

Institutional Dynamics Impact the Response of Urban Socio-Hydrologic Systems to Supply Challenges

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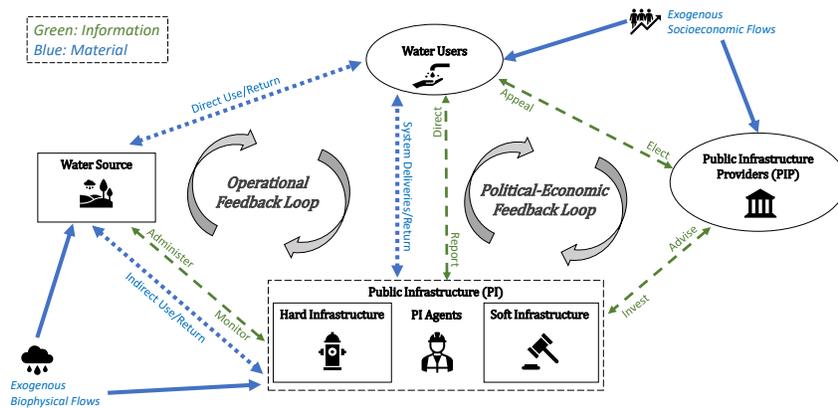
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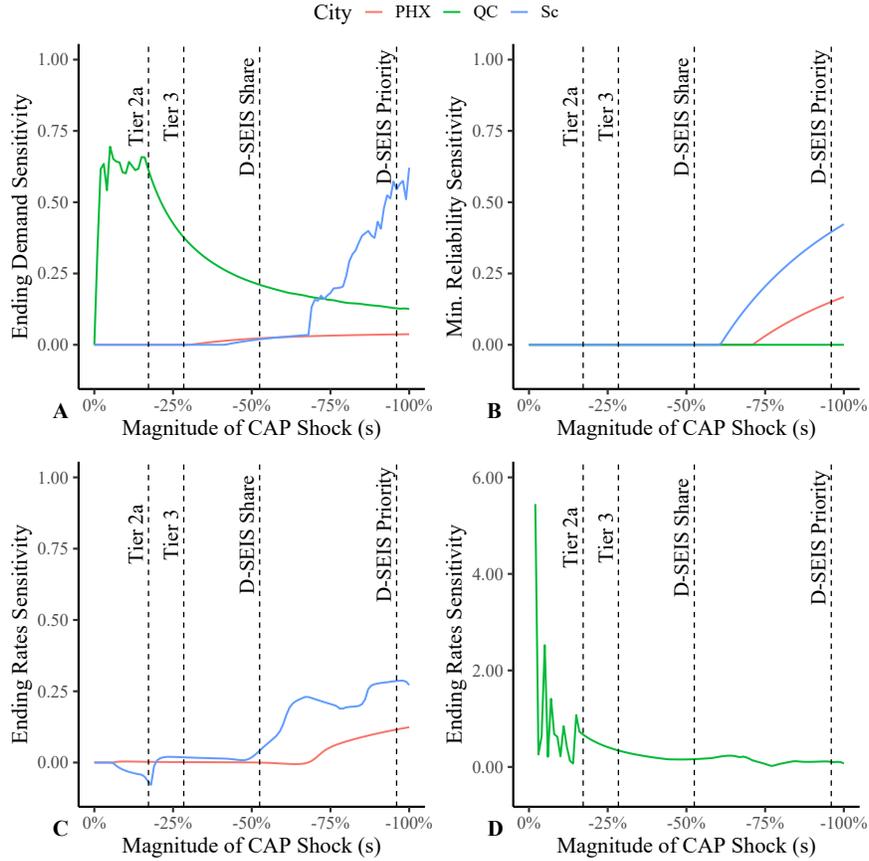
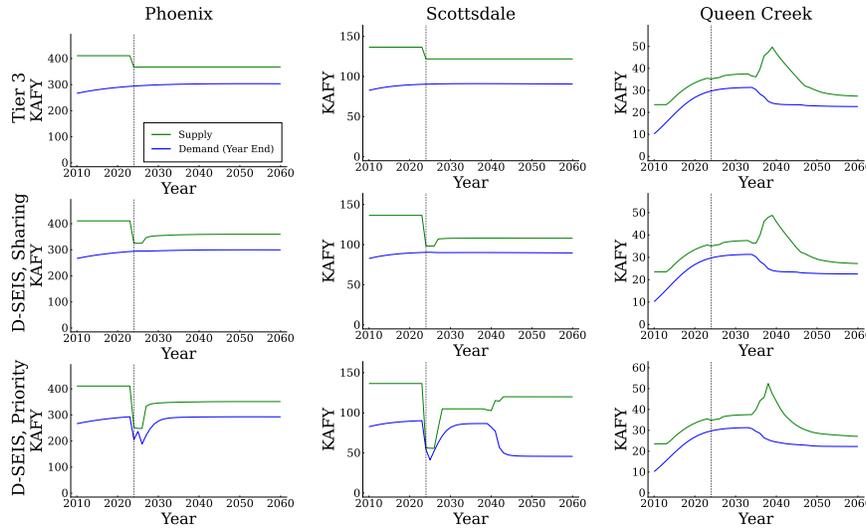
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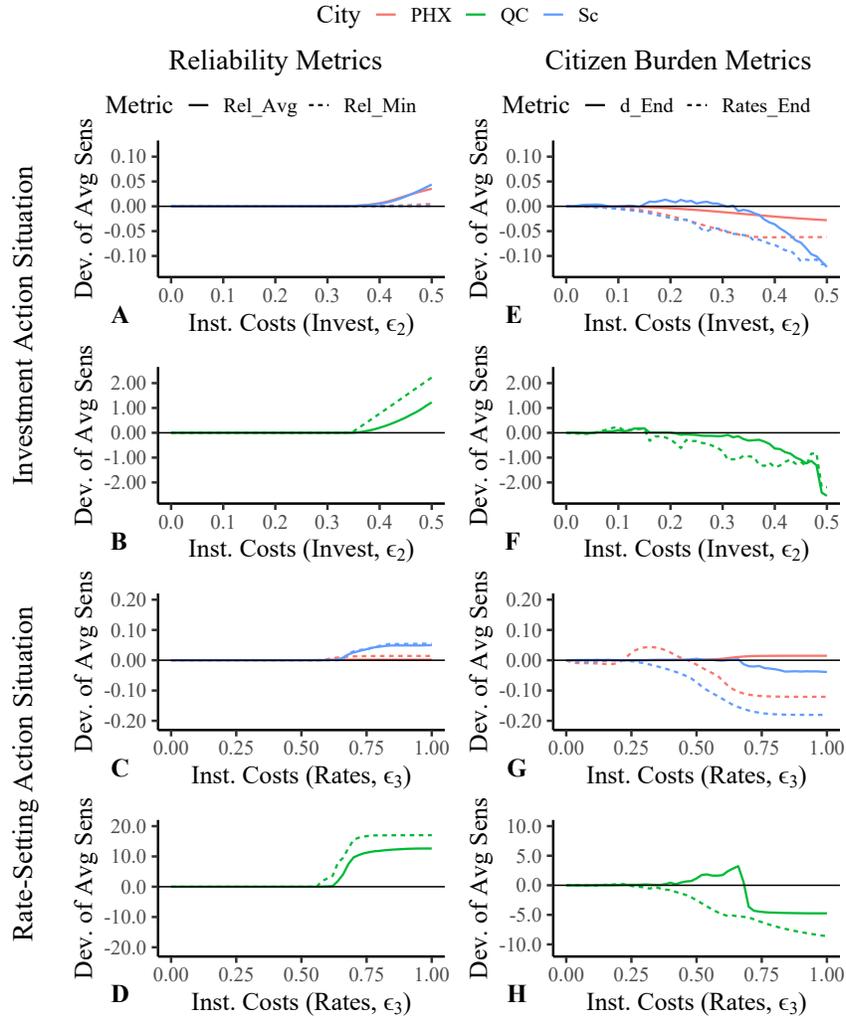
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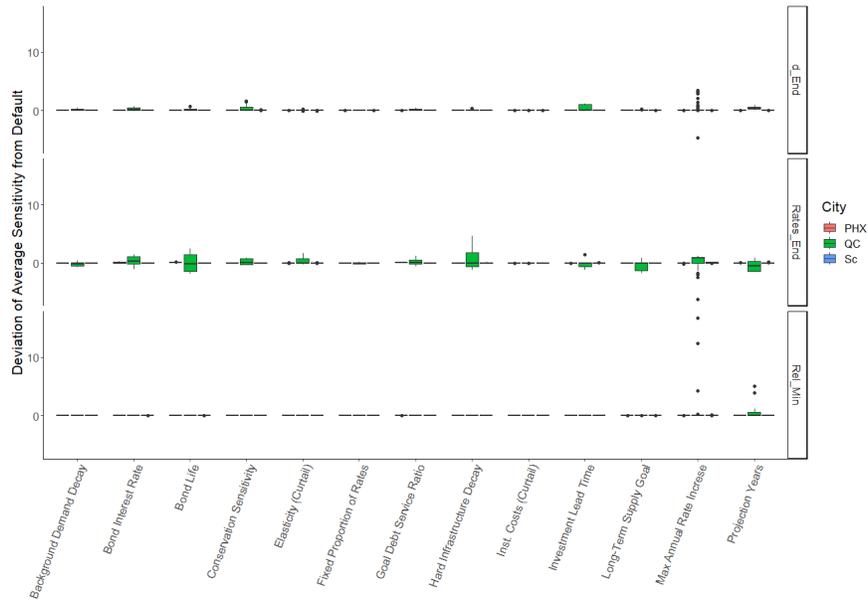
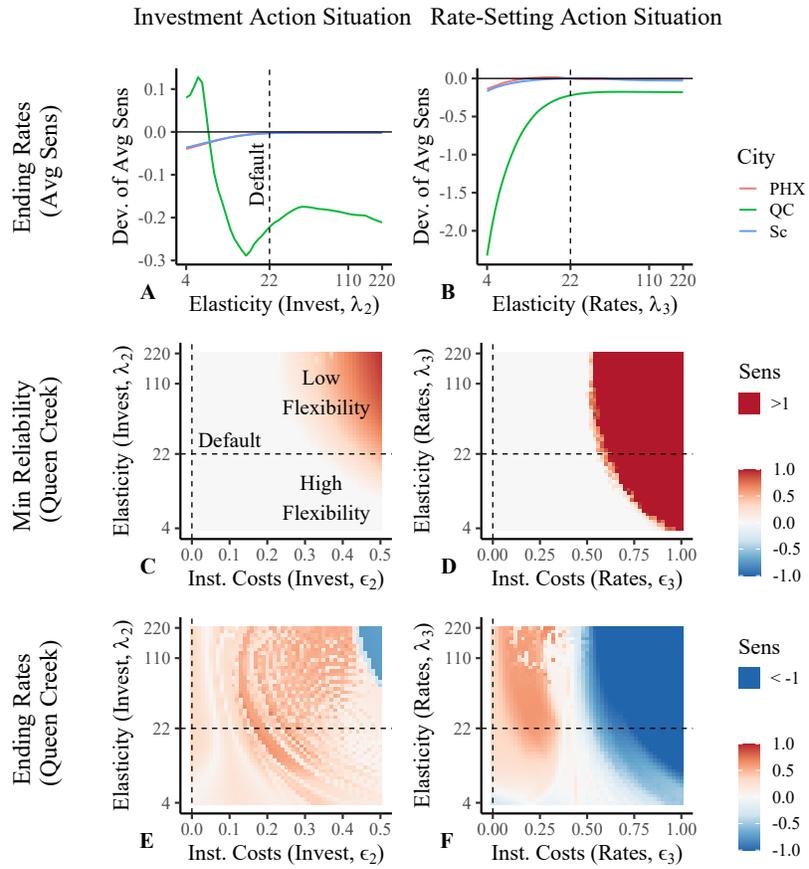
Abstract

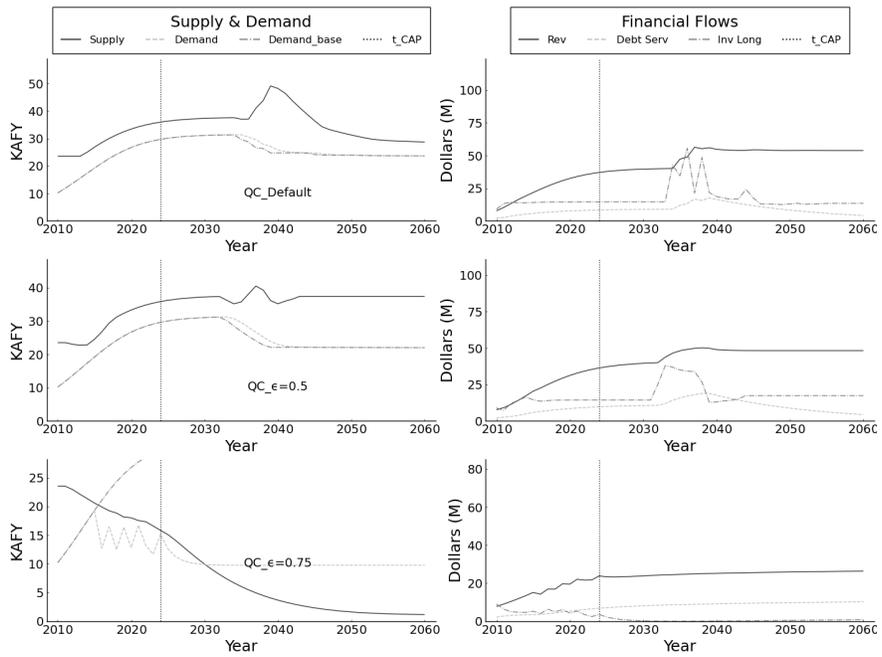
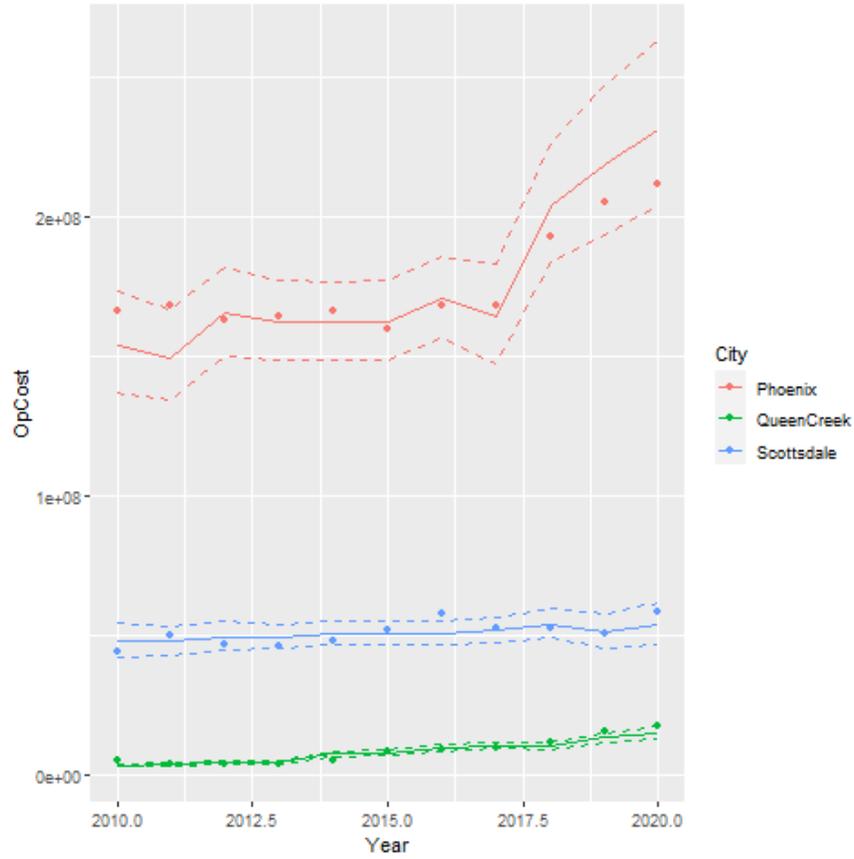
Designing urban water systems to respond to the accelerating and unpredictable changes of the Anthropocene will require changes not only to built infrastructure and operating rules, but also to the governance arrangements responsible for investing in them. Yet, inclusion of this political-economic feedback in dynamic models of infrastructure systems and socio-hydrology has significantly lagged behind operational feedback concerns. We address this gap through a dynamical systems application of the Coupled Infrastructure Systems (CIS) Framework, which provides the conceptual building blocks for analyzing social-ecological systems through various classes of infrastructure and the flows of material and information among them. In the model, political-economic feedback involves three decisions - infrastructure investment, rate-setting, and short-term demand curtailment - and each decision is constrained by institutional friction, the aggregation of decision and transaction costs associated with taking action. We apply the model to three cities in the Phoenix Metropolitan Area to compare how institutional friction interacts with a city's water resource portfolio and financial position to determine its sensitivity to reductions in Colorado River water availability. We find that the slowing effect of institutional friction on investment and rate-setting decisions can increase the sensitivity of a city's supply, but it can also promote objectives that compete with over-response (e.g., rate burden). The effect is dependent on the initial operating capacity of the CIS and flexibility within the institutions, highlighting the need to consider political-economic and operational feedback together when evaluating infrastructure systems.











1 **Institutional Dynamics Impact the Response of Urban**
2 **Socio-Hydrologic Systems to Supply Challenges**

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8 **Key Points:**

- 9
- 10 • Having sufficient supply redundancy can outweigh the negative effect of slow in-
 - 11 stitutions on a city's ability to address supply shocks.
 - 12 • The supplies of cities with institutions that require more stress to act are more
 - 13 sensitive to shocks, but their rates are less sensitive.
 - 14 • Adding flexibility to institutions can ease the burden of large investments on ratepay-

Abstract

Designing urban water systems to respond to the accelerating and unpredictable changes of the Anthropocene will require changes not only to built infrastructure and operating rules, but also to the governance arrangements responsible for investing in them. Yet, inclusion of this *political-economic feedback* in dynamic models of infrastructure systems and socio-hydrology has significantly lagged behind *operational* feedback concerns. We address this gap through a dynamical systems application of the Coupled Infrastructure Systems (CIS) Framework, which provides the conceptual building blocks for analyzing social-ecological systems through various classes of infrastructure and the flows of material and information among them. In the model, political-economic feedback involves three decisions - infrastructure investment, rate-setting, and short-term demand curtailment - and each decision is constrained by institutional friction, the aggregation of decision and transaction costs associated with taking action. We apply the model to three cities in the Phoenix Metropolitan Area to compare how institutional friction interacts with a city's water resource portfolio and financial position to determine its sensitivity to reductions in Colorado River water availability. We find that the slowing effect of institutional friction on investment and rate-setting decisions can increase the sensitivity of a city's supply, but it can also promote objectives that compete with over-response (e.g., rate burden). The effect is dependent on the initial operating capacity of the CIS and flexibility within the institutions, highlighting the need to consider political-economic and operational feedback together when evaluating infrastructure systems.

