# Empirical Dynamic Modeling Reveals Complexity of Methane Fluxes in a Temperate Salt Marsh

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#### Abstract

Methane dynamics within salt marshes are complex because vegetation types, temperature, oscillating water levels, and changes in salinity and redox conditions influence CH4 production and emission. These non-linear and complex interactions among variables affect the traditionally expected functional relationships and present challenges for interpretation and developing process-based models. We employ empirical dynamic modeling (EDM) and convergent cross mapping (CCM) as a novel approach for characterizing seasonal/multiday and diurnal CH4 dynamics by identifying causal variables, lags, and interconnections among multiple biophysical variables within a temperate salt marsh using five years of eddy covariance data. EDM/CCM is a nonparametric approach capable of quantifying the coupling between variables while determining time scales where variable interactions are most relevant. We found that gross primary productivity, tidal creek dissolved oxygen, and temperature were important for seasonal/multiday dynamics (rho=0.73-0.80), while water level was most important for diurnal dynamics during both the growing and dormancy phenoperiods (rho=0.72 and 0.56, respectively). Lags for top causal variables (gross primary productivity, tidal creek dissolved oxygen, temperature, water level) occurred between 1-5 weeks at the seasonal scale and 1-24 hours at the diurnal scale. The EDM had high prediction capabilities for intra-/inter-seasonal patterns and annual CH4 sums but with limitations to represent large infrequent fluxes. Results highlight the importance of non-linearity, causal drivers, lag times, and interconnections among multiple biophysical variables that regulate CH4 fluxes in tidal wetlands. This study presents a new dimension for analyzing CH4 fluxes, which will prove helpful to test current paradigms in wetlands and other ecosystems.

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# 1 Empirical Dynamic Modeling Reveals Complexity of Methane

# 2 Fluxes in a Temperate Salt Marsh

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### 17 Abstract

18 Methane dynamics within salt marshes are complex because vegetation types, temperature, oscillating 19 water levels, and changes in salinity and redox conditions influence CH4 production and emission. These 20 non-linear and complex interactions among variables affect the traditionally expected functional 21 relationships and present challenges for interpretation and developing process-based models. We employ 22 empirical dynamic modeling (EDM) and convergent cross mapping (CCM) as a novel approach for 23 characterizing seasonal/multiday and diurnal  $CH_4$  dynamics by identifying causal variables, lags, and 24 interconnections among multiple biophysical variables within a temperate salt marsh using five years of 25 eddy covariance data. EDM/CCM is a nonparametric approach capable of quantifying the coupling 26 between variables while determining time scales where variable interactions are most relevant. We found 27 that gross primary productivity, tidal creek dissolved oxygen, and temperature were important for 28 seasonal/multiday dynamics (rho=0.73-0.80), while water level was most important for diurnal dynamics 29 during both the growing and dormancy phenoperiods (rho=0.72 and 0.56, respectively). Lags for top 30 causal variables (gross primary productivity, tidal creek dissolved oxygen, temperature, water level) 31 occurred between 1-5 weeks at the seasonal scale and 1-24 hours at the diurnal scale. The EDM had high 32 prediction capabilities for intra-/inter-seasonal patterns and annual CH<sub>4</sub> sums but with limitations to 33 represent large infrequent fluxes. Results highlight the importance of non-linearity, causal drivers, lag 34 times, and interconnections among multiple biophysical variables that regulate CH<sub>4</sub> fluxes in tidal 35 wetlands. This study presents a new dimension for analyzing  $CH_4$  fluxes, which will prove helpful to test 36 current paradigms in wetlands and other ecosystems.

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38 Keywords: Methane flux, Saltmarsh, Nonlinear dynamics, Methane prediction, Empirical dynamic
 39 modeling, Tidal wetland

# 41 Plain Language Summary

42 The movement of methane gas in salt marshes is complex and influenced by various factors such 43 as plant types, temperature, water level, and changes in water salinity and oxygen levels. These 44 factors interact in intricate ways, making it difficult to predict the outcome of methane 45 production and movement. We applied a new method of studying methane dynamics using 46 Empirical Dynamic Modeling (EDM) and Convergent Cross Mapping (CCM). After analyzing five years of ecosystem-scale measurements of methane fluxes, we show that the amount of 47 48 oxygen in the water, the temperature, and the amount of light received by plants are crucial for 49 understanding regulating methane fluxes within days and across seasons. The interconnections 50 among these variables are complex, and methane fluxes may have delayed responses, which 51 highlight the importance of these interactions. This research improves our comprehension of how 52 environmental factors interact to affect methane fluxes in wetlands.

# 54 Key points

- 55 Methane dynamics in salt marshes are complex and influenced by multiple variables.
- 56 Empirical dynamic modeling (EDM) and convergent cross mapping (CCM) are novel
- 57 approaches to characterizing CH<sub>4</sub> dynamics.
- 58 Results highlight non-linearity, causal drivers, lag times, and interconnections among multiple
- 59 biophysical variables for CH<sub>4</sub> dynamics.

62 Methane (CH<sub>4</sub>) is the second most potent greenhouse gas and a crucial atmospheric trace gas with 63 a warming potential 25-38 times greater than carbon dioxide (CO<sub>2</sub>) (Wuebbles and Hayhoe 2002, 64 Bridgham et al. 2013, Neubauer and Megonigal 2019). Following a largely unreconciled plateau during a 65 stabilized phase from 2000-2006, there has been a steady global rise, with wetlands likely responsible for 66 most natural contributions (Jackson et al. 2020). While there is evidence that this increase is the result of 67 both anthropogenic and biogenic emissions (Stavert et al. 2022), there is no easy mitigation option for 68 biogenic sources, and climate change could increase these emissions resulting in positive feedback for the 69 global carbon cycle (Dean et al. 2018, Zhang et al. 2017). This is especially important for coastal 70 wetlands where warming and sea level rise are expected to impact  $CH_4$  fluxes directly through changes in 71 microbial metabolism (Yvon-Durocher et al. 2014), and indirectly from effects on co-dependent factors 72 such as plant productivity (GPP) and sediment redox conditions (Strom et al. 2015, Liu et al. 2019, 73 Seyfferth et al. 2020). 74 Mechanisms driving  $CH_4$  fluxes are complex within coastal wetlands with many potential 75 regulators (Huertas et al. 2019, Vázquez-Lule and Vargas 2021). Water level and waterlogged conditions 76 have been identified as a substantial control for methanogenesis (Li 2007), but these ecosystems have

been considered low CH<sub>4</sub> emitters despite prevailing anoxic conditions (Conrad 2020, Bartlett et al. 1987,

78 Wang, Zeng and Patrick 1996, Borges and Abril 2011). This paradigm relies on the idea that acetoclastic

79 or hydrogenotrophic methanogenesis are the dominant biogeochemical pathways and prevailing sulfate-

80 reducing bacteria outcompete methanogens (Poffenbarger, Needelman, and Megonigal 2011). However,

81 new evidence indicates coastal wetlands could have high CH<sub>4</sub> fluxes where other biogeochemical

82 pathways, such as methylotrophic methanogenesis, play essential roles (Al-Haj and Fulweiler 2020,

83 Seyfferth et al. 2020, Conrad 2020). In addition, it has been recognized that multiple environmental

84 factors (e.g., water level, temperature, GPP, salinity) regulate CH<sub>4</sub> fluxes at different temporal scales with

85 potential nonlinear interactions (Huertas et al. 2019, Reid et al. 2013, Li et al. 2018). Consequently, it is

86 imperative that we accurately understand how biophysical controls regulate  $CH_4$  fluxes in coastal 87 wetlands to better understand ecosystem response to environmental change.

88 Our past knowledge of CH<sub>4</sub> dynamics in tidal wetlands has vastly relied on chamber-based 89 manual measurements. This information has influenced the development of most functional relationships 90 (e.g., salinity thresholds or temperature dependency) and, subsequently their incorporation into models 91 used to predict CH<sub>4</sub> fluxes. However, chambers often face logistical challenges and are sporadic in time 92 and space, restricting our understanding by delivering limited or incomplete information (Hill and Vargas 93 2022, Kim 2007, Yang et al. 2021). Technological advances have allowed us to collect higher frequency 94 CH<sub>4</sub> fluxes at the ecosystem scale using the eddy covariance technique (EC) (Morin 2019, Baldocchi 95 2003). This data can be leveraged to explore relationships in greater detail, providing a wealth of 96 information regarding the dynamics of driving mechanisms. Previous studies have analyzed how 97 biophysical variables regulate ecosystem-scale CH<sub>4</sub> fluxes in salt marshes (Vázquez-Lule and Vargas 98 2021, Hill and Vargas 2022, Huertas et al. 2019, Li et al. 2018, Reid et al. 2013). Other studies outside 99 salt marshes have used machine learning techniques (Rey-Sanchez et al. 2018, Zaki and Abdul-Aziz 100 2022) or time series analysis and information theory to identify the dominant controls of ecosystem-scale 101 CH<sub>4</sub> fluxes (Knox et al. 2021, Sturtevant et al. 2016). Together, these studies provide insights into the 102 complex mechanisms and hint that nonlinear dynamics may be more relevant for explaining ecosystem-103 scale CH<sub>4</sub> fluxes. Therefore, there is a need to identify potential lags, interactions, and interconnections to 104 provide insights about causality and biogeochemical mechanisms controlling CH<sub>4</sub> fluxes across wetland 105 types.

106 The complexity of  $CH_4$  dynamics in coastal wetlands results from interacting variables that 107 cannot be readily isolated to identify independent functional relationships (Morin 2019; Vazquez-Lule 108 and Vargas 2021). This complexity also brings an opportunity to test alternative methods for data 109 analysis. As a novel approach, we employed a form of nonlinear state space reconstruction referred to as 110 empirical dynamic modeling (EDM; Sugihara and May 1990, Sugihara et al. 2012). We propose EDM as 111 an alternative to techniques such as Granger Causality, wavelets, and information theory to unravel

112 complex non-linear CH<sub>4</sub> dynamics using the wealth of information from continuous time series generated 113 from EC data (Schafer, Tripathee et al. 2014, Sturtevant, Ruddell et al. 2016, Li, Dai et al. 2018, Knox, 114 Bansal et al. 2021). This method falls within nonlinear dynamical systems but differs from other widely 115 used prediction-based methods, such as Granger Causality, which holds the criteria of separability 116 (Granger 1969) and may be less effective in identifying weak-moderate couplings (Guo et al. 2022). The 117 premise of EDM relies upon a simplex projection algorithm, the nearest neighbor forecasting method, that 118 tracks the evolution of nearby points within a lagged coordinate state space reconstruction or embedding 119 (Hsieh et al. 2005). In other words, simplex projection attempts to predict future values based on when 120 similar patterns were observed in the past, with more similar past patterns assigned a higher weight for 121 calculating the mean location of the predicted point (Petchey, 2016). In this study, we implement a form 122 of EDM known as convergent cross mapping (CCM) as a novel approach and alternative technique to 123 disentangling CH<sub>4</sub> dynamics with EC data within a temperate coastal salt marsh (Munch, Rogers, and 124 Sugihara 2022). We chose the CCM approach because it can describe the complex interrelations in a 125 dynamic ecological system where dependencies cannot be evaluated independently. 126 Here we examine five years of ecosystem-scale CH<sub>4</sub> fluxes using the EC technique and CCM to

127 study seasonal and diurnal CH<sub>4</sub> fluxes (i.e., dependent variable) by quantifying the coupling between 128 independent variables (i.e., predictors) while considering time lags and interconnections. Briefly, CCM 129 uses state space reconstruction methods that allow us to analyze how variables interact and change with 130 the flow of time by examining past variable states to determine if similar dynamics are embedded in the 131 presently observed target variable (i.e., CH<sub>4</sub>). This method is appropriate for informing causality within 132 complex nonlinear systems where lags or inconsistent relationships commonly manifest (McGowan et al. 133 2017), and works within the time domain, which is fundamentally different from wavelet analysis which 134 focuses on the frequency domain (Vargas et al. 2010). CCM can also identify complex interactions where 135 the indirect effect of one variable is relevant via a second variable that directly influences the dependent 136 variable. Consequently, CCM can provide interpretable empirical models that can be used to develop 137 predictions for gap filling of data or forecasting applications. Because CH<sub>4</sub> dynamics exhibit complex

patterns generated by multiple interactions, we propose that CCM is best suited to identify the nonlinearities and potential causal relationships that regulate  $CH_4$  dynamics in wetlands and other terrestrial ecosystems.

