A multi-chemistry modelling framework to enable flexible and reproducible water quality simulations in existing hydro-models: 1. The OpenWQ concept and the water quality modelling lab

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Abstract

This work advances the incorporation and cross-model deployment of multi-biogeochemistry and ecological simulations in existing process-based hydro-modelling tools. It aims to transform the current practice of water quality modelling as an isolated research effort into a more integrated and collaborative activity between science communities. Our approach, which we call "Open Water Quality" (OpenWQ), enables existing hydrological, hydrodynamic, and groundwater models to extend their capabilities to water quality simulations, which can be set up to examine a variety of water-related pollution problems. OpenWQ's objective is to provide a flexible biogeochemical model representation that can be used to test different modelling hypotheses in a multi-disciplinary co-creative process. In this paper, we introduce the general approach used in OpenWQ. We detail aspects of its architecture that enable its coupling with existing models. This integration enables water quality models to benefit from advances made by hydrologic- and hydrodynamic-focused groups, strengthening collaboration between the hydrological, biogeochemistry, and soil science communities. We also detail innovative aspects of OpenWQ's modules that enable biogeochemistry lab-like capabilities, where modellers can define the pollution problem(s) of interest, the appropriate complexity of the biogeochemistry routines, and test different modelling hypotheses. In a companion paper, we demonstrate how OpenWQ has been coupled to two hydrological models, the "Structure for Unifying Multiple Modelling Alternatives" (SUMMA) and the "Cold Regions Hydrological Model" (CRHM), demonstrating the innovative aspects of OpenWQ, the flexibility of its couplers and internal spatiotemporal data structures, and the versatile eco-modelling lab capabilities that can be used to study different pollution problems.

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11 Abstract

This work advances the incorporation and cross-model deployment of multi-12 biogeochemistry and ecological simulations in existing process-based hydro-modelling tools. 13 It aims to transform the current practice of water quality modelling as an isolated research 14 effort into a more integrated and collaborative activity between science communities. Our 15 approach, which we call "Open Water Quality" (OpenWQ), enables existing hydrological, 16 hydrodynamic, and groundwater models to extend their capabilities to water quality sim-17 ulations, which can be set up to examine a variety of water-related pollution problems. 18 19 OpenWQ's objective is to provide a flexible biogeochemical model representation that can be used to test different modelling hypotheses in a multi-disciplinary co-creative process. In 20 this paper, we introduce the general approach used in OpenWQ. We detail aspects of its 21 architecture that enable its coupling with existing models. This integration enables water 22 quality models to benefit from advances made by hydrologic- and hydrodynamic-focused 23 groups, strengthening collaboration between the hydrological, biogeochemistry, and soil sci-24 ence communities. We also detail innovative aspects of OpenWQ's modules that enable 25 biogeochemistry lab-like capabilities, where modellers can define the pollution problem(s) 26 of interest, the appropriate complexity of the biogeochemistry routines, and test different 27 modelling hypotheses. In a companion paper, we demonstrate how OpenWQ has been 28 coupled to two hydrological models, the "Structure for Unifying Multiple Modelling Alter-29 natives" (SUMMA) and the "Cold Regions Hydrological Model" (CRHM), demonstrating 30 the innovative aspects of OpenWQ, the flexibility of its couplers and internal spatiotemporal 31 data structures, and the versatile eco-modelling lab capabilities that can be used to study 32 different pollution problems. 33

³⁴ 1 Introduction

Societies are becoming increasingly aware of the importance of harmonizing economic development with thriving aquatic ecosystems (Keith et al., 2022; Frank & Schlenker, 2016). This has led to an increase in the use of ecological and water quality models to study management solutions to pollution problems such as nutrient excess due to the use of fertilizers or microbiological contamination due to the discharge of untreated wastewater in rivers.

Over the past decades, substantial progress has been made in computational hydrology, 40 producing many powerful generic and fit-for-purpose modeling tools. Some of these models 41 have been particularly tailored to specific environments, such as the Cold Regions Hydrolog-42 ical Model (J. Pomeroy et al., 2022; J. W. Pomeroy et al., 2007), the Wetland DEM Ponding 43 Model (Shook et al., 2013), and the FLUXOS-Overland model (Costa, Shook, et al., 2020), 44 that have been developed to deal with the specific hydrological challenges of simulating the 45 Canadian Prairie pothole region due to the lack of well-defined river networks. Unfortu-46 nately, the science and modelling progress achieved by such hydrological model development 47 communities rarely transfers to (and translates into) water quality modeling improvements 48 because hydrological and water quality modeling tools are typically developed in isolation, 49 a problem that becomes clearly noticeable when examining the hydrological calculations 50 embedded in popular water quality modeling tools, such as SWAT (Arnold et al., 1998), 51 HYPE (Lindström et al., 2010; Arheimer et al., 2012), and INCA (Whitehead et al., 1998; 52 A. J. Wade et al., 2002; Jackson-Blake et al., 2016), that often rely on simplifications that 53 may be limiting for some regions and applications (Costa, Baulch, et al., 2020a). 54

Besides the issues with the communication and transfer of knowledge between these two scientific communities, other scientific and technical biogeochemistry-specific challenges remain critical for reliable simulations across climate regions, landscapes, and seasons (Costa, Baulch, et al., 2020a; Fu et al., 2019; Wellen et al., 2015). Some key challenges include (1) the adequate representation of the complexity and heterogeneity of biogeochemical processes and their dependency on weather, soil, and sediment characteristics; (2) the dependency on accurate hydrological drivers to reliably track the vertical and lateral movement of chemical constituents (being at the end of the climate-hydrology-ecology modeling chain invariably
 propagates uncertainties into water quality simulations); and (3) the often limited range of
 applicability of models to specific pollution problems and landscape conditions due to their
 rigid and heavily parameterized representation of water quality processes.

Underlying all of these issues are foundational questions of scientific reproducibility 66 in computational hydrology (C. Hutton et al., 2016), as well as considerable challenges in 67 defining the appropriate complexity, scale, and scope of water quality models (Costa, Baulch, 68 et al., 2020a; Moore et al., 2006; Mekonnen, 2016; Shoemaker, 1997). This challenge is in part 69 70 due to uncertainties associated with the prediction of many hydrological and biogeochemical responses at different spatial scales and geographical settings, which are aggravated by the 71 often sparse water quality data available for testing and validation of the models. This raises 72 critical questions for the design, application, and benefit of such modeling tools (Beck, 1987; 73 Moore et al., 2006; A. Wade et al., 2008; Costa, Baulch, et al., 2020a). 74

Improving process-based hydro-biogeochemical models for terrestrial hydrological simu-75 lations requires effective and continuous integration of progress across several research areas, 76 including hydrology, biogeochemistry, and soil science. Models must combine practical as-77 pects related to model application with theoretical scientific insights at various degrees of 78 scientific maturity and geographical applications. Such cross-disciplinary modeling efforts 79 require modelers to make model development decisions based on several considerations com-80 mon to hydrological models (Clark et al., 2011) that include (1) model fidelity, complexity 81 and practicality, (2) scientific reproducibility and transparency, and (3) data availability. 82

It has been recognized that more flexible water quality modelling approaches are needed 83 to address these issues (Yang et al., 2022; Costa, Shook, et al., 2020). Some approaches 84 have emerged that address some of the challenges and improve transparency and engage-85 ment among hydrologists, biogeochemists, soil scientists, and decision-makers, such as the 86 STELLA (Structural Thinking and Experiential Learning Laboratory with Animation) soft-87 ware (Richmond, 2003) and the MIKE Eco Lab, which is a water quality modelling addition 88 to the MIKE tools (Refsgaard & Storm, 1995). However, such tools are not suitable for 89 integration into modelling platforms developed by various hydrological communities around 90 the world. In the case of STELLA, the tool is standalone and more suitable for applica-91 tions where simpler batch-reactor assumptions are applicable, such as for small reservoirs 92 and wastewater treatment plants. The MIKE tools from DHI now provide an Eco-Lab 93 module for more flexible water quality and ecological simulations, but this module is a pro-94 prietary tool integrated into the MIKE ecosystem of tools, so it does not benefit the wider 95 hydrology and modelling communities, and it is constrained by the chemical constituents and types of biogeochemical processes that are explicitly introduced by the MIKE model 97 developers. More recently, Yang et al. (2022) proposed a new model named HiWaQ for 98 flexible catchment water quality assessments with compatibility for multiple hydrological 99 model structures. This is a valuable contribution toward a more unified hydrological-water 100 quality modelling approach, but the current version is limited to the nitrogen cycle. 101

¹⁰² 2 Current modelling capabilities and needs

Widely used process-based catchment nutrient models such as SWAT, HYPE, and 103 INCA have been at the forefront of innovation in water quality modeling, paving a way 104 for researchers and decision-makers around the world to investigate solutions to a variety of 105 pollution problems, particularly related to nutrient pollution. These models have become 106 increasingly complex and heavily parameterized, but they remain a limited representation of 107 reality because hydro-biogeochemical processes are highly complex in natural environments 108 (A. Wade et al., 2008; Beck, 1987; Costa, Baulch, et al., 2020b). These models typically 109 simulate a series of biogeochemical processes conceptualized to address particular pollutions 110 problems, and processes are represented through a combination of empirical and physico-111 chemically based methods, often leading to many calibration parameters and thus increasing 112

the risk of parameter equifinality (Costa, Baulch, et al., 2020b). Sparse and sporadic water
 quality measurements, which are common problems for most water bodies, limit the further
 application and development of these water quality models.