Plain Language Summary

Urban water systems must grapple with accelerating social and environmental change that requires them to not only consider future infrastructure needs, but also, the configuration of decisions responsible for infrastructure investment. Unfortunately, inclusion of *political-economic feedback* has lagged behind *operational* feedback in models that examine water systems response to changing environments. We present a modeling approach to trace the flow of water, information, and investment in a general urban water system that must make three annual decisions: infrastructure investment, rate-setting, and short-term demand curtailment. Each decision is influenced by costs to taking action and the flexibility involved in setting action magnitudes. We apply the model to three cities in the Phoenix Metropolitan Area to compare how these institutional constraints interact with existing infrastructure and finances to affect their sensitivity to reductions in Colorado River water. We find that when institutional barriers to action increase, cities are more sensitive to supply shocks, but such barriers can benefit other objectives like rate burden. The effect is dependent on the presence of redundant supplies, demand growth, and decision-making flexibility, highlighting the need to consider both political-economic and operational concerns when evaluating water systems.

1 Introduction

Urban water systems seek to provide a stable supply of water at low costs to their users. Like most infrastructure systems, to achieve these goals, they must address variation of multiple environmental (e.g., streamflow change) and socioeconomic (e.g., population growth) drivers (Chester et al., 2020; Treuer et al., 2017). Many cities have coped with these variations through a combination of built infrastructure and soft infrastructure, i.e., institutions (Padowski et al., 2016; Hoffmann et al., 2020), but growing urban populations and climate change threaten their ability to continue meeting demand (Gleick, 2018; Larsen et al., 2016; McDonald et al., 2014).

Yet, transitioning urban water systems is difficult. Many systems were designed during a century of relative stability (Chester et al., 2019) and empirical studies confirm that the actors governing urban water systems are hesitant to change given significant

65 social and climatic stressors (Hornberger et al., 2015; R. Brown et al., 2011; Hughes &
 66 Mullin, 2018). For decades, experts have offered alternative investment pathways, includ-
 67 ing moving from “supply-focused,” hard infrastructure towards “soft paths” (e.g., Her-
 68 ington et al., 2013; Christian-Smith et al., 2012; Gleick, 2002; Hornberger et al., 2015), but
 69 without well-developed understanding of the political-economic system responsible for
 70 bringing about desired changes, water management transition can resemble a Sisyphean
 71 task (van de Meene & Brown, 2009; Pahl-Wostl et al., 2010; Brelsford et al., 2020; Winz
 72 et al., 2009).

73 Working towards this broader perspective, socio-hydrology has elevated the “two-
 74 way feedback” between social and hydrological systems (Di Baldassarre et al., 2013; Siva-
 75 palan et al., 2012), but due to its epistemological grounding in natural science, it has strug-
 76 gled to fully capture the layers of feedback *within* human systems (Xu et al., 2018; Lu
 77 et al., 2018; Blair & Buytaert, 2016). In this paper, we contribute an additional distinc-
 78 tion in human water-systems by introducing both operational and political-economic feed-
 79 back. Using the Coupled Infrastructure Systems (CIS) Framework (Anderies et al., 2016),
 80 we conceptualize urban water systems as arrangements of multiple infrastructures (hard
 81 and soft) that process flows of water and information. All CISs, must deal with varia-
 82 tion in key inflows through the interaction between infrastructures and human agents,
 83 creating multiple forms of feedback control to govern the system (Rodriguez et al., 2011;
 84 Anderies, 2015a). We group these feedback processes into *operational feedback*, which
 85 concerns the behavior of humans and infrastructures within a specific infrastructure sys-
 86 tem (e.g., operating a reservoir, implementing use restrictions, etc.), and *political-economic*
 87 *feedback*, which concerns the investment in or alteration of the infrastructure system.

88 Analyzing the two forms of feedback provides deeper insight into the *robustness* of
 89 urban water systems. By robustness, we refer to the preservation of a system’s desired
 90 performance to a specific input disturbance (the opposite of sensitivity) (Anderies et al.,
 91 2013). Robust control is a common design approach in water systems engineering (e.g.,
 92 Herman et al., 2015; Kasprzyk et al., 2013; Herman et al., 2020; Trindade et al., 2020),
 93 but a gap remains regarding the inclusion of political-economic feedback processes. While
 94 Herman et al. (2020) recognize a considerable degree of “endogenous uncertainty” as-
 95 sociated with human-environmental system response that can significantly alter the find-
 96 ings of dynamic models, they situate this uncertainty in the way the human-environmental
 97 system responds to a selected policy or disturbance (operational feedback), not the pol-
 98 icy selection process itself (political-economic feedback).

99 Current approaches treat outputs from political-economic processes as (i) param-
 100 eters to be sampled (e.g., Krueger et al., 2019; Gober et al., 2010; Rehan et al., 2015;
 101 Koutiva & Makropoulos, 2016), (ii) solutions to be optimized (e.g., Kasprzyk et al., 2013;
 102 Cohen & Herman, 2021; Trindade et al., 2020), or when they are endogenous, (iii) out-
 103 puts of strict decision rules that implicitly assume rational actor theory (e.g., Kanta &
 104 Zechman, 2014; Munepeerakul & Anderies, 2020; Bakarji et al., 2017; Baeza et al., 2019)
 105 or (iv) outputs of a single dynamical equation that ignores the layered networks of in-
 106 stitutions responsible for filtering and translating information into action (e.g., Di Bal-
 107 dassarre et al., 2013; Elshafei et al., 2014; Mazzoleni et al., 2021; Garcia et al., 2016).
 108 Meanwhile, ample empirical evidence suggests that these outputs endogenously evolve
 109 with the socio-hydrologic system (e.g., Sullivan et al., 2017; Garcia et al., 2019; Treuer
 110 et al., 2017; R. R. Brown et al., 2009) and are not necessarily reflective of rational, long-
 111 term, goal-directed behavior (e.g., Hansen & Mullin, 2022; Mullin & Hansen, 2022; Horn-
 112 berger et al., 2015; Winz et al., 2009).

113 In this manuscript, we investigate the scaffolding of political-economic feedback,
 114 institutions. Institutions mediate the identification of perceived problems and their trans-
 115 lation into actions by constraining the space of possible actions (Ostrom, 2011). Just as
 116 networks of hard infrastructure steer the flow of water, so too do networks of soft infras-
 117 tructure steer the flow of information and actions at both the operational (e.g., use re-

118 strictions) and political-economic level (e.g., city goals) (Pahl-Wostl et al., 2010; Anderies
 119 et al., 2016; Brelsford et al., 2020). Socio-hydrologists and others concerned with human-
 120 water systems recognize the importance of institutions and have developed ways to in-
 121 corporate their consideration into water system operation (e.g., Brelsford et al., 2020;
 122 Lund, 2015; Konar et al., 2019; Yu et al., 2017; Herman et al., 2020; Trindade et al., 2020),
 123 but the way institutions shape information processing in political economic processes re-
 124 mains under-studied.

125 One tool from the field of policy studies used to address the role of institutions on
 126 information processing is the concept of *institutional friction*. Institutional friction refers
 127 to the emergent information, decision, and transaction costs within a policy system that
 128 incentivize inaction among relevant decision-making actors even when error, the gap be-
 129 tween goals and reality, exists (Workman et al., 2009; Jones & Baumgartner, 2005a). While
 130 institutional friction has been used in abstract dynamical system models of information
 131 processing (Jones & Baumgartner, 2005a), it has not previously been applied in a model
 132 of infrastructure investment.

133 This manuscript presents a novel approach to incorporate institutional friction into
 134 analysis of urban water coupled infrastructure systems implemented in the Urban Wa-
 135 ter Infrastructure Investment Model (UWIIM). Coupled to a simple urban water dynam-
 136 ical system that captures operational feedback in balancing supply and demand, we de-
 137 fine three information processing nodes in the political-economic system as closed-loop
 138 controllers (Anderies et al., 2007) whose attention is distorted by institutional friction.
 139 These nodes are long-term infrastructure investment, short-term demand curtailment,
 140 and rate-making.

141 To demonstrate the UWIIM in context, we compare the urban water systems of
 142 three cities in the Phoenix Metropolitan Area (PMA): Phoenix, Scottsdale, and Queen
 143 Creek. The area faces a common hydrologic challenge: declining availability of Colorado
 144 River water due to basin-wide aridification and chronic over-allocation (Udall & Over-
 145 peck, 2017; Overpeck & Udall, 2020), and rapid population growth (Gober et al., 2016;
 146 Healy, 2021). We examine the following question: *how does institutional friction in in-*
 147 *vestment and rate-making decisions interact with a city’s water resource portfolio and*
 148 *financial position to affect the sensitivity of its water supply, demand, and ratepayer bur-*
 149 *den to sudden, long-term reductions in water availability.*

150 2 Model

151 The Coupled Infrastructure Systems (CIS) Framework (Anderies et al., 2016) con-
 152 ceptually grounds the model. Nodes in a CIS are actors, resource users or public infras-
 153 tructure providers (PIPs), and infrastructures of multiple types, including natural in-
 154 frastructure (e.g., a watershed), and public infrastructures (hard/built and soft/institutional).
 155 With Yoon et al’s (2022) typology, PIPs are governing actors, users are utilizing actors,
 156 and public infrastructure (PI) agents are provisioning actors. The PI agents implement
 157 directives from PIPs and operate the infrastructure system (Anderies et al., 2019). The
 158 CIS allows us to map flows of material (e.g., water) and information between actors and
 159 infrastructures to visualize feedback loops of interest that govern system dynamics (Anderies,
 160 2015b).

161 *Action situations* are spaces of interaction between actors, institutions, and their
 162 environment that produce relevant outcomes (Ostrom, 2011), and they are the primary
 163 units of human information processing in a CIS (Anderies et al., 2016). Taken together,
 164 the CIS is an information and material processing network that endogenously controls
 165 responses to variation in exogenous flows. This is what the CIS Framework defines as
 166 *governance* (Anderies, 2015b) and it consists of two loops of interest: operational feed-
 167 back and political-economic feedback.

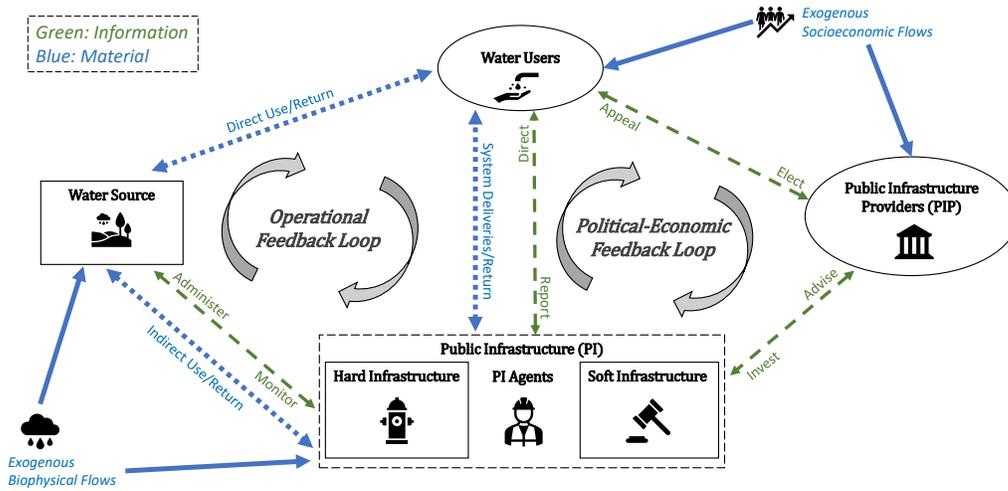


Figure 1. Summary of actors (ovals), infrastructure (rectangles), and flows (blue dotted and green dashed are internal material and information flows, respectively) in a general urban water coupled infrastructure system (UW-CIS). Exogenous material flows are depicted by solid blue lines. Two overall feedback loops govern the system: Operational Loop and Political-Economic Loop.

168 The UW-CIS (Figure 1) begins with a portfolio of water sources, often natural in-
 169 frastructure, that catch, transmit, and store precipitation and runoff. In urban systems,
 170 characterized by high population density and occupational specialization (West, 2017),
 171 users are supplied through large public infrastructure systems. These public infrastruc-
 172 ture systems store, treat, and distribute water and are operated by PI agents. Of course,
 173 this operation is quite complex and requires the PI agents and users to process informa-
 174 tion from the infrastructure system given PIP directives. This *operational feedback loop*
 175 is often the focus of conventional urban water models.

176 The political-economic loop can take on many forms given the structure of system
 177 ownership (Deslatte et al., 2021). Because the infrastructure system is shared, a PIP (city
 178 council, special district, company, etc.) is responsible for investing in the system. With
 179 their authority, the PIP makes investment decisions taking into account information from
 180 the PI agents and users. Together, the operational and political-economic feedback loops
 181 allow the UW-CIS to process material and information to ensure desired performance
 182 goals are met despite biophysical and socioeconomic variation.

183 2.1 Model Overview

184 The Urban Water Infrastructure Investment Model (UWIIM) is a discrete time,
 185 dynamical system that models the flow of water, investment, and information in a styl-
 186 ized UW-CIS with annual time steps (Figure 2a). Refer to the Supporting Information
 187 for detailed definition.

188 The system consists of a homogenous user population of size P_t with per capita de-
 189 mand d_t and a network of infrastructures that use water from available sources (indexed
 190 by i) to meet demand. The number of sources included in an application of the UWIIM
 191 is flexible to allow for case-specific representations (e.g., our PMA cases have access to

three possible sources, so $i \in [1, 3]$). Each year, the UW-CIS receives water inflows, $Q_{i,t}$, from source i that can be stored, used, or released back into the environment.

Each source is characterized by two time-varying parameters: mean annual streamflow ($\mu_{i,t}$) and its coefficient of variation ($c_{i,t}^v$). In our simple demonstration case, we are only interested in mean changes, so we set $c_{i,t}^v$ to near zero. For each source, the system has storage infrastructure of volumetric capacity, $\bar{V}_{i,t}$. $V_{i,t}$ refers to the volume of water stored from i at the beginning of t ($V_{i,t} \leq \bar{V}_{i,t}$). Outflows from source i , $O_{i,t}$, come in two forms (indicated by superscript): city withdrawals to satisfy demand, $O_{i,t}^d$, and releases (e.g., flood control), $O_{i,t}^f$. Two types of hard infrastructure determine how much of $O_{i,t}^d$ can be delivered to users: processing and delivery infrastructure. Processing capacity is source-specific, $w_{i,t}$, and is defined as the proportion of the city's maximum available water (i.e., total volumetric capacity plus expected annual inflow, $\bar{V}_{i,t} + \mu_{i,t}$) that it can process in year t , taking into account pumping and treatment infrastructure. Delivery efficiency, η_t , is a proportional coefficient that refers to, on average, the amount of demand the city can meet per unit of processed water from all of its sources, taking into account lost water and re-use. In essence, we treat re-use as an increase to delivery efficiency because the city is able to deliver more water to users per unit of water extracted from a surface or groundwater source (if net re-used, $\eta_t > 1$). Two forms of soft infrastructure operate the system, demand management and rate policy, and two exogenous drivers, defined by the model user, can disrupt the UW-CIS: population growth and water inflow. The UWIM stores the system's state variables in vector x_t .

The UWIM models the political-economic feedback loop through a network of controller feedback loops that make changes to the infrastructure system to meet supply and financial goals (2b). We model three representative action situations as controllers: short-term curtailment, long-term investment, and rate-setting. Each converts perceived error, $e_{j,t}$, between the desired system state, γ_j , and x_t (assuming perfect measurement) into actions, $u_{j,t}$, through an algorithm, $G_j(e_{j,t}, x_t)$. We account for the distorting influence of institutional friction (Jones & Baumgartner, 2005a) through the use of an attention filter before $G_j(e_{j,t}, x_t)$ (Figure 2b).

2.2 Operational Feedback Loop

The operational feedback loop seeks to meet demand with available supply, S_t . S_t is the annual volume of water that can be delivered to users and is the product of water available in a given year, A_t , and η_t . A_t is the sum of available water from each source i in t , $A_{i,t}$. Each $A_{i,t}$ is the minimum of the amount of water from source i the city is legally entitled to use in year t , $A_{i,t}^l$, and the water that can be technically processed by the city from source i in year t , $A_{i,t}^w$. $A_{i,t}^l$ translates the water from source i in storage, $V_{i,t}$, and flowing in, $Q_{i,t}$, into legal availability. We assume each city has legal access to proportions a_i^v and a_i^q of $V_{i,t}$ and $Q_{i,t}$, respectively. $A_{i,t}^w$ converts the processing capacity, $w_{i,t}$, back into volumetric units by multiplying it by its normalizing denominator, $(\bar{V}_{i,t} + \mu_{i,t})$ (Table 1).

