A potential source of mismatch between bottom-up and top-down estimates is the source of information used to estimate prior fluxes for the GHG inversion systems. It is possible that if prior flux ensembles consistently included all of the land cover types and processes included in bottom-up estimates, they would be more similar. For instance, the global inversions do not include emissions from inland water as prior knowledge for the inversion system (supplemental table S3 and references 898 therein). For CO2 the prior fluxes are based on land surface models and fire inventories as well as 899 anthropogenic emissions, with systems optimized to estimate fluxes from the natural ecosystems 900 (Friedlingstein et al., 2022). For CH₄, the prior fluxes included only emissions estimated from 901 process-based models for wetlands and anthropogenic sources (Saunois et al., 2020). Inland waters 902 are important sources of all GHGs, with a combined emission of 330 (180, 500) Tg CO₂-Ceq yr⁻¹ 903 (Table 6). Adding inland water emissions to the prior flux ensemble may help reconcile bottom-up 904 and top-down estimates. We note that the budgets estimated by the inversion ensembles are relatively 905 similar to the budgets estimated process-based models which have been used as priors. The NEE as 906 well as wetland CH₄ fluxes derived from process-model ensembles are similar to the atmospheric inverse model CO₂ and CH₄ budgets, respectively (-420 and -587 Tg CO₂-C y⁻¹ as well as 12 and 15 907 908 Tg CH₄-C yr-1, respectively). It is possible that in regions with limited atmospheric observations to 909 correct concentration, the posterior inverse model fluxes do not deviate much from the priors. This 910 problem may be particularly important during the long cold season. Shoulder-seasons and winter 911 fluxes of both CO_2 and CH_4 are significant parts of the annual budgets in field measurements across 912 multiple sites but process-based models (used as inversion priors) capture these emissions poorly 913 (Natali et al., 2019; Treat, et al., 2018b). This bias may be further exacerbated by systematic lack of 914 observational constraints during winter. The observational networks for GHG fluxes in the permafrost 915 region during winter are very sparse, especially in Canada and Russia (Pallandt et al., 2022). Further, 916 the inverse model systems that use satellite instruments need enough insolation to measure the CO_2 or 917 CH_4 atmospheric columns. In practice, this restriction means that the high latitudes are rarely sampled 918 by satellites around the winter hemisphere.

919 But even with updated prior fluxes, it is probable that estimates from bottom-up approaches would 920 still be higher than the atmospheric constraint and the balance of global GHG budgets. This suggests a 921 need for further revising the bottom-up upscaling methods in parallel with development of top-down 922 methods and datasets. The spatial products used to delineate the surface area of land cover types for 923 this study remain coarse (Olefeldt et al., 2021), both in terms of spatial resolution and the diversity of 924 land form classes (Ramage et al., in prep). The next generation of remote sensing products are likely 925 to improve estimates of areas of different important land cover classes in the near future. In addition 926 to uncertain areal coverage of key land cover classes, the bottom-up budgets for all GHGs are strongly 927 affected by average fluxes from individual landcover classes. Because of the heterogeneous nature of 928 GHG fluxes in both space and time, it is challenging to generate datasets with unbiased and accurate 929 annual GHG budgets (Rößger et al., 2019; Treat, et al., 2018a). There is a risk of biases toward 930 measurements of high fluxes associated with spatial and temporal variability both within and across 931 landcover types. There is also a risk for higher reporting prevalence from sites with high emissions, or 932 that land cover types that may be weak sinks of GHGs (but may cover large areas) are under-reported. 933 For example, the importance of spatially widespread but weak soil CH_4 sinks in determining the full 934 landscape CH₄ budget has been shown for Arctic tundra in Greenland and Siberia (Juncher Jørgensen