#### 141 **2.** Materials and Methods

142 **2.1 Study Site** 

143 This study was performed at the St. Jones Reserve as part of the Delaware National Estuarine 144 Research Reserve System (DNERR). The site is representative of a mid-Atlantic salt marsh with a mean 145 elevation of  $0.60 \pm 0.26$  m relative to the NAVD88 datum (McKenna et al., 2018), yet is still influenced 146 by semi-diurnal tidal activity and site hydrology (i.e., riverine and groundwater flows). Vegetation is 147 dominated by a monoculture of short-form S. alterniflora (~66%). The remaining cover (~33%) is 148 associated with tall S. alterniflora, S. cynosuroides, and P. australis, which flank tidal creeks, and S. 149 patens and P. australis along the upland terminus (Vázquez-Lule and Vargas 2021). Previous studies 150 conducted in this wetland suggest the presence of methylotrophic methanogenesis responsible for very 151 high CH<sub>4</sub> concentrations (>200  $\mu$ M) within the sediments (Seyferth et al. 2020) but with low CH<sub>4</sub> 152 sediment-atmosphere emissions (Capooci and Vargas 2022) These findings challenge the current 153 paradigm that low methanogenesis is expected in tidal wetlands and consequently new approaches are 154 needed to identify the complexity of the underlying biophysical drivers.

# 155 2.2 Data Acquisition

# 156 2.2.1 Eddy Covariance Measurements

157 The eddy covariance (EC) technique was used to measure the ecosystem-scale net exchange of

158 CH<sub>4</sub> with an open path near-infrared gas analyzer (Li-7700, Licor, Lincoln, NE, USA) and ecosystem-

159 scale net exchange of CO<sub>2</sub> (NEE) with an enclosed path infrared gas analyzer (Li-7200, Licor, Lincoln,

160 NE, USA), and wind components with a 3D sonic anemometer (Gill Windmaster Pro, Gill Instruments,

161 Lymington, UK) recording measurements at 10 Hz. Preprocessing was completed in Eddy Pro (version 162 7.0.6), which consisted of time lag compensations, double coordinate rotation of wind components, and 163 Reynolds block averaging to calculate 30-minute fluxes of  $CH_4$  and  $CO_2$ . Expanded preprocessing and 164 tower set-up descriptions can be found elsewhere (Vázquez-Lule and Vargas 2021, Hill and Vargas 165 2022). During post-processing, several standardized Ameriflux QA/QC procedures were applied, 166 including removing values flagged for low quality during pre-processing, calculating nighttime storage fluxes determined by Eddy Pro, and range filtering of CO<sub>2</sub> to remove outliers falling beyond +/- 50 umol 167 m<sup>-2</sup> s<sup>-1</sup>. A range filter was not applied to CH<sub>4</sub> data as spikes since ebullition could occur, and the goal was 168 169 to incorporate all CH<sub>4</sub> dynamics within the 30-minute block averaging period. In addition, we used an optimal friction velocity ( $u^*$ ) threshold of 0.069 m s<sup>-1</sup> to remove low turbulence conditions and applied a 170 171 fetch/footprint filter to exclude fluxes originating from forested regions beyond the marsh terminus 172 (Vázquez-Lule and Vargas 2021).

173 Data gaps occurring in NEE were filled using marginal distribution sampling (MDS) with 174 customized site variables consisting of energy fluxes (sensible and latent heat; H and LE), radiation, air 175 temperature (Tair), soil temperature (Tsoil), relative humidity (RH) and vapor pressure deficit (VPD) 176 (Vázquez-Lule and Vargas 2021). Partitioning of NEE into component fluxes of gross primary 177 productivity (GPP) was completed with the REddyProc R package (Reichstein et al., 2005; Wutzler et al., 178 2018) based on the standard nighttime method (Reichstein et al., 2005). To fill gaps in  $CH_4$  fluxes, we 179 applied a random forest technique with the Caret R package and used a full suite of relevant site variables 180 (Fig. S2) (Kim et al., 2020; Kuhn et al. 2016). This method was selected because it is less biased when 181 identifying predictors of CH<sub>4</sub> flux and because CCM requires continuous gap-filled data (Kim et al., 182 2020; Chang, Ushio, and Hsieh 2017).

# 183 2.2.2 Meteorology and Water Quality Measurements

184 Meteorological variables included air temperature (Tair) and relative humidity (RH, HC2-S3,
185 Campbell Scientific, Logan, UT), precipitation (Precip) (TE 525, Tipping Bucket Rain Gauge, Campbell

186 Scientific, Logan, UT), photosynthetically active radiation (PAR) (SQ-110, quantum sensor, Apogee, 187 Logan, UT), air pressure (Patm) (CS-106, Vaisala, Vantaa, Finland) and wind speed (WSpd) (05103-L 188 Wind Monitor, Campbell Scientific, Logan, UT). VPD was calculated using Tair and RH based on the 189 Tetens formula (Murray 1967). Water quality data were collected from the tidal creek adjacent to the EC 190 tower using a YSI EXO2 sonde outfitted with EXO sensors. Measured variables included water 191 temperature (Twater), water level (Level), salinity (Sal), and dissolved oxygen (DO). Meteorological and 192 water quality data were averaged for 30 minutes to align with flux data. Minor gaps in meteorological 193 data were filled using available on-site sensors. Gaps in water quality data, which mainly occurred during 194 the dormancy phenoperiod, were filled using empirical relationships from another YSI EXO2 sonde 195 located approximately 2.0 km upstream within the St. Jones River, which shares connectivity to the tidal 196 channel. Meteorological and water quality data were collected under the National Estuarine Research 197 Reserve (NERR) Centralized Data Management Protocol (Kennish, 2019) and can be accessed from the 198 NERR System Centralized Data Management Office (NERR CDMO; station: delsjmet-p).

# 199 **2.3 Data Selection and Phenology**

200 For examining seasonal CH<sub>4</sub> dynamics, we selected five years spanning 2017-2021 to calculate 201 daily means of non-gap-filled observations. This period was chosen as all years contained minimal 202 continuous gaps (<30 days). To examine diurnal dynamics, we used data only from 2020-21, the period 203 with the most complete record (i.e., containing 15% of data gaps) and when no substantial water surges or 204 storm events occurred. To delineate these periods, we determined phenology dates with the Phenopix R 205 package using the greenness chromatic coordinate (GCC) from a site phenocam (Filippa et al. 2016, Hill, 206 Vázquez-Lule and Vargas 2021). Season start and end dates were calculated based on the upturn and 207 recession dates of the annual GCC curve from daily midday images (Gu et al., 2009). To avoid an artifact 208 of calendar years that would result in discontinuous time series for dormancy data, we included the tail 209 end of 2020 and excluded the tail end of 2021.

#### 210 2.4 Convergent Cross Mapping

While previous studies have described in detail the methodology behind CCM (Ushio and
Kawatsu 2020, Sugihara and May 1990, Sugihara et al. 2012, Chang, Ushio and Hsieh 2017, Tsonis et al.
2018, Wang et al. 2018), we provide a brief overview of the implementation of CCM and how we
interpreted results as they pertain to the goals of the current study (Fig. S1). All CCM analysis was
completed within the rEDM R package (Ye et al. 2016).

216 To evaluate the causality between variables, we applied CCM between CH<sub>4</sub> and all independent 217 variables via the function CCM (Ye et al. 2016). The CCM function is essentially a wrapper for the 218 simplex algorithm but accesses the level improvement in nearest neighbor predictions within the state 219 space as the data sample size increases. Specifically, this tests if lags of a predictor variable can be used to 220 predict a target variable (e.g., does GPP have a causal relationship with  $CH_4$ ) (Tsonis et al. 2018, 221 Schiecke et al. 2015, Sugihara et al. 2012). It is a standard procedure also to test the opposite cross 222 mapping (i.e., does CH<sub>4</sub> have a causal relationship with GPP). In a dynamic system, information about the 223 causal variable becomes embedded within the target variable (i.e., information about past causal variables 224 is observed in the present target variable). CCM is carried out using successively larger data samples or 225 libraries. Predictive skill is evaluated with the output parameter rho (the overall skill of the cross-226 mapping) and is expected to increase with increasing library (data) size (i.e., convergence) subsampled 227 randomly from the main dataset as the first criterion for causation (Chang et al. 2017). The predictive skill 228 at maximum library size (the complete data set) is further compared to results generated from a surrogate 229 dataset containing a preserved seasonal phase amplitude with the original data randomized in time. Cross 230 mappings need a higher final predictive skill than the final surrogate predictive skill as a second criterion 231 for causation via this significance test (Tsonis et al. 2018). Aside from verifying convergence with 232 increasing data libraries, all figures in this study present results from the full data library size for both 233 surrogate and real data, representing 5 complete years of data.

234 Two steps are applied before CCM, which provide information about the time series of interest 235 (CH<sub>4</sub>) (Li et al. 2021). The initial step is determining the embedding dimension (e) via the function 236 *EmbedDimension*, which iteratively takes successively longer lags of a set time period defined by the 237 parameter Tau (T). For this study, we set T to the default value of T=1 to represent one day for seasonal 238 analyses and one hour for diurnal analyses. The lagged time series are then projected into an e-239 dimensional state space, forming a shadow manifold or state space projection (Ye et al. 2016) (Fig. S3). 240 The resulting manifolds are then projected back onto a coordinate axis with the best embedding 241 dimension resulting in a nearly 1:1 mapping of the original time series (Fig. S4). The number of 242 dimensions (e) required to recreate the original time series is both a metric of complexity and an a priori 243 requirement for CCM. The second step is confirming nonlinear dynamics by constructing sequential 244 locally weighted global linear maps (S-map). S-maps are like nearest neighbor predictions from the 245 simplex algorithm, but instead of considering the average localization of the nearest neighbors, all 246 neighboring points are regarded with more weight given to closer neighbors through an exponential 247 localization function (Sugihara et al. 1994). This is done via the function PredictNonlinear, which 248 determines if predictions made by S-maps depend on the local state of the variable being predicted as 249 defined by the parameter theta  $(\Theta)$  in a locally weighted linear regression function. If predictive skill 250 increases at  $\Theta > 0$ , we assume that predictions are highly dependent on ecosystem state space (i.e., variable 251 interactions change with time or under specific conditions) and the hallmark of a nonlinear dynamical 252 system (Chang et al. 2017).