For population growth, we opt for a basic logistic definition with carrying capacity κ and intrinsic growth rate r (Garcia et al., 2016; Liu et al., 2015; Elshafei et al., 2014). d_t is the per capita demand at the start of t before taking into account curtailment enacted in t , $u_{1,t}$. \tilde{d}_t is the actual per capita demand after subtracting $u_{1,t}$. When calculating the per capita demand for the next year, d_{t+1} , we account for conservation rebound, the return of demand to baseline demand, \bar{d}_t , a separate state variable, with a common equation in socio-hydrology that takes a community sensitivity parameter α (Garcia et al., 2016; Gonzales & Ajami, 2017). By baseline ($\bar{d}_t \geq d_t$), we refer to the demand if no curtailments had ever occurred, reflecting hard infrastructure (e.g, efficiency improvements) and long-term policies (e.g., building codes) (Garcia et al., 2016). We use the general infrastructure dynamical equation for \bar{d}_t (Equation 1) where δ represents the back-

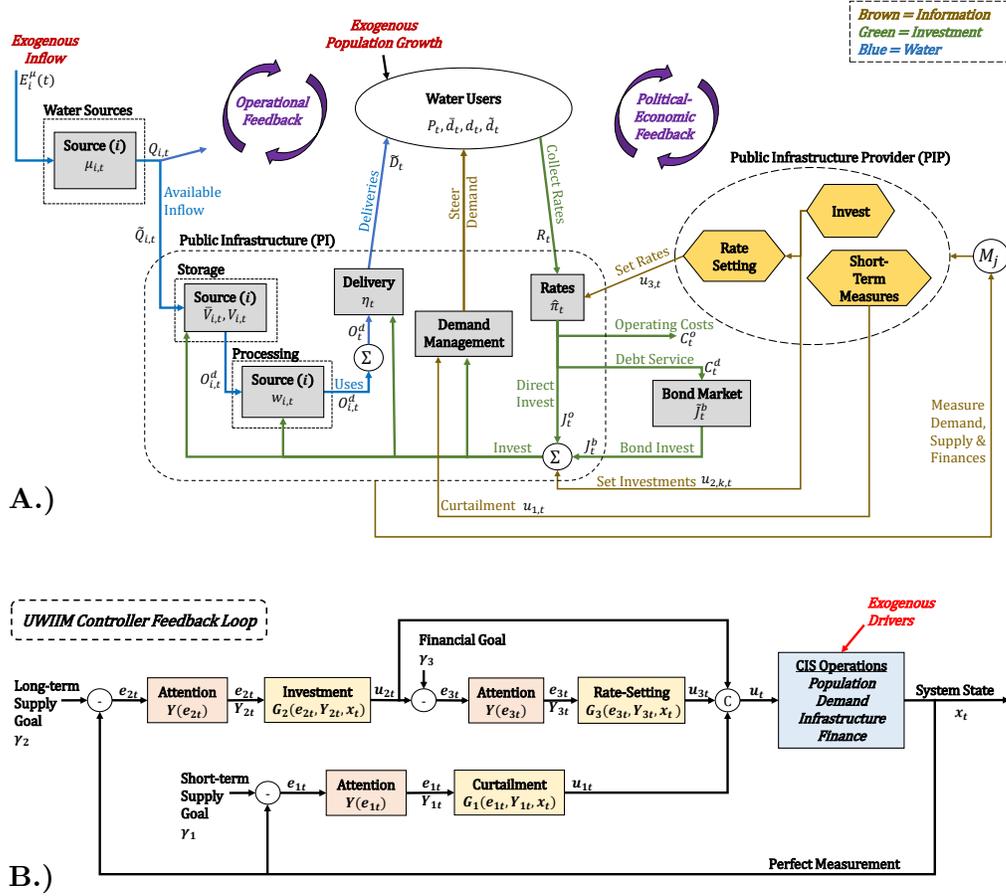


Figure 2. Overview of the Urban Water Infrastructure Investment Model (UWIIM). CIS representation (a) maps flows of information, water, and investment. Grey boxes depict infrastructure and yellow hexagons represent action situations. See Table 1 for variable definitions. The network (b) plots the order of information processing in the political-economic feedback loop with the attention (pink) and response (yellow) steps of the controllers governing the CIS (blue).

243 ground conservation rate (Garcia et al., 2016). Each per capita demand variable can be
 244 translated to total system demand by multiplying it by P_t (Table 1).

245 **2.2.1 Infrastructure**

246 The vector $I_{k,t}$ contains the capacities for each infrastructure, k . Its size depends
 247 on the case being modeled (e.g., in our PMA cases, $k \in [1, 7]$). Some infrastructures
 248 like w are source specific, in which case, there is a unique capacity for each source. $I_{k,t}$
 249 decays each year according to a decay parameter, δ_k , and increases with investment, $u_{2,k,t}$
 250 (Muneepeerakul & Anderies, 2020, 2017). We add $E_k(t)$ to capture exogenous drivers
 251 that can be defined by the model user in scenario definition (e.g., our Colorado River
 252 scenarios). The general difference equation is as follows,

$$I_{k,t+1} = (1 - \delta_k)I_{k,t} + H_k(u_{2,k,t}, x_t; \tau_k^i) + E_k(t) \quad (1)$$

253 We represent investments as increases to capacity rather than dollar amounts. To
 254 keep units common, $u_{2,k,t}$, is in volumetric flow units (e.g., volume of water that can be
 255 processed in a given year). Each variable stored in $I_{k,t}$ is not given in the same units,
 256 so the function $H_k(u_{2,k,t}, x_t)$ (Supporting information) transforms the volumetric $u_{2,k,t}$
 257 into the relevant units for $I_{k,t}$. Because infrastructure investments often carry lag time
 258 between investment and implementation, the τ_k^i parameter specifies the delay between
 259 investment and infrastructure change.

260 **2.2.2 Water Balance**

261 The stored water from each source i at time t , $V_{i,t}$, increases with the city’s legally
 262 entitled inflow, $a_i^q Q_{i,t}$, and decreases with city withdrawals, $O_{i,t}^d$, and releases, $O_{i,t}^f$ (Ta-
 263 ble 1). $O_{i,t}^d$ reflects source priority (in order i) to calculate the water needed to meet de-
 264 mand with,

$$O_{i,t}^d = \min \left[\left(1 - \sum_{i' > i} \theta_{i'} \right) \left(\frac{\tilde{D}_t}{\eta_t} - \sum_{i' < i} O_{i',t}^d \right), A_{i,t} \right] \quad (2)$$

265 where θ_i specifies the minimum proportion of demand that must be met with source i
 266 regardless of higher priority availability. We calculate $O_{i,t}^f$ such that $V_{i,t} \leq \bar{V}_{i,t}$.

267 **2.2.3 Finances**

268 The PIP can only raise revenue through rates paid by users. The UWIIM rate struc-
 269 ture allows for per-user (connection) charges and per-use (volumetric) charges, based on
 270 \tilde{d}_t . When setting rates, we assume the PIP knows the population but not future curtail-
 271 ments, so it predicts demand will resemble the baseline, \bar{d}_t , in its target per-capita rev-
 272 enue, $\hat{\pi}_t$. To calculate the actual revenue generated, R_t , we simplify the split between
 273 volumetric and connection components with the assumption that the PIP seeks the same
 274 proportion of revenue from connection charges, β_p^π (Table 1).

275 R_t covers operating costs (C_t^o), debt service requirements (C_t^d), and a portion of
 276 infrastructure investment (J_t^o). Operating costs, C_t^o , account for volumetric and user economies
 277 of scale with fitted exponents, z_d and z_p respectively, and normalizing constant g_o . We
 278 use \bar{D}_t in the equation because the infrastructure needed to meet baseline demand is not
 279 abandoned during curtailment (Table 1).

280 J_t^o refers to investments that come directly from R_t as opposed to bonds, J_t^b . We
 281 assume bond life, τ_b , and interest rate, i_b , remain constant over time and that the PIP
 282 pays off debt in equal increments, so to keep track of accrued debt, we use a state vari-

Table 1. Dynamic Variables in the UWIIM

Name	Symbol	Definition
<i>Demand</i>		
*Population	P_t	$P_{t+1} = P_t + r(1 - \frac{P_t}{\kappa_t})$
*Per Capita Demand (Year Start)	d_t	$\tilde{d}_t \left[1 + \alpha \left(1 - \frac{\tilde{d}_t}{d_t} \right) \right]$
Per Capita Demand (Year End)	\tilde{d}_t	$\tilde{d}_t = d_t - u_{1,t}$
Total System Demand	$\bar{D}_t, D_t, \tilde{D}_t$	$\bar{D}_t = \bar{d}P_t, D_t = d_tP_t, \tilde{D}_t = \tilde{d}_tP_t$
<i>Supply & Water Balance</i>		
Supply	S_t	$S_t = \eta_t \sum_i A_{i,t}$
Available Water (Total)	$A_{i,t}$	$A_{i,t} = \min(A_{i,t}^l, A_{i,t}^w)$
Available Water (Legal)	$A_{i,t}^l$	$A_{i,t}^l = a_i^v V_{i,t} + a_i^q Q_{i,t}$
Available Water (Tech.)	$A_{i,t}^w$	$A_{i,t}^w = w_{i,t}(\bar{V}_{i,t} + \mu_{i,t})$
City Withdrawal	$O_{i,t}^d$	Equation 2
Flood Release	$O_{i,t}^f$	$O_{i,t}^f = \max[V_{i,t} + a_i^q Q_{i,t} - O_{i,t}^d - \bar{V}_{i,t}, 0]$
*Stored Water	$V_{i,t}$	$V_{i,t+1} = V_{i,t} + a_i^q Q_{i,t} - O_{i,t}^d - O_{i,t}^f$
*Source Inflow	$Q_{i,t}$	Scenario Definition
<i>Infrastructure: $I_{k,t} = (\bar{d}_t, \eta_t, \bar{V}_{i,t}, w_{i,t}, \mu_{i,t})$</i>		
*Per Capita Demand (Baseline)	\bar{d}_t	Equation 1
*Delivery Efficiency	η_t	Equation 1
*Storage Capacity	$\bar{V}_{i,t}$	Equation 1
*Processing Capacity	$w_{i,t}$	Equation 1
*Mean Inflow	$\mu_{i,t}$	Equation 1 & Scenario
<i>Finances</i>		
*Per Capita Rate Policy	$\hat{\pi}_t$	$\hat{\pi}_{t+1} = \hat{\pi}_t + u_{3,t}$
Revenue	R_t	$R_t = P_t \hat{\pi}_t (\beta_p^\pi + (1 - \beta_p^\pi) \frac{\tilde{d}_t}{d_t})$
Operating Costs	C_t^o	$C_t^o = g_o P_t^{z_p} \bar{D}_t^{z_d}$
Debt Service	C_t^d	$C_t^d = (1 + \tau_b i_b) \tilde{J}_t^b$
*Average Bond-Sourced Investment	\tilde{J}_t^b	$\tilde{J}_{t+1}^b = \frac{1}{\tau_b} [(\tau_b - 1) \tilde{J}_t^b + J_t^b]$
Bond-Sourced Investment	J_t^b	$J_t^b = J_t - J_t^o$
Direct-Sourced Investment	J_t^o	$J_t^o = \min(J_t, R_t - C_t^o - C_t^d)$
Investment (Dollars)	J_t	$J_t = \sum_k F_k(u_{2,k,t})$
<i>Action Situation Controllers</i>		
Signal (Curtailment)	M_1	$M_{1,t} = \frac{S_t}{\tilde{D}_t}$
Signal (Investment)	M_2	$M_{2,t} = \frac{\hat{S}_{t+\tau_p}}{\tilde{D}_{t+\tau_p}}$
Signal (Rates)	M_3	$M_{3,t} = \frac{\hat{R}_{t+1} - \hat{C}_{t+1}^o}{\hat{C}_{t+1}^d}$
Error	$e_{j,t}$	$e_{j,t} = \gamma_j - M_{j,t}$
Attention	$Y_{j,t}$	$Y_{j,t} = \frac{1}{1 + \exp(-\lambda_j(e_{j,t} - \epsilon_j))}$
Response (Curtailment)	$u_{1,t}$	$u_{1,t} = d_t \alpha \left(1 - \frac{d_{min}}{d_t} \right) Y_{1,t}$
Response (Investment)	$u_{2,k,t}$	$\hat{u}_{2,k,t} = \beta_{k,t} e_{2,t} Y_{2,t} \tilde{D}_{t+\tau_p} + \hat{H}_{k,t}^m$, Check $J_t \leq \tilde{J}_t$
Response (Rates)	$u_{3,t}$	$\hat{u}_{3,t} = \frac{\epsilon_{3,t} Y_{3,t} \hat{C}_{t+1}^d}{\hat{P}_{t+1}}$, Check $u_{3,t} \leq \psi_r \hat{\pi}_t$

*Indicates a model state variable stored in x_t through model runs. Superscripts on variables indicate type (e.g., in J_t^b , b indicates investment dollars sourced from *bonds*).

able \tilde{J}_t^b to keep a τ_b -year moving average of the bond-sourced investment. The debt service paid in a particular year, C_t^d is then a function of \tilde{J}_t^b (Table 1).

Decisions to make capital investments can be quite complex because project details and city context vary greatly. We choose the following simplified treatment in our model. For each infrastructure type k , the cost, $J_{k,t}$, of increasing its volumetric capacity an amount, $u_{2,k,t}$, follows a power law relationship contained in $F_k(u_{2,k,t})$ and defined as,

$$J_{k,t} = F_k(u_{2,k,t}) = \begin{cases} g_k(\eta_t u_{2,k,t})^{z_k} & \text{delivery efficiency} \\ g_k(u_{2,k,t})^{z_k} & \text{all other infrastructure} \end{cases} \quad (3)$$

Because the marginal costs of delivery efficiency investments increase with added efficiency, $F_k(\cdot)$ for η uses the interaction $\eta_t u_{2,k,t}$ as the exponential base. If the PIP wishes to maintain $I_{k,t}$, the maintenance investment needed, $\hat{H}_{k,t}^m$ is calculated with $H_k^{-1}(\cdot)$, the inverse of the $H_k(\cdot)$ functions (Supporting Information).

2.3 Political-Economic Feedback Loop

In the the political-economic loop PIPs take actions to track goals. We model this through controller feedback loops (CFBLs), a helpful tool for modeling action situations in social-ecological systems (e.g., Anderies et al., 2007; Rodriguez et al., 2011; Anderies, 2015b). A CFBL consists of (i) a dynamical system that outputs state variables of interest (x_t) and (ii) a mechanism of closed-loop feedback control that measures the output, compares it to a goal, and responds with action back into the system (Anderies et al., 2007).

We model three representative action situations (Figure 2b): short-term curtailment measures, long-term investment, and rate-setting. The first two occur in parallel, and rates are set with investment information. Each, in reality, is a collection of many networked action situations (Deslatte et al., 2021; McGinnis, 2011), but we aggregate the “representative” action situation based on the emergent policy output of interest (e.g., an investment plan).

2.3.1 Goals, Signals & Error

$M_j(x_t)$ translates x_t into the relevant metric for each goal, γ_j (Table 1). Curtailment and investment examine the ratio of supply to demand. Curtailment only considers the current state, and investment makes projections τ_p years into the future, considering future population, groundwater volume, and already planned investments (Supporting Information). We assume the PIP has perfect information on population growth and approximates per capita demand at baseline, maintained infrastructure, and inflow at mean. For rate-setting, the PIP calculates the expected debt service capacity ratio of the next year given the current rate policy, expected operating costs, and planned investments. Error is the difference between γ_j and $M_j(x_t)$.

2.3.2 Controller Design & Representative Action Situations

All controllers follow two steps to produce an action, $u_{j,t}$, given $e_{j,t}$ and x_t (Figure 2b). The first step generates attention attributable to $e_{j,t}$, and the second step responds to the attention-mediated error with an action akin to a proportional controller (Rodriguez et al., 2011).

Each controller, being human-driven, is subject to disproportionate information processing challenges where actions pursued are not necessarily proportional to the actual scale of the error (Workman et al., 2009). To address this, socio-hydrologic models (e.g.,

326 Yu et al., 2017; Garcia & Islam, 2021; Garcia et al., 2016; Di Baldassarre et al., 2013;
 327 Mazzoleni et al., 2021) have often quantified the salience of water issues in a social sys-
 328 tem. Our attention variable is similar, but we (i) focus the attention relevant to a par-
 329 ticular action situation as opposed to the social system at-large and (ii) incorporate in-
 330 stitutional friction. Regarding institutional friction, our approach builds on the origi-
 331 nal Jones and Baumgartner (2005b) model but differs in two ways: (i) our function is
 332 continuous over error and (ii) response, $u_{j,t}$, is now a nonlinear transformation of error,
 333 $G_j(\cdot)$. We model attention, $Y_{j,t}$, as the proportion of error actually registered by the con-
 334 troller. If $Y_{j,t} = 1$, the error is fully perceived by the controller and a proportional re-
 335 sponse will be calculated. If $Y_{j,t} = 0$, there will be no response (Table 1).

336 Institutional information processing can be disproportionate because actions carry
 337 costs associated with gathering information, searching for a solution, coming to an agree-
 338 ment, and implementing the solution (Jones & Baumgartner, 2012; Jones et al., 2003;
 339 Workman et al., 2009). We aggregate this into an *institutional costs* parameter, ϵ_j . We
 340 use a *response elasticity* parameter, λ_j , to define the slope of the sigmoid as it converges
 341 to full attention (higher λ means steeper slope). In this sense, λ_j can represent the in-
 342 stitutional ambiguity or *flexibility* present within the system that distorts actor percep-
 343 tion of the system’s proximity to its action threshold, ϵ_j .

344 $G_j(e_{j,t}Y_{j,t}, x_t)$ has two parts: (i) calculate a potential response, $\hat{u}_{j,t}$ and (ii) account
 345 for saturation constraints, $sat_j(\hat{u}_{j,t}, x_t)$. Short-term curtailment follows prior socio-hydrology
 346 models with a logistic decay function to account for diminishing conservation returns as
 347 the system approaches minimum per capita demand, d_{min} (Garcia et al., 2016; Mazzoleni
 348 et al., 2021; Gonzales & Ajami, 2017). Because \hat{u}_{1t} accounts for d_{min} , there are no other
 349 saturation checks.

350 The investment response, $u_{2,k,t}$, is a vector of investments for each infrastructure
 351 k in volumetric units. $\hat{u}_{2,k,t}$ is a function of the perceived supply gap and maintenance
 352 needs, $\hat{H}_{k,t}^m$, which we assume will occur regardless of attention. The supply gap refers
 353 to needed expansion. We distribute the supply gap across k according to coefficients $\beta_{k,t}$,
 354 the infrastructure investment strategy of the city. Starting β_k values are parameters, but
 355 $\beta_{k,t}$ can vary in a given year if investments need to be re-distributed from infrastructures
 356 that have reached their maximum capacities (Supporting Information). Each year has
 357 a maximum investment capacity, \bar{J}_t , that takes into account the debt service implica-
 358 tions for the following year (Supporting Information). The saturation function compares
 359 the total investment dollars needed to fund $\hat{u}_{2,k,t}$, \hat{J}_t , to \bar{J}_t and outputs the minimum
 360 converted back into volumetric units as $u_{2,k,t}$.