935 et al., 2015; Juutinen et al., 2022). Larger regional gaps in observational flux networks may also be 936 important (Pallandt et al., 2022; Virkkala et al., 2018), because different regions can have highly 937 diverse environmental conditions and rates of warming and permafrost thaw which can impact GHG 938 fluxes in various ways. The formation and past history of permafrost deposits in different regions also 939 influence the potential permafrost GHG feedback strengths (Jones et al., 2023), which may bias 940 estimates if field data is unevenly spaced. Better representation is also needed from sites affected by 941 different disturbance regimes. This includes both fluxes from disturbance events as well as post-942 disturbance trajectories, such as CO2 sinks after fire (Walker et al., 2019) or biological pest outbreaks 943 (Lund et al., 2017). If the assemblage of sites used for calculating average GHG fluxes is not broadly 944 representative of areas in long-term recovery trajectories, it may well underestimate C sinks in post-945 disturbance ecosystems. Another large source of uncertainty is how abrupt permafrost thaw affects 946 GHG budgets. In our budgets, it is partly included in the data-driven upscaling (via specific land 947 cover or lake types types in BAWLD), but entirely missing in the process models. The spatial extent 948 and annual fluxes of abrupt thaw landforms remain poorly constrained (Turetsky et al., 2020). In 949 addition, abrupt thaw landforms and other disturbed soils may emit N₂O (Voigt et al., 2020), but this 950 is not included in any of our budgeting approaches. Further improvements to data-driven bottom-up 951 budgets could be made by i) an increased number of observations, including more spatially distributed 952 data and non-growing season measurements, ii) consistent reporting of net-zero or negative fluxes to 953 prevent biased site selection and reporting in published literature and iii) upscaling using techniques 954 that can simultaneously consider several environmental conditions and their variability across the 955 entire permafrost region (Hugelius et al., 2020; Natali et al., 2019; Virkkala et al., 2021)

956 Conclusions

957 We present the first synthesis of GHG budgets for CO₂, CH₄ and N₂O as well lateral fluxes of C and 958 N across the northern terrestrial permafrost region using bottom-up (ecosystem flux upscaling and 959 process-based models) and top-down (atmospheric inversion models) approaches for the period 2000-960 2020. In comparison, bottom-up approaches consistently yield estimates of stronger GHG sources 961 compared to top-down. Both approaches show a net sink of CO_2 in natural ecosystems, but they 962 diverge by several hundred Tg CO₂-C yr-1 (-31 (-667, 559) and -587 (-862, -312), respectively). The 963 Boreal biome, especially Boreal forest land cover, is a stronger net sink while the tundra biome is 964 neutral, or even a source when accounting for fluxes from inland waters. Bottom-up and top-down 965 approaches both show sources of CH₄, but the 95% CI ranges do not overlap (38 (23, 53) and 15 (11, 966 18) Tg CH4-C yr-1, respectively). The strongest sources of CH_4 are permafrost-free wetlands, and 967 inland waters. Estimates of N₂O are highly uncertain, but both methods estimate sources to the 968 atmosphere (0.6 (0.03, 1.2) and 0.09 (-0.19, 0.37) Tg N₂O-N yr-1). Anthropogenic emissions from the 969 region are 73 (56, 89) Tg CO₂-C yr-1 and 5.4 (3.8, 7.1) Tg CH₄-C yr-1, in both cases dominated by

- 970 combustion of fossil fuels (estimates not available for N2O). Assuming equal weight to bottom-up
- 971 and top-down budgets, the combined global warming potential at a 100 year timescale (GWP100) is a
- net GHG source of 147 (-492, 759) Tg CO₂-Ceq yr-1, The CO₂ sink is more than offset by the CH₄
- source, with a small source contribution from N_2O (-285, 343 and 40 Tg CO_2 -Ceq yr-1, respectively).
- 974 When calculating global warming potential over decadal time-scales (GWP20) both bottom-up and
- top-down approaches show net sources of GHGs. The estimated total annual budgets of C and N,
- 976 when anthropogenic and lateral fluxes are included (but not accounting for N deposition or N2
- 977 fluxes), are -189 (-669, 270) Tg C yr–1 and 2.9 (1.8, 4.4) Tg N yr-1.
- 978 Inverse model datasets have not been extensively used for studies of the permafrost region, but are 979 highly useful for broad-scale budgets and should be utilized to a greater degree in future studies. The 980 consistently lower land to atmosphere fluxes from top-down inversions compared to bottom up points 981 to potential systematic biases in both methods. Future efforts should focus on improved observational 982 networks to support atmospheric inversions in the region and comparison of spatial patterns within 983 atmospheric inversion models, including analysis of how prior fluxes affect the posterior budgets. 984 Data-driven bottom-up estimates are still data-limited and further refinement of the spatial resolution 985 and GHG balances for individual classes could improve estimates. Process-based model estimates are highly useful and complementary to other budgeting approaches. With future addition or 986 987 improvement of key processes, such as fire, abrupt thaw and inland water dynamics, it is likely that 988 budgets from process-based models would be similar to data-driven upscaling. If it can be shown that 989 process-based models mimic data-constrained estimates for present day budgets, it increases 990 confidence in using models for projections of future GHG dynamics.