The combination of process-representation methods with varying degrees of empiricism 116 is in part due to knowledge gaps in understanding the drivers and controls of hydrological 117 and biogeochemical responses at various spatial scales and across different landscapes and 118 climate zones. In regions where relatively uncommon processes may play an important role 119 in the overall water quality dynamics, the use of these models becomes problematic because 120 121 there is little flexibility for adjusting conceptual models, adding or removing processes, or testing different modeling hypotheses. For example, research has shown that in some 122 cold regions, processes such as preferential infiltration of hydrochemical into frozen soils 123 (Lilback & Pomerov, 2007), preferential elution of hydrochemicals from melting snowpacks 124 (Davies et al., 1987; Marsh & Pomeroy, 1999; Costa & Pomeroy, 2019), microbial uptake 125 and fixing of nitrogen in melting snowpack (Jones, 1999), and volatilization of nitrogen 126 during snow redistribution and sublimation (J. W. Pomeroy et al., 1991; J. Pomeroy et al., 127 1999) may affect water quality, but they are not represented in most popular models. This 128 lack of representation may compromise their use in such regions and calls for a more flexible 129 approach to water quality modeling that enables a systematic and controlled approach 130 for the addition and removal of processes as needed to reflect (1) regional and climate 131 characteristics, (2) data limitations, and (3) objectives of the study. 132

3 The OpenWQ concept

3.1 Overview

134

OpenWQ is a coupler-modelling framework designed to provide portable and customiz-135 able multi-chemistry modelling capabilities to existing hydro-models. The vision is to create 136 a tool that could plug into existing process-based hydrological, hydrodynamic, and ground-137 water models to extend their capabilities to environmental and ecological studies. The 138 approach allows tailoring chemical-microbiological constituents and biogeochemistry-cycling 139 processes to enable the representation of different pollution problems and landscapes, as well 140 as to compare modelling hypotheses. This framework is a much-needed effort to bring the 141 hydrology and biogeochemistry communities together, optimizing research and investment 142 efforts. The work stems from previous model developments by Environment and Climate 143 Change Canada and the University of Saskatchewan, particularly CRHM-WQ (Cold Regions 144 Hydrological Model - Water Quality) (Costa et al., 2021) that extends the original CRHM 145 platform (hydrology) model to nitrogen and phosphorus simulations (J. W. Pomeroy et al., 146 2007), the WINTRA framework (Costa et al., 2017), the multiphase multilayer PULSE snow 147 hydrochemistry model (Costa et al., 2018), and the FLUXOS-OVERLAND model for wa-148 tershed hydrodynamic-water quality simulations suitable for Prairie regions (Costa, Shook, 149 et al., 2020). 150

OpenWQ aims to address three main challenges with existing water quality models: (1) 151 structural rigidity in the representation of chemical constituents and biogeochemical pro-152 cesses, (2) over-simplification and limitations of hydro-flux calculations, and (3) inadequacy 153 for testing different modelling hypotheses for proper quantification of structural uncertainty. 154 Structural rigidity is perhaps the key factor that hinders the effective use of models across 155 landscapes and in complex, diverse environments (e.g., permafrost, peatlands, variable con-156 tributing areas) that require more investigative, open-ended, and interactive simulation 157 approaches. A flexible environment that enables integrating new methods and concepts 158 from complementary disciplines and experts (e.g., limnology, soil science, biogeochemistry) 159 is critical to advance science and promote meaningful and impactful cross-disciplinary col-160 laborations. The static, hard-coded implementation of biogeochemical reaction-network 161 transformations limits their suitability for a wide range of environmental problems. It also 162 hinders multi-disciplinary co-creation efforts because models provide little flexibility for 163

changing, expanding, and testing different strategies for biogeochemical cycling representa tion. Finally, the hydro-flux calculations that are embedded in water quality modelling tools
 are often outdated or limited compared to dedicated, disciplinary hydro-models (e.g., hydro logical, hydrodynamic, and groundwater modelling tools). Uncertainty in model structure,
 process representation, and future scenarios (e.g., climate change) cannot be adequately
 quantified without flexible and transparent modelling structures.

170 3.2 General design

This section describes how OpenWQ was designed to address the key challenges in 171 water quality modelling described before. First, focus is given to describe how OpenWQ 172 tackles the need for more robust hydro-flux and associated solute transport calculations 173 through integration within existing hydro-models. This includes details on how OpenWQ 174 can be linked to existing hydro-models considering key practical aspects of version control 175 (Section 3.2.1), portability (Section 3.2.2), code integration (Sections 3.2.3), and state vari-176 ables and spatiotemporal discretization (Section 3.2.4). Second, focus is given to describe 177 how OpenWQ tackles the need for more flexible representation of physical and biochemical 178 processes, which includes aspects of model input (Section 3.3), model structure to enable 179 testing modelling hypotheses and quantifying structural uncertainty (Sections 3.4 and 3.5). 180

181

3.2.1 Obtaining OpenWQ

OpenWQ can be obtained from the official GitHub repository: https://github.com/ 182 ue-hydro/openwq. The recommended method for obtaining OpenWQ is to clone the repos-183 itory with "git clone". Downloading the repository is also possible, but it is less advisable 184 because OpenWQ is continuously updated and improved. OpenWQ is designed for the 185 user to update their copy quickly and efficiently with "git pull" instead of requiring them 186 to download each new version. The cloning (or download) should be performed inside the 187 host hydro-model code directory as shown in Fig. 1. It will create a new folder named 188 "openwq" with the entire repository. The "main" branch should be used because it will 189 contain the latest official release version. Compiling OpenWQ is carried out with CMake 190 (Kitware, 2022) and the provided "CMakeLists" file. When new versions of OpenWQ are 191 made available, they will only require the user to recompile using the same "CMakeLists" 192 file. Once the coupling of OpenWQ to a hydro-model is completed, carrying out updates to 193 OpenWQ's source code will not break OpenWQ's API calls and interface/coupler functions 194 in the host hydro-model. 195

196

3.2.2 Portability via internal dynamic coupler

In order to optimize the implementation of OpenWQ in hydro-models, the model has 197 been designed as an internal coupler module (Fig. 2). External coupling (i.e., OpenWQ 198 reading output files from the host-model and running standalone) was not a viable option 199 because hydro-models, particularly hydrological models, often deal with many water fluxes 200 moving around vertically and horizontally within and across hydrological compartments 201 (e.g., snow, soil), which are not always possible to export and disentangle. Even in cases 202 where models allow exporting all water fluxes separately, mapping those in OpenWQ and 203 204 harmonizing units to correctly compute the corresponding solute mass transport would be extremely difficult for most applications. Focusing on a flexible full-coupling approach with 205 minimal code re-engineering allowed us to address this problem, with OpenWQ being specif-206 ically designed to adapt its internal structure (e.g., spatial domain, temporal resolution) to 207 that of the host hydro-model with the support of interface and coupler functions that estab-208 lish one-way, plug-in-type communications between the two simulation systems. Although 209 these interface and coupler functions may need adjustments to harmonize particular aspects 210 of the host hydro-models, a generic "hydro-link" file is provided with a template for these 211



Figure 1: General model structure used in hydrological model

functions. It is possible to create such a template because most hydro-models follow the general model architecture depicted in Fig. 3.



Figure 2: OpenWQ concept as a coupler

The integration of OpenWQ into existing hydro-models is carried out through four 214 coupler functions that are responsible for (1) converting datatypes and data structures be-215 tween OpenWQ and the "host" hydro-model, (2) passing hydro-fluxes into OpenWQ, and 216 (3) calling OpenWQ's APIs (Fig. 4). These four coupler functions are invoked through inter-217 face routines contained within a C++ file with the default name "OpenWQ_hydrolink.cpp". 218 This is a one-way communication from the host hydro-model to OpenWQ, so no information 219 is returned to the host model. The interface functions are generic and serve as templates 220 (coupling recipes) that have been optimized to streamline the coupling procedure. The ad-221 justments needed in the interface functions are to ensure that the data types and structure 222



Figure 3: General model structure commonly used in dynamic hydrological, hydrodynamic, and groundwater model

used in the hydro-model are properly translated into OpenWQ's own data structure and conventions so that OpenWQ's APIs can perform adequately.

The following general steps can be performed to couple OpenWQ to a hydro-model: 225 (STEP 1) identify critical elements of hydro-model structure (see Fig. 3), (STEP 2) git-clone 226 OpenWQ, (STEP 3) create the C++ OpenWQ classes and objects to be loaded as modules 227 or libraries in the hydro-model, (STEP 4) identify the appropriate places in the hydro-228 model to call OpenWQ's coupler functions and APIs (see Fig. 4), (STEP 5) materialize 229 such calls, (STEP 6) adjust OpenWQ's coupler function to adapt to hydro-model data 230 structures (referred to "COUPLER CODE", see below), and (STEP 7) compile the new 231 coupled model. Fig. 5 shows the general structure of each of the four coupler functions, 232 where the location and purpose of the "COUPLER CODE" block are also explained: 233

- Coupler function 1 (*openwq::decl*) invokes a series of API calls that handle tasks associated with the initial configuration of the model, initialization of variables, and pre-processing of the input data;
 Coupler function 2 (*openwq::run_time_start*) invokes a series of API calls that handle tasks required at the start of each time step;
 Coupler function 3 (*openwq::run_space*) contains a series of API calls that handle tasks related to the spatial domain of the model; and
 Coupler function 4 (*openwq::run_space*)
- 4. Coupler function 4 (*openwq::run_time_end*) contains a series of API calls that handle
 tasks required at the end of each time step.