361 Rate-making calculates the change in per capita rate policy needed to address the
 362 perceived revenue gap, $\hat{u}_{3,t}$. The saturation function checks that $\hat{u}_{3,t}$ is less than ψ_r , the
 363 maximum rate increase allowed by the socio-political system, to get $u_{3,t}$.

364 2.4 Performance Metrics & Sensitivity

365 We measure three performance metrics: (i) *reliability* (Ajami et al., 2008), or the
 366 proportion of \bar{D}_t met by S_t , (ii) rates ($\hat{\pi}_t$), and (iii) demand (\bar{d}_t). Over a model run, we
 367 aggregate the reliability, rate, and demand time series in four ways: mean reliability, min-
 368 imum reliability, rates at the end of the model run, and demand at the end of the model
 369 run. To measure robustness, we create sensitivity measures (Anderies et al., 2007; Ro-
 370 driguez et al., 2011), which are ratios of percent change in a performance metric to per-
 371 cent change in an input, for each of the four performance metric aggregations.

3 Case: Phoenix Metropolitan Area

We use the UWIIM to compare the sensitivity of water systems within the Phoenix Metropolitan Area (PMA) to change in Colorado River availability. Cities in the arid PMA rely on a mix of heavily regulated surface water delivered via canals and local groundwater. The Salt and Verde Rivers drain a watershed contained within central and east Arizona and combine into the Salt River before entering the PMA where a network of canals managed by the Salt River Project (SRP) distribute flows to city water treatment plants. SRP water follows a land-based seniority system, where it can only be delivered to lands demarcated by the Kent Decree in 1910 (Feller, 2007; Phillips et al., 2009; Salt River Project, 2017). Water from the Colorado River is conveyed to the PMA through the Central Arizona Project (CAP) canal. The Arizona Department of Water Resources (ADWR) regulates groundwater use within Arizona. The cities now face a situation where most renewable water has already been allocated, if not over-allocated, while experiencing a prodigious population and economic growth (Healy et al., 2021).

The city definitions used in this study are adaptations of a complex water resources planning environment onto the general structure we use in our simple model. This study is not intended to predict the future state of the three real cities or provide decision support, but to understand how institutions can alter the sensitivity of various operational contexts to environmental changes. The baseline parameter and initial conditions provide three different artificial testbeds to demonstrate this relationship, not replicate an actual city water system in the PMA.

3.1 PMA Model Definition

The three cities differ in their infrastructure capacity, financial position, demand profile, and water entitlements (Table 2). Designations of Assured Water Supply (DAWS), completed for most PMA cities in 2010, provide commonly formatted water resource information. We, thus, set 2010 as the start year for all model runs and use DAWS estimations to set model parameters and initial conditions related to the water resource portfolio of Scottsdale (ADWR, 2013) and Phoenix (ADWR, 2010). Queen Creek does not have a DAWS, but they have a Certificate of Assured Water Supply (CAWS) for their 2011 system (ADWR, 2011b) and the H2O, Inc. system (ADWR, 2011a) that they acquired in 2013. We triangulated these official filings with published supply and demand data from ADWR (ADWR, 2022a), the CAP sub-contract registry (CAP, 2022a), and city plans (City of Phoenix, 2011, 2021; Sunrise Engineering, Inc., 2017; Scottsdale Water, 2021). See Supporting Information for details.

We define legal and technical availability parameters for SRP ($i = 1$), CAP ($i = 2$), and groundwater ($i = 3$) with a few additional case-specific considerations. η_t accounts for re-use water (Supporting Information). We account for the land-specific nature of SRP rights by only allowing the city to use their base SRP allocation for a portion of their demand, ξ_1 , but additional SRP rights accrued after the Kent Decree can be used throughout the city. We account for the priority of CAP water entitlements possessed by each city in our CAP shock scenarios (Supporting Information). We assume surface water storage is negligible, and Phoenix and Scottsdale can fully process their available surface water. We set Queen Creek’s surface processing capacity to near zero to calibrate the cost functions.

We estimate initial groundwater storage (V_0^g) with ADWR allocations, provided as a total volume to be used over 100 years, and long-term storage credits. Currently, Queen Creek depends on recharge support from the Central Arizona Groundwater Replenishment District (CAGRDR) to offset its groundwater use in excess of its allowance. However, modeling the future of CAGRDR is beyond the scope of this exploratory study (Ferris & Porter, 2019), so to examine Queen Creek’s non-CAGRDR potential, we add to

Table 2. City Variation in the PMA UWIIM Adaptation

Variable	PHX Value	Sc Value	QC Value	Units	Source
<i>Water Resources</i>					
CAP (Low) Approx.	37280	3306	4162	AFY	CAP
CAP (High) Approx.	147426	77794	495	AFY	CAP
SRP (Base) Approx.	220647	19000	0	AFY	D/CAWS
SRP (Added) Approx.	57300	8600	0	AFY	D/CAWS
Proportion of Demand on SRP Eligible Land	0.5	0.17	0	N/A	WRP
Groundwater Allowance	3699500	1290528	1474600	AF	D/CAWS
Groundwater Credits (Banked)	240989	93846	289535	AF	D/CAWS, QC (2017)
Processing Capacity (Surface)	Full	Full	Negligible	N/A	WRP
Processing Capacity (Ground)	43000	34827	24520	AFY	WRP
Delivery Efficiency	0.9742	1.0562	0.9339	N/A	ADWR
<i>Demand</i>					
Population	1458275	217943	32197	persons	CAFR
Population Intrinsic Growth Rate	0.088	0.143	0.240	N/A	CAFR
Population Carrying Capacity	1686528	242300	101553	persons	CAFR
Per Capita Demand	163.17	338.75	282.85	GPCD	ADWR
<i>Finances</i>					
Per Capita Rates	238.36	398.11	241.06	\$/person	CAFR
Average Debt Service	111.5	20	5.7	\$M	CAFR
Proportion of Rates from Fixed Fees	0.6322	0.3586	0.5801	N/A	UNC EFC

WRP = Water Resource Plan; CAFR = Comprehensive Annual Financial Report; D/CAWS = Designation/Certificate of Assured Water Supply; CAP = Central Arizona Project; ADWR = Arizona Department of Water Resources; UNC EFC = (UNC Environmental Finance Center, 2022).

422 their V_0^g the recent purchase of 175 KAF of groundwater credits and credits equal to their
423 estimated excess pumping prior to 2024 (Supporting Information).

424 We use city Comprehensive Annual Financial Reports (CAFRs) 2010-2021 trian-
425 gulated with the EPA's Safe Drinking Water Information System (US EPA, 2022), ADWR
426 reports (ADWR, 2021), and the Water Rates Dashboard for Arizona (UNC Environmen-
427 tal Finance Center, 2022) to fit population and financial parameters (Supporting Infor-
428 mation). Investment cost function parameters assume (i) investment dollars are split ac-
429 cording to average proportions of Phoenix Capital Improvement Plans (CIPs) 2010-2021
430 (City of Phoenix, 2022), (ii) $J_0^b \sim \bar{J}_0^b$, and (iii) $J_0^o \sim (\gamma_3 - 1)C_0^d$. The latter assump-
431 tion accounts for the fact that some cities do not start 2010 at γ_3 , so the J_0^o requirement
432 forces early rate increases in the model to bring the system to γ_3 . Demand management
433 is not reported in Phoenix CIPs, so we rely on the reported cost of the Southern Nevada
434 Water Authority's rebate program to infer its cost function (Supporting Information).
435 τ_k^i , has a baseline value of 3 years for hard infrastructure and 1 year for demand man-
436 agement.

β_k only relates to investments that expand infrastructure capacity, not maintenance. We assume Scottsdale and Phoenix do not need additional surface water treatment capacity ($\beta_{w_s} = 0$), and given that much of their reclaimed water is already committed to exchange agreements, we assume the first infrastructure priority would be groundwater processing followed by delivery efficiency and demand management (City of Phoenix, 2021; City of Scottsdale, 2022). Queen Creek has the opportunity to gain surface processing, but their recent plans suggest that they are leveraging almost exclusively groundwater (Sunrise Engineering, Inc., 2017).

Sensitivity analysis of the empirically grounded parameters indicate that the results remain consistent for reasonable uncertainties in parameter values (Supporting Information).

3.2 PMA Colorado River Availability Scenarios

Multiple scholars (Wheeler et al., 2022), the Bureau of Reclamation (DOI, 2022), and water users (Goddard & Atkins, 2022), argue that the existing water distribution rules are not enough to address the Colorado River’s pressing state, which may require a 2-4 MAFY basin-wide cut in use (DOI, 2022). The Bureau has undergone a Supporting Environmental Impact Statement (SEIS) process to reform the 2007 guidelines (Bureau of Reclamation, 2022) and published a draft (D-SEIS) in April (Bureau of Reclamation, 2023b) that includes two alternatives to share needed cuts by the traditional priority system or adopt a use-proportional sharing approach.

Our CAP scenarios assume 2024 will be a year of considerable change in basin operations and require an immediate, long-term decrease in use. We note that drought-induced shortages are not necessarily long-term, so these scenarios are meant to be exploratory of hypothetical long-term cuts to river use, not policy diagnoses.

The CAP has a tiered system of sub-contracts that we simplify into low and high priority sub-contracts. High priority includes Municipal and Industrial (M&I), Indian, and P3 rights (CAP, 2022b). Low priority use includes Non-Indian Agriculture (NIA) and excess water use. The Supporting Information details the method we use to convert a possible shortage in Arizona’s allocation (Δ_{AZ}) to the shortage for PMA users (Δ_{PMA}). $\bar{\mu}_2$ is the annual water (AFY) that PMA cities, together, are entitled to use in a non-shortage year, and it totals 448,663 AFY. We define $E_2^\mu(\cdot)$ (from Equation 1) to take a parameter s , which is the proportional change in PMA CAP availability ($\frac{\Delta_{PMA}}{\bar{\mu}_2}$), as,

$$E_2^\mu(\bar{\mu}_2, t) = \begin{cases} s\bar{\mu}_2 & t = t^* \\ 0 & \text{otherwise} \end{cases} \quad (4)$$

The May 24-month study projects a most-probable Lake Mead level consistent with a Tier 1 shortage (Bureau of Reclamation, 2023a), but Arizona leaders have cautioned that 2024 cuts could resemble a Tier 3 shortage ($s = -0.284$), if additional cuts are added by the D-SEIS (CAP, 2023). In May, the basin states released a compromise proposal for the Bureau to consider in its D-SEIS, along with the two alternatives, that includes total cuts of 3 MAF by 2026 (Buchatzke et al., 2023), but the details regarding how the cut will be shared among users like the CAP remains unknown.

3.3 Running the Model

We run each configuration of the model for 50 years because the shock is a simple step disturbance. Because the model is deterministic, multiple runs of the same configuration were not needed. Investments and rate-setting are driven by a goal supply buffer of $\gamma_2 = 1.2$ and a minimum debt service coverage ratio of $\gamma_3 = 2$ (Rafelis, 2021), respectively. The most Phoenix has raised rates within a year is 13%, so we set $\psi_r = 0.15$.

482 The institutional friction parameters are difficult to define for a specific city, so we
 483 define the baseline values in the following way. For ϵ_j , we assume no costs at baseline
 484 ($\epsilon_j = 0$). The response elasticity, λ_j , refers to the degree of flexibility in the action sit-
 485 uation, altering the amount of partial response generated in the neighborhood of ϵ_j . Higher
 486 λ_j implies little partial response, closely resembling a binary switch near ϵ_j . We, thus,
 487 relate λ_j to the range of error values, Δe_j , around ϵ_j that will generate a significant at-
 488 tention response ($Y > 0.1$). We use a baseline Δe_j of 0.1 for investment and rate-setting
 489 and 0.02 for short-term measures, assuming short-term measures are taken with less un-
 490 certainty because the stressors are presently felt, which implies λ_j values of 22 and 110,
 491 respectively (Supporting Information).

492 3.4 Exploration of Institutional Friction’s Effect

493 We examine the effect of varying institutional friction parameters, ϵ_j and λ_j , on
 494 the cities’ sensitivity to Colorado River restrictions. We tested each institutional fric-
 495 tion parameter setting on the full range of s values in $[0, 1]$. We vary investment insti-
 496 tutional costs, ϵ_2 from its no-costs baseline up to $\epsilon_2 = 0.5$, which implies that the PIP
 497 would not act until there is only supply for 60% of projected demand. Given the very
 498 conservative, reliability driven nature of water utilities, this is a very unlikely formal thresh-
 499 old, but the institutional costs of major investment can be an overwhelming informal bar-
 500 rier to glaringly needed supply or demand improvements (Muller, 2018). For rate-setting,
 501 we vary it up to $\epsilon_3 = 1$, where the utility only acts once debt requirements cover all
 502 net revenue. We choose the λ_j range based on the extreme cases of $\Delta e_j \in [0.01, 0.5]$,
 503 which implies $\lambda_j \in [4, 220]$ (Supporting Information). We vary λ_j on a log scale due
 504 to its position as a coefficient within an exponential expression.

505 4 Results

506 We present the results of running the UWIIM for each PMA city under the base-
 507 line assumptions (4.1) to understand their responses to various magnitudes of CAP short-
 508 age. Then, with this baseline established, we analyze the effect of varying institutional
 509 friction on the city response (4.2).

510 4.1 Baseline Operational Capacities of Each City

511 Absent consideration of institutional variation, the unique infrastructure, demand,
 512 and financial context of each city shapes their response to CAP shortages. To contex-
 513 tualize our shortage approximations, we highlight the s values equivalent to a Tier 2a
 514 and Tier 3 shortage and the two alternatives proposed in the April D-SEIS (Support-
 515 ing Information).

516 Phoenix and Scottsdale are primarily dependent on surface water, but they differ
 517 in how they leverage non-CAP supplies to respond to a CAP shortage. Phoenix relies
 518 on their much larger SRP allocation while Scottsdale relies on its higher groundwater
 519 processing capacity. Both sources provide a reliable buffer to the cities for the Tier 3 and
 520 sharing-based D-SEIS scenarios (Figure 3). We identify the “break away” points where
 521 the CAP shock overcomes the city’s operational capacity and compromises supply reli-
 522 ability at s values of -61% and -72% for Scottsdale and Phoenix, respectively (Fig-
 523 ure 4b). However, Scottsdale’s reliance on non-renewable groundwater to buffer against
 524 higher CAP shocks causes them to run out of surplus credits near 2045, requiring ad-
 525 ditional investments in re-use and ultimately, serious demand reductions that Phoenix
 526 can avoid (Figure 3, 4a). The additional investments pertaining to Scottsdale’s ground-
 527 water reliance makes their rates more sensitive than Phoenix’s rates (Figure 4c).

528 Queen Creek is a groundwater dependent city, but the substantial amount of ground-
 529 water credits it has secured, without CAGR, can only hold the city until around 2040.

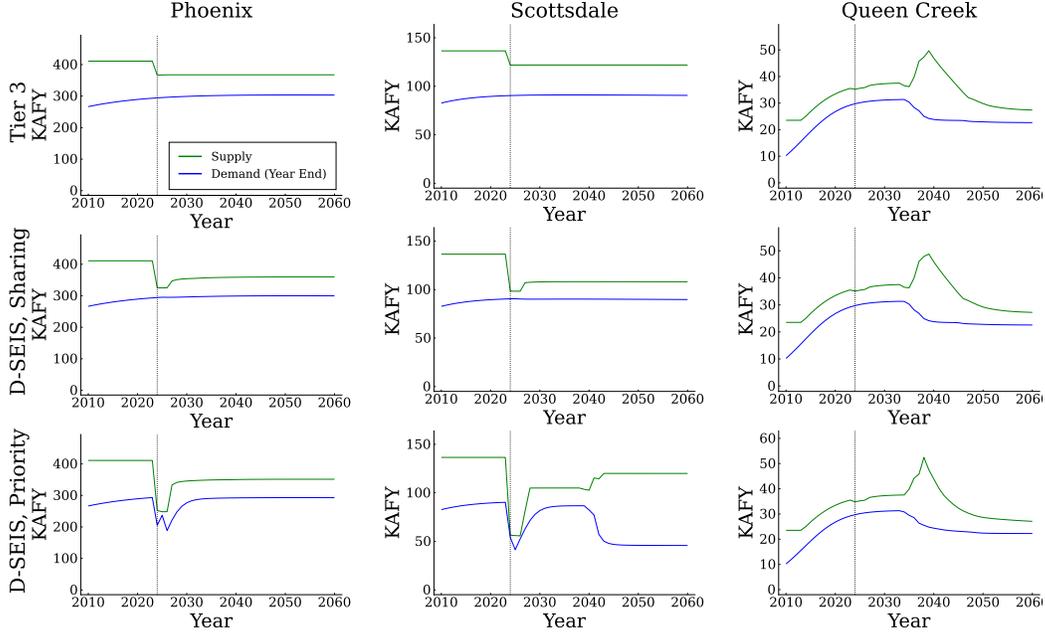


Figure 3. Supply and demand time series for each city given a long-term cut to CAP availability in 2024 equivalent to a Tier 3 shortage ($s = -0.284$), the priority-based ($s = -0.957$), and the sharing-based ($s = -0.525$) D-SEIS alternatives. D-SEIS approximations assume the 2.083 MAF basin-wide shortage scenario.

530 In the baseline model runs, Queen Creek registers that its credits will run out in the PIP’s
 531 five-year projection window (τ_p) and responds with investments in re-use, demand man-
 532 agement, and surface water processing (if CAP water is available) that prevent it from
 533 experiencing any reliability problems (Figure 3, 4b). Prior to 2024, Queen Creek experi-
 534 ences rapid demand increase driven by population growth but is able to maintain sup-
 535 ply reliability through groundwater processing investments. At baseline parameter set-
 536 tings, s determines how much available CAP water they can leverage. With less CAP
 537 availability, they have to cut demand more and pass higher costs to their ratepayers. Queen
 538 Creek’s small and mainly low priority CAP allocation makes these two measures very
 539 sensitive to s values smaller than -20% (near a Tier 2a shortage) (Figures 4a,d), after
 540 which, they lose most of their allocation regardless of s , and their sensitivity curves de-
 541 cay to zero. With the high sensitivity region, for each 1% drop in CAP availability, Queen
 542 Creek must cut demand by an additional $0.53 - 0.69\%$. Rate sensitivity, in this zone,
 543 is much more volatile (Figure 4) due to the nonlinear interaction between s , groundwa-
 544 ter use, and shifting investment priorities ($\beta_{k,t}$), which each have their own cost curve.