991 In summary, we cannot currently reconcile bottom-up and top-down GHG budgets for the permafrost 992 region. The bottom-up budget may be biased in ways that increase estimated fluxes to the atmosphere 993 from high-emitting land cover types while top-down atmospheric inversion budgets may be biased in 994 ways that decrease fluxes to the atmosphere by not including ecosystem types that are known net 995 GHG sources in prior flux estimates. Considering these constraints, a mean between the integrated bottom-up and top-down budget approaches can be seen as the most robust best estimate under the 996 997 current state of knowledge. We conclude that while uncertainties remain, the budgets are sufficiently 998 well constrained to shows the northern permafrost region as a net sink of organic C, but a net source 999 of combined global warming potential over decadal to century time-scales. The boreal biome is likely 1000 a GHG sink, or neutral, but the tundra biome is a GHG source. Ongoing and projected future 1001 permafrost thaw as well as intensification of disturbance regimes, including droughts, storms, pests 1002 and fires are likely to strengthen GHG sources across the whole region.

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1452 Supplementary text and figures



1454 Figure S1. The spatial extent of permafrost defined in this study - the BAWLD-RECCAP2 permafrost

- region shown in relation other RECCAP2 regions as well as the full permafrost extent. The spatial extent of the permafrost region defined in this study as an overlap of the RECCAP2 permafrost extent
- 1457 and the Boreal Arctic Wetlands and Lakes Dataset (BAWLD, Olefeldt et al., 2021). The colours show
- regional overlap with different RECCAP2 regions (Ciais et al., 2022). The grey shades show the full

northern permafrost extent as defined by the combination of data from Obu et al. (2021) plus areas
that recorded mean annual air temperature (MAAT) below 0°C between 1990 and 2000, which is
consistent with the area used in the ISIMIP3 model intercomparison.

1462

1463 Detailed method description for CARDAMOM

1464 The DALEC intermediate complexity model for terrestrial ecosystems simulates the C stocks, inputs 1465 (photosynthesis), outputs (respiration, fire and removals) and internal flows (allocation / turnover) of 1466 4 biomass (labile, foliage, roots and wood) and two dead organic matter (litter and soil) pools (Bloom 1467 & Williams 2015; Smallman & Williams 2019). Photosynthesis is simulated by the ACM-GPP-ET 1468 model which estimates photosynthesis as a function of available CO₂, temperature, absorbed 1469 shortwave radiation, leaf area. Available CO_2 is a function of stomatal conductance which is itself 1470 determined by opening the stomatal until the additional photosynthate gained falls below a critical 1471 threshold or the limits of water supply from the soil via the roots is reached (for details see Smallman 1472 & Williams 2019). Photosynthate is allocated to autotrophic respiration and plant tissues based on 1473 fixed fractions. Allocation of labile to the canopy and canopy senescence are determined as a function 1474 of day of year. Turnover of wood and fine roots follows first order kinetics. Decomposition of litter to 1475 soil, and heterotrophic respiration from both litter and soil mineralisation follow first order kinetics 1476 modified by an exponential temperature response function. Each pool, and flux within DALEC and its 1477 exchanges with its environment are governed by parameters retrieved by CARDAMOM for each 1478 location independently, but as a function of local information.

1479

1480 CARDAMOM analysed terrestrial C-cycling at a monthly time step and 0.5 x 0.5 degree spatial 1481 resolution for 19 years (2001-2019). CARDAMOM combines information contained within the 1482 DALEC model structure, DALEC's drivers, the assimilated observations and ecological knowledge to 1483 estimate ensembles of local parameters. From these ensembles of parameters we can explicitly 1484 quantify at pixel scale uncertainty in both the underlying parameters and C-cycling. Meteorological 1485 drivers are drawn from the GSWP3-W5E5 dataset, fire is imposed as a function of MODIS burned 1486 area and forest loss is imposed using global forest watch (Hansen et al., 2013). Assimilated 1487 information are time series estimates of leaf area index (Copernicus Service Information 2021), 1488 woody biomass for 2017 and 2018 (Santoro et al., 2021), and net biome exchange of CO2 (Koren 1489 2020). NCSCD provides pixel specific prior for the initial soil C content while a globally applied prior 1490 for the ratio of autotrophic respiration and photosynthesis (of 0.46 +/- 0.12, Collalti & Prentice 2019). 1491 Ecological knowledge is applied using ecological and dynamical constraints (EDCs) which ensure 1492 rejection of parameter combinations which are ecologically unrealistic, such as wood turnover being 1493 faster than fine root or inappropriate exponential changes in C stocks (for details see Bloom et al., 1494 2016).