Figure 4: General placement of the calls for OpenWQ's coupler functions

243 3.2.3 General architecture

OpenWQ has been created in a way that separates the physics and biochemistry cal-244 culations from the numerical implementation. Such an approach was implemented in the 245 SUMMA model (Clark et al., 2015a,b), and we adopted it here to the development of 246 OpenWQ's core structure to improve scalability. This approach addresses a major problem 247 with many hydrological, hydrodynamic, and water quality models where the specification of 248 the model equations is intertwined with their numerical solution (Clark & Kavetski, 2010). 249 This specification complicates the selection and assessment of different model representa-250 tions (hypotheses) and makes introducing and evaluating alternative numerical methods 251 challenging. As such, the state variables of OpenWQ are only updated inside OpenWQ's 252 own numerical solver based on rates of chemical mass changes (time and space derivatives) 253 caused by different physical and biogeochemical processes. These processes and "rates of 254 change" are computed separately, in process-specific routines, and then passed into the nu-255 merical solver for a controlled and contained update of OpenWQ's state variables, which is 256 currently performed as a finite volume problem solved via a simple forward Euler method. 257 The separation of the numerical solver from the physics-biogeochemistry calculations will 258 enable improving the robustness of the numerical implementation in a contained and con-259 trolled manner in the future. 260

The modules in OpenWQ are divided into four groups. Each of these groups is respon-261 sible for the calculation of a chemical mass change (time derivative) driven by a particular 262 process (or phenomenon) or group of processes (or phenomena), specifically, (1) initial con-263 ditions (dm_ic) , (2) sinks and sources of chemical load (i.e., chemical mass entering or exiting 264 the model domain) (dm_ss) , (3) biogeochemical processes $(dm_dt_chemistry)$, and (4) phys-265 ical transport of chemical constituents with water flow $(dm_{-}dt_{-}transport)$ (Fig. 6). Each of 266 the modules, in turn, enables a series of modeling options that can be explored, but only 267 one module can be activated for each of the four spatiotemporal-derivative calculations. In 268 other words, the first layer of decision in representing a process pertains to the selection of 269 the key modules that will be responsible for computing each of these four derivatives. 270











CLASS OpenWQ_couplercall

OpenWQ_couplercalk

CALL

START

CLASS OpenWQ_couplerc

CLASS OpenWQ

CALL

START

Before space loop

SPACE

CLASS OpenWC

(BR



Figure 6: General modular architecture of OpenWQ that separates the different modules and methods available

3.2.4 State variables & Model-adaptive spatial and temporal discretization

271

OpenWQ was designed to automatically adapt to the hydro-model spatial and tempo-272 ral discretization structure, which can include Hydrological Response Units (HRUs), 1D–3D 273 spatial distributed meshes, structured or unstructured meshes, and multiple domains or hy-274 drological compartments (e.g., snow, soil, groundwater). OpenWQ's state variables (column 275 1 in Fig. 7), such as solute mass, are stored in a hierarchical data structure organized with 276 the following nested fields: (1) domain that refers to hydrological compartments (column 2) 277 in Fig. 7), (2) chemical species (column 3 in Fig. 7), and (3) sub-domain that refers to the 278 internal 1D–3D spatial discretization of the domains or hydrological compartments (column 279 4 in Fig. 7). The hierarchical data structures are built via the Armadillo C++ library for 280 linear algebra and scientific computing (Sanderson & Curtin, 2016, 2018). 281

The model interface and coupler are responsible for passing information about the 282 spatiotemporal configuration of the host hydro-model into OpenWQ. Such domain config-283 uration options are specific to each host hydro-model and depend on its particular domain 284 discretization scheme and model decisions specified by the users of the hydro-model. Once 285 that information is digested by the couplers and passed on to OpenWQ, a series of API 286 calls dynamically create the corresponding hierarchical data structures and sub-structures 287 that match those in the host model. Each of the data structures stores information about 288 a state variable or a supporting non-state variable (e.g., water fluxes, time derivatives). 289 The state-variable data structures record the spatiotemporal evolution of the mass of the 290 different chemical species tracked in each model domain or hydrological compartment (e.g., 291 snow, soil, lake). 292

STATE VARIABLES

arma::field(arma:field(arma::cube))



Figure 7: State variables are stored in OpenWQ via dynamic hierarchical data structures

293

3.3 Model setup and configuration: Inputs and Outputs

The configuration of OpenWQ is provided via JavaScript Object Notation (JSON) files. JSON is an ideal format for large data inputs that have a hierarchically structured relationship. JSON files are composed of key/value pairs, e.g., "RUN_MODE_DEBUG: TRUE". OpenWQ requires four JSON files to run. Each of these files deals with a particular aspect of the model setup and is given one of the following designations (1) "runManagement", (2) "Biogeochemistry", (3) "Configuration", and (4) "Source/Sink".

The "RunManagement" file is the entry point to OpenWQ. It provides the basic instructions, simulation and model output decisions, and the full or relative paths to the other JSON files needed to run a model. Currently, OpenWQ supports HD5F and CSV output file formats. The "Biogeochemistry" file is where the chemical species and biogeochemical cycling frameworks are created and characterized. It uses a standard template structure for the characterization transformations and the chemical species involved. Different "Biogeochemistry" files can be prepared to simulate different pollution problems, which can be readily loaded simultaneously (or used separately or swapped as needed) into a simulation

via the "RunManagement" file. The "Biogeochemistry" file is also where the contaminant 308 species available for the simulation are defined and given a "tag" that can be invoked in 309 other input files. All cycling frameworks and their respective transformations are also pro-310 vided with a unique "tag" that can be used to flexibly set up the model as desired. These 311 cycling framework tags are particularly important in the "Configuration" file to assign the 312 desired transformations to each domain or hydrological compartment of the model (e.g., 313 snow, soil, groundwater). The "Configuration" file also includes information about the ini-314 tial conditions. A "Sink/Source" file provides information about the mass exchange with 315 regions outside the model domain. This information can include external chemical mass 316 loading into (source) or loss from (sink), the model domain that can be associated with 317 particular external water fluxes, as in the case of precipitation, or mass inputs that can 318 be independent of the hydrological cycle. Typical examples of such mass inputs can in-319 clude fertilizer application, which is a relatively localized source, or atmospheric deposition, 320 which is a relatively distributed source. There can be multiple "Sink/Source" files in a given 321 OpenWQ configuration. 322

323

3.4 Native Modules and Process representation

OpenWQ contains four main groups of modules, each producing rates of chemical mass change (time derivatives) associated with specific processes, phenomena, and model aspects, which are passed into the numerical solver for updating the state-variables throughout the simulation (see Fig. 6). The modules deal with (1) initial conditions, (2) sinks and sources, (3) transport with water, and (4) biogeochemistry.

The first two groups of modules deal with initial conditions and sinks and external 329 sources with a focus on translating user inputs into the simulation. For example, the sink 330 and source module applies chemical load as prescribed by the users that can include (1) 331 continuous load arising from sources like atmospheric deposition or the outlet of a wastewater 332 treatment plant or (2) episodic/instantaneous loads arising from sources such as fertilizer 333 and chemical spills. The remaining two groups of modules focus respectively on (1) the 334 physical transport of chemical constituents as water moves through the system and (2) 335 biogeochemical transformations. 336

337

3.4.1 Transport Module

Currently, there are two options available for computing the physical transport of dis-338 solved solutes and fully suspended sediments. The first option accounts for both advection 339 and dispersion and solves the hyperbolic-parabolic advection-diffusion partial differential 340 equation (PDE) in up to 3 dimensions depending on the host model spatial discretiza-341 tion scheme (Equation 1); the second option only accounts for advection (Equation 2). In 342 both cases, the PDEs are solved inside OpenWQ's numerical solver as described in Sec-343 tion 3.2.3 using the specific hydro-fluxes and corresponding source and recipient domain 344 cells prescribed by the host models. This means that for each water flux computed by the 345 host-model (at each grid cell and timestep), OpenWQ calculates the corresponding solute 346 mass transported, which will be intimately linked to the internal spatio-temporal water-flux 347 representations of the host-model (i.e., if the host model represents snow as a 1 vertical layer 348 domain (lumped), OpenWQ will compute solute concentrations at that spatial resolution; 349 if the host model runs at daily time steps, OpenWQ will calculate solute concentrations at 350 that temporal resolution). The physical transport can be between computation elements 351 (e.g., HRUs, grid cells) and across domains or hydrological compartments (e.g., runoff, soil 352 saturated, soil unsaturated, canopy), as prescribed by the host model. 353

$$\frac{\partial (Vc_s)}{\partial t} + \nabla \cdot (V\vec{u}c_s) = \nabla \cdot (V\mathbf{E} \cdot \nabla c_s) + S, \tag{1}$$

$$\frac{\partial (Vc_s)}{\partial t} + \nabla \cdot (V\vec{v}c_s) = S, \tag{2}$$

where V is the volume of the computational element/cell $[L^3]$, c_s is the concentration of a given dissolved substance $[ML^{-3}]$; $\vec{u} = (u_x, u_y, u_z)$ is the velocity in the x, y, and z directions $[LT^{-1}]$; **E** is the (diagonal) diffusivity tensor $[L^2T^{-1}]$; and S is a source term $[MT^{-1}]$ that is linked to the chemistry modules. The **E** term accounts for the combined effect of different mixing phenomena:

$$\mathbf{E} = \mathbf{E}^{turb} + \mathbf{E}^{tsd} + \mathbf{E}^{sgt} + \mathbf{E}^{d},\tag{3}$$

where \mathbf{E}^{turb} is the turbulent or eddy diffusivity, which is a complex phenomenon with multi-359 fractal behaviour dominated by friction forces, \mathbf{E}^{tsd} is the so-called Taylor shear dispersion 360 that arises from the unresolved vertical variation of the horizontal flow, \mathbf{E}^{sgt} accounts for 361 the sub-grid eddy viscosity and diffusivity arising from unresolved mixing occurring at sub-362 grid scale, and \mathbf{E}^d is the background molecular diffusion resulting from the probabilistic 363 Brownian motion concept occurring at particle scales. E is approximated to the dominant 364 turbulent dispersion based on the eddy viscosity concept: $\mathbf{E}^{turb} = \boldsymbol{\nu}_t / \sigma$, where σ is the 365 Prandtl-Schmidt number. Turbulent viscosity (ν_t) depends on shear velocity (\vec{u}^*) and on 366 a turbulent length scale $(l_t = 0.07h)$. In this model, this relationship is approximated by 367 an algebraic expression: $\nu_t \sim k \vec{u}^* l_t$, where k is a user-defined scaling factor to account for 368 sub-grid scale eddies (Costa et al., 2016). 369