545 4.2 Institutional Friction Effects

546 The two institutional friction components, costs and flexibility, alter the response
 547 of PMA cities in different ways. We focus on the role institutional friction plays in each
 548 city’s *vulnerable range* of s , identified in the baseline analysis (Figure 4). For Phoenix
 549 and Scottsdale, this range is $s < -0.6$, and for Queen Creek it is $s > -0.2$.

550 4.2.1 Institutional Costs

551 Across all three cities, increasing institutional costs, ϵ_j , in both the investment and
 552 rate-making action situations increases the sensitivity of the city’s supply to its vulner-

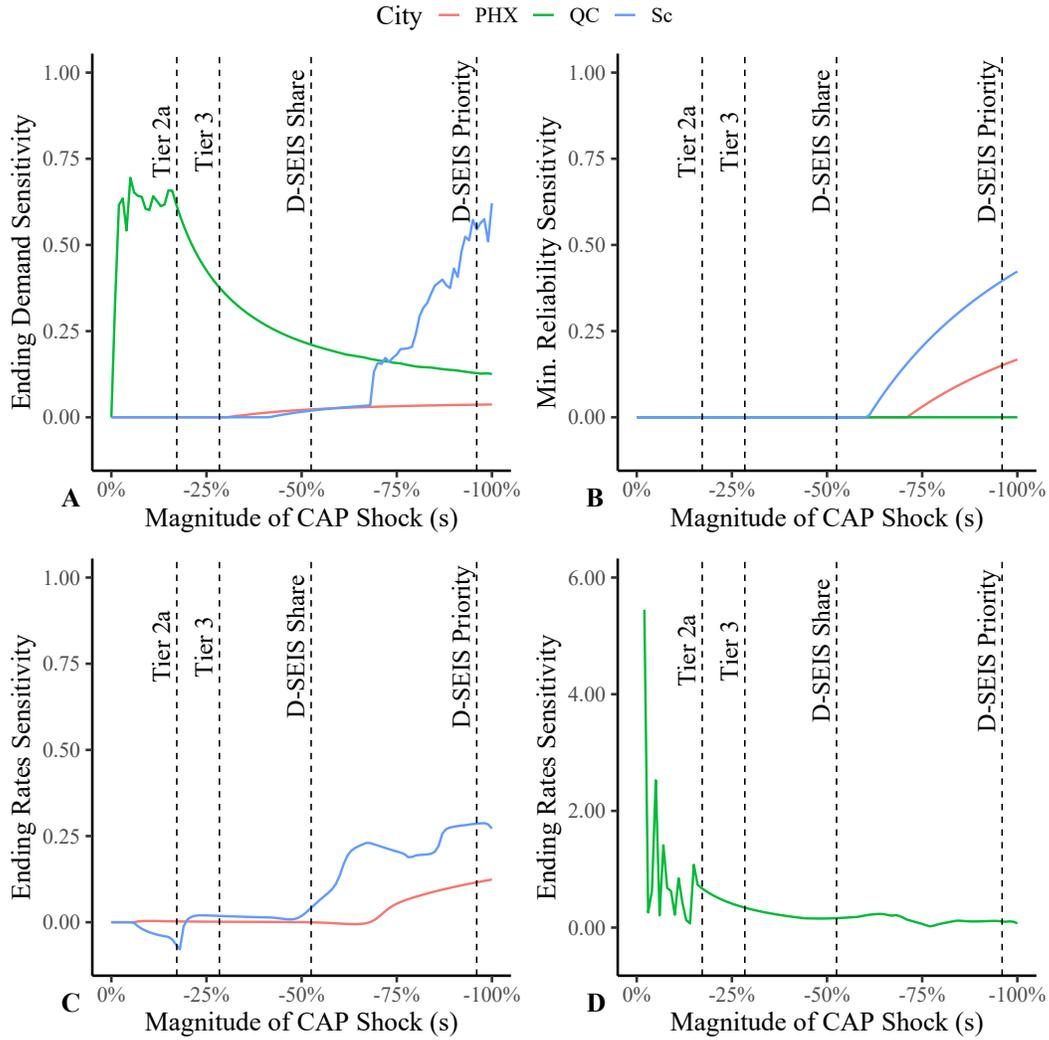


Figure 4. Effect of CAP shock magnitude (s) on the sensitivity of ending demand (a), minimum reliability (b), and ending rate pressure (c-d) experienced by each city with other parameters at the baseline setting. Mean reliability experienced minimal change and is therefore, not depicted here. We distinguish Queen Creek’s rate sensitivity (d) due to its larger scale.

553 able range of s . With more institutional costs, the PIP allows more error to accumulate
 554 before acting.

555 Low institutional costs in both rate-setting (e.g., $\epsilon_3 < 0.6$) and investment (e.g.,
 556 $\epsilon_2 < 0.35$) do not significantly impact reliability (Figures 5a-d). For Phoenix and Scotts-
 557 dale, investment institutional costs increase the sensitivity of average reliability more than
 558 minimum reliability (Figure 5a) because the CAP shock is sudden and the investment
 559 lag time (τ_k^i) is independent of ϵ_2 , making the “worst year” primarily dependent on s .
 560 ϵ_2 , thus, affects how quickly the system bounces back to a desirable reliability, which is
 561 reflected in the average reliability metric.

562 Queen Creek’s precarious position, explained above, necessitates major investment
 563 and rate increases, so slowing their response increases their reliability sensitivity by or-
 564 ders of magnitude more than Phoenix and Scottsdale (Figures 5b,d). In the rate-setting
 565 action situation (Figure 5d), high ϵ_3 places Queen Creek in a demand curtailment trap
 566 where they must continuously cut demand as they fail to keep up with 2010-2020 growth
 567 and are left with not enough revenue to fund needed investments (illustration in Sup-
 568 porting Information).

569 The reliability sensitivity curves for investment and rate-setting have opposing con-
 570 cavities (Figures 5a-d). For investment, the concave up shape of the average reliability
 571 sensitivity curve reflects the fact that increasing ϵ_2 increases the range of perceived short-
 572 ages that will *not* generate action from the PIP, so each marginal increase in ϵ_2 has a
 573 larger effect on average reliability (Figures 5a-b). For rate-setting, the concave down shape
 574 in both reliability curves suggests that when the city is more hesitant to increase rates
 575 but still willing to invest, the city funds partial investments within the existing rate struc-
 576 ture (Figures 5c-d). These sensitivity curves eventually converge to a “pass through” level
 577 where the shortage signal affects the city’s reliability at the same rate (recall, sensitiv-
 578 ity is a slope of output change to input change).

579 The slowing effect of institutional costs, in both rate-setting and investment, can
 580 benefit the citizen burden metrics: demand and rates (Figures 5e-h). Slowing investment
 581 prevents the implementation of demand management programs and prevents the need
 582 to increase rates (Figures 5e-f). Slowing rate increases does not affect demand sensitiv-
 583 ity in Phoenix and Scottsdale significantly (Figure 5g), but intuitively, decreases rate sen-
 584 sitivity.

585 However, this trade-off between reliability and citizen burdens is not one-for-one.
 586 For each city, there are ϵ_j values that decrease rate sensitivity with minimal impact on
 587 reliability (e.g., $\epsilon_3 = 0.3$ for Scottsdale). Rate-setting institutional costs, though, can
 588 be detrimental for citizen burden metrics. When $\epsilon_3 \in [0.25, 0.45]$, Phoenix experiences
 589 higher rates because Phoenix was slow to increase rates during the 2010-2020 period, leav-
 590 ing them with more debt by 2024 and therefore, in need of a larger rate increase to cover
 591 shock-induced investments (Figure 5g). For Queen Creek, $\epsilon_3 \in [0.5, 0.7]$ forces partial
 592 investments before the groundwater crisis, which then requires higher demand cuts when
 593 they run out of credits (Figure 5h).

594 **4.2.2 Institutional Flexibility**

595 Varying the response elasticity, λ_j , did not significantly alter reliability or the end-
 596 ing demand when institutional costs, ϵ_j , are kept at their baseline no-costs level. This
 597 is likely because the CAP shock and Queen Creek’s groundwater crisis are at high enough
 598 magnitudes to produce error that exceeds Δe_j , leaving no concern for partial response
 599 when $\epsilon_j = 0$. However, when these high magnitude events occur, adding flexibility to
 600 rate-making (lowering λ_3), can decrease the ending rates sensitivity (Figure 6b) by low-
 601 ering the immediate rate increase, which can temporarily hurt the debt service goal but
 602 allows the PIP to gradually raise rates as the operational context evolves. Lowering λ_3

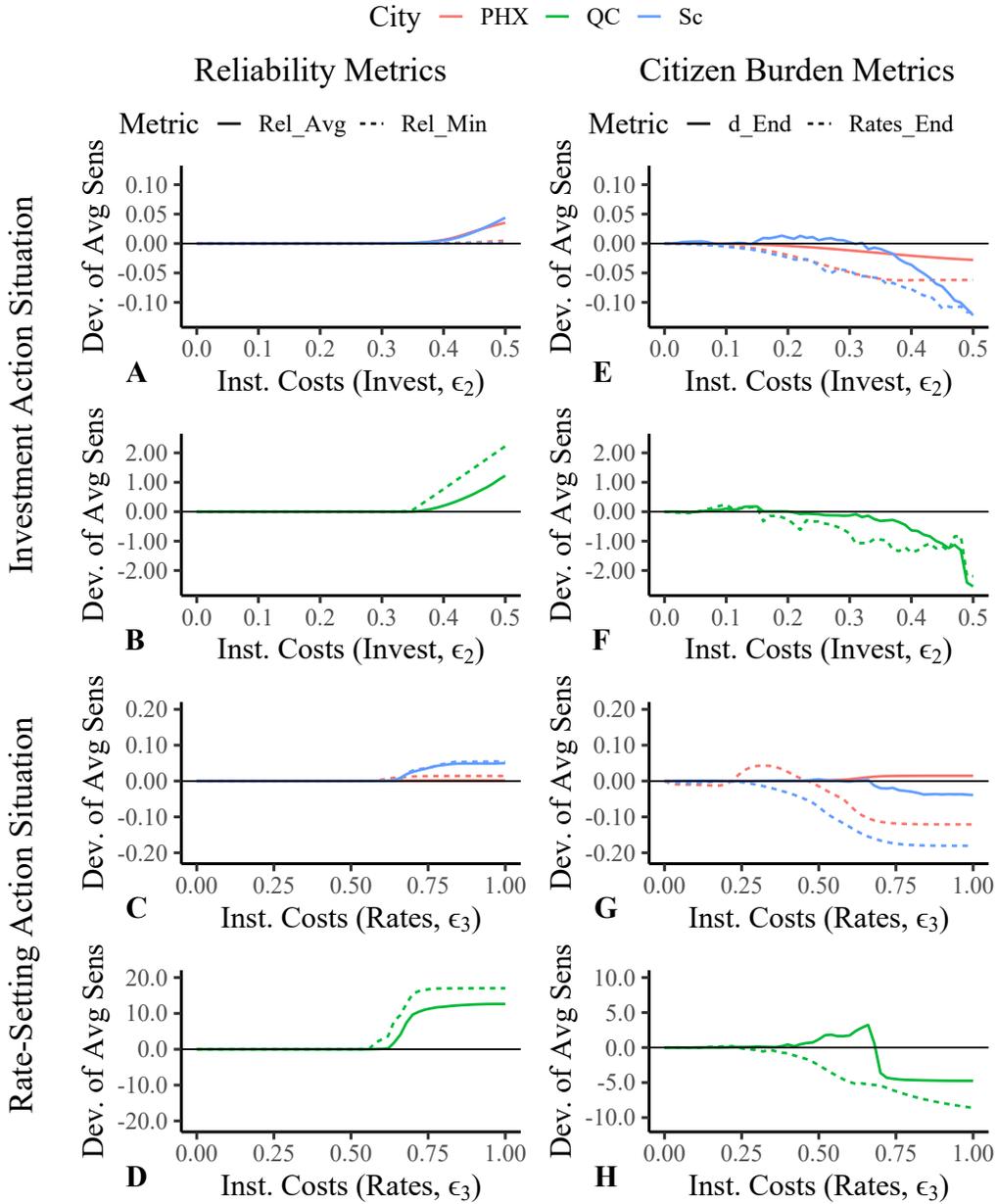


Figure 5. Sensitivity analysis results for varying institutional costs in investment (ϵ_2) and rate-making (ϵ_3) for each city. Sensitivity of each performance metric is presented on the y axis by the deviation of the average sensitivity from the average sensitivity associated with the baseline setting ($\epsilon_{2,3} = 0$) over all CAP shock magnitudes where $s < -0.6$ for Phoenix and Scottsdale and $s > -0.2$ for Queen Creek.

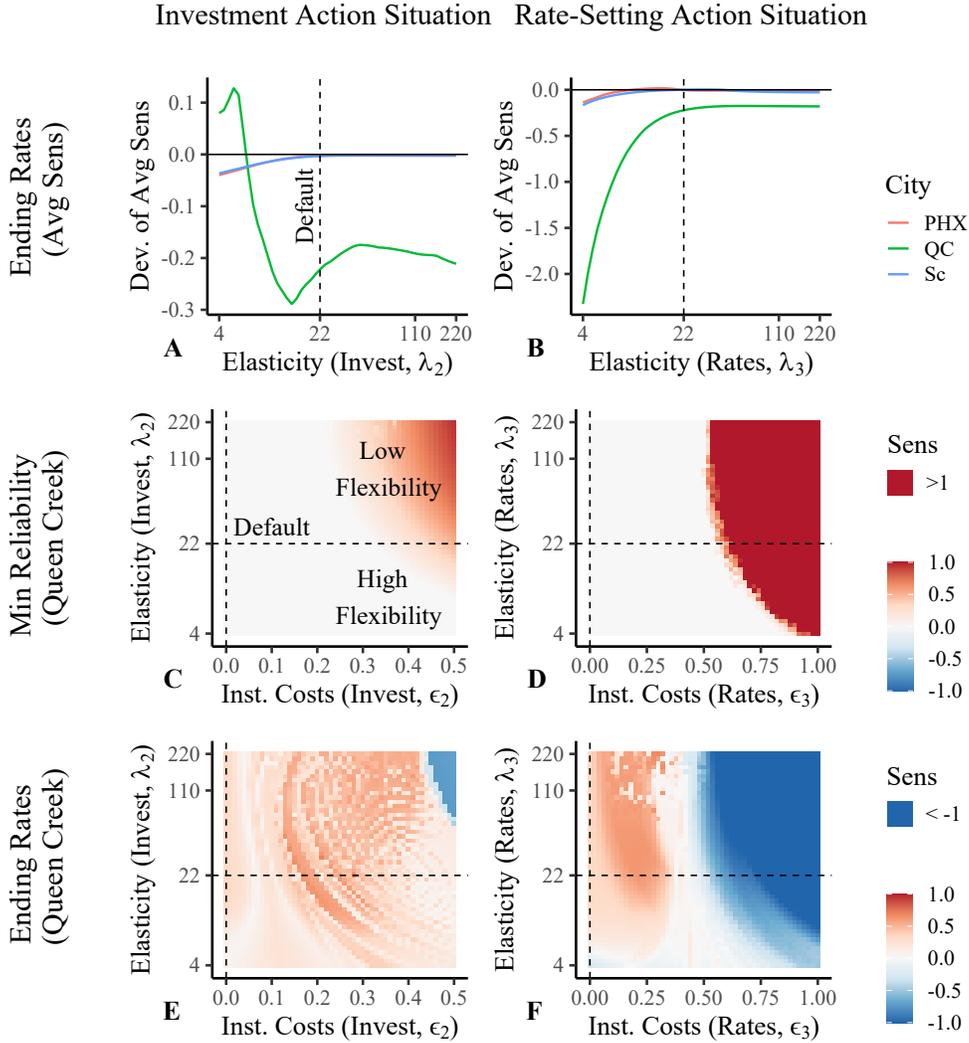


Figure 6. Sensitivity analysis results for varying institutional flexibility in investment (λ_2) and rate-making (λ_3) for each city. First, (a-b) the deviation of the average sensitivity of ending rates from the baseline setting ($\lambda_{2,3} = 22$) over all CAP shock magnitudes where $s < -0.6$ for Phoenix and Scottsdale and $s > -0.2$ for Queen Creek. Then, the sensitivity of minimum reliability (c-d) and ending rates (e-f) at the Tier 3 CAP shock ($s = -0.284$) given variation in institutional costs (ϵ) and flexibility (λ) in both action situations.

603 also allows the PIP to lower rates because the partial response region (Δe_j) exists on
 604 both sides of ϵ_j . Queen Creek’s rates sensitivity curve for the investment action situa-
 605 tion does turn positive when $\lambda_2 < 6$ because the partial investments caused by high
 606 flexibility ultimately sum up to a higher revenue need (Figure 6a).

607 Additionally, we investigate the interactive effect of λ_j with varying ϵ_j , keeping $s =$
 608 -0.284 (Tier 3 shortage). At this level, Phoenix and Scottsdale have sufficient opera-
 609 tional capacity to buffer the shock, but Queen Creek must take investment action (Fig-
 610 ure 4). In each action situation for Queen Creek, the slowing effect of institutional costs
 611 on reliability can be overcome with added flexibility (Figures 6c,d), which add needed
 612 partial responses when the action threshold would otherwise preclude the PIP from re-
 613 sponding. On the other hand, adding flexibility to low-costs rate-setting situations (e.g.,
 614 $\epsilon_3 \in [0, 0.25]$) can ease the ending rate pressure through gradual rate increases (Fig-
 615 ure 6f). There are complex interactions between investment and rate-making where par-
 616 tial responses may increase overall investment and revenue needs in the long-term (es-
 617 pecially at the borders of the color regions in Figures 6e,f), but complete investigation
 618 of this interactive effect is beyond the scope of our initial study.

619 5 Discussion

620 Our application of the UWIIM yields three important takeaways regarding the in-
 621 teraction of operational and political-economic feedback in the response of coupled in-
 622 frastructure systems to changing environmental inflows.

623 5.1 The Inescapable Role of Redundant Operational Capacity

624 Even though political-economic features, like institutions, can influence the way
 625 an urban water system responds to environmental changes, the operational capacity and
 626 in particular, the presence of redundancy, sets the bandwidth of disturbance that the sys-
 627 tem can tolerate *before* requiring political-economic action. Redundancy is a well-established
 628 feature of robust systems (Anderies & Janssen, 2013; Biggs et al., 2012), and our base-
 629 line analysis of the PMA cities before institutional friction variation highlights the need
 630 to clarify the operational context before investigating political-economic considerations.