1497 Figure S2: Steady state C-budget for RECCAP2-Permafrost. Numbers show median estimate of fluxes 1498 (alongside arrows) and of stocks (in boxes). Units are MgC ha-1 for stocks and MgC ha-1 yr-1 for fluxes. 95% 1499 confidence intervals are shown in a fractional form with 2.5 and 97.5 percentile as numerator and denominator. 1500 Black fluxes are biogenic, including net primary production (NPP), mortality (Mort), autotrophic respiration 1501 (Ra), and heterotrophic respiration (SHR). NEE = Ra + SHR - GPP. NBE = NEE + Etotal. Red fluxes are fire-1502 driven emissions (E).

1503

1504 Process based model result separated to the Boreal and Tundra biomes





1507 Figure S3.Carbon fluxes, stocks and relevant ecosystem properties from the process-based models listed in supplemental 1508 table S1 for the boreal region. The top row shows the following simulated multi-annual mean C fluxes (left to right) -

1509 heterotrophic respiration (SHR); gross primary productivity (GPP); net primary productivity (NPP); autotrophic respiration 1510 (R_a) all in Pg CO₂-C y⁻¹. Also shown on the top row are the carbon use efficiency (CUE, dimensionless) and the fire C flux (F_{fires} in Pg CO₂-C y⁻¹). The bottom row shows C stocks (soil and litter carbon and vegetation carbon, both in Pg C), the net 1511 ecosystem productivity (NPP - SHR in Pg CO2-C y-1); and the net biome productivity (NPP - SHR - F_{fires} in Pg CO2-C y-1) 1512 1513 for the models that include fire emissions. The final plot at the bottom right shows the mean residence time of dead organic 1514 matter (MRT_{SHR}; years). The model ensemble is divided into two sub ensembles depending on whether they have a 1515 representation of permafrost carbon. In each subplot the left hand box plot ("bulk C", n=55) represents models without 1516 permafrost carbon representation and the right hand box plot ("layered C", n=18) represents models which include 1517 permafrost carbon. The grey shading represents the likely range estimated by the observationally-informed CARDAMOM 1518 analysis. The solid grey line indicates the 50 % quantile, i.e. most likely estimate. The dark grey zone defines the 50 % 1519 confidence interval around the 50 % quantile while the light grey zone is the 95 % confidence interval also around the 50 % 1520 quantile. In the (NPP - SHR) and (NPP - SHR - Ffires) plots the red line is at zero and positive values are a net uptake of 1521 carbon.





Table 1. Summary of annual budgets from process-based model ensembles and the model data-fusion approach CARDAMOM. For CO2 the mean from the full ensemble of models that include fire is presented as well as the annual budgets from CARDAMOM. For CH4 the annual fluxes from wetlands are reported. No process-based model results are available for N2O

P	rocess-base	ed models, a	innual budgets			
	CO2	2.5% CI	97.5% CI	CH4	2.5% CI	97.5% CI
	T	g CO2-C yr-	1	-	Гg CH4-C yr	-1
Model ensemble: NPP- SHR - Ffire	-340	-930	90	NA		
Model ensemble: NPP - SHR	-420	-1020	-80			
Model ensemble: Ffire	120	10	460			
CARDAMOM: NPP- SHR - Ffire	-870	-1780	160	NA		
CARDAMOM: NPP - SHR	-960	-1880	20			
CARDAMOM: Ffire	60	50	100			
Model ensemble wetland CH4 flux	NA			12	8.6	5 16