We applied year-round daily data for seasonal dynamics with a minimum library size of n=15 days and a maximum of n=1,800 days. For diurnal dynamics, hourly data was used from the selected growing and dormancy phenoperiods using a minimum library size of n=15 hours and a maximum of n=2,900 and n=2,700 hours, respectively. CCM was carried out using 100 sample runs with the default time period for lags set to 1 unit (Tp=1 day or hour). The execution radius, which ignores nearest neighbors within the state space projection, was set at one day or 24 hours to help eliminate influence from temporal autocorrelation. To generate surrogate datasets, the original data was randomized in time

with the seasonal phase amplitude preserved for daily data by setting the method to 'seasonal'. For hourlydata, the method was set to 'random' (Deyle et al. 2016).

#### 262 **2.5 Model Interpretation with Extended CCM and Causal Network Maps**

263 Conventional CCM provides a robust means of identifying causality between variable pairs, yet it 264 fails to capture the optimal time period of the causal influence. Extended CCM is a computationally 265 intensive technique that involves applying CCM iteratively across multiple time period lags to determine 266 time periods where cross-mapping skills are maximized. The procedure is also used to help resolve issues 267 of generalized synchrony where cross-mapping is significant in both directions (i.e., Tair causality on 268 CH<sub>4</sub> and CH<sub>4</sub> causality on Tair) (Example: Fig. S7 panel C) as the most probable causal relationship will 269 generally have a maximum predictive skill which peaks within the negative time domain (i.e., negative 270 lags) (Fig. S8) (Ye et al. 2015b). Causal influence can occur synchronously, over a day, several hours, or 271 an extended period, several consecutive or separate days (Sun et al. 2021). Data were binned into 7-day 272 periods to examine seasonal dynamics, and iterations were made spanning +/- 91 days at weekly intervals. 273 For diurnal dynamics, hourly data was used directly, and iterations were made spanning +/- 24 hours. 274 By considering both the optimum identified time lags and the overall strength of the causal 275 relationship, network maps can be constructed to assess how the full suite of variables interact to 276 influence  $CH_4$  fluxes. Conceptually, this is achieved by summarizing the strength of coupling and optimal 277 lag times between all tested variable pairs, including all ancillary variables with only positively identified 278 (significant) connections mapped. In general, direct connectivity is exhibited when the predictive skill is 279 high and lag times are short and indirect connectivity is exhibited when the prediction skill is lower and 280 lag times are long (Fan et al. 2020). We selected ranking categories of high, moderate, and low based on 281 rho values of 0.60-1.0, 0.30-0.59, and 0-0.29, respectively. This technique allows variables with indirect 282 links that may appear to influence CH<sub>4</sub> fluxes to be sorted from variables with highly confident direct 283 causal links, yet it remains difficult to distinguish consistent directionality (i.e., sign positive or negative) 284 of the connection as relationships can vary in time (i.e., state dependence) (Deyle et al., 2016).

#### 285 **2.6 Multivariate Predictive Modeling**

286 As an additional feature and promising application of EDM, we also made model predictions 287 using a newly constructed multivariate EDM model. We used an out-of-sample validation method with 288 2017-2019 as the training interval and 2020-2021 as the prediction interval. Only data from the training 289 period was used to make predictions. Predictor data was scaled to reduce dimensional distortion in the 290 state space. The provided prediction variance based on nearest neighbors in the state space was used to 291 calculate a standard deviation as a metric of uncertainty. We included all variables identified as coupled 292 to  $CH_4$  fluxes by daily CCM analysis: temperature components of soil, water, and air, DO, GPP,  $H_2O$ 293 flux, Level, LE, WSpd, PAR, Patm, RH, and Precip. While issues of cross-correlation would potentially 294 violate assumptions for linear models, each of these variables causally impacts CH<sub>4</sub> fluxes in different 295 ways as determined by information embedded in the CH<sub>4</sub> time series. Thus, the inclusion of variables that 296 exhibit cross-correlation is possible with EDM.

**3. Results** 

# 298 3.1 Site Characteristics, Phenology, and CH<sub>4</sub> Fluxes

299 During the study period (2017-2021), the annual average air temperature (Tair), salinity (Sal), and 300 water level (Level) were  $13.9 \pm 9.1$  °C,  $8.9 \pm 4.1$  ppt and  $0.25 \pm 0.2$  m above sea level, respectively with a 301 tidal range of 1.2 m and maximum daily salinity as high as 19.5 ppt. The average annual precipitation 302 (Precip) ranged from 567-779 mm, which is received evenly across the year (Fig. 1) (Tables S1-S3). The 303 growing phenoperiod typically begins with green-up events occurring in April or early May, and 304 dormancy commences with complete vegetation senescence in November. For the reference year utilizing hourly data (2020-21), the dormancy phenoperiod began on October 25<sup>th</sup>. It ended at the start of the 2021 305 growing phenoperiod on April 26<sup>th</sup>, which continued until the subsequent dormancy phenoperiod on 306 October 23<sup>rd</sup>. Despite high levels of GPP, which ranged from 1,412-1,609 g C m<sup>-2</sup> yr<sup>-1</sup>, the site was a net 307

308 carbon source for four out of the five years (NEE range: -126 to 221 g C m<sup>-2</sup> yr<sup>-1</sup>) and was a constant 309 annual source of CH<sub>4</sub> with emissions ranging from 15.4-16.5 g C m<sup>-2</sup> yr<sup>-1</sup> (Fig. 1).



310

Fig. 1: Time series of mean daily data for top identified predictive drivers and CH<sub>4</sub> flux from 2017-21. a)
Air, soil, and water temperature (green, brown, and blue, respectively), b) level (gray), c and d) salinity
(pink) and dissolved oxygen (DO) (teal), d) atmospheric pressure (Patm) (purple), e) gross primary
productivity (GPP) (blue), and f) CH<sub>4</sub> flux (black).

#### 316 **3.2 CH<sub>4</sub> Time Series Characteristics**

We determined that  $CH_4$  is a highly dimensional independent variable that required at least ten dimensions to recreate the original dynamics (Fig. S5). Despite a slightly lower peak at four dimensions, we selected the dimensionality with the greatest predictive skill to capture a full range of potential variable dynamics; because CCM is expected to identify coupling and quantify the overall strength of all possible relationships. Consequently, we confirmed that the response of  $CH_4$  flux was highly dependent on state space or transient ecosystem conditions (i.e., dynamical and nonlinear) (Fig. S6).



323

**Fig. 2:** Results from Convergent cross mapping (CCM) and extended CCM for seasonal dynamics from daily data spanning 2017-21. a) Identification of strongest predictor variables for  $CH_4$  flux using CCM. Red points represent the maximum predictive skill achieved with simulated surrogate data. Black points indicate prediction skill of real data. ns denotes non-significant cross-mapping due to a lack of convergence or overlap with surrogate data results. Outlined variables exhibit a prediction skill > 0.6. b) Optimal time lags (in number of days) determined from extended CCM.

330

# 331 3.3 Causality Analysis with CCM

Examination of daily data with CCM revealed that the top causal variables at the

333 seasonal/multiday scale (rho>0.70) consisted of temperature components (Tsoil, Twater, Tair), DO, and

| 334 | GPP (Fig. 2a). These select variables were a distinct grouping, with all other variables falling below a        |
|-----|-----------------------------------------------------------------------------------------------------------------|
| 335 | threshold of rho=0.50. The remaining variables (H <sub>2</sub> O flux, Level, LE, PAR, WSpd, Patm, RH, Precip;  |
| 336 | sorted from higher to lower coupling strength) were identified as causal (rho=0.10-0.50) except for Sal,        |
| 337 | VPD, and H, which did not meet convergence criteria or overlapped with surrogate data (Fig. S7). The            |
| 338 | growing phenoperiod was complex regarding the number of predictors and lags, yet physical factors also          |
| 339 | played an important role. Examination of hourly data for the diel scale analysis from the growing               |
| 340 | phenoperiod revealed Level as the top causal variable (rho~0.70), followed by a grouping that included          |
| 341 | Patm, Sal, DO, and Tsoil (rho=0.30-0.50) (Fig. 3a). The remaining variables (PAR, Twater, RH, GPP, H,           |
| 342 | H <sub>2</sub> O flux, WSpd, LE, VPD, Tair; sorted from higher to lower coupling strength) also had a degree of |
| 343 | coupling (rho=0.15-0.25), except Precip (Fig. S8). Examination of hourly data from the dormancy                 |
| 344 | phenoperiod revealed Level, DO, and Sal as top causal variables (rho=0.55-0.30) (Fig. 3c). Most                 |
| 345 | remaining variables (Patm, RH, WSpd, Tair, VPD, PAR, Precip; sorted from higher to lower coupling               |
| 346 | strength) also exhibited some coupling (rho=0.10-0.20), yet this period contained the largest number of         |
|     |                                                                                                                 |

347 non-significant variables, which included GPP, Tsoil, Twater, H, LE, and water fluxes (Fig. S9).



349 Fig. 3: Results from Convergent cross mapping (CCM) and extended CCM for diurnal/seasonal dynamics 350 from select hourly data spanning 2020-21. a) Growing phenoperiod CCM at maximum library size 351 between all variables and CH<sub>4</sub> flux, b) Growing phenoperiod optimal time lags for extended CCM 352 represented by the number of days (x-axis), c) Dormancy phenoperiod CCM at maximum library size 353 between all variables and CH<sub>4</sub> flux. d) Dormancy phenoperiod optimal time lags for extended CCM 354 represented by the number of days (x-axis). Red points represent the maximum predictive skill achieved 355 with simulated surrogate data. Black points indicate prediction skill of real data. ns denotes non-356 significant cross-mapping due to a lack of convergence or overlap with surrogate results. Outlined 357 variables exhibit a prediction skill > 0.3.

#### 359 3.4 Optimal Lags with Extended CCM

360 Implementation of extended CCM revealed that the influence of some variables is more 361 immediate while others are associated with a lag. While any identified lags indicate the optimal time 362 period of causal influence, these lags are not necessarily related to the overall strength of the causal 363 relationship. However, longer lags can mean more indirect relationships within the context of causal 364 network mapping. When applying daily data to unravel seasonal/multiday dynamics, we found that PAR, 365 LE, and H<sub>2</sub>O fluxes are associated with the most extended lag times occurring between 49-70 days 366 (seasonal/multiday scale) (moderate causal strength). These were followed by a grouping of top predictor 367 variables identified by the previous CCM analysis (Tsoil, Twater, Tair, GPP) (high causal strength), 368 which were most influential between 21-35 days (monthly scale). The remaining variables (DO, Level, 369 WSpd, Patm, RH, Precip) exerted a close temporal relationship (0 days) at a range of causal strengths, but 370 DO was important over a more extended period (0-3 weeks), followed by Level, Patm, and RH (0-1 371 week) and finally WSpd and Precip (0 days). However, Precip had the weakest coupling (Fig. 2b) (Fig. 372 S10).

373 When applying hourly data from the growing phenoperiod to unravel diurnal dynamics, we found 374 that lags were more complex. Both Level and Sal (strong and moderate causal strength, respectively) had 375 peaks occurring between 0-5 hours and again between 19-23 hours, but the influence of Level was also 376 persistent throughout the day (denoted as a gray bar in Fig. 3b). Temperature components had 377 successively shorter lags moving from Twater (9-13 hours) to Tsoil (5-7 hours) and finally Tair (3-6 378 hours) (all moderate causal strength). The remaining variables were influential immediately (0 hours), 379 with GPP extending the longest (0-9 hours), followed by Patm (0-5 hours), LE and water fluxes (0-3 380 hours), RH (0-1 hour) and finally DO, VPD, and H (0 hours) (moderate causal strength) (Fig. 3b) (Fig. 381 S11). Lag times were much shorter for the dormancy phenoperiod. Almost all causal variables (Level, 382 DO, Sal, Patm, RH, Tair, VPD, Tsoil) exerted a close temporal relationship (0 hours) with DO, VPD, 383 Tsoil, Patm, and Level active over the most extended period (0-10 hours). Precip was influential between

7-8 hours, yet again had the lowest causal strength while PAR, RH, WSpd, and Tair all exerted influence
between 0-5 hours) (moderate causal strength) (Fig. 3d) (Fig. S12).