3.4.2 Biogeochemistry module

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Biogeochemical cycling frameworks are characterized in the appropriate biogeochem-371 istry JSON file. Each cycling framework is given a "tag" that is used to load it into simu-372 lations. The characterization of each of these cycling frameworks involves the definition of 373 the associated biogeochemical transformations and the chemical species involved. In turn, 374 the characterization of each of these transformations requires information about the chem-375 ical species consumed and produced and the reaction kinetics. It allows for the creation of 376 reaction parameters that can be loaded into the kinetics solver, which relies on the com-377 prehensive C++ Mathematical Expression Toolkit Library (*ExprTk*) developed by Arash 378 Partow (1999–2020) (Partow, 1999). The implementation of ExpTk in OpenWQ is simple 379 to use and provides an efficient run-time mathematical expression parser and evaluation 380 engine. ExprTk supports numerous forms of functional, logical, and vector processing se-381 mantics and is easily extendible. The equations can be written with (1) multiple chemical 382 species, (2) user-defined parameters, and (3) built-in hydro-model dependencies. These 383 model dependencies are tailored to each hydro-model, but they usually include variables like soil moisture and air and soil temperature. 385

The biogeochemistry JSON file can be prepared manually following the appropriate 386 OpenWQ JSON structure (i.e., key-value pairs). Alternatively, the cycling frameworks can 387 be characterized via a diagram drawn using a *GraphML* editor. *GraphML* is an XML-based 388 file format for graphs, and there are several free GraphML editors, such as yEd. After 389 the diagrams have been drawn in GraphML format, they can be converted into OpenWQ's 390 JSON format using a Python script available on OpenWQ's GitHub repository and copied 391 locally during the coupling step defined in Section 3.2.1. This graphical option has been 392 developed to enable collaboration and co-creation of water quality models through a more 393 visual interaction, which can be particularly helpful for water quality modeling activities 394 that often involve interaction between different disciplines that may be less familiar with 395 JSON files. 396

The expressions used to represent the different reaction kinetics in the model via the biogeochemistry JSON files can take many forms and are deployed in OpenWQ via the ExprTk integration. However, these expressions often take the form of sequences of reaction networks involving single or multiple chemical species governed by first-, second-, or third-order kinetics (Eq. 4, Eq. 5, Eq. 6), respectively:

$$\frac{dc_A}{dt} = -k\lambda c_A,\tag{4}$$

$$\frac{dc_A}{dt} = -k\lambda c_A^2,\tag{5}$$

$$\frac{dc_A}{dt} = -k\lambda c_A^2 c_B,\tag{6}$$

where c_A and c_B [ML⁻³] are the concentrations of chemical species A and B, parameter/variable λ represents weather/hydrological dependencies (such as soil moisture and temperature), and k is the reaction rate [ML⁻³T⁻¹]. The reaction rate k can be provided as the reaction rate using standard maximum at a reference temperature (often 20°C) or using expressions that can include relationships with the hydrological/weather dependent variables/parameters.

3.5 Eco-modelling lab and Cross-model deplyoment: Benefits and Innova tion

OpenWQ's concept as a coupler and customizable biogeochemistry modeling frame-405 work allows for cross-model deployment of eco-lab, co-creation modeling capabilities, pro-406 viding a long-awaited transformative direction for innovation in water quality modeling that 407 tackles the inherent challenges of being at the intersection between several scientific fields, 408 including biogeochemistry, soil science, hydrology, hydrodynamics, and hydrogeology. This 100 deployment is an essential step to enhancing collaborative efforts and streamlining knowl-410 edge/innovation transfer between the different disciplines involved, ultimately benefiting 411 the entire environmental and ecological research and management community. The ability 412 to deploy OpenWQ across different hydro-models provides the following transformational 413 changes in the current paradigm of water quality modeling. 414

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 1. Optimization of investment because focus is given to a single biogeochemical tool,
 416 thus reducing code maintenance,
- Maximization of benefit because it allows (a) any existing hydro-tool to extend its
 capabilities to water quality and ecological studies, (b) progress achieved by the bio geochemistry and soil-science research community to be transferred into OpenWQ
 and automatically benefits multiple hydro-models via update installs, (c) OpenWQ's
 input files to be transferred across hydro-models, providing experience and knowledge
 transfer between environmental projects and research communities, and
 - 3. Reproducibility is reinforced because OpenWQ's input files are transferable across hydro-models, allowing for more rigorous cross-model comparisons.

⁴²⁵ Once different biogeochemical modeling hypotheses have been identified and set up ⁴²⁶ for testing in one hydro-model-OpenWQ coupled model, they can be easily transferred for ⁴²⁷ simulation in any other hydro-amodel that has been coupled to OpenWQ (Fig. 8).

$_{428}$ 4 Discussion

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4.1 Unifying Different Approaches to Water Quality Modelling

430 OpenWQ provides a unifying modeling framework to deploy different approaches to 431 physiochemically based soil and water quality modeling across existing hydrological, hydro-



Figure 8: Cross-model deployment of biogeochemistry modelling setup, hypothesis and scenários

dynamic, and hydrogeological models. Here, we emphasize how OpenWQ can help unify different modeling approaches to simulate different landscapes and pollution problems.

1. Sub-catchment versus catchment-scale modeling. OpenWQ provides a unique oppor-434 tunity to harmonize spatial scales in water quality modeling. Water quality models 435 usually belong to one of two scales regarding the spatial domain, (a) sub-catchment 436 and (b) catchment. Sub-catchment water quality models, including river-reach mod-437 els such as WASP (Wool et al., 2020; Di Toro et al., 1983) and QUAL2E (Brown 438 & Barnwell, 1987), lake models such as MyLake (Saloranta & Andersen, 2007) and 439 Delft3D (Lesser et al., 2004), and aquifer models such as MODFLOW-MT3D (Har-440 baugh, 2005; Bedekar et al., 2016) and FEFLOW (Trefry & Muffels, 2007), require 441 the characterization of both horizontal and vertical boundary conditions because they 442 do not represent entire closed systems such a river basin. Instead, they focus on sub-443 regions within such closed systems. On the other hand, catchment models such as 444 INCA (Jackson-Blake et al., 2016; A. J. Wade et al., 2002; Whitehead et al., 1998) 445 and SWAT (Arnold et al., 2012) only require vertical boundary conditions (e.g., pre-446 cipitation) because they focus on the larger closed system that contains the entire 447 basin area. Accordingly, these two scales of models tend to focus on different pol-448 lution problems. Sub-catchment models tend to focus more on point sources, such 449 as wastewater discharge, and catchment models often look at diffuse pollution (e.g., 450 agriculture nutrients and fertilizer use). These differences also result in different 451 chemical species and biogeochemical cycles of focus, e.g., river models often address 452 problems related to low Dissolved Oxygen (DO) levels caused by biomass decomposi-453 tion and BOD (Biological Oxygen Demand) arising from wastewater discharges and 454 other point sources discharging directly in river and lakes. 455

OpenWQ enables multi-scale chemistry simulations that can be integrated into 456 hydro-models of sub-catchment and catchment scales; thus, it can help bridge 457 the gap between these two approaches, as well as allow for better integration of 458 cross-dependent biogeochemical cycles. For example, whereas the DO and nitrogen 459 cycles are often simulated in river-reach models, catchment models tend to focus 460 more on the nitrogen cycle and assume that there is an unlimited amount of DO avail-461 able. Although this assumption could be considered valid in many cases for surface 462 runoff, it is certainly erroneous for water bodies such as rivers, lakes, and groundwater. 463

- 2. Simple versus detailed biogeochemistry representation. Water quality models have 465 been developed with varying degrees of detail in the representation of biogeochemical 466 processes (Costa, Baulch, et al., 2020a). For example, while INCA and HYPE 467 (Lindström et al., 2010) provide simpler biogeochemical modeling approaches for 468 the nitrogen and phosphorous cycles, requiring a smaller number of reaction-kinetic 469 parameters to calibrate them may be more suitable for data-scarce regions; more 470 complex biogeochemical models like HSPF ("Hydrological Simulation Program-471 Fortran, User's manual for version 11: U.S. Environmental Protection Agency" 472 1997) arguably provide higher model fidelity but may only be applicable in data-rich 473 environments. However, when selecting a model for a particular region, modelers 474 often have to make compromises with process representation because some models 475 have more detailed physics-based coupled water-energy balance computations such 476 as AnnAGNPS (Bosch et al., 1998) but may offer more limited biogeochemical 477 capabilities (Costa, Baulch, et al., 2020a). OpenWQ enables addressing this issue 478 through its flexible eco-modelling lab, which allows testing different biogeochemical 479 conceptual models and modelling hypotheses, from simple biogeochemical cycles and 480 transformations to more complex and intertwined reaction networks involving dozens 481 or hundreds of chemical species. 482
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3. Unconnected versus interconnected biogeochemical cycling representation. Most catchment models simulate biogeochemical cycles in isolation. For example, popular

models such as SWAT and HYPE simulate the phosphorous and nitrogen cycles 486 without integrating them with the DO-BOD cycle, despite them being strongly 487 interrelated because DO is used in both cycles. In the nitrogen cycle, DO is used 488 in nitrification, where ammonia is oxidated into nitrite (NO_3) and nitrate (NO_2) . 489 In the DO-BOD cycle, DO is used in the biological, aerobic decomposition of 490 organic matter. OpenWQ aims to allow for a biogeochemistry representation that 491 is less compartmentalized, making interactions between cycling frameworks more 492 fluid and closer to reality (i.e., model fidelity). In reality, DO dynamics affect the 493 cycling of many chemical constituents, from nutrients to heavy metals. OpenWQ's 494 eco-modelling lab addresses this issue through its open reaction-network solver, 495 which provides the flexibility to deploy any number of biogeochemical cycling 496 representations of any number of chemical species simultaneously, which can be 497 connected or unconnected. 498