631 In the PMA case, the presence of non-CAP supplies, SRP water and groundwa-
 632 ter, defined the range of CAP shortage cities could manage before requiring additional
 633 investments and rate increases, and in the current PMA management context, this is a
 634 prominent focus. Phoenix, for instance, has invested in flexible delivery system infras-
 635 tructure to ensure that they can use their additional SRP water throughout their ser-
 636 vice area. The non-renewable nature of groundwater, though, has placed Scottsdale in
 637 a more sensitive position, assuming they can distribute the groundwater throughout their
 638 service area, to buffer CAP shocks. In our scenarios, large CAP shocks beyond the D-
 639 SEIS shortage sharing option force Scottsdale to ultimately use up their groundwater
 640 credits within thirty years, which would then require more serious action, potentially in-
 641 cluding major demand cuts. Queen Creek is groundwater dependent, but they are pur-
 642 suing alternative renewable water sources to achieve this needed redundancy (ADWR,
 643 2022b).

644 5.2 CIS Robustness & Political-Economic Feedback

645 When the system’s operational capacity cannot tolerate a disturbance, political-
 646 economic action is necessary. Given our critical infrastructure systems, like urban wa-
 647 ter, face an uncertain future in the Anthropocene (Chester et al., 2019), analysis must
 648 consider the multiple layers of governance, including, as presented here, the differing roles
 649 of operational and political-economic feedback. The UWIIM demonstrates the poten-

650 tial to weave political-economic considerations into the dynamics of human-water sys-
651 tems.

652 One political-economic consideration elevated by this manuscript is the dynamic
653 implication of institutional friction (Jones & Baumgartner, 2005a) on CIS sensitivity to
654 environmental shocks. Across all cities examined, adding institutional costs tended to
655 increase the sensitivity of supply reliability when the shock necessitated political-economic
656 action (otherwise, the existing operational capacity was able to cope), but our analysis
657 points to city-specific ways that this complex relationship between institutional costs and
658 operational context behaves. For instance, Phoenix and Scottsdale’s higher operational
659 capacity allowed them to maintain reliability when faced with low-to-moderate institu-
660 tional costs (5a), but Queen Creek’s immediate pressure to keep up with rising popu-
661 lation, move away from CAGR, and find alternative sources before it ran out of ground-
662 water credits, necessitated major early action, making it more sensitive to anything in
663 the institutional context that slowed the response. In fact, high institutional costs in rate-
664 setting decisions, created a demand curtailment trap, identified in other socio-hydrologic
665 studies of urban water systems (Rachunok & Fletcher, 2023; Kenney, 2014) where the
666 city must continually cut demand to ration supply, but then loses potential revenue to
667 support needed investments.

668 While institutional costs can make reliability more sensitive, they can benefit po-
669 tentially competing objectives relating to externalities of investments like the rates or
670 demand pressure placed on ratepayers. In fact, consideration of such externalities is likely
671 one of the main reasons such institutional costs exist in the decision-making process (Des-
672 latte et al, in press). Such performance-vulnerability trade-offs are a feature of complex
673 social-ecological systems (Anderies et al., 2007; Homayounfar et al., 2018), and incor-
674 porating institutional information processing into human-water system models is a promis-
675 ing way to make these trade-offs apparent in the system’s dynamics rather than serv-
676 ing as static decision parameters manipulated exogenously by the modeler.

677 **5.3 The Interactive Effect of Institutional Flexibility**

678 Even if political-economic actors are aware of the institutional costs associated with
679 taking action, institutions often fail to specify exactly how an actor should respond to
680 a particular problem, perhaps even providing conflicting direction. These situations can
681 be termed “institutional voids” (Mesdaghi et al., 2022), and we attempt to capture their
682 dynamic role through the response elasticity parameter.

683 For all cities, adding flexibility did not impact reliability significantly when there
684 are no institutional costs, but, particularly when added to rate-setting decisions, flex-
685 ibility can ease the burden placed on ratepayers. It generates partial rate increases that
686 may temporarily violate debt coverage goals but decrease the rate burden over time if
687 reliability is sufficiently managed and more shocks do not occur. Proportional control
688 is often associated with abrupt, over-responsive action (Rodriguez et al., 2011), and our
689 findings demonstrate that institutional flexibility can be conceptualized as a potentially
690 helpful *integral* element to the control scheme to encourage a smoother response to sud-
691 den supply shocks. Additionally, when a system must deal with high institutional costs
692 to taking major action, adding flexibility can improve reliability and rate pressure by open-
693 ing up opportunities for much needed partial responses. We emphasize that this com-
694 plex interaction between institutional costs, flexibility, and operating capacity speaks to
695 the importance of considering institutional friction in the analysis of coupled infrastruc-
696 ture systems.

5.4 Limitations & Opportunities

The UWIIM offers multiple opportunities for additional research on the relationship between operational and political economic feedback in human-water systems. Our focus on institutional friction bridges policy process theory with human-water systems, but additional policy process concepts can be investigated. Other iterations can consider competition between belief systems from the Advocacy Coalition Framework (Weible et al., 2018), richer institutional characteristics with the Institutional Grammar Tool (Siddiki et al., 2019), or the way actor participation in multiple action situations creates institutional cost spillovers with the Ecology of Games Framework (Berardo & Lubell, 2019). Once actor heterogeneity is considered, an agent-based model may be a more effective tool than our aggregate dynamical system.

The study did not consider the interaction effects between institutional friction and other institutional design parameters like goals (e.g., debt service coverage ratio) and choice constraints (e.g., maximum annual rate increase allowed). Analysis of these concerns, particularly the selection of supply and financial goals, may benefit from the risk of failure threshold choice analysis performed by multiple water resource models (Trindade et al., 2020; Zeff et al., 2016; Palmer & Characklis, 2009).

Finally, even though the UWIIM is not designed to be a decision support or predictive model, we defined the PMA cities to reflect the critical ways they vary in their operational capacity. We do not account for the complex firming contracts the cities may have to weather additional CAP shortages and any sudden water transfers from actors that have unused portions of their CAP allocation. Additionally, we do not incorporate new water sources like importing desalinated water.

Political-economic feedback provides a valuable opportunity for human-water systems modelers to enrich their understanding of information processing in complex social systems like cities and deepen the interdisciplinary ethos of socio-hydrology with insight from fields like policy process theory. In the Anthropocene, it is no longer sufficient to treat the political-economic processes that govern a CIS as exogenous parameters or endogenous rational entities that can be modeled in the same way as a pipe network. The UWIIM is one attempt to operationalize political-economic and operational feedback in a coupled infrastructure systems model, and we encourage adding additional considerations or the development of alternative models to further this much needed discussion.

Open Research Section

This version of the Urban Water Infrastructure Investment Model (UWIIM), written in Julia, all output data generated from model runs, and the R scripts used to create the published figures, will be published and made available on GitHub and HydroShare repositories once review comments are incorporated.

All data used to define parameters and initial conditions for the studied cases were calculated from publicly available databases (Supporting Information), which can be obtained from the citations noted in-text.

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Figure 1.

Green: Information
Blue: Material

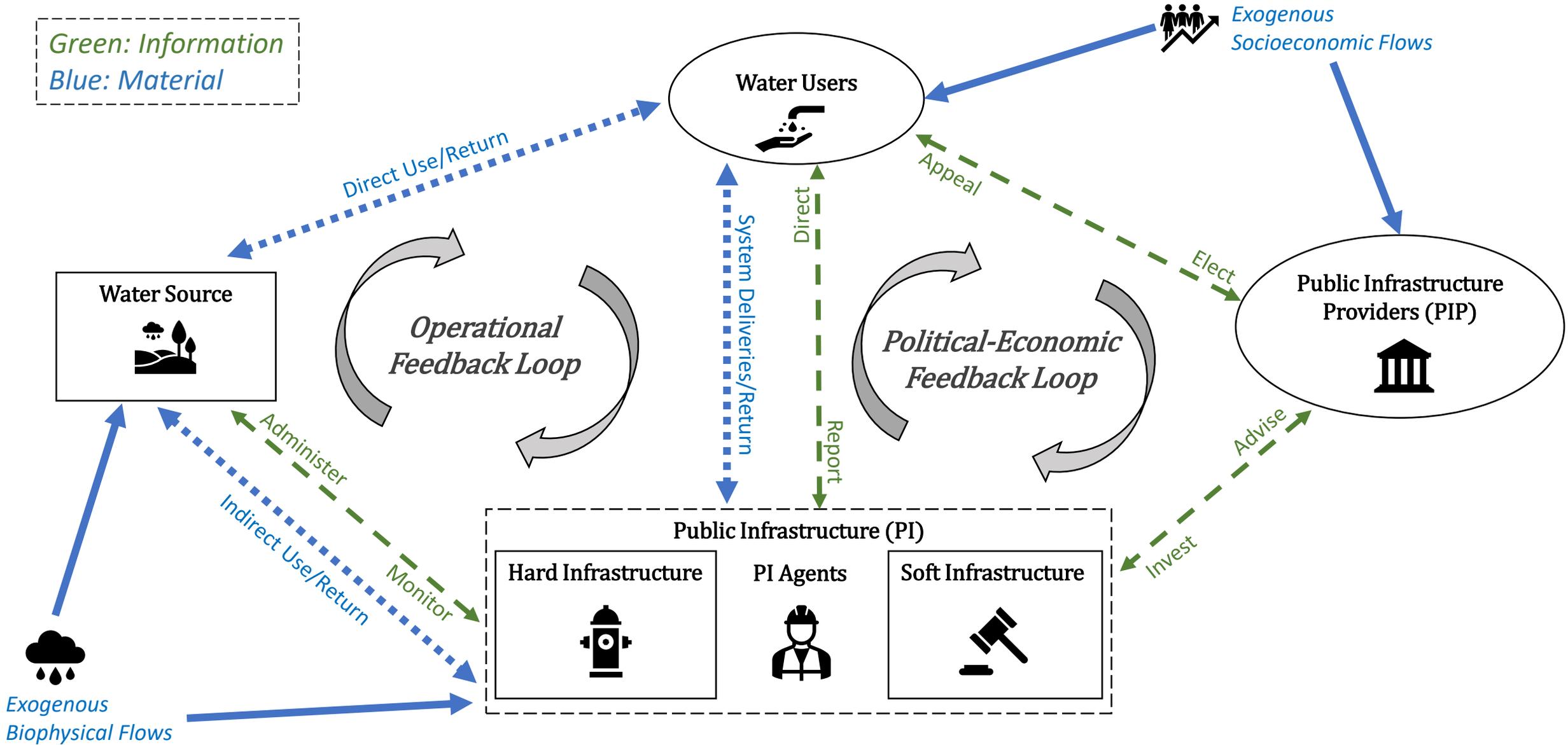
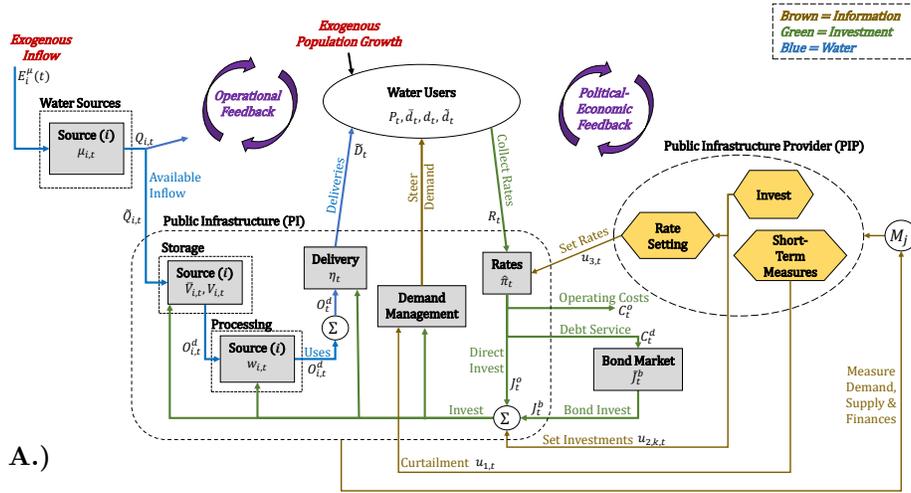
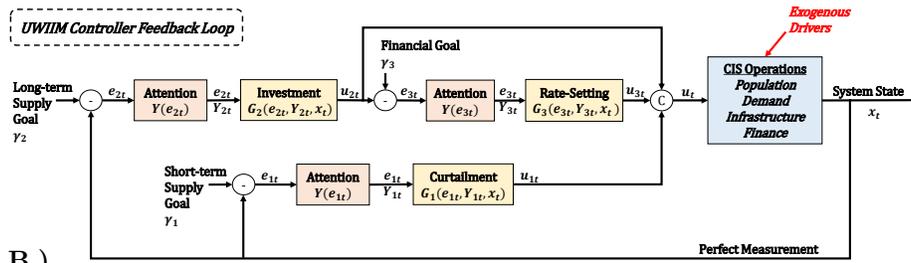


Figure 2.



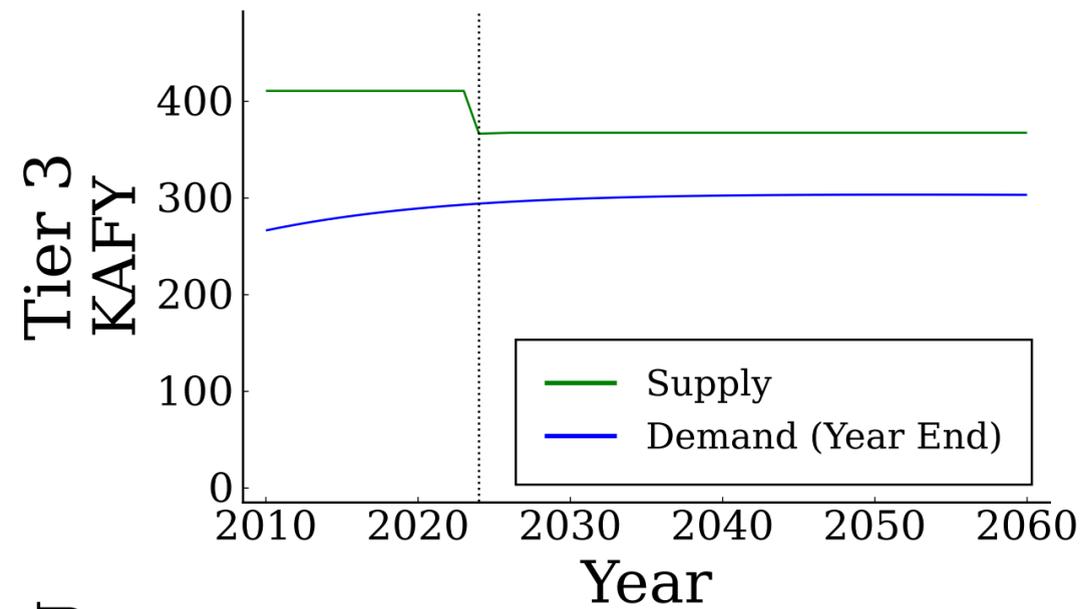
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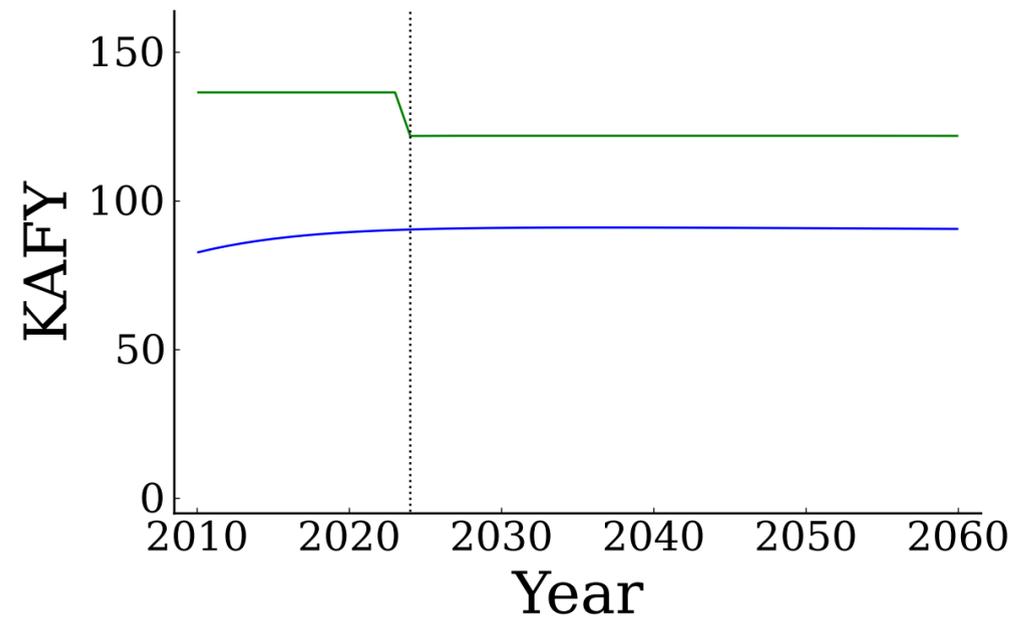
B.)

Figure 3.