## 386 **3.5 Causal Network Mapping**

387 Causal network maps reveal high interconnection among variables at multiple scales (seasonal, 388 multiday, and diurnal) and indicate that fluxes result from a suite of factors. When using daily data, we 389 discovered several interesting patterns (Fig. 4). First, PAR is important for CH<sub>4</sub> (rho=0.66), but its action 390 is likely indirect via greater influence on other relevant variables (e.g., H<sub>2</sub>O fluxes and temperature 391 components) (rho=0.65 and 0.90). The strong relationship between GPP and CH<sub>4</sub> fluxes (rho=0.86) was 392 influenced by VPD, Tair, Sal, and DO (rho=0.52-0.82), which are in turn influenced by RH, PAR, and 393 Level, respectively (rho=0.41-0.71). Other variables, such as RH and WSpd, were not causally connected 394 to any variables examined.





Fig. 4: Causal network map showing seasonal dynamics between all relevant variables for daily data spanning 2017-21. Numbers represent predictive skill followed by the median value of the optimum time lag. Colors represent the strength of coupling, green=strong (0.6-1.0), orange=moderate (0.30-0.59), and red (weak) (0-0.29). Solid lines represent direct relationships, and dashed lines represent indirect relationships. Variables outlined in bold represent the top causal variables identified by CCM analysis.

When applying hourly data for the growing phenoperiod, many highly connected variables identified from the seasonal analysis (Temperature components, GPP, DO) exert far less influence (Fig. 5a). The water level was the most directly important variable (rho=0.72) and was highly connected to both Sal and DO (rho-0.90-92), which exerted moderate influence on  $CH_4$  fluxes (rho=0.35-0.38). Over seven additional variables provided moderate-low influence, but all showed a similar predictive skill (rho=0.17-0.41), highlighting the complexity associated with the sub-daily growing phenoperiod. Hourly data from the dormancy phenoperiod exhibited the most straightforward network connectivity with the

- 408 least interdependencies (Fig. 5b). As with growing phenoperiod data, Level was the most directly
- 409 important variable (rho=0.60) and exerted a strong influence on Sal and DO (rho=0.81-89), which exerted
- 410 moderate influence on CH<sub>4</sub> fluxes (rho=0.38-0.49). Additional variables were associated with weaker
- 411 casual relationships (rho=0.12-0.22), and only temperature components, VPD, Patm, Sal, and DO were
- 412 dependent on the activity of other variables during this phenoperiod.





414 Fig. 5: Causal network maps showing diurnal/seasonal dynamics between all relevant variables for select

- 415 hourly data spanning 2020-21 for a) Growing phenoperiod and b) Dormancy phenoperiod. Numbers
- 416 represent predictive skill followed by the median value of the optimum time lag. Colors represent the
- 417 strength of coupling, green=strong (0.6-1.0), orange=moderate (0.30-0.59), and red (weak) (0-0.29). Solid

418 lines represent direct relationships, and dashed lines represent indirect relationships. Variables outlined in419 bold represent the top causal variables identified by CCM analysis.

#### 420 **3.6 Modeling Output**

421 The predictive multivariate EDM model fell within 1% of the expected 2-year annual sum derived from the gap-filled EC benchmark for CH<sub>4</sub> flux (31.6 g C m<sup>-2</sup>) (Fig. 7). This model incorporated 422 423 all of the positively coupled variables identified by CCM analysis including temperature components of 424 soil, water, air, GPP, DO, H<sub>2</sub>O flux, Level, LE, PAR, WSpd, Patm, RH, and Precip. This generated a 2year sum of 31.8 g C m<sup>-2</sup> (0.63% difference; slope 0.57) with an annual uncertainty of +/-29.3 g C m<sup>-2</sup>, 425 426 calculated from the state space. This approach had challenges representing high fluxes during senescence 427 and despite good overall predictions, carried a high overall uncertainty and small over-prediction bias for 428 lower magnitude fluxes ( $< 50 \text{ nmol m}^{-2} \text{ s}^{-1}$ ) (Fig. 6b). It should be noted that the uncertainty estimated for 429 the EDM model is based on the nearest neighbor distance weights within the projected state space and 430 although we scaled predictor variables, including more variables can distort the state space to a greater 431 extent.



Fig. 6: a) Multivariate EDM modeling results at the daily time step for prediction years 2020-21 and b)
linear regression comparisons. Black lines represent daily CH<sub>4</sub> flux measured by eddy covariance (EC) as
a benchmark +/- daily sd (grey shading), and red lines represent daily modeled CH<sub>4</sub> flux +/- prediction
variance sd (pink shading). The reported black sum is the 2-year total carbon budget from CH<sub>4</sub> calculated

437 with EC data (benchmark), and the red sum is the 2-year total carbon budget from CH<sub>4</sub> estimated with

438 modeled data and percent difference (gray). The blue line on regression plots is a 1:1 line.

439 **4.** Discussion

#### 440 4.1 Comparisons with Previous Wetland CH<sub>4</sub> Studies

441 There are limited CH<sub>4</sub> studies in salt marshes but it has been reported by linear methods that 442 temperature, PAR, plant productivity, or some proxy (e.g., GPP, biomass, chlorophyll concentration), 443 water level, and salinity are important predictors for CH<sub>4</sub> fluxes (Abdul-Aziz et al. 2018, Huertas et al. 444 2019, Martin and Moseman-Valtierra 2017, Poffenbarger et al. 2011.) Despite differing temporal 445 averaging periods (daytime, nighttime, year-round, multi-year) and different model structures (linear 446 regression, multivariate canonical correlation, multivariate generalized least squares), multiple approaches 447 have also been able to identify the importance of temperature, DO, and Patm for CH<sub>4</sub> dynamics 448 (Vázquez-Lule and Vargas 2021, Hill and Vargas 2022). Other multi-site studies have also identified 449 similar predictor variables with both linear and other nonlinear methods, especially temperature 450 components, in non-tidal freshwater wetlands (Knox et al, 2020). The mechanisms supported by 451 additional variables such as water level and GPP are more complex, with influence often manifesting 452 across several temporal scales (Sturtevant et al. 2016). Our results indicate that the effects of GPP are 453 highly dependent on the time domain at multiple scales (seasonal and diurnal) and multivariate 454 interactions among variables. While linear models may be capable of extracting similar variables in many 455 cases, these methods fail to identify dependencies, lags, and complex interrelationships. Thus, linear 456 approaches may still limit our interpretations because the influence of one variable can be the result of 457 indirect influence from a third variable, a phenomenon that can be indirectly inferred when applying 458 CCM.

| 459 | Several studies based on non-tidal freshwater wetlands have successfully applied alternative                        |
|-----|---------------------------------------------------------------------------------------------------------------------|
| 460 | nonlinear analysis methods (Knox et al. 2021, Sturtevant et al. 2016) or machine learning approaches                |
| 461 | such as ANN's or random forest techniques (Rey-Sanchez et al. 2018, Zhu et al. 2013, Abbasi et al. 2019,            |
| 462 | Zaki and Abdul-Aziz 2022, Morin et al. 2017). In these non-tidal freshwater wetlands, interactions                  |
| 463 | regulating fluxes do not include the same dynamics found within tidal wetlands including frequent                   |
| 464 | oscillation of water level and subsequent modulation of sediment salinity, dissolved oxygen, and redox              |
| 465 | conditions (Vazquez-Lule and Vargas 2021). Further, dominant vegetation species differ, which can                   |
| 466 | influence the biogeochemical conditions of the sediments and CH <sub>4</sub> production (Seyfferth et al. 2020,     |
| 467 | Yuan et al. 2016, Gao et al. 2018). Despite differences, these studies also show that the strongest relevant        |
| 468 | drivers are identified by both linear and nonlinear approaches, yet incomplete results from solely applying         |
| 469 | linear methods limit the full mechanistic understanding of the system (Sturtevant et al. 2016).                     |
| 470 | Specifically, issues of cross-correlation can restrict which variables are included to satisfy model                |
| 471 | requirements, and important lags and interrelationships fail to be identified. In fact, tidal saltwater             |
| 472 | wetlands have been purposely excluded from several FLUXNET- $CH_4$ analyses because they do not                     |
| 473 | follow the general pattern of linear relationships defined by freshwater wetlands, and an overall lack of           |
| 474 | data from these highly dynamic ecosystems creates difficulties when attempting to generalize across                 |
| 475 | multiple sites (Chang et al. 2021, Delwiche et al. 2021, Knox et al. 2021).                                         |
| 476 | In limited cases where nonlinear methods are applied, findings largely support our results,                         |
| 477 | demonstrating lags with GPP, temperature, and water level, but with temporal synchrony with VPD and                 |
| 478 | energy fluxes (Delwiche et al. 2021). For example, in a multi-site synthesis study of freshwater sites,             |
| 479 | median seasonal lags associated with water level extended 17±11 days, 8±16, and 5±14 days for Tair and              |
| 480 | Tsoil, respectively, and were synchronous for Patm while at the diel scale, GPP was lagged up to 4 hours            |
| 481 | and energy fluxes and VPD were synchronous (<1 hour) (Knox et al. 2021). A similar study from a                     |
| 482 | forested/shrub wetland found CH <sub>4</sub> lags associated with GPP extending up to 60 days (Turner et al. 2021), |
| 483 | potentially indicating different time dependencies based on vegetation type and structure. There is                 |
|     |                                                                                                                     |

484 overwhelming evidence supporting that temporal lags exist between multiple environmental variables and

485  $CH_4$  fluxes across different types of wetlands. Thus, it is important to include methods such as CCM to 486 identify such lags in the context of causality to properly represent the complexity of how biophysical 487 drivers control  $CH_4$  fluxes

# 488 4.2 Seasonal CH<sub>4</sub> Dynamics

489 Seasonal CH<sub>4</sub> dynamics are strongly controlled by temperature components, GPP, and DO, 490 indicating reliance on biological factors. Previous studies showed that predictors such as Level, Patm, and 491 VPD are important in a linear multivariate context but rank lower in causality in the current study 492 (Vázquez-Lule and Vargas 2021, Hill and Vargas 2022). This is likely because these variables are less 493 strongly coupled with CH<sub>4</sub> activity at this scale due to: a) limited occurrence as with transient low-494 pressure systems, which operate synchronously with CH<sub>4</sub> emissions and rarely exert influence on scales 495 extending beyond several days (Fig. 2b) (Mønster, Kjeldsen, and Scheutz 2019, Tokida et al. 2007, Knox 496 et al. 2021); or b) indirect mechanisms as with Patm and pressurized gas flow (Björn et al. 2022, Zhang 497 and Ding 2011), and VPD limiting stomatal conductance and the possible decrease in plant-mediated 498 transport. At the seasonal scale, temperature is a strong predictor in temperate ecosystems because it 499 regulates the window of active plant growth and belowground microbial metabolic rates (Zhu et al. 2019, 500 Yvon-Durocher et al. 2014), influencing below-ground storage pools and, ultimately, sediment-501 atmosphere fluxes (Reid et al. 2013, Capooci and Vargas 2022). GPP also has strong coupling and may 502 suggest there is connectivity or flow between plant photoassimilates and methanogens in the rhizosphere, 503 which increases with seasonal biomass development. We observed a 32-day median lag time between 504 GPP and effects on CH<sub>4</sub> flux, similar to the lag identified for temperature (35 days), which may indicate 505 structural changes in the vegetation are responsible for the observed seasonal increase in CH<sub>4</sub> flux. 506 Weaker couplings were associated with physical variables such as Patm and lag times for these variables 507 were much shorter. Thus, it appeared that biological vs. physical processes operate at different temporal 508 scales.