- 4. Background chemical transport driver. Typically, hydro-models are developed and 500 maintained by research communities with specific research motivations driven by a 501 regional context. These modeling tools often become highly specialized for particular 502 environments and applications, and modelers commonly find it hard to find a water 503 quality modeling tool that integrates such regionally important hydro-transport mod-504 eling capabilities. There is also in-house expertise in such communities and research 505 groups that is passed on over the years between elements of the research community, 506 which can make them reluctant to switch to other modeling tools because it may 507 involve steep learning curves. OpenWQ aims to address this issue by enabling exten-508 sion to water quality modeling capabilities directly on those models so that modelers 509 can continue using the hydro-modeling tool that they consider more suitable to the 510 environments on which they focus. 511
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4.2 Advancing Current Model Development Paradigms

Our focus on providing a flexible biogeochemical model representation framework that 513 is deployable on any process-based hydro-models and offers a unique opportunity to trans-514 form the way environmental studies involving water quality modeling are carried out. It 515 also provides a concrete way for streamlining collaboration, co-creation, and knowledge 516 and innovation transfer between the hydrological and biogeochemistry/soil science com-517 518 munities. Such streamlining is materialized by OpenWQ's coupler interface, which links developments/progress in the hydro-modeling tool (hydro-modeling community) with de-519 velopments/progress in OpenwQ (biogeochemistry communities). Another major issue with the current water quality modeling paradigm is that most models offer little flexibility to 521 adapt the representation of biogeochemistry processes to (1) local/region context, (2) data 522 availability, and (3) application(s) of interest, resulting in the need to make difficult choices 523 when selecting a modeling tool because compromises between the representation of hydro-524 fluxes and biogeochemistry are often required (Costa, Baulch, et al., 2020a). 525

526 4.3 Limitations

In OpenWQ, the physical transport of chemical constituents relies on the quality of the water fluxes passed by the host hydro-model. Although this may be an advantage because it allows the user to choose the hydro-model that best suits the modeling needs, it can become an issue if the calculated fluxes in the hydro-model are not accurate. Modelers should select the hydro-model coupled to OpenWQ that is the most suitable for the application at hand or pursue the coupling of OpenWQ to another hydro-model.

The native biogeochemistry module of OpenWQ that provides water quality-lab capabilities assumes that biogeochemical cycling can be represented via a series of sequential and parallel reaction networks. Although this representation is true for most chemical constituents, pollution problems, and environmental studies, some biogeochemistry may involve

formulations that rely on variable dependencies that may not be available in a particular cou-537 pled model. For example, simulating microbiological pollution in lakes, rivers, and beaches, 538 such as contamination with fecal coliforms originating from leaking septic tanks and wastew-539 ater discharges, requires the simulation of die-off rates due to exposure to solar radiation. 540 In the case of hydrodynamic models, such information may not be available. In the case of 541 hydrological models, which typically deal with such data, that variable dependency may not 542 have been passed into OpenWQ during the development of the coupler interface, so updates 543 to the coupler may be needed. OpenWQ's coupler functions have been designed to make 544 the addition of new dependency variables straightforward. 545

The portability of OpenWQ is materialized through a series of coupler functions and 546 wrapper interface functions that enable its coupling to hydro-models written in C++ or 547 Fortran. However, the use of a more standardized model interface framework, such as 548 the Basic Model Interface (BMI Peckham et al., 2013; E. W. Hutton et al., 2020), could 549 help streamline further the coupling process through the use of standard control and query 550 functions. This could make that model both easier to learn and easier to couple with other 551 software elements. BMI, for example, currently supports five languages: C, C++, Fortran, 552 Java, and Python. 553

554 5 Conclusions

This paper describes a unified framework for enabling multi-biogeochemical modeling capabilities in existing hydro-models. The framework, which we call Open Water Quality (OpenWQ), was designed with both coupler and water quality-lab modeling capabilities to enable (1) flexible co-creation and testing of biogeochemistry modeling representations, (2) systematic implementation and evaluation of alternative modeling approaches for process representation, and (3) identification of specific causes of model weaknesses.

This work arises from the recognition that the hydrological transport of contaminants 561 in the environment strongly affects their concentrations in aquatic ecosystems, but the inter-562 action between the biogeochemical and hydrological-hydrodynamic-hydrogeological commu-563 nities lacks the appropriate mechanisms for an efficient transfer of knowledge and innovation 564 between the two communities. Many hydro-models are developed and maintained by spe-565 cific research communities addressing particular regions and climate zones. The investment 566 in such tools creates invaluable in-house expertise, making them less likely to switch to 567 other modeling tools and hindering their expansion to water quality and ecological studies. 568 OpenWQ aims to address this problem and provide a concrete direction for innovation in 569 570 connecting communities through an optimized plug-in-like water quality model that can be coupled to existing hydro-modeling tools, extending the value of these tools for water 571 quality and ecological studies in their region of focus. Different process representations and 572 different spatial configurations can be integrated into the structural model core, which en-573 ables users to decompose the modeling problem into the individual decisions made as part 574 of model development and evaluate different "fine-grained" model development decisions in 575 a systematic and controlled way. 576

OpenWQ can provide the necessary model flexibility to progress toward answering the 577 following fundamental modeling questions and challenges: (1) which hydro-biogeochemical 578 processes should be represented explicitly in different environmental settings, and, corre-579 spondingly, which processes can be ignored or greatly simplified; (2) what modeling ap-580 proaches should be used to represent the dominant biogeochemistry at the spatial scale 581 of the model discretization; (3) how should heterogeneity in pollution and biogeochemical 582 processes be represented across spatial scales, including the complexity of transport across 583 landscapes; and (4) how can we provide insights into the sources of model uncertainty. The 584 companion paper describes the integration of OpenWQ into two hydro-models, SUMMA and 585 CRHM, describing how coupling interfaces between the two models have not only enabled 586 water quality simulation capacities in these host hydro-models but, even more importantly, 587

established a direct and permanent link for the transfer of innovation between the associated modeling communities, promoting cooperation and co-creation. Example applications of pollution studies enabled by the coupling of the tools are also provided to begin to address some of these fundamental modeling challenges.

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598 References

- Arheimer, B., Dahné, J., Donnelly, C., Lindström, G., & Strömqvist, J. (2012). Water
 and nutrient simulations using the HYPE model for Sweden vs. the Baltic Sea basin
 influence of input-data quality and scale. *Hydrology Research*, 43(4), 315-329. doi:
 10.2166/nh.2012.010
- Arnold, J. G., Moriasi, D. N., Gassman, P. W., Abbaspour, K. C., White, M. J., Srinivasan,
 R., ... Jha, M. K. (2012). SWAT: Model Use, Calibration, and Validation. *Trans.*ASABE, 55(4), 1491–1508.
- Arnold, J. G., Srinivasan, R., Muttiah, R. S., & Williams, J. R. (1998). Large area hydrologic
 modeling and assessment Part I: Model Development. JAWRA Journal of the American
 Water Resources Association, 34(1), 73-89. doi: https://doi.org/10.1111/j.1752-1688
 .1998.tb05961.x
- Beck, M. B. (1987). Water quality modeling: A review of the analysis of uncertainty. *Water Resources Research*, 23(8), 1393–1442. doi: 10.1029/WR023i008p01393
- Bedekar, V., Morway, E. D., Langevin, C. D., & Tonkin, M. J. (2016). MT3D-USGS version
 1: A U.S. Geological Survey release of MT3DMS updated with new and expanded transport capabilities for use with MODFLOW (Tech. Rep.). Reston, VA. doi: 10.3133/tm6A53
- Bosch, D., Theurer, F., Bingner, R., & Felton, G. (1998). Evaluation of the AnnAGNPS
 water quality model. Agricultural Non-Point Source Water Quality Models: Their Use
 and Application; John, EP, Daniel, LT, Rodney, LH, Eds, 45–54.
- Brown, L., & Barnwell, T. (1987). The enhanced stream water quality models qual2e and qual2e-uncas: documentation and user manual. *Environmental Protection Agency*.
- Clark, M. P., & Kavetski, D. (2010). Ancient numerical daemons of conceptual hydrological
 modeling: 1. fidelity and efficiency of time stepping schemes. Water Resources Research,
 46(10). doi: https://doi.org/10.1029/2009WR008894
- Clark, M. P., Kavetski, D., & Fenicia, F. (2011). Pursuing the method of multiple working
 hypotheses for hydrological modeling. *Water Resources Research*, 47(9). doi: https://
 doi.org/10.1029/2010WR009827
- Clark, M. P., Nijssen, B., Lundquist, J. D., Kavetski, D., Rupp, D. E., Woods, R. A., ...
 Rasmussen, R. M. (2015a). A unified approach for process-based hydrologic modeling: 1.
 modeling concept. *Water Resources Research*, 51 (4), 2498-2514. Retrieved from https://
 agupubs.onlinelibrary.wiley.com/doi/abs/10.1002/2015WR017198
 doi.org/10.1002/2015WR017198
- Clark, M. P., Nijssen, B., Lundquist, J. D., Kavetski, D., Rupp, D. E., Woods, R. A.,
 Marks, D. G. (2015b). A unified approach for process-based hydrologic model ing: 2. model implementation and case studies. Water Resources Research, 51(4), 2515 2542. Retrieved from https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1002/
 2015WR017200 doi: https://doi.org/10.1002/2015WR017200
- Costa, D., Baulch, H., Elliott, J., Pomeroy, J., & Wheater, H. (2020a). Modelling nutrient
 dynamics in cold agricultural catchments: A review. *Environmental Modelling & Software*,

- ⁶³⁸ *124*, 104586. doi: https://doi.org/10.1016/j.envsoft.2019.104586
- Costa, D., Baulch, H., Elliott, J., Pomeroy, J., & Wheater, H. (2020b). Modelling nutrient
 dynamics in cold agricultural catchments: A review. *Environmental Modelling & Software*,