Phoenix



Scottsdale



Queen Creek

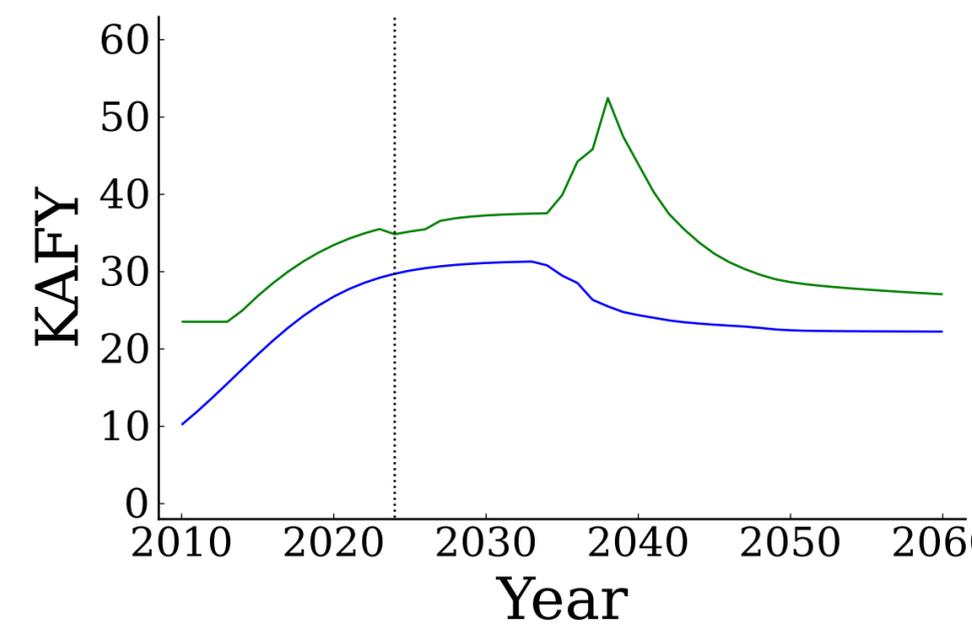
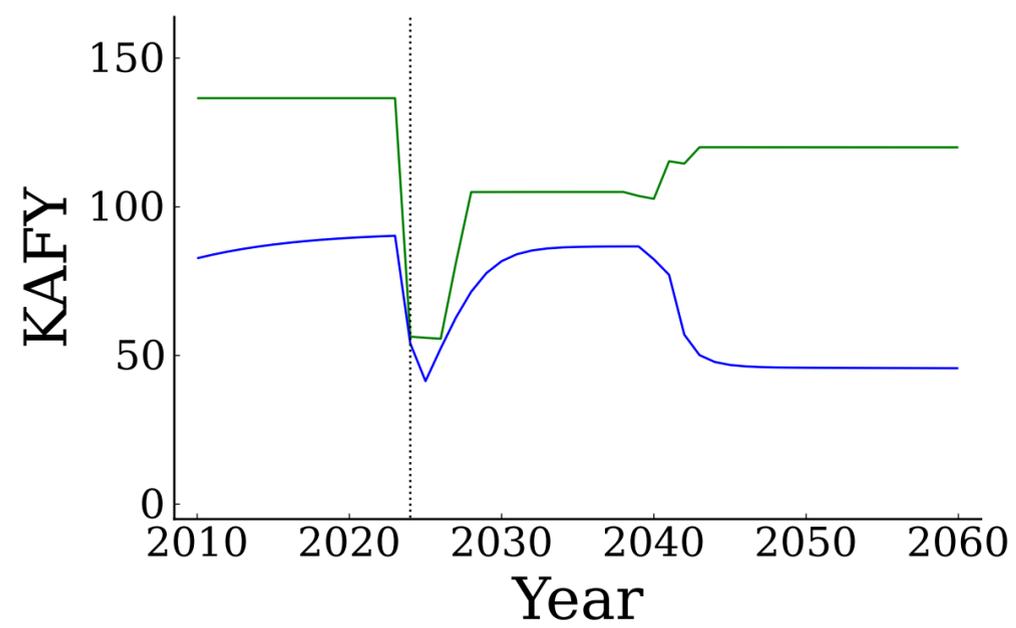
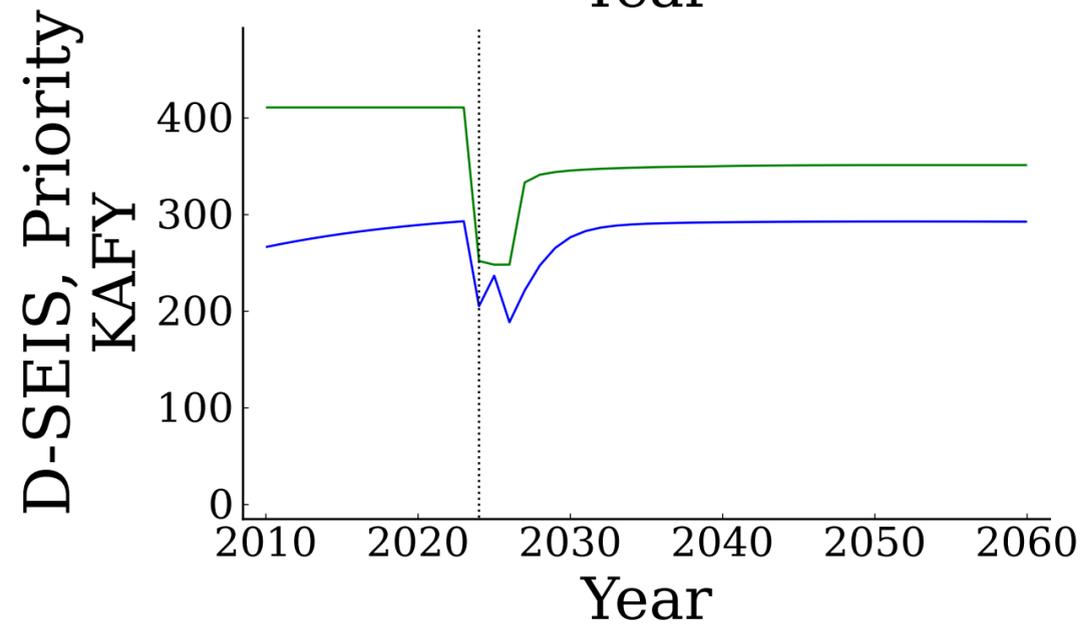
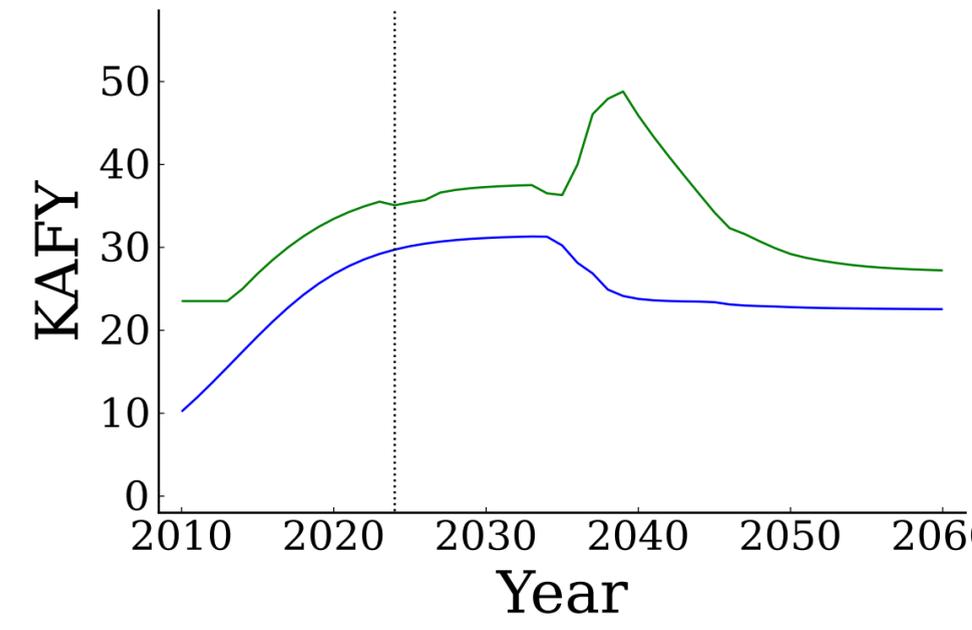
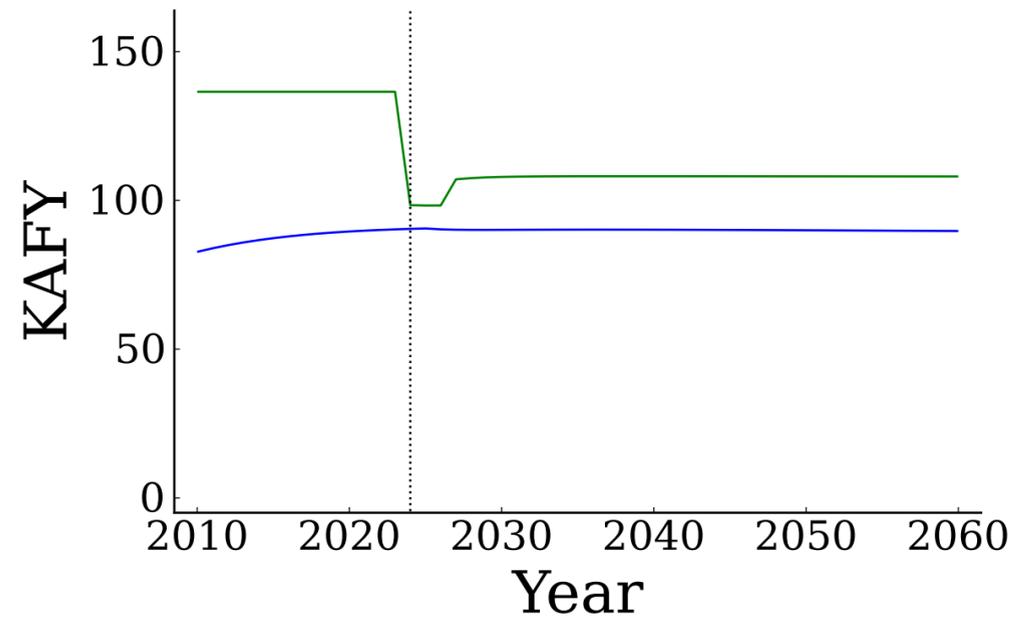
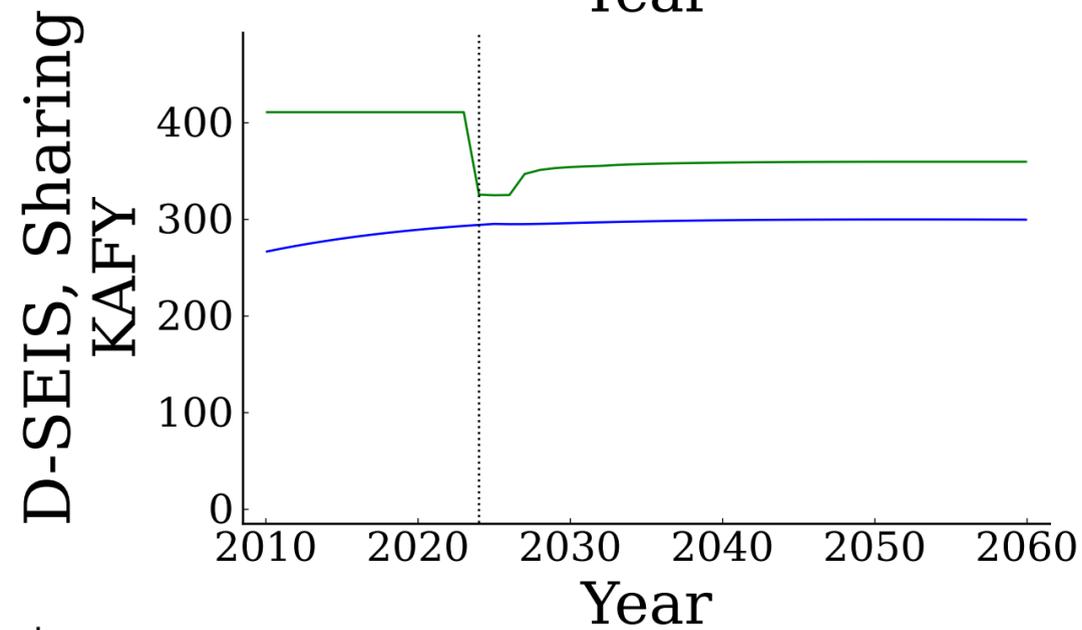
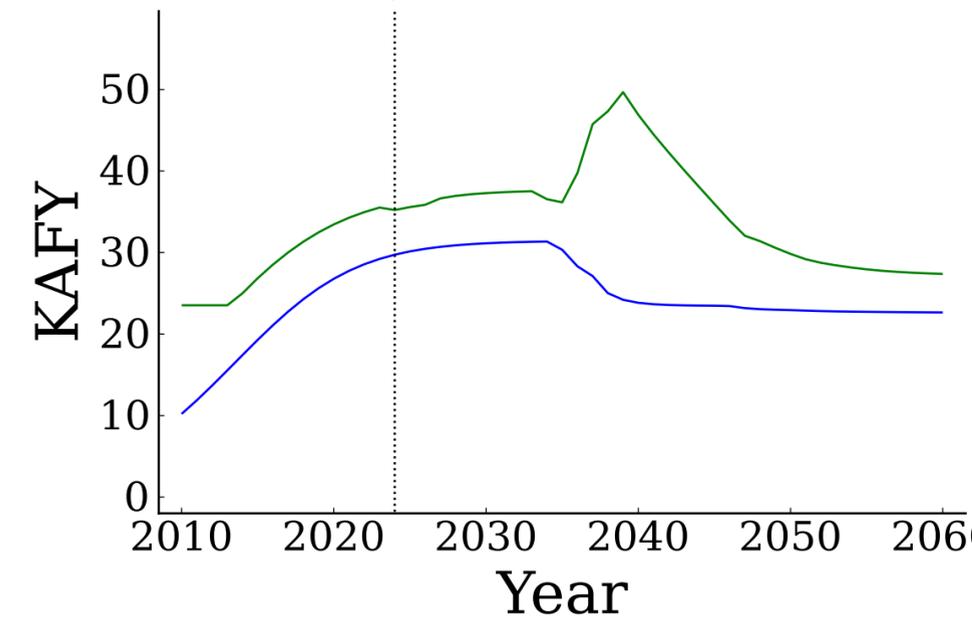


Figure 4.