509 Salinity has previously been identified as an important daily-scale predictor by linear modeling 510 but had no direct causal influence on seasonal CH<sub>4</sub> fluxes in this study. While salinity causes direct 511 inhibitory effects for methanogens by limiting the availability of sulfate substrates via microbial 512 competition in most wetlands (Reddy and DeLaune 2008), there is evidence for the presence of 513 methylotrophic bacteria in salt marshes which produce CH<sub>4</sub> from non-competitive substrates resulting in 514 elevated emissions despite the presence of sulfate (Seyfferth et al. 2020, Capooci and Vargas 2022). 515 Evidence to support this phenomenon exists from a variety of wetland studies that either lack S. 516 alterniflora and thus more closely follow the paradigm of CH<sub>4</sub> inhibition (Chmura, Kellman, and 517 Guntenspergen 2011, Krauss et al. 2016, Poffenbarger et al. 2011), or contain S. alterniflora and emit 518 methane above or at the high-end of a widely accepted CH<sub>4</sub>-sulfate threshold (Poffenbarger et al2011, 519 Oremland et al. 1982, Huertas et al. 2019, Rosentreter et al. 2021). We propose that the elevated 520 emissions, especially during senescence, are tied to the release of secondary plant compounds (osmolytes) 521 associated with S. alterniflora, including glycine betaine and dimethyl sulfoniopropionate which are 522 broken down via fermentation to trimethylamine and dimethyl sulfide and are exclusively utilized by 523 methylotrophic methanogens (Husband and Kiene 2007, Jones et al. 2019). It is also important to note 524 that this marsh rarely experiences salinity extremes over 20ppt, thus sites influenced by higher salinity 525 may experience different characteristics.

# 526 4.3 Diurnal CH<sub>4</sub> Dynamics

527 Diurnal CH<sub>4</sub> dynamics during the growing phenoperiod are controlled by multiple factors with 528 many interdependencies among variables, yet the most important causal variables are Level, Patm, Sal, 529 DO, and Tsoil. This indicates that CH<sub>4</sub> fluxes over the diurnal course are controlled by both biochemical 530 factors likely tied to plant and microbial activity (Sal, DO, Tsoil) and physical factors likely tied to 531 emissions (Level, Patm) (Moore and Roulet 1993, Reid, et al 2015, Knox et al. 2021, Liu et al. 2019). 532 While several studies have made attempts at reducing this level of complexity by applying parsimonious 533 models that only incorporate the most important drivers (via linear methods) (Levy et al. 2012, Abdul-

Aziz et al. 2018, Baird et al. 2019), this may only be useful for generating coarse-scale information. Other studies have applied more complex process-based models, although these require an extensive mechanistic understanding and often experience difficulties when moving across study sites (Zhang et al. 2020, Li et al. 2016), where interacting variables and driving processes can differ considerably (Krauss et al. 2016, Rosentreter et al. 2021). This can limit the range of potential applications and may be less relevant for coastal wetlands where relationships can differ greatly from site to site.

540 The dormancy phenoperiod is comparatively simple, with the most important causal variables 541 consisting of Level, DO, and Sal. This is also supported by considerably less interconnection between 542 variables. Considering the lack of active vegetation, we can assume mechanisms related to these 543 predictors change in relation to the growing or year-round daily averaging periods (Bansal et al. 2020, 544 Reid et al. 2013). Although the relationship is weak, coupling between  $CH_4$  and Sal increases following 545 the end of senescence and the preceding seasonal peak in emissions. Because vegetation is no longer 546 present, the flow of photoassimilates stops, and availability of osmolyte byproducts likely decreases 547 within sediments. Thus, Sal may become more relevant during dormancy as competition between sulfate-548 reducing bacteria and methanogens likely increases (Derby et al. 2021, Poffenbarger et al. 2011), while 549 the spatial separation between production zones at depth may be less exposed to sulfate fluctuations from 550 tidal exchange (Koebsch et al. 2019). Further, plant-mediated transport is curtailed, leading to greater 551 importance of conditions at the sediment interface (Wang and Han 2005). This is supported by the lack of 552 causal influence from associated variables such as GPP, LE, and H<sub>2</sub>O flux. This also likely explains the 553 importance of DO in tidal waters, which reaches a concentration peak during dormancy, likely 554 influencing  $CH_4$  oxidation rates. In either case, the top hourly predictor for both growing and dormancy 555 phenoperiods is Level that oscillates between low and high tides at six-hour intervals but also shifts on a 556 14-day spring/neap tide and is highly tied to fluctuations of DO and Sal (Vázquez-Lule and Vargas 2021). 557 Our results show that while all these variables are related and often highly cross-correlated with linear 558 methods, each of these variables exerts a unique influence on CH<sub>4</sub> dynamics, yet there is also strong

interdependence with how these variables change in time, highlighting the importance of incorporatingtrue nonlinear analysis methods.

# 561 4.4 Relevance of Time Lags

562 On a seasonal/multiday scale, almost all causally significant variables were associated with lag 563 times where influence is exerted across several consecutive days or weeks. The most important causal 564 variables (temperature components and GPP) had medium lag times (i.e., "slow") near 3-5 weeks, likely 565 associated with the possible threshold effect with the temperature-driven initiation of plant growth 566 (O'Connell, Alber and Pennings 2020), or increased production and transport of photoassimilates to the 567 rhizosphere (Knox et al. 2021). This lag time (3 weeks) was also associated with the effect of salinity on 568 GPP, likely indicating increases in plant osmolyte production coinciding with initial spring biomass 569 accumulation (Husband and Kiene 2007, Mulholland and Otte 2000). It is important to recognize that we 570 are measuring  $CH_4$  fluxes (the physical process of release); thus, these observed lags are also likely 571 associated with the disconnection between production and observed fluxes. The shortest lags exert 572 immediate (<1 week) (i.e., "fast") but also prolonged impact extending up to 1 week for Level, Patm, and 573 RH and up to 4 weeks for DO. This effect from Level is likely a result of high/low water levels tied to 574 spring/neap tide cycles, which oscillate on a 6-7 day period. Yet, there are also fluctuations at shorter time 575 periods, both of which impact soil water content and biogeochemical conditions of the sediments 576 (Seyfferth et al. 2020). Tidal waters also directly influence salinity and DO levels within sediments, but 577 the effect of DO is prolonged and likely tied to its strong influence on GPP, as shown by our results (Fig. 578 4). The effects of Patm and RH are likely tied to mesoscale circulation and air mass transit which modify 579 the air column and provide a physical forcing mechanism (Tokida et al. 2007). The longest lag times were 580 associated with LE, H<sub>2</sub>O flux, and PAR, which likely represents a major seasonal oscillation driven by 581 solar angle and changes in available radiation (Guo et al. 2010). While only weakly causal, there were no 582 lags associated with WSpd or Precip, indicating an effect exclusively at the sub-daily level as, in most 583 cases, the occurrence of wind or rain rarely extends into multi-day events.

584 During the growing phenoperiod, top causal variables identified by CCM analysis (Level, Patm, 585 Sal, DO, Tsoil) all operated at different peak lag times, highlighting the complexity of observing and 586 measuring fluxes during this time. The most important causal variable (Level) exerts influence over the 587 entire 24-hour period yet peaks in causal effect at 0-5 hours and 20-24 hours, near when Sal (moderately 588 casual) was influential, which is likely tied to semi-diurnal tidal cycling and water levels which are 589 typically higher during nighttime high tide (Forbrich and Giblin 2015). Influence over the entire 24-hour 590 period may be due to a smearing-out effect as the tide shifts its range an hour each day. Patm and DO 591 (moderately causal) were associated with a zero-lag time but the effect of Patm extended up to 5 hours. 592 These lags are likely related to fast (immediate) oxidation activity across sediment, daily pressure 593 fluctuations, and convective flow-through transport of  $CH_4$  via plant-mediated transport driven by solar 594 activity (Kim, Verma and Billesbach 1999, Ding and Cai 2007). Although the overall causality was 595 weaker and distributed across variables, many of the fast variables during the growing phenoperiod are 596 related to energy exchange dynamics (LE and H) and tied to changes in evapotranspiration (H<sub>2</sub>O flux, 597 VPD, RH). This is likely explained by combined evaporation from saturated sediments where CH<sub>4</sub> has 598 accumulated in porewater, evaporation from areas of open water that receive a lateral influx of dissolved 599 CH<sub>4</sub>, and plant-mediated transport factors reliant on stomatal activity, which channel dissolved CH<sub>4</sub> in the 600 transpiration stream (Megonigal and Guenther 2008, Sun et al. 2012, Trifunovic et al. 2020). GPP is both 601 a fast and slow variable indicating a supply of photoassimilates is triggered upon initiation of morning 602 photosynthesis but with lags extending up to 8 hours when photosynthesis gears down in the evening. 603 During the dormancy phenoperiod, top causal variables identified by CCM analysis (Level, Sal, 604 DO) all had immediate influence but were influential at lag times extending up to 5-10 hours, highlighting 605 the importance of tidal activity and modulation of associated variables (Sal and DO). While the remaining 606 variables were only weakly causal, several interesting patterns exist. In contrast to the slower effect of 607 Tair during the growing phenoperiod, Tair was a fast variable with short lag times in dormancy, while the 608 effect of Tsoil had longer lags, up to 9 hours, indicating a connection to microbial activity and 609 temperature thresholds associated with CH<sub>4</sub> production (Chadburn et al. 2020). All top causal variables

610 were related to tidal activity (Level, DO, Sal), indicating that hydrological factors and the riverine-ocean 611 connectivity become crucial during the dormancy phenoperiod. The shorter lag times during this 612 phenoperiod may also indicate more direct coupling with physical factors and less complex interactions 613 involving the presence of vegetation, such as substrate supply and plant-mediated transport (Vázquez-614 Lule and Vargas 2021). Taken together, these results highlight the multi-mechanistic activities and 615 processes which regulate CH<sub>4</sub> fluxes.

# 616 4.5 Multivariate EDM Modeling

617 This study demonstrates high complexity in the biophysical controls of CH<sub>4</sub> fluxes in tidal 618 wetlands. These results are based on high-frequency measurements and long-term observations (5 years) 619 that contribute to our understanding of CH<sub>4</sub> dynamics. The multivariate EDM mode provided predictions 620 within 1% of the CH<sub>4</sub> carbon sum measured from the EC benchmark. This model captured most of the 621 higher magnitude fluctuations and high emission peaks during the growing phenoperiod, yet omitted 622 higher magnitude fluctuations during senescence. Despite good performance, this model still considerably 623 underpredicted during senescence and overpredicted during green-up. Consequently, these transitions 624 represent a challenge as rapid metabolic and structural changes happen in the plant canopy (Vazquez-Lule 625 et al 2022). Although we included the full suite of identified causal variables from CCM, this could 626 indicate that additional variables omitted or not measured (e.g., sediment redox potential, biomass, 627 sediment substrate availability) are essential to explain these unusually high fluxes during senescence.