124, 104586. doi: https://doi.org/10.1016/j.envsoft.2019.104586

641

- Costa, D., Burlando, P., & Liong, S.-Y. (2016). Coupling spatially distributed river and
 groundwater transport models to investigate contaminant dynamics at river corridor
 scales. *Environmental Modelling & Software*, 86, 91-110. doi: https://doi.org/10.1016/
 j.envsoft.2016.09.009
- Costa, D., Pomeroy, J., & Wheater, H. (2018). A numerical model for the simulation of snowpack solute dynamics to capture runoff ionic pulses during snowmelt: The
 pulse model. Advances in Water Resources, 122, 37-48. doi: https://doi.org/10.1016/
 j.advwatres.2018.09.008
- Costa, D., & Pomeroy, J. W. (2019). Preferential meltwater flowpaths as a driver of prefer ential elution of chemicals from melting snowpacks. Science of The Total Environment,
 662, 110–120. doi: 10.1016/J.SCITOTENV.2019.01.091
- Costa, D., Pomeroy, J. W., Brown, T., Baulch, H., Elliott, J., & Macrae, M. (2021).
 Advances in the simulation of nutrient dynamics in cold climate agricultural basins:
 Developing new nitrogen and phosphorus modules for the Cold Regions Hydrological
 Modelling Platform. Journal of Hydrology, 603, 126901. doi: https://doi.org/10.1016/
 j.jhydrol.2021.126901
- ⁶⁵⁸ Costa, D., Roste, J., Pomeroy, J., Baulch, H., Elliott, J., Wheater, H., & Westbrook, C.
 (2017). A modelling framework to simulate field-scale nitrate response and transport during snowmelt: The wintra model. *Hydrological Processes*, 31(24), 4250-4268. doi: https://doi.org/10.1002/hyp.11346
- Costa, D., Shook, K., Spence, C., Elliott, J., Baulch, H., Wilson, H., & Pomeroy, J. W.
 (2020). Predicting variable contributing areas, hydrological connectivity, and solute
 transport pathways for a canadian prairie basin. Water Resources Research, 56(12),
 e2020WR027984. doi: https://doi.org/10.1029/2020WR027984
- Davies, T. D., Brimblecombe, P., Tranter, M., Tsiouris, S., Vincent, C. E., Abrahams,
 P., & Blackwood, I. L. (1987). *The Removal of Soluble Ions from Melting Snowpacks*(H. G. Jones & W. J. Orville-Thomas, Eds.). Dordrecht: Springer Netherlands. doi:
 10.1007/978-94-009-3947-9_20
- Di Toro, D. M., Fitzpatrick, J. J., Thomann, R. V., & Hydroscience, I. (1983). Documentation For Water Quality Analysis Simulation Program (WASP) And Model Verification Program (MVP) (Tech. Rep.).
- Frank, E. G., & Schlenker, W. (2016). Balancing economic and ecological goals. *Science*, 353(6300), 651-652. doi: 10.1126/science.aaf9697
- Fu, B., Merritt, W. S., Croke, B. F., Weber, T. R., & Jakeman, A. J. (2019). A review of catchment-scale water quality and erosion models and a synthesis of future prospects. *Environmental Modelling & Software*, 114, 75–97. doi: 10.1016/J.ENVSOFT.2018.12
 .008
- Harbaugh, A. W. (2005). MODFLOW-2005, the US Geological Survey modular groundwater model: the ground-water flow process (Vol. 6). US Department of the Interior, US
 Geological Survey Reston, VA, USA.
- Hutton, C., Wagener, T., Freer, J., Han, D., Duffy, C., & Arheimer, B. (2016). Most
 computational hydrology is not reproducible, so is it really science? Water Resources
 Research, 52(10), 7548–7555. doi: https://doi.org/10.1002/2016WR019285
- Hutton, E. W., Piper, M. D., & Tucker, G. E. (2020). The basic model interface 2.0: A
 standard interface for coupling numerical models in the geosciences. Journal of Open
 Source Software, 5(51), 2317. doi: 10.21105/joss.02317
- Hydrological Simulation Program-Fortran, User's manual for version 11: U.S. Environmental Protection Agency. (1997). National Exposure Research Laboratory, Athens, Ga., EPA/600/R-97/080, 755 p.
- Jackson-Blake, L. A., Wade, A. J., Futter, M. N., Butterfield, D., Couture, R.-M., Cox,

B. A., ... Whitehead, P. G. (2016). The INtegrated CAtchment model of phosphorus dynamics (INCA-P): Description and demonstration of new model structure
and equations. *Environmental Modelling & Software*, 83, 356–386. doi: 10.1016/
J.ENVSOFT.2016.05.022

- ⁶⁹⁶ Jones, H. G. (1999). The ecology of snow-covered systems: a brief overview of nutrient ⁶⁹⁷ cycling and life in the cold. *Hydrological Processes*, 13(14-15), 2135–2147. doi: 10.1002/ ⁶⁹⁸ (SICI)1099-1085(199910)13:14/15(2135::AID-HYP862)3.0.CO;2-Y
- Keith, D. A., Ferrer-Paris, J., Nicholson, E., Bishop, M. J., Polidoro, B. A., Ramirez-Llodra,
 E., ... Kingsford, R. T. (2022). A function-based typology for earth's ecosystems. *Nature*,
 610(7932), 513–518. doi: 10.1038/s41586-022-05318-4
- ⁷⁰² Kitware. (2022). *CMake: Reference documentation*.
- Lesser, G., Roelvink, J., van Kester, J., & Stelling, G. (2004). Development and validation of
 a three-dimensional morphological model. *Coastal Engineering*, 51(8), 883-915. (Coastal
 Morphodynamic Modeling) doi: https://doi.org/10.1016/j.coastaleng.2004.07.014
- Lilbaek, G., & Pomeroy, J. W. (2007). Modelling enhanced infiltration of snowmelt ions into frozen soil. *Hydrological Processes*, 21(19), 2641–2649. doi: 10.1002/hyp.6788
- Lindström, G., Pers, C., Rosberg, J., Strömqvist, J., & Arheimer, B. (2010). Development and testing of the HYPE (Hydrological Predictions for the Environment) water quality model for different spatial scales. *Hydrology Research*, 41 (3-4), 295. doi: 10.2166/nh.2010
 .007
- Marsh, P., & Pomeroy, J. W. (1999). Spatial and temporal variations in snowmelt runoff
 chemistry, Northwest Territories, Canada. Water Resources Research, 35(5), 1559–1567.
 doi: 10.1029/1998WR900109
- Mekonnen, B. A. (2016). Modeling and management of water quantity and quality
 in cold-climate Prairie Watersheds (Unpublished doctoral dissertation). University of
 Saskatchewan.
- Moore, D. S., Bingner, R. L., & Theurer, F. D. (2006). ANNAGNPS: Accounting for
 Snowpack, Snowmelt, and Soil Freeze-Thaw. In *Eighth federal interagency sedimentation conference (8thfisc)*, (pp. 475–482). Reno, NV, USA.
- Partow, A. (1999). ExprTk C++ Mathematical Expression Library.
- Peckham, S. D., Hutton, E. W., & Norris, B. (2013). A component-based approach to
 integrated modeling in the geosciences: The design of csdms. *Computers & Geosciences*,
 53, 3–12.
- Pomeroy, J., Brown, T., Fang, X., Shook, K., Pradhananga, D., Armstrong, R., ... Lopez
 Moreno, J. (2022). The cold regions hydrological modelling platform for hydrological diagnosis and prediction based on process understanding. *Journal of Hydrology*, 615, 128711. doi: https://doi.org/10.1016/j.jhydrol.2022.128711
- Pomeroy, J., Davies, T., Jones, H., & Marsh, P. (1999). Transformations of snow chemistry
 in the boreal forest: accumulation and volatilization. *Hydrological Processes*, 13(1415),
 2257–2273.
- Pomeroy, J. W., Davies, T. D., & Tranter, M. (1991). The Impact of Blowing Snow on
 Snow Chemistry. In Seasonal snowpacks: Processes of compositional change (pp. 71–113).
 Springer, Berlin, Heidelberg. doi: 10.1007/978-3-642-75112-7_4
- Pomeroy, J. W., Gray, D. M., Brown, T., Hedstrom, N. R., Quinton, W. L., Granger, R. J.,
 & Carey, S. K. (2007). The cold regions hydrological model: A platform for basing process
 representation and model structure on physical evidence. *Hydrological Processes*, 21(19),
 2650–2667. doi: {10.1002/hyp.6787}
- Refsgaard, J., & Storm, B. (1995). Mike she.[in:] singh vp (ed.), computer models of
 watershed hydrology. *Water Resources Publication, Colorado*, 809–847.
- Richmond, B. (2003). STELLA: An Introduction to Systems Thinking.
- Saloranta, T. M., & Andersen, T. (2007). Mylake—a multi-year lake simulation model code
- suitable for uncertainty and sensitivity analysis simulations. *Ecological Modelling*, 207(1),
 45-60. (Uncertainty in Ecological Models) doi: https://doi.org/10.1016/j.ecolmodel.2007
- .03.018