City — PHX — QC — Sc

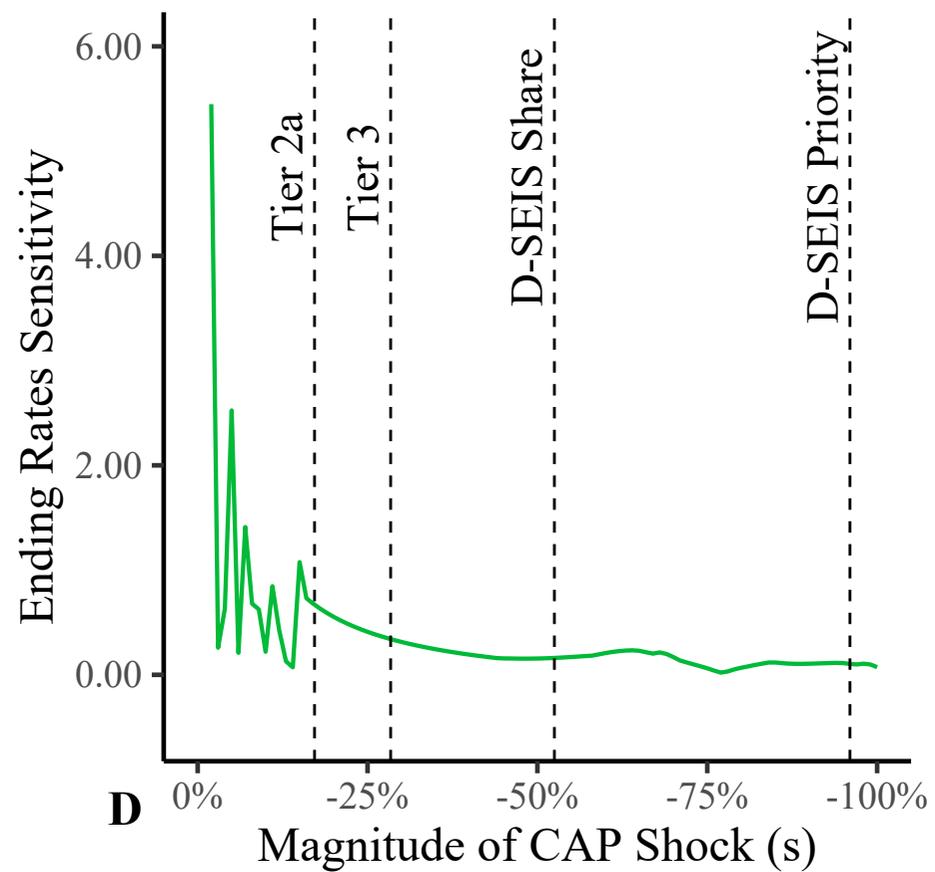
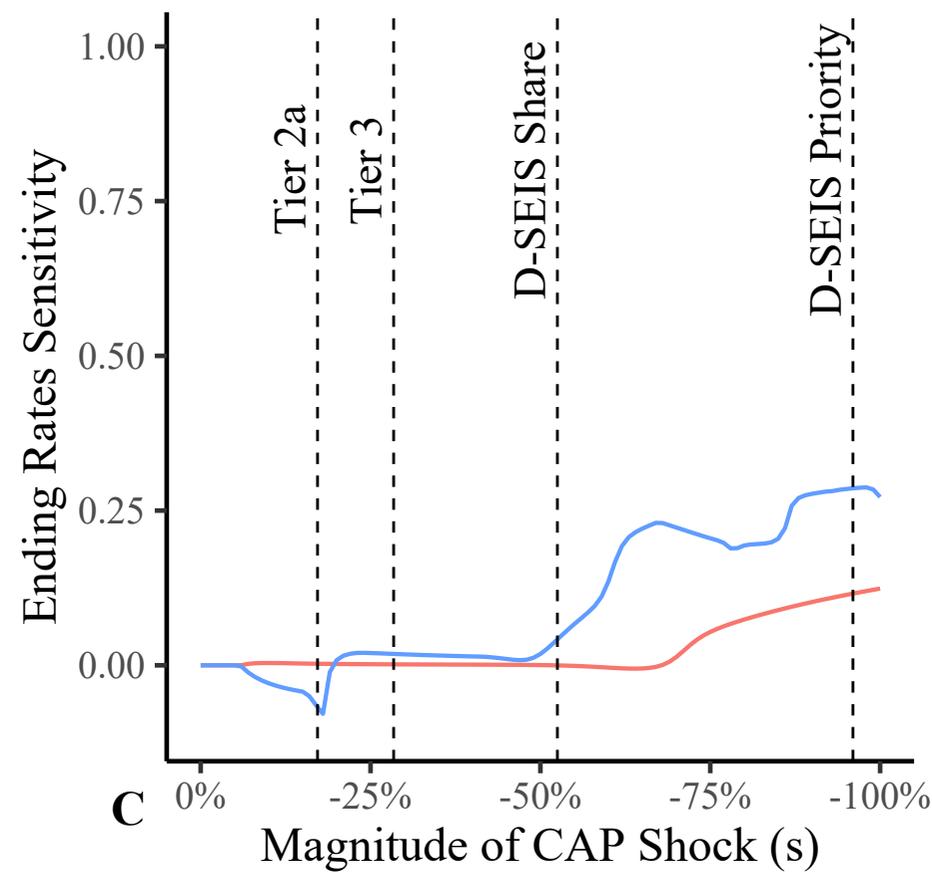
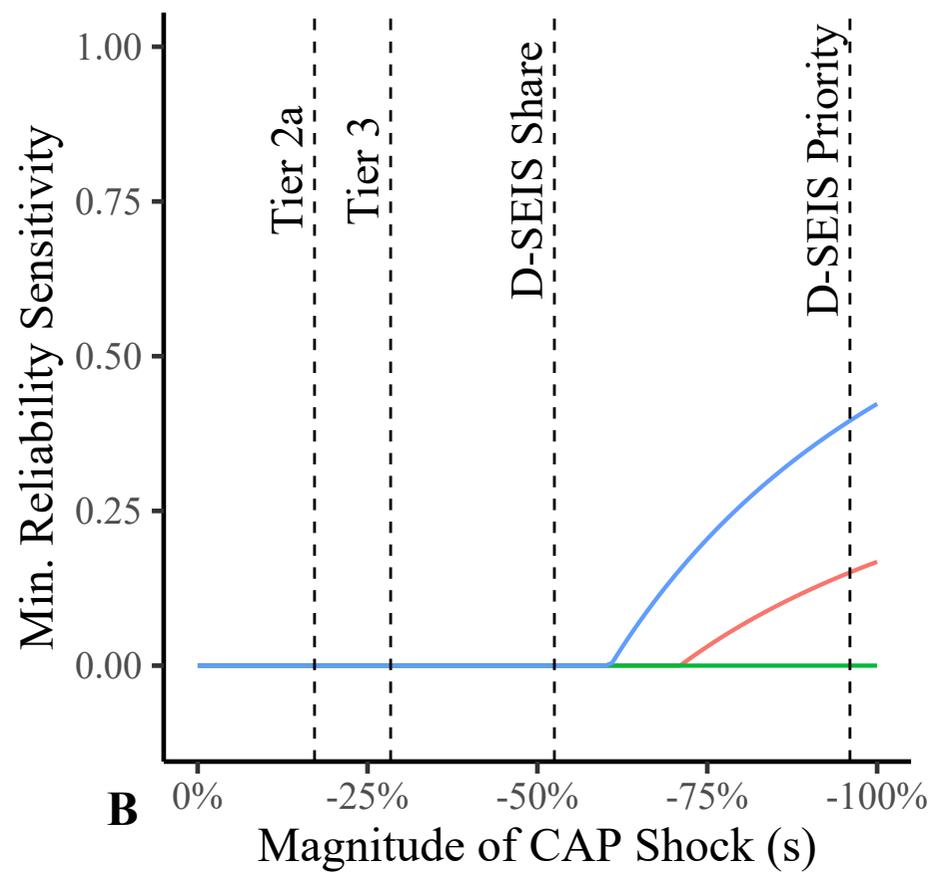
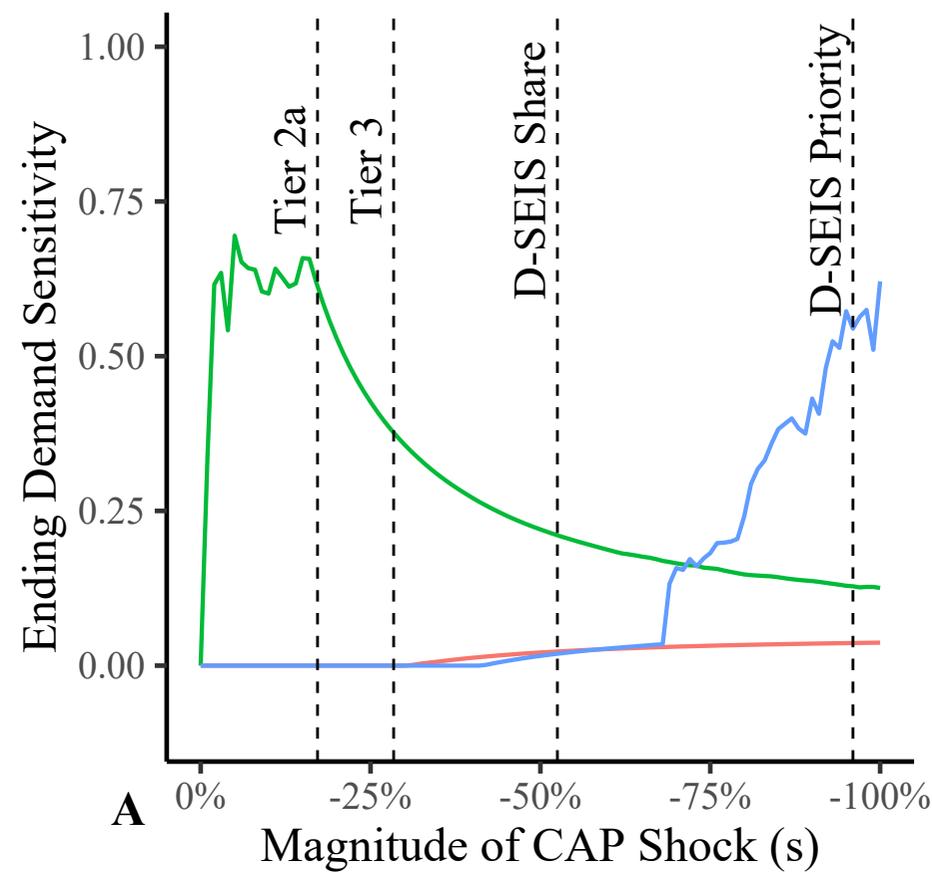


Figure 5.

City — PHX — QC — Sc

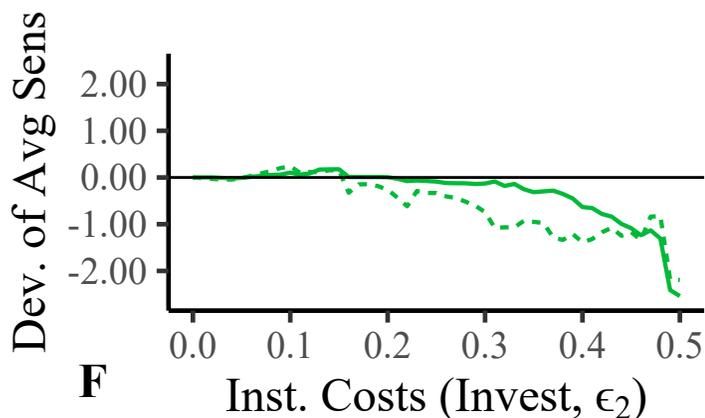
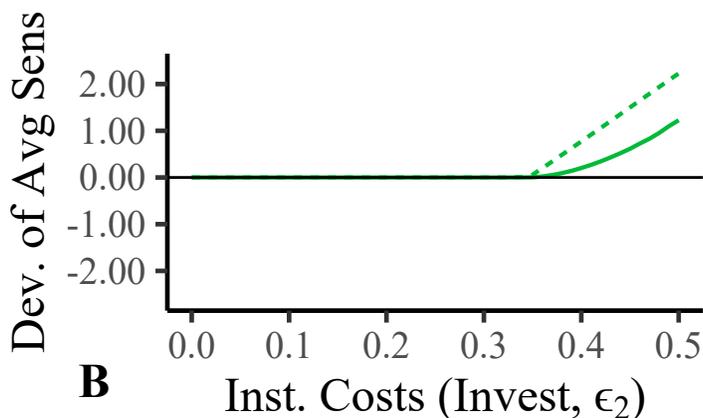
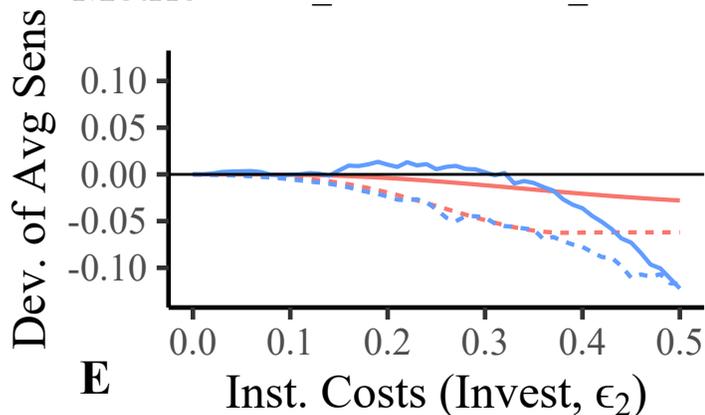
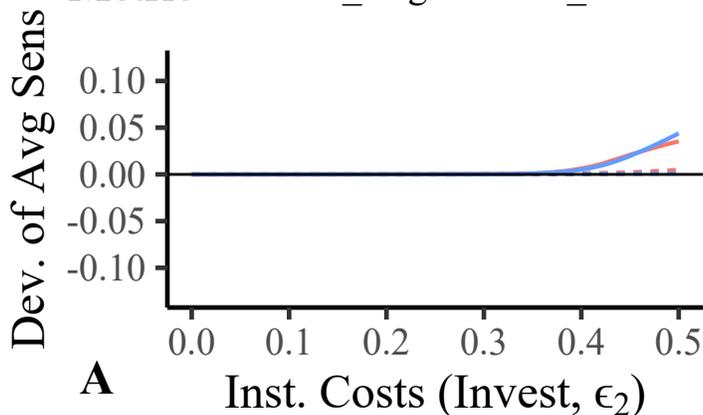
Reliability Metrics

Citizen Burden Metrics

Metric — Rel_Avg - - - Rel_Min

Metric — d_End - - - Rates_End

Investment Action Situation



Rate-Setting Action Situation

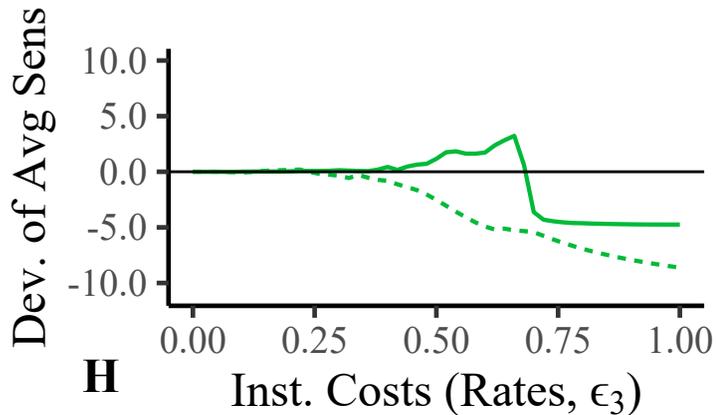
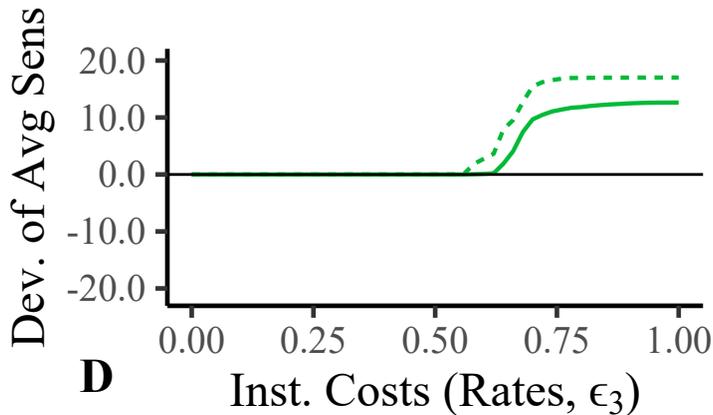
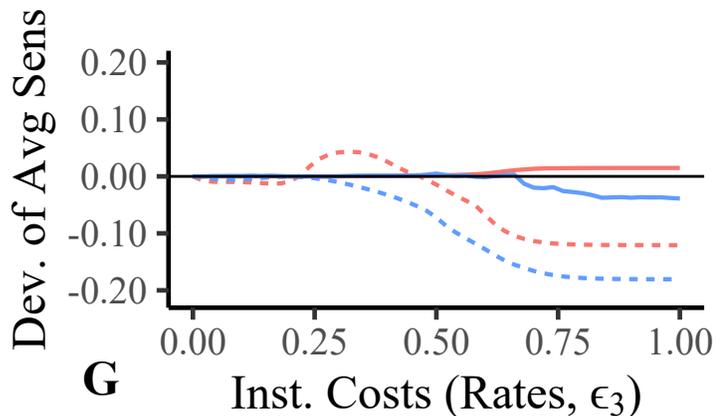
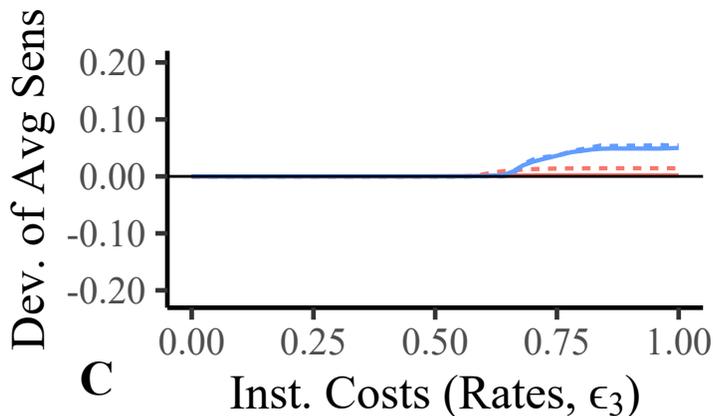
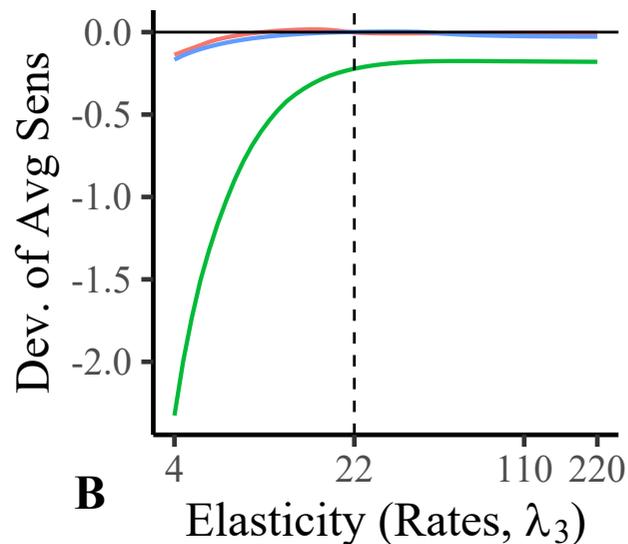
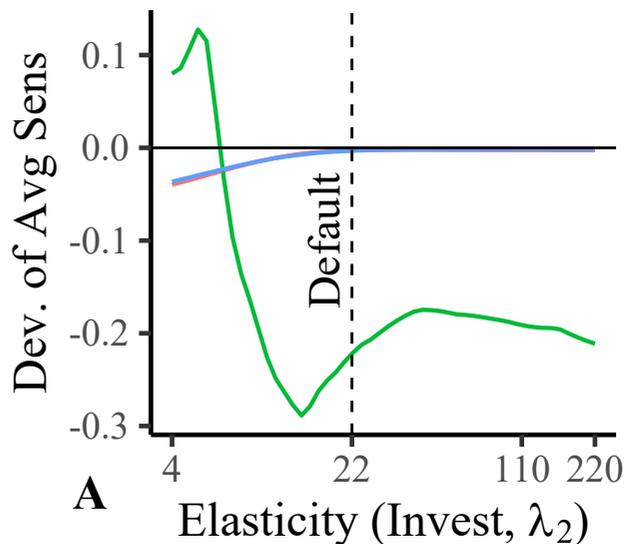


Figure 6.

Investment Action Situation Rate-Setting Action Situation

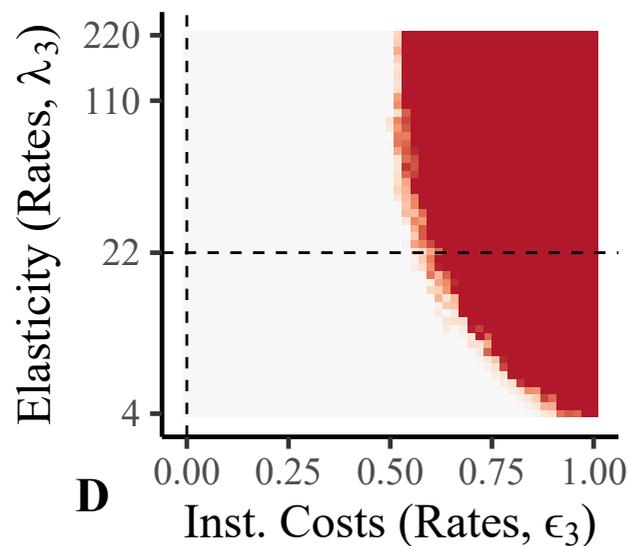
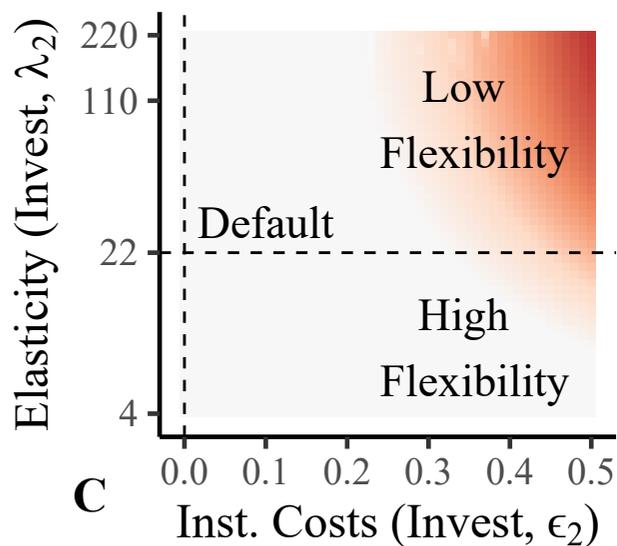
Ending Rates
(Avg Sens)



City

- PHX
- QC
- Sc

Min Reliability
(Queen Creek)



Ending Rates
(Queen Creek)