628 **5.** Conclusion

We demonstrate that nonlinear state space methods uncover the complex dynamics and interdependencies among predictors and their influence on  $CH_4$  flux. The technique can tease apart not only the strength of causal variables but also identify lags, and complex interconnections among the full suite of measured variables as past information about driving variables becomes embedded within the presently observed  $CH_4$  time series. Overall, CCM was able to identify and rank multiple causal variables while providing a holistic picture of variable interactions and lagged relationships, which is critical for an

accurate mechanistic understanding of system behavior. We highlight that CCM is a powerful technique
that provides mechanistic information regarding causality, lags, and interconnectivity among an extensive
collection of potentially important variables. Therefore, by complementing CCM with multivariate EDM
we can generate interpretable empirical models that provide insights into controlling mechanisms of
ecosystem processes.

640

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647

## 648 **Open Research**

649 Meteorological (station: delsjmet-p) and water quality (station: Aspen Landing) data are

available from the National Estuarine Research Reserve's Centralized Data Management Office

651 (CDMO) at https://cdmo.baruch.sc.edu/. Phenological data are available from the PhenoCam

652 network (site: stjones) at https://phenocam.sr.unh.edu/webcam/sites/stjones/. Eddy covariance

data are available via AmeriFlux (US-StJ; https://ameriflux.lbl.gov/sites/siteinfo/US-StJ).

# 655 References

- Abbasi, T., T. Abbasi, C. Luithui & S. A. Abbasi (2019) Modelling Methane and Nitrous Oxide
- Emissions from Rice Paddy Wetlands in India Using Artificial Neural Networks (ANNs). *Water*, 11.
- Abdul-Aziz, O. I., K. S. Ishtiaq, J. W. Tang, S. Moseman-Valtierra, K. D. Kroeger, M. E. Gonneea, J.
- 659 Mora & K. Morkeski (2018) Environmental Controls, Emergent Scaling, and Predictions of Greenhouse
- 660 Gas (GHG) Fluxes in Coastal Salt Marshes. Journal of Geophysical Research-Biogeosciences, 123, 2234-
- 661 2256.
- Al-Haj, A. N. & R. W. Fulweiler (2020) A synthesis of methane emissions from shallow vegetated
  coastal ecosystems. *Global Change Biology*, 26, 2988-3005.
- Baird, A., S. Green, E. Brown & G. Dooling (2019) Modelling time-integrated fluxes of CO2 and CH4 in
   peatlands: A review. *Mires and Peat*, 24.
- Baldocchi, D. D. (2003) Assessing the eddy covariance technique for evaluating carbon dioxide exchange
  rates of ecosystems: past, present and future. *Global Change Biology*, 9, 479-492.
- Bansal, S., O. F. Johnson, J. Meier & X. Zhu (2020) Vegetation Affects Timing and Location of Wetland
  Methane Emissions. *Journal of Geophysical Research: Biogeosciences*, 125, e2020JG005777.
- 670 Barba, J., A. Cueva, M. Bahn, G. A. Barron-Gaffordd, B. Bond-Lamberty, P. J. Hanson, A. Jaimes, L.
- 671 Kulmala, J. Pumpanen, R. L. Scott, G. Wohlfahrt & R. Vargas (2018) Comparing ecosystem and soil
- respiration: Review and key challenges of tower-based and soil measurements. *Agricultural and Forest Meteorology*, 249, 434-443.
- Barton, K. & M. K. Barton (2015) Package 'mumin'. Version, 1, 439.
- 675 Bridgham, S. D., H. Cadillo-Quiroz, J. K. Keller & Q. L. Zhuang (2013) Methane emissions from
- 676 wetlands: biogeochemical, microbial, and modeling perspectives from local to global scales. *Global*
- 677 *Change Biology*, 19, 1325-1346.
- 678 Capooci, M. & R. Vargas (2022) Trace gas fluxes from tidal salt marsh soils: implications for carbon-679 sulfur biogeochemistry. *Biogeosciences*, 19, 4655-4670.
- 680 Chadburn, S. E., T. Aalto, M. Aurela, D. Baldocchi, C. Biasi, J. Boike, E. J. Burke, E. Comyn-Platt, A. J.
- 681 Dolman, C. Duran-Rojas, Y. Fan, T. Friborg, Y. Gao, N. Gedney, M. Göckede, G. D. Hayman, D. Holl,
- 682 G. Hugelius, L. Kutzbach, H. Lee, A. Lohila, F.-J. W. Parmentier, T. Sachs, N. J. Shurpali & S.
- 683 Westermann (2020) Modeled Microbial Dynamics Explain the Apparent Temperature Sensitivity of
- 684 Wetland Methane Emissions. *Global Biogeochemical Cycles*, 34, e2020GB006678.
- Chang, C.-W., M. Ushio & C.-h. Hsieh (2017) Empirical dynamic modeling for beginners. *Ecological Research*, 32, 785-796.
- 687 Christensen, T. R., A. Ekberg, L. Ström, M. Mastepanov, N. Panikov, M. Öquist, B. H. Svensson, H.
- 688 Nykänen, P. J. Martikainen & H. Oskarsson (2003) Factors controlling large scale variations in methane
- 689 emissions from wetlands. *Geophysical Research Letters*, 30.

- 690 Chu, X. J., G. X. Han, Q. H. Xing, J. Y. Xia, B. Y. Sun, J. B. Yu & D. J. Li (2018) Dual effect of
- 691 precipitation redistribution on net ecosystem CO2 exchange of a coastal wetland in the Yellow River 692 Delta. *Agricultural and Forest Meteorology*, 249, 286-296.
- 693 Dean, J. F., J. J. Middelburg, T. Röckmann, R. Aerts, L. G. Blauw, M. Egger, M. S. M. Jetten, A. E. E. de
- Jong, O. H. Meisel, O. Rasigraf, C. P. Slomp, M. H. in't Zandt & A. J. Dolman (2018) Methane
- 695 Feedbacks to the Global Climate System in a Warmer World. *Reviews of Geophysics*, 56, 207-250.
- 696 Delwiche, K. B., S. H. Knox, A. Malhotra, E. Fluet-Chouinard, G. McNicol, S. Feron, Z. Ouyang, D.
- 697 Papale, C. Trotta, E. Canfora, Y. W. Cheah, D. Christianson, M. C. R. Alberto, P. Alekseychik, M.
- 698 Aurela, D. Baldocchi, S. Bansal, D. P. Billesbach, G. Bohrer, R. Bracho, N. Buchmann, D. I. Campbell,
- 699 G. Celis, J. Chen, W. Chen, H. Chu, H. J. Dalmagro, S. Dengel, A. R. Desai, M. Detto, H. Dolman, E.
- 700 Eichelmann, E. Euskirchen, D. Famulari, K. Fuchs, M. Goeckede, S. Gogo, M. J. Gondwe, J. P.
- 701 Goodrich, P. Gottschalk, S. L. Graham, M. Heimann, M. Helbig, C. Helfter, K. S. Hemes, T. Hirano, D.
- Hollinger, L. Hörtnagl, H. Iwata, A. Jacotot, G. Jurasinski, M. Kang, K. Kasak, J. King, J. Klatt, F.
- 703 Koebsch, K. W. Krauss, D. Y. F. Lai, A. Lohila, I. Mammarella, L. Belelli Marchesini, G. Manca, J. H.
- 704 Matthes, T. Maximov, L. Merbold, B. Mitra, T. H. Morin, E. Nemitz, M. B. Nilsson, S. Niu, W. C.
- 705 Oechel, P. Y. Oikawa, K. Ono, M. Peichl, O. Peltola, M. L. Reba, A. D. Richardson, W. Riley, B. R. K.
- 706 Runkle, Y. Ryu, T. Sachs, A. Sakabe, C. R. Sanchez, E. A. Schuur, K. V. R. Schäfer, O. Sonnentag, J. P.
- 707 Sparks, E. Stuart-Haëntjens, C. Sturtevant, R. C. Sullivan, D. J. Szutu, J. E. Thom, M. S. Torn, E. S.
- 708 Tuittila, J. Turner, M. Ueyama, A. C. Valach, R. Vargas, A. Varlagin, A. Vazquez-Lule, et al. (2021)
- FLUXNET-CH4: a global, multi-ecosystem dataset and analysis of methane seasonality from freshwater wetlands. *Earth Syst. Sci. Data*, 13, 3607-3689.
- 711 Deyle, E. R., May, R. M., Munch, S. B., & Sugihara, G. (2016). Tracking and forecasting ecosystem
- 712 interactions in real time. *Proc Biol Sci*, 283(1822).
- Derby, R. K., B. A. Needelman, A. A. Roden & J. P. Megonigal (2021) Stratifying by Vegetation and
   Hydrology Improves Tidal Marsh Methane Emission Accounting.
- 715 Fan, H., Y. Meng, D. Wang, Y. Zhao & C. Zhao. 2020.. In 2020 35th Youth Academic Annual
- 716 *Conference of Chinese Association of Automation (YAC)*, 347-352. IEEE.
- 717 Filippa, G., E. Cremonese, M. Migliavacca, M. Galvagno, M. Forkel, L. Wingate, E. Tomelleri, U. Morra
- di Cella & A. D. Richardson (2016) Phenopix: A R package for image-based vegetation phenology.
   *Agricultural and Forest Meteorology*, 220, 141-150.
- 720 Gao, G. F., P. F. Li, Z. J. Shen, Y. Y. Qin, X. M. Zhang, K. Ghoto, X. Y. Zhu & H. L. Zheng (2018)
- Exotic Spartina alterniflora invasion increases CH4 while reduces CO2 emissions from mangrove wetland
   soils in southeastern China. *Scientific Reports*, 8.
- 723 Geoghegan, E. K., J. S. Caplan, F. N. Leech, P. E. Weber, C. E. Bauer & T. J. Mozdzer (2018) Nitrogen
- enrichment alters carbon fluxes in a New England salt marsh. *Ecosystem Health and Sustainability*, 4,
- 725 277-287.
- 726 Glagolev, M. V., O. R. Kotsyurbenko, A. F. Sabrekov, Y. V. Litti & I. E. Terentieva (2021)
- Methodologies for Measuring Microbial Methane Production and Emission from Soils-A Review.
   *Microbiology*, 90, 1-19.
- 729 Granger, C. W. (1969) Investigating causal relations by econometric models and cross-spectral methods.
- 730 Econometrica: journal of the Econometric Society, 424-438.