(p. cp/.													
Data-driven upscaling (All numbers from Ramage et al. in prep)													
	Surface area	CO2	2.5% CI 9	7.5% CI	CH4	2.5% CI	97.5% CI	N2O	2.5% CI	97.5% CI			
	106 km2	Τg	Tg CO2-C yr-1			g CH4-C yr-	1	Tg N2O-N yr-1					
Total budget	17.05	0.4	-620	652	37.7	21.3	52.8	0.62	0.03	1.19			
Sum all terrestrial land cover:		-340	-836	156	25.6	14.69	36.4	0.55	-0.03	1.14			
Sum uplands	14.2	-267	-687	153	1	-2.6	4.5	0.37	0.03	0.72			
Boreal forests	s 9	-270	-540	- 1	-1.1	-2.2	0	0.14	-0.01	0.3			
Dry tundra	a 5.2	2.9	-148	154	2.1	-0.4	4.5	0.23	0.04	0.42			
Sum wetlands	2.8	-72	-148	4	24.6	17.29	31.9	0.18	-0.06	0.42			
Non-permafrost wetlands	s 1.6	-69	-125	-14	21	14	27	0.07	-0.03	0.17			
Permafrost bogs	s 0.86	-0.05	-0.8	0.73	0.7	0.29	1.1	0.1	-0.03	0.23			
Tundra wetlands	s 0.4	-2.7	-23	17	3.3	2.7	3.9	0.01	0	0.02			
Fires	1.1	109) 84	135	1.2	0.93	1.45	0.07	0.057	0.083			
Inland waters	1.4	231	132	360	9.4	4.5	13.1	0.002	0.001	0.003			
Geological emissions		NA	NA	NA	1.5	1.2	1.8	NA					

Table 2. Summary of all main budget posts for the three GHGs from bottom-up ecosystem upscaling as presented by Ramage et al., (in prep).

Table 3. Summary of the different budget components used to generate an integrated bottom-up budget for all three GHGs.

		CO2	2.5% CI	97.5% CI	CH4	2.5% CI	97.5% CI	N2O	2.5% CI	97.5% CI
		Т	g CO2-C yr-	1	Т	g CH4-C yr-	1	Τg	g N2O-N yr-	1
Data-driven synthesis	Total budget	0.4	-620	652	38	21	53	0.6	0.03	1.2
	Sum upland landcover types	-267	-687	153	1	-2.6	4.5	0.37	0.03	0.72
	Sum wetlands	-72	-148	4	25	17	32	0.18	-0.06	0.42
	Fires	109	84	135	1.2	0.9	1.5	0.07	0.06	0.08
	Inland waters	231	132	360	9.4	4.5	13	0.002	0.001	0.003
Abrupt thaw wetland flux	(not included in data-driven upscaling total)	19	13	26	19	12	26			
Process-based models	Model ensemble NEE (NPP- SHR - Ffire)	-340	-930	90	NA			NA		
	Model ensemble CH4 wetland flux	NA			12	8.6	16	NA		
Integrated bottom-up budget	Natural ecosystems (including fires)	-31	-667	559	38	23	53	0.6	0.03	1.2
	Mean vegetated upland+wetland ecosystems (with fire)	-262	-799	200	29	18	40	NA		
	Data-driven, sum upland+wetland landcover types	-230	-757	292	27	15.6	37.9	0.6	0	1.2
	Model ensemble NEE plus and wetland abrupt thaw	-293	-842	109	NA			NA		
	Model ensemble wetland CH4 flux and abrupt wetland thaw				31	21	42			
	Inland waters (from data-driven synthesis)	231	132	360	9.4	4.5	13	0.002	0.001	0.003

Table 4: Summary table of GHG emissions (annual mean and 95% CI) from the RECCAP2 permafrost domain from atmospheric inversion models. The reported inverse model fluxes are the ecosystem fluxes, not including anthropogenic emissions. Sub-fluxes from fires (already included in the total) are also shown for CO2 and CH4.

			CH4	N2O					
	Tg CO2-C yr-1	2.5% CI	97.5% Cl	Tg CH4-C yr-1	2.5% CI	97.5% CI Tg I	N2O-N yr-1	2.5% CI	97.5% CI
Inversion ensemble 2000-2009	* -526	-802	-250	14	11	18	0.01	-0.24	0.27
Fire sub-flux**	72.8	48.048	97.552	1.235	1	1.47 NA			
Inversion ensemble 2010-2020	* -643	-917	-369	15	12	18	0.16	-0.14	0.46
Fire sub-flux**	82.1	54.186	110.014	1.6	1.4	1.7 NA			
Inversion ensemble 2000-2020	* -587	-862	-312	15	11	18	0.09	-0.19	0.37
Fire sub-flux**	77.7	51.282	104.118	1.4	1.2	1.6 NA			

* GHG are reported for different periods: 2000-2020 for CO2, 2000-2017 for CH4, and 2000-2019 for N2O **Fire sub-flux of CO2 from Zheng et al., (2023) CI range assumed to be proportional to reported CI for trends in that paper. Fire sub-flux of CH4 is biomass and biofuel burning extracted by mask from global methane budget datasets used in Saunois et al (2020).