- Sanderson, C., & Curtin, R. (2016). Armadillo: a template-based c++ library for linear
 algebra. Journal of Open Source Software, 1(2), 26. doi: 10.21105/joss.00026
- Sanderson, C., & Curtin, R. (2018). A user-friendly hybrid sparse matrix class in c++. , 422–430. doi: $10.1007/978-3-319-96418-8_50$
- Shoemaker, L. (1997). Compendium of tools for watershed assessment and tmdl development.
- Shook, K., Pomeroy, J. W., Spence, C., & Boychuk, L. (2013). Storage dynamics simulations in prairie wetland hydrology models: evaluation and parameterization. *Hydrological Processes*, 27(13), 1875–1889. doi: 10.1002/hyp.9867
- Trefry, M. G., & Muffels, C. (2007). Feflow: A finite-element ground water flow and transport modeling tool. *Groundwater*, 45(5), 525-528. doi: https://doi.org/10.1111/ j.1745-6584.2007.00358.x
- Wade, A., Jackson, B., & Butterfield, D. (2008). Over-parameterised, uncertain 'mathematical marionettes' How can we best use catchment water quality models? An example of
 an 80-year catchment-scale nutrient balance. Science of The Total Environment, 400(1-3),
 52–74. doi: 10.1016/j.scitotenv.2008.04.030
- Wade, A. J., Durand, P., Beaujouan, V., Wessel, W. W., Raat, K. J., Whitehead, P. G.,
 ... Lepisto, A. (2002). A nitrogen model for european catchments: Inca, new model
 structure and equations. *Hydrology and Earth System Sciences*, 6(3), 559–582. doi: 10.5194/hess-6-559-2002
- Wellen, C., Kamran-Disfani, A. R., & Arhonditsis, G. B. (2015). Evaluation of the current state of distributed watershed nutrient water quality modeling. *Environmental Science and Technology*, 49(6), 3278–3290. doi: 10.1021/es5049557
- Whitehead, P., Wilson, E., Butterfield, D., & Seed, K. (1998). A semi-distributed integrated flow and nitrogen model for multiple source assessment in catchments (inca): Part ii application to large river basins in south wales and eastern england. Science of The Total Environment, 210-211, 559-583. doi: https://doi.org/10.1016/S0048-9697(98)00038-2
- Wool, T., Ambrose, R. B., Martin, J. L., & Comer, A. (2020). Wasp 8: The next generation
 in the 50-year evolution of usepa's water quality model. *Water*, 12(5). doi: 10.3390/
 w12051398
- Yang, X., Tetzlaff, D., Soulsby, C., & Borchardt, D. (2022). Hiwaq v1.0: A flexible catchment water quality assessment tool with compatibility for multiple hydrological model structures. *Geoscientific Model Development Discussions*, 2022, 1–26. doi: 10.5194/gmd-2022-239

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A multi-chemistry modelling framework to enable flexible and reproducible water quality simulations in existing hydro-models: 1. The OpenWQ concept and the water quality modelling lab

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11 Abstract

This work advances the incorporation and cross-model deployment of multi-12 biogeochemistry and ecological simulations in existing process-based hydro-modelling tools. 13 It aims to transform the current practice of water quality modelling as an isolated research 14 effort into a more integrated and collaborative activity between science communities. Our 15 approach, which we call "Open Water Quality" (OpenWQ), enables existing hydrological, 16 hydrodynamic, and groundwater models to extend their capabilities to water quality sim-17 ulations, which can be set up to examine a variety of water-related pollution problems. 18 19 OpenWQ's objective is to provide a flexible biogeochemical model representation that can be used to test different modelling hypotheses in a multi-disciplinary co-creative process. In 20 this paper, we introduce the general approach used in OpenWQ. We detail aspects of its 21 architecture that enable its coupling with existing models. This integration enables water 22 quality models to benefit from advances made by hydrologic- and hydrodynamic-focused 23 groups, strengthening collaboration between the hydrological, biogeochemistry, and soil sci-24 ence communities. We also detail innovative aspects of OpenWQ's modules that enable 25 biogeochemistry lab-like capabilities, where modellers can define the pollution problem(s) 26 of interest, the appropriate complexity of the biogeochemistry routines, and test different 27 modelling hypotheses. In a companion paper, we demonstrate how OpenWQ has been 28 coupled to two hydrological models, the "Structure for Unifying Multiple Modelling Alter-29 natives" (SUMMA) and the "Cold Regions Hydrological Model" (CRHM), demonstrating 30 the innovative aspects of OpenWQ, the flexibility of its couplers and internal spatiotemporal 31 data structures, and the versatile eco-modelling lab capabilities that can be used to study 32 different pollution problems. 33

³⁴ 1 Introduction

Societies are becoming increasingly aware of the importance of harmonizing economic development with thriving aquatic ecosystems (Keith et al., 2022; Frank & Schlenker, 2016). This has led to an increase in the use of ecological and water quality models to study management solutions to pollution problems such as nutrient excess due to the use of fertilizers or microbiological contamination due to the discharge of untreated wastewater in rivers.

Over the past decades, substantial progress has been made in computational hydrology, 40 producing many powerful generic and fit-for-purpose modeling tools. Some of these models 41 have been particularly tailored to specific environments, such as the Cold Regions Hydrolog-42 ical Model (J. Pomeroy et al., 2022; J. W. Pomeroy et al., 2007), the Wetland DEM Ponding 43 Model (Shook et al., 2013), and the FLUXOS-Overland model (Costa, Shook, et al., 2020), 44 that have been developed to deal with the specific hydrological challenges of simulating the 45 Canadian Prairie pothole region due to the lack of well-defined river networks. Unfortu-46 nately, the science and modelling progress achieved by such hydrological model development 47 communities rarely transfers to (and translates into) water quality modeling improvements 48 because hydrological and water quality modeling tools are typically developed in isolation, 49 a problem that becomes clearly noticeable when examining the hydrological calculations 50 embedded in popular water quality modeling tools, such as SWAT (Arnold et al., 1998), 51 HYPE (Lindström et al., 2010; Arheimer et al., 2012), and INCA (Whitehead et al., 1998; 52 A. J. Wade et al., 2002; Jackson-Blake et al., 2016), that often rely on simplifications that 53 may be limiting for some regions and applications (Costa, Baulch, et al., 2020a). 54

Besides the issues with the communication and transfer of knowledge between these two scientific communities, other scientific and technical biogeochemistry-specific challenges remain critical for reliable simulations across climate regions, landscapes, and seasons (Costa, Baulch, et al., 2020a; Fu et al., 2019; Wellen et al., 2015). Some key challenges include (1) the adequate representation of the complexity and heterogeneity of biogeochemical processes and their dependency on weather, soil, and sediment characteristics; (2) the dependency on accurate hydrological drivers to reliably track the vertical and lateral movement of chemical constituents (being at the end of the climate-hydrology-ecology modeling chain invariably
 propagates uncertainties into water quality simulations); and (3) the often limited range of
 applicability of models to specific pollution problems and landscape conditions due to their
 rigid and heavily parameterized representation of water quality processes.

Underlying all of these issues are foundational questions of scientific reproducibility 66 in computational hydrology (C. Hutton et al., 2016), as well as considerable challenges in 67 defining the appropriate complexity, scale, and scope of water quality models (Costa, Baulch, 68 et al., 2020a; Moore et al., 2006; Mekonnen, 2016; Shoemaker, 1997). This challenge is in part 69 70 due to uncertainties associated with the prediction of many hydrological and biogeochemical responses at different spatial scales and geographical settings, which are aggravated by the 71 often sparse water quality data available for testing and validation of the models. This raises 72 critical questions for the design, application, and benefit of such modeling tools (Beck, 1987; 73 Moore et al., 2006; A. Wade et al., 2008; Costa, Baulch, et al., 2020a). 74

Improving process-based hydro-biogeochemical models for terrestrial hydrological simu-75 lations requires effective and continuous integration of progress across several research areas, 76 including hydrology, biogeochemistry, and soil science. Models must combine practical as-77 pects related to model application with theoretical scientific insights at various degrees of 78 scientific maturity and geographical applications. Such cross-disciplinary modeling efforts 79 require modelers to make model development decisions based on several considerations com-80 mon to hydrological models (Clark et al., 2011) that include (1) model fidelity, complexity 81 and practicality, (2) scientific reproducibility and transparency, and (3) data availability. 82

It has been recognized that more flexible water quality modelling approaches are needed 83 to address these issues (Yang et al., 2022; Costa, Shook, et al., 2020). Some approaches 84 have emerged that address some of the challenges and improve transparency and engage-85 ment among hydrologists, biogeochemists, soil scientists, and decision-makers, such as the 86 STELLA (Structural Thinking and Experiential Learning Laboratory with Animation) soft-87 ware (Richmond, 2003) and the MIKE Eco Lab, which is a water quality modelling addition 88 to the MIKE tools (Refsgaard & Storm, 1995). However, such tools are not suitable for 89 integration into modelling platforms developed by various hydrological communities around 90 the world. In the case of STELLA, the tool is standalone and more suitable for applica-91 tions where simpler batch-reactor assumptions are applicable, such as for small reservoirs 92 and wastewater treatment plants. The MIKE tools from DHI now provide an Eco-Lab 93 module for more flexible water quality and ecological simulations, but this module is a pro-94 prietary tool integrated into the MIKE ecosystem of tools, so it does not benefit the wider 95 hydrology and modelling communities, and it is constrained by the chemical constituents and types of biogeochemical processes that are explicitly introduced by the MIKE model 97 developers. More recently, Yang et al. (2022) proposed a new model named HiWaQ for 98 flexible catchment water quality assessments with compatibility for multiple hydrological 99 model structures. This is a valuable contribution toward a more unified hydrological-water 100 quality modelling approach, but the current version is limited to the nitrogen cycle. 101

¹⁰² 2 Current modelling capabilities and needs

Widely used process-based catchment nutrient models such as SWAT, HYPE, and 103 INCA have been at the forefront of innovation in water quality modeling, paving a way 104 for researchers and decision-makers around the world to investigate solutions to a variety of 105 pollution problems, particularly related to nutrient pollution. These models have become 106 increasingly complex and heavily parameterized, but they remain a limited representation of 107 reality because hydro-biogeochemical processes are highly complex in natural environments 108 (A. Wade et al., 2008; Beck, 1987; Costa, Baulch, et al., 2020b). These models typically 109 simulate a series of biogeochemical processes conceptualized to address particular pollutions 110 problems, and processes are represented through a combination of empirical and physico-111 chemically based methods, often leading to many calibration parameters and thus increasing 112

the risk of parameter equifinality (Costa, Baulch, et al., 2020b). Sparse and sporadic water
 quality measurements, which are common problems for most water bodies, limit the further
 application and development of these water quality models.