- 731 Granse, D., J. Titschack, M. Ainouche, K. Jensen & K. Koop-Jakobsen (2022) Subsurface aeration of
- tidal wetland soils: Root-system structure and aerenchyma connectivity in Spartina (Poaceae). Science of
   *The Total Environment*, 802, 149771.
- Guo, H. Q., B. Zhao, J. Q. Chen, Y. E. Yan, B. Li & J. K. Chen (2010) Seasonal Changes of Energy
  Fluxes in an Estuarine Wetland of Shanghai, China. *Chinese Geographical Science*, 20, 23-29.
- Guo, Z., M. Hao, B. Yu & B. Yao (2022) Detecting delay propagation in regional air transport systems
- 737 using convergent cross mapping and complex network theory. *Transportation Research Part E: Logistics*
- and Transportation Review, 157, 102585.
- Helfter, C., M. Gondwe, M. Murray-Hudson, A. Makati, M. F. Lunt, P. I. Palmer & U. Skiba (2022)
- 740 Phenology is the dominant control of methane emissions in a tropical non-forested wetland. *Nature*
- 741 *Communications*, 13, 133.
- Hill, A. & R. Vargas. 2021. Methane and Carbon Dioxide Fluxes in a Temperate Salt Marsh:
  Comparisons Between Plot, Ecosystem and Component Measurements. B25C-1465.
- Hill, A. C. & R. Vargas (2022) Methane and Carbon Dioxide Fluxes in a Temperate Tidal Salt Marsh:
- Comparisons Between Plot and Ecosystem Measurements. *Journal of Geophysical Research: Biogeosciences*, 127, e2022JG006943.
- Hsieh, C.-h., S. M. Glaser, A. J. Lucas & G. Sugihara (2005) Distinguishing random environmental
  fluctuations from ecological catastrophes for the North Pacific Ocean. *Nature*, 435, 336-340.
- Huertas, I. E., M. de la Paz, F. F. Perez, G. Navarro & S. Flecha (2019) Methane Emissions From the Salt
- 750 Marshes of Donana Wetlands: Spatio-Temporal Variability and Controlling Factors. *Frontiers in Ecology*
- and Evolution, 7.
- Husband, J. D. & R. P. Kiene (2007) Occurrence of dimethylsulfoxide in leaves, stems, and roots of
  Spartina alterniflora. *Wetlands*, 27, 224-229.
- Jackson, R. B., M. Saunois, P. Bousquet, J. G. Canadell, B. Poulter, A. R. Stavert, P. Bergamaschi, Y.
- 755 Niwa, A. Segers & A. Tsuruta (2020) Increasing anthropogenic methane emissions arise equally from
- agricultural and fossil fuel sources. *Environmental Research Letters*, 15, 071002.
- Jones, H. J., Kröber, E., Stephenson, J., Mausz, M. A., Jameson, E., Millard, A., Purdy, K. J., & Chen, Y.
- 758 (2019) A new family of uncultivated bacteria involved in methanogenesis from the ubiquitous osmolyte 759 glycine betaine in coastal saltmarsh sediments. *Microbiome*, 7(1), 120.
- Kim, D. S. (2007) Greenhouse Gas (CH4, CO2, N2O) Emissions from Estuarine Tidal and Wetland and
- 761 Their Characteristics. Journal of Korean Society for Atmospheric Environment, 23, 225-241.
- 762 Kim, Y., M. S. Johnson, S. H. Knox, T. A. Black, H. J. Dalmagro, M. Kang, J. Kim & D. Baldocchi
- 763 (2020) Gap-filling approaches for eddy covariance methane fluxes: A comparison of three machine
- learning algorithms and a traditional method with principal component analysis. *Global Change Biology*,
  26, 1499-1518.
- 766 Knox, S. H., S. Bansal, G. McNicol, K. Schafer, C. Sturtevant, M. Ueyama, A. C. Valach, D. Baldocchi,
- K. Delwiche & A. R. Desai (2021) Identifying dominant environmental predictors of freshwater wetland
   methane fluxes across diurnal to seasonal time scales. *Global change biology*.

- 769 Knox, S. H., J. H. Matthes, C. Sturtevant, P. Y. Oikawa, J. Verfaillie & D. Baldocchi (2016) Biophysical
- controls on interannual variability in ecosystem-scale CO2 and CH4 exchange in a California rice paddy.
- 771 *Journal of Geophysical Research-Biogeosciences*, 121, 978-1001.
- 772 Koebsch, F., M. Winkel, S. Liebner, B. Liu, J. Westphal, I. Schmiedinger, A. Spitzy, M. Gehre, G.
- Jurasinski, S. Köhler, V. Unger, M. Koch, T. Sachs & M. E. Böttcher (2019) Sulfate deprivation triggers
- high methane production in a disturbed and rewetted coastal peatland. *Biogeosciences*, 16, 1937-1953.
- 775 Krauss, K. W., G. O. Holm, B. C. Perez, D. E. McWhorter, N. Cormier, R. F. Moss, D. J. Johnson, S. C.
- 776 Neubauer & R. C. Raynie (2016) Component greenhouse gas fluxes and radiative balance from two
- deltaic marshes in Louisiana: Pairing chamber techniques and eddy covariance. *Journal of Geophysical Research-Biogeosciences*, 121, 1503-1521.
- 778 *Research-Diogeosciences*, 121, 1505-1521.
- Kuhn, M. (2008) Building Predictive Models in R Using the caret Package. *Journal of Statistical Software*, 28, 1 26.
- 781 Lakomiec, P., J. Holst, T. Friborg, P. Crill, N. Rakos, N. Kljun, P. O. Olsson, L. Eklundh, A. Persson & J.
- Rinne (2021) Field-scale CH4 emission at a subarctic mire with heterogeneous permafrost thaw status.
   *Biogeosciences*, 18, 5811-5830.
- 784 Levy, P. E., A. Burden, M. D. Cooper, K. J. Dinsmore, J. Drewer, C. Evans, D. Fowler, J. Gaiawyn, A.
- Gray & S. K. Jones (2012) Methane emissions from soils: synthesis and analysis of a large UK data set.
   *Global Change Biology*, 18, 1657-1669.
- Li, C. (2007) Quantifying greenhouse gas emissions from soils: Scientific basis and modeling approach. *Soil Science and Plant Nutrition*, 53, 344-352.
- Li, T. T., B. H. Xie, G. C. Wang, W. Zhang, Q. Zhang, T. Vesala & M. Raivonen (2016) Field-scale
   simulation of methane emissions from coastal wetlands in China using an improved version of
- simulation of methane emissions from coastal wetlands in China using an improved version of
   CH4MOD(wetland). *Science of the Total Environment*, 559, 256-267.
- Li, Y., D. Wang, Z. Chen, J. Chen, H. Hu & R. Wang (2021) Methane Emissions during the Tide Cycle of a Yangtze Estuary Salt Marsh. *Atmosphere*, 12, 245.
- Liu, L. J., D. Q. Wang, S. Chen, Z. J. Yu, Y. K. Xu, Y. Li, Z. M. Ge & Z. L. Chen (2019) Methane
  Emissions from Estuarine Coastal Wetlands: Implications for Global Change Effect. *Soil Science Society*
- 796 *of America Journal*, 83, 1368-1377.
- Martin, R. M. & S. Moseman-Valtierra (2017) Different short-term responses of greenhouse gas fluxes
   from salt marsh mesocosms to simulated global change drivers. *Hydrobiologia*, 802, 71-83.
- 799 McKenna, T., J. Callahan, C. Medlock & N. Bates (2018) Creation of improved accuracy LiDAR-based
- 800 digital elevation models for the St. Jones River and Blackbird Creek Watersheds. *Newark: Delaware*
- 801 Geologic Survey, Technical Report, 1-30.
- 802 Mitra, B., K. Minick, G. Miao, J.-C. Domec, P. Prajapati, S. G. McNulty, G. Sun, J. S. King & A.
- 803 Noormets (2020) Spectral evidence for substrate availability rather than environmental control of methane 804 emissions from a coastal forested wetland. *Agricultural and Forest Meteorology*, 291, 108062.
- 805 Mønster, J., P. Kjeldsen & C. Scheutz (2019) Methodologies for measuring fugitive methane emissions
- 806 from landfills A review. *Waste Management*, 87, 835-859.

- 807 Moore, T. R. & N. T. Roulet (1993) Methane flux: Water table relations in northern wetlands.
- 808 *Geophysical Research Letters*, 20, 587-590.
- 809 Morin, T. H. (2019) Advances in the Eddy Covariance Approach to CH4 Monitoring Over Two and a
- 810 Half Decades. Journal of Geophysical Research-Biogeosciences, 124, 453-460.
- 811 Morin, T. H., G. Bohrer, L. Naor-Azrieli, S. Mesi, W. T. Kenny, W. J. Mitsch & K. V. R. Schafer (2014)
- 812 The seasonal and diurnal dynamics of methane flux at a created urban wetland. *Ecological Engineering*,
- 813 72, 74-83.
- 814 Morin, T. H., G. Bohrer, K. C. Stefanik, A. C. Rey-Sanchez, A. M. Matheny & W. J. Mitsch (2017)
- 815 Combining eddy-covariance and chamber measurements to determine the methane budget from a small, 816 heterogeneous urban floodplain wetland park. *Agricultural and Forest Meteorology*, 237, 160-170.
- 817 Mulholland, M. M., & Otte, M. L. (2000) Effects of varying sulphate and nitrogen supply on DMSP and 818 glycine betaine levels in Spartina anglica. *Journal of Sea Research*, *43*(3-4), 199-207.
- 819 O'Connell, J. L., M. Alber & S. C. Pennings (2020) Microspatial Differences in Soil Temperature Cause 820 Phenology Change on Par with Long-Term Climate Warming in Salt Marshes. *Ecosystems*, 23, 498-510.
- 821 Oikawa, P. Y., G. D. Jenerette, S. H. Knox, C. Sturtevant, J. Verfaillie, I. Dronova, C. M. Poindexter, E.
- 822 Eichelmann & D. D. Baldocchi (2017) Evaluation of a hierarchy of models reveals importance of
- 823 substrate limitation for predicting carbon dioxide and methane exchange in restored wetlands. *Journal of*
- 824 *Geophysical Research-Biogeosciences*, 122, 145-167.
- Paparella, F. (2005) Filling gaps in chaotic time series. *Physics Letters A*, 346, 47-53.
- Pinheiro, J., D. Bates, S. DebRoy, D. Sarkar, S. Heisterkamp, B. Van Willigen & R. Maintainer (2017)
  Package 'nlme'. *Linear and nonlinear mixed effects models, version,* 3.
- Poffenbarger, H. J., B. A. Needelman & J. P. Megonigal (2011) Salinity Influence on Methane Emissions
  from Tidal Marshes. *Wetlands*, 31, 831-842.
- 830 Reichstein, M., E. Falge, D. Baldocchi, D. Papale, M. Aubinet, P. Berbigier, C. Bernhofer, N. Buchmann,
- 831 T. Gilmanov, A. Granier, T. Grünwald, K. Havránková, H. Ilvesniemi, D. Janous, A. Knohl, T. Laurila,
- A. Lohila, D. Loustau, G. Matteucci, T. Meyers, F. Miglietta, J.-M. Ourcival, J. Pumpanen, S. Rambal, E.
- 833 Rotenberg, M. Sanz, J. Tenhunen, G. Seufert, F. Vaccari, T. Vesala, D. Yakir & R. Valentini (2005) On
- the separation of net ecosystem exchange into assimilation and ecosystem respiration: review and
- 835 improved algorithm. *Global Change Biology*, 11, 1424-1439.
- Reid, M. C., D. S. Pal & P. R. Jaffe (2015) Dissolved gas dynamics in wetland soils: Root-mediated gas transfer kinetics determined via push-pull tracer tests. *Water Resources Research*, 51, 7343-7357.
- 838 Reid, M. C., R. Tripathee, K. V. R. Schafer & P. R. Jaffe (2013) Tidal marsh methane dynamics:
- B39 Difference in seasonal lags in emissions driven by storage in vegetated versus unvegetated sediments.
- 840 Journal of Geophysical Research-Biogeosciences, 118, 1802-1813.
- 841 Rey-Sanchez, A. C., T. H. Morin, K. C. Stefanik, K. Wrighton & G. Bohrer (2018) Determining total
- 842 emissions and environmental drivers of methane flux in a Lake Erie estuarine marsh. *Ecological* 843 *Engineering*, 114, 7-15.