		CO2	2.5% Cl	97.5%	CI	CH4 2	.5% C	97.5% CI	N2O	2.5% CI	97.5% CI
		Tg CO2-C yr-1				Тg	CH4-C	yr-1	Tg N2O-N yr-1		
Bottom-up integrated budget	All ecosystems (including fires)	-31	-66	67 55	59	38	23	53	0.6	0.03	1.2
Top-down atmosperic inversions	All ecosystems (including fires)	-587	-86	62 -31	2	15	11	18	0.09	-0.19	0.37
Anthropogenic emissions	Total budget	73	Ę	56 8	89	5.4	3.8	7.1	NA		
	Fossil fuels	73	5	6 8	9	4.9	3.3	6.5	NA		
	Agriculture and waste	NA				0.54	0.44	0.64	NA		
		Carbo					Nitrog	gen stock ch	anges		
		Tg C yr-1	2.5% CI	97.5%	CI				Tg N yr-1	2.5% CI	97.5% CI
Mean gas C (CO2+CH4) and N (N2O)	budge1Bottom-up	7	-64	14 61	2				0.6	0.03	1.2
	Top-down	-573	-85	51 -29	94				0.09	-0.19	0.37
Lateral flux C and N budgets		94	7	79 11	1				2.6	1.9	3.6
	Riverine flux	78	7	70 8	7				1.0	0.9	1.1
	Coastal erosion	15	9.	2 2	4				1.6	1.0	2.5
Sum C and N changes	Bottom-up	101	-56	65 72	23				3.2	1.93	4.8
	Top-down	-479	-77	72 -18	33				2.7	1.7	4.0
	Mean bottom-up/top-down	-189	-66	69 27	0				2.9	1.8	4.4

Table 5. Summary of all main budget posts for the three GHGs, including anthropogenic fluxes as well as lateral fluxes and total sum changes of C and N.

Table 6. Summary of annual GHGs budgets for all main budget posts, converted to CO2-equivalents (Tg CO2-Ceq, with 95% CI) using a 100-year Global Warming Potential (GWP-100) from bottom-up approaches, top-down atmospheric inversion models and from anthropogenic fluxes. A GWP-100 of 29.8 for CH4 and 273 or N2O relative to CO2 was used (IPCC, 2023; Table 7.15).

	GWP100	CO2	2.5% CI	97.5% CI	CH4	2.5% CI	97.5% CI	N2O	2.5% CI	97.5% CI	Total GHGs	2.5% CI	97.5% CI
			Tg CO2-Ce yr-1		Тg	CO2-Ce y	r-1	Тд	CO2-Ce yr	-1	Τg	CO2eq yr	-1
Integrated bottom-up budge Natural ecosystems (including fires)		-31	-667	559	410) 248	572	70	3.5	140	449	-415	1272
	Mean vegetated upland+wetland ecosyste	-262	-799	200	313	194	432	NA			52	-605	632
	Data-driven, sum upland+wetland landcov	-230	-757	292	292	168	409	70	0	140	132	-588	841
	Model NEE and wetland abrupt thaw	-293	-842	109	NA			NA			-293	-842	109
	Model wetland CH4 flux and abrupt wetlan	d thaw			335	227	454				335	227	454
	Inland waters (from data-driven synthesis)	230	132	359	102	49	140	0.23	0.12	0.35	332	181	500
	Fires (from data-driven synthesis)	109	80	136	13	8 10	16	8.2	7.0	9.3	130	97	162
Top-down atmospheric	inverNatural ecosystems (including fires)	-587	-862	-312	162	2 119	194	11	-22	43	-414	-765	-74
	Fires (from one inversion system)	78	51	104	NA			NA			78	51	104
Anthropogenic fluxes	Fossil fuel, agriculture and waste, cement	73	56	89	57	7 41	72				129	98	161
Summary of all fluxes	Bottom-up plus anthropogenic	41	-611	648	467	290	644	70	3.5	140	579	-317	1432
	Top-down plus anthropogenic	-514	-806	-223	219	9 160	266	11	-22	43	-285	-668	86
	Mean bottom-up/top-down plus anthropogenic	-237	-708	212	343	3 225	455	40	-9.3	92	147	-492	759