- 844 Rosentreter, J. A., A. N. Al-Haj, R. W. Fulweiler & P. Williamson (2021) Methane and Nitrous Oxide
- 845 Emissions Complicate Coastal Blue Carbon Assessments. *Global Biogeochemical Cycles*, 35,
  846 e2020GB006858.
- 847 Schiecke, K., B. Pester, M. Feucht, L. Leistritz & H. Witte. 2015. Convergent Cross Mapping: Basic
- 848 concept, influence of estimation parameters and practical application. In 2015 37th Annual International
- 849 *Conference of the IEEE Engineering in Medicine and Biology Society (EMBC)*, 7418-7421.
- 850 Seyfferth, A. L., F. Bothfeld, R. Vargas, J. W. Stuckey, J. Wang, K. Kearns, H. A. Michael, J. Guimond,
- X. Yu & D. L. Sparks (2020) Spatial and temporal heterogeneity of geochemical controls on carbon
- 852 cycling in a tidal salt marsh. *Geochimica et Cosmochimica Acta*, 282, 1-18.
- 853 Stavert, A. R., M. Saunois, J. G. Canadell, B. Poulter, R. B. Jackson, P. Regnier, R. Lauerwald, P. A.
- Raymond, G. H. Allen, P. K. Patra, P. Bergamaschi, P. Bousquet, N. Chandra, P. Ciais, A. Gustafson, M.
  Ishizawa, A. Ito, T. Kleinen, S. Maksyutov, J. McNorton, J. R. Melton, J. Müller, Y. Niwa, S. Peng, W. J.
- Riley, A. Segers, H. Tian, A. Tsuruta, Y. Yin, Z. Zhang, B. Zheng & Q. Zhuang (2022) Regional trends
- 857 and drivers of the global methane budget. *Global Change Biology*, 28, 182-200.
- 858 Strom, L., J. M. Falk, K. Skov, M. Jackowicz-Korczynski, M. Mastepanov, T. R. Christensen, M. Lund &
   859 N. M. Schmidt (2015) Controls of spatial and temporal variability in CH4 flux in a high arctic fen over
- three years. *Biogeochemistry*, 125, 21-35.
- Sugihara, G., R. May, H. Ye, C.-h. Hsieh, E. Deyle, M. Fogarty & S. Munch (2012) Detecting causality
  in complex ecosystems. *science*, 338, 496-500.
- Sugihara, G. & R. M. May (1990) Nonlinear forecasting as a way of distinguishing chaos from
  measurement error in time series. *Nature*, 344, 734-741.
- Sun, X., W. Fang, X. Gao, S. An, S. Liu & T. Wu (2021) Time-varying causality inference of different
   nickel markets based on the convergent cross mapping method. *Resources Policy*, 74, 102385.
- 867 Swain, C. K., P. Bhattacharyya, A. K. Nayak, N. R. Singh, S. Neogi, D. Chatterjee & H. Pathak (2018)
- Byvani, C. Ri, F. Bhataenaryya, R. R. Payaa, W. R. Singa, S. Peogr, D. Chaterjee & H. Fathar (2010)
   Dynamics of net ecosystem methane exchanges on temporal scale in tropical lowland rice. *Atmospheric Environment*, 191, 291-301.
- Takens, F. 1981. Detecting strange attractors in turbulence. In *Dynamical systems and turbulence*, *Warwick 1980*, 366-381. Springer.
- Tokida, T., T. Miyazaki, M. Mizoguchi, O. Nagata, F. Takakai, A. Kagemoto & R. Hatano (2007) Falling
  atmospheric pressure as a trigger for methane ebullition from peatland. *Global Biogeochemical Cycles*,
  21.
- Tong, C., W. Q. Wang, C. S. Zeng & R. Marrs (2010) Methane (CH4) emission from a tidal marsh in the
- 876 Min River estuary, southeast China. Journal of Environmental Science and Health Part a-
- 877 *Toxic/Hazardous Substances & Environmental Engineering*, 45, 506-516.
- 878 Tsonis, A. A., E. R. Deyle, H. Ye & G. Sugihara. 2018. Convergent Cross Mapping: Theory and an
- 879 Example. In Advances in Nonlinear Geosciences, ed. A. A. Tsonis, 587-600. Cham: Springer
- 880 International Publishing.

- 881 Turner, J., A. R. Desai, J. Thom & K. P. Wickland (2021) Lagged Wetland CH4 Flux Response in a
- Historically Wet Year. *Journal of Geophysical Research: Biogeosciences*, 126, e2021JG006458.
- 883 Ushio, M. & K. Kawatsu. 2020. Forecasting Ecological Time Series Using Empirical Dynamic Modeling:
- A Tutorial for Simplex Projection and S-map. In *Diversity of Functional Traits and Interactions:*
- 885 Perspectives on Community Dynamics, ed. A. Mougi, 193-213. Singapore: Springer Singapore.
- Vázquez-Lule, A. & R. Vargas (2021) Biophysical drivers of net ecosystem and methane exchange across
   phenological phases in a tidal salt marsh. *Agricultural and Forest Meteorology*, 300, 108309.
- 888 Vázquez-Lule, A., A. L. Seyfferth, M. A. Limmer, P. Mey, M. Guevara, and R. Vargas. 2022.
- Hyperspectral reflectance for measuring canopy-level nutrients and photosynthesis in a salt marsh.
  Journal of geophysical research. *Biogeosciences*, 127.
- Villa, J. A., Y. Ju, T. Stephen, C. Rey-Sanchez, K. C. Wrighton & G. Bohrer (2020) Plant-mediated
- methane transport in emergent and floating-leaved species of a temperate freshwater mineral-soil wetland.
   *Limnology and Oceanography*, 65, 1635-1650.
- Villa, J. A., Y. Ju, C. Vines, C. Rey-Sanchez, T. H. Morin, K. C. Wrighton & G. Bohrer (2019)
- 895 Relationships Between Methane and Carbon Dioxide Fluxes in a Temperate Cattail-Dominated
- 896 Freshwater Wetland. *Journal of Geophysical Research-Biogeosciences*, 124, 2076-2089.
- Wang, D. Q., Z. L. Chen & S. Y. Xu (2009) Methane emission from Yangtze estuarine wetland, China. *Journal of Geophysical Research-Biogeosciences*, 114.
- Wang, Y., J. Yang, Y. Chen, P. De Maeyer, Z. Li & W. Duan (2018) Detecting the Causal Effect of Soil
  Moisture on Precipitation Using Convergent Cross Mapping. *Scientific Reports*, 8, 12171.
- Wang, Z. P. & X. G. Han (2005) Diurnal variation in methane emissions in relation to plants and
   environmental variables in the Inner Mongolia marshes. *Atmospheric Environment*, 39, 6295-6305.
- 903 Ward, N. D., J. P. Megonigal, B. Bond-Lamberty, V. L. Bailey, D. Butman, E. A. Canuel, H.
- 904 Diefenderfer, N. K. Ganju, M. A. Goñi, E. B. Graham, C. S. Hopkinson, T. Khangaonkar, J. A. Langley,
- N. G. McDowell, A. N. Myers-Pigg, R. B. Neumann, C. L. Osburn, R. M. Price, J. Rowland, A.
- 906 Sengupta, M. Simard, P. E. Thornton, M. Tzortziou, R. Vargas, P. B. Weisenhorn & L. Windham-Myers
- 907 (2020) Representing the function and sensitivity of coastal interfaces in Earth system models. *Nature* 908 *Communications*, 11, 2458.
- Wuebbles, D. J. & K. Hayhoe (2002) Atmospheric methane and global change. *Earth-Science Reviews*,
  57, 177-210.
- 911 Yang, B., X. Li, S. Lin, C. Jiang, L. Xue, J. Wang, X. Liu, M. Espenberg, J. Pärn & Ü. Mander (2021)
- 912 Invasive Spartina alterniflora changes the Yangtze Estuary salt marsh from CH4 sink to source.
- 913 Estuarine, Coastal and Shelf Science, 252, 107258.
- 914 Ye, A. Clark, E. Deyle & G. Sugihara (2016) rEDM: an R package for empirical dynamic modeling and 915 convergent cross-mapping. *cran. r-project. org.*
- 916 Ye, H., R. J. Beamish, S. M. Glaser, S. C. H. Grant, C.-h. Hsieh, L. J. Richards, J. T. Schnute & G.
- 917 Sugihara (2015a) Equation-free mechanistic ecosystem forecasting using empirical dynamic modeling.
- 918 *Proceedings of the National Academy of Sciences*, 112, E1569-E1576.

- 919 Ye, H., E. R. Deyle, L. J. Gilarranz & G. Sugihara (2015b) Distinguishing time-delayed causal 920 interactions using convergent cross mapping. *Scientific Reports*, 5, 14750.
- 921 Yuan, J. J., W. X. Ding, D. Y. Liu, H. Kang, J. Xiang & Y. X. Lin (2016) Shifts in methanogen
- 922 community structure and function across a coastal marsh transect: effects of exotic Spartina alterniflora
   923 invasion. *Scientific Reports*, 6.
- 924 Yuan, J. J., D. Y. Liu, Y. Ji, J. Xiang, Y. X. Lin, M. Wu & W. X. Ding (2019) Spartina alterniflora
- 925 invasion drastically increases methane production potential by shifting methanogenesis from
  926 hydrogenotrophic to methylotrophic pathway in a coastal marsh. *Journal of Ecology*, 107, 2436-2450.
- 927 Yvon-Durocher, G., A. P. Allen, D. Bastviken, R. Conrad, C. Gudasz, A. St-Pierre, N. Thanh-Duc & P.
- A. del Giorgio (2014) Methane fluxes show consistent temperature dependence across microbial to
   ecosystem scales. *Nature*, 507, 488-491.
- Zaki, M. T. & O. I. Abdul-Aziz (2022) Predicting greenhouse gas fluxes in coastal salt marshes using
   artificial neural networks. *Wetlands*, 42.
- 2 Zhang, Q., L. Ting-Ting, Q. Zhang, W. Guo-Cheng, Y. Li-Jun, G. Bin & H. Peng-Fei (2020) Accuracy
  analysis in CH4MODwetland in the simulation of CH4 emissions from Chinese wetlands. *Advances in*
- 934 Climate Change Research, 11, 52-59.
- 235 Zhang, Z., N. E. Zimmermann, A. Stenke, X. Li, E. L. Hodson, G. F. Zhu, C. L. Huang & B. Poulter
- 936 (2017) Emerging role of wetland methane emissions in driving 21st century climate change. *Proceedings*
- 937 of the National Academy of Sciences of the United States of America, 114, 9647-9652.
- 938 Zhao, M. L., G. X. Han, J. Y. Li, W. M. Song, W. D. Qu, F. Eller, J. P. Wang & C. S. Jiang (2020)
- Responses of soil CO2 and CH4 emissions to changing water table level in a coastal wetland. *Journal of Cleaner Production*, 269.
- 941 Zhu, X. D., L. X. Meng, Y. H. Zhang, Q. H. Weng & J. Morris (2019) Tidal and Meteorological
- 942 Influences on the Growth of Invasive Spartina alterniflora: Evidence from UAV Remote Sensing. *Remote*
- 943 Sensing, 11.
- 244 Zhu, X. D., Q. L. Zhuang, Z. C. Qin, M. Glagolev & L. L. Song (2013) Estimating wetland methane
- 945 emissions from the northern high latitudes from 1990 to 2009 using artificial neural networks. *Global*946 *Biogeochemical Cycles*, 27, 592-604.