The combination of process-representation methods with varying degrees of empiricism 116 is in part due to knowledge gaps in understanding the drivers and controls of hydrological 117 and biogeochemical responses at various spatial scales and across different landscapes and 118 climate zones. In regions where relatively uncommon processes may play an important role 119 in the overall water quality dynamics, the use of these models becomes problematic because 120 121 there is little flexibility for adjusting conceptual models, adding or removing processes, or testing different modeling hypotheses. For example, research has shown that in some 122 cold regions, processes such as preferential infiltration of hydrochemical into frozen soils 123 (Lilback & Pomerov, 2007), preferential elution of hydrochemicals from melting snowpacks 124 (Davies et al., 1987; Marsh & Pomeroy, 1999; Costa & Pomeroy, 2019), microbial uptake 125 and fixing of nitrogen in melting snowpack (Jones, 1999), and volatilization of nitrogen 126 during snow redistribution and sublimation (J. W. Pomeroy et al., 1991; J. Pomeroy et al., 127 1999) may affect water quality, but they are not represented in most popular models. This 128 lack of representation may compromise their use in such regions and calls for a more flexible 129 approach to water quality modeling that enables a systematic and controlled approach 130 for the addition and removal of processes as needed to reflect (1) regional and climate 131 characteristics, (2) data limitations, and (3) objectives of the study. 132

3 The OpenWQ concept

3.1 Overview

134

OpenWQ is a coupler-modelling framework designed to provide portable and customiz-135 able multi-chemistry modelling capabilities to existing hydro-models. The vision is to create 136 a tool that could plug into existing process-based hydrological, hydrodynamic, and ground-137 water models to extend their capabilities to environmental and ecological studies. The 138 approach allows tailoring chemical-microbiological constituents and biogeochemistry-cycling 139 processes to enable the representation of different pollution problems and landscapes, as well 140 as to compare modelling hypotheses. This framework is a much-needed effort to bring the 141 hydrology and biogeochemistry communities together, optimizing research and investment 142 efforts. The work stems from previous model developments by Environment and Climate 143 Change Canada and the University of Saskatchewan, particularly CRHM-WQ (Cold Regions 144 Hydrological Model - Water Quality) (Costa et al., 2021) that extends the original CRHM 145 platform (hydrology) model to nitrogen and phosphorus simulations (J. W. Pomeroy et al., 146 2007), the WINTRA framework (Costa et al., 2017), the multiphase multilayer PULSE snow 147 hydrochemistry model (Costa et al., 2018), and the FLUXOS-OVERLAND model for wa-148 tershed hydrodynamic-water quality simulations suitable for Prairie regions (Costa, Shook, 149 et al., 2020). 150

OpenWQ aims to address three main challenges with existing water quality models: (1) 151 structural rigidity in the representation of chemical constituents and biogeochemical pro-152 cesses, (2) over-simplification and limitations of hydro-flux calculations, and (3) inadequacy 153 for testing different modelling hypotheses for proper quantification of structural uncertainty. 154 Structural rigidity is perhaps the key factor that hinders the effective use of models across 155 landscapes and in complex, diverse environments (e.g., permafrost, peatlands, variable con-156 tributing areas) that require more investigative, open-ended, and interactive simulation 157 approaches. A flexible environment that enables integrating new methods and concepts 158 from complementary disciplines and experts (e.g., limnology, soil science, biogeochemistry) 159 is critical to advance science and promote meaningful and impactful cross-disciplinary col-160 laborations. The static, hard-coded implementation of biogeochemical reaction-network 161 transformations limits their suitability for a wide range of environmental problems. It also 162 hinders multi-disciplinary co-creation efforts because models provide little flexibility for 163

changing, expanding, and testing different strategies for biogeochemical cycling representa tion. Finally, the hydro-flux calculations that are embedded in water quality modelling tools
 are often outdated or limited compared to dedicated, disciplinary hydro-models (e.g., hydro logical, hydrodynamic, and groundwater modelling tools). Uncertainty in model structure,
 process representation, and future scenarios (e.g., climate change) cannot be adequately
 quantified without flexible and transparent modelling structures.

170 3.2 General design

This section describes how OpenWQ was designed to address the key challenges in 171 water quality modelling described before. First, focus is given to describe how OpenWQ 172 tackles the need for more robust hydro-flux and associated solute transport calculations 173 through integration within existing hydro-models. This includes details on how OpenWQ 174 can be linked to existing hydro-models considering key practical aspects of version control 175 (Section 3.2.1), portability (Section 3.2.2), code integration (Sections 3.2.3), and state vari-176 ables and spatiotemporal discretization (Section 3.2.4). Second, focus is given to describe 177 how OpenWQ tackles the need for more flexible representation of physical and biochemical 178 processes, which includes aspects of model input (Section 3.3), model structure to enable 179 testing modelling hypotheses and quantifying structural uncertainty (Sections 3.4 and 3.5). 180

181

3.2.1 Obtaining OpenWQ

OpenWQ can be obtained from the official GitHub repository: https://github.com/ 182 ue-hydro/openwq. The recommended method for obtaining OpenWQ is to clone the repos-183 itory with "git clone". Downloading the repository is also possible, but it is less advisable 184 because OpenWQ is continuously updated and improved. OpenWQ is designed for the 185 user to update their copy quickly and efficiently with "git pull" instead of requiring them 186 to download each new version. The cloning (or download) should be performed inside the 187 host hydro-model code directory as shown in Fig. 1. It will create a new folder named 188 "openwq" with the entire repository. The "main" branch should be used because it will 189 contain the latest official release version. Compiling OpenWQ is carried out with CMake 190 (Kitware, 2022) and the provided "CMakeLists" file. When new versions of OpenWQ are 191 made available, they will only require the user to recompile using the same "CMakeLists" 192 file. Once the coupling of OpenWQ to a hydro-model is completed, carrying out updates to 193 OpenWQ's source code will not break OpenWQ's API calls and interface/coupler functions 194 in the host hydro-model. 195

196

3.2.2 Portability via internal dynamic coupler

In order to optimize the implementation of OpenWQ in hydro-models, the model has 197 been designed as an internal coupler module (Fig. 2). External coupling (i.e., OpenWQ 198 reading output files from the host-model and running standalone) was not a viable option 199 because hydro-models, particularly hydrological models, often deal with many water fluxes 200 moving around vertically and horizontally within and across hydrological compartments 201 (e.g., snow, soil), which are not always possible to export and disentangle. Even in cases 202 where models allow exporting all water fluxes separately, mapping those in OpenWQ and 203 204 harmonizing units to correctly compute the corresponding solute mass transport would be extremely difficult for most applications. Focusing on a flexible full-coupling approach with 205 minimal code re-engineering allowed us to address this problem, with OpenWQ being specif-206 ically designed to adapt its internal structure (e.g., spatial domain, temporal resolution) to 207 that of the host hydro-model with the support of interface and coupler functions that estab-208 lish one-way, plug-in-type communications between the two simulation systems. Although 209 these interface and coupler functions may need adjustments to harmonize particular aspects 210 of the host hydro-models, a generic "hydro-link" file is provided with a template for these 211



Figure 1: General model structure used in hydrological model

functions. It is possible to create such a template because most hydro-models follow the general model architecture depicted in Fig. 3.



Figure 2: OpenWQ concept as a coupler

The integration of OpenWQ into existing hydro-models is carried out through four 214 coupler functions that are responsible for (1) converting datatypes and data structures be-215 tween OpenWQ and the "host" hydro-model, (2) passing hydro-fluxes into OpenWQ, and 216 (3) calling OpenWQ's APIs (Fig. 4). These four coupler functions are invoked through inter-217 face routines contained within a C++ file with the default name "OpenWQ_hydrolink.cpp". 218 This is a one-way communication from the host hydro-model to OpenWQ, so no information 219 is returned to the host model. The interface functions are generic and serve as templates 220 (coupling recipes) that have been optimized to streamline the coupling procedure. The ad-221 justments needed in the interface functions are to ensure that the data types and structure 222



Figure 3: General model structure commonly used in dynamic hydrological, hydrodynamic, and groundwater model

used in the hydro-model are properly translated into OpenWQ's own data structure and conventions so that OpenWQ's APIs can perform adequately.

The following general steps can be performed to couple OpenWQ to a hydro-model: 225 (STEP 1) identify critical elements of hydro-model structure (see Fig. 3), (STEP 2) git-clone 226 OpenWQ, (STEP 3) create the C++ OpenWQ classes and objects to be loaded as modules 227 or libraries in the hydro-model, (STEP 4) identify the appropriate places in the hydro-228 model to call OpenWQ's coupler functions and APIs (see Fig. 4), (STEP 5) materialize 229 such calls, (STEP 6) adjust OpenWQ's coupler function to adapt to hydro-model data 230 structures (referred to "COUPLER CODE", see below), and (STEP 7) compile the new 231 coupled model. Fig. 5 shows the general structure of each of the four coupler functions, 232 where the location and purpose of the "COUPLER CODE" block are also explained: 233

- Coupler function 1 (*openwq::decl*) invokes a series of API calls that handle tasks associated with the initial configuration of the model, initialization of variables, and pre-processing of the input data;
 Coupler function 2 (*openwq::run_time_start*) invokes a series of API calls that handle tasks required at the start of each time step;
 Coupler function 3 (*openwq::run_space*) contains a series of API calls that handle tasks related to the spatial domain of the model; and
 Coupler function 4 (*openwq::run_space*)
- 4. Coupler function 4 (*openwq::run_time_end*) contains a series of API calls that handle
 tasks required at the end of each time step.



Figure 4: General placement of the calls for OpenWQ's coupler functions

243 3.2.3 General architecture

OpenWQ has been created in a way that separates the physics and biochemistry cal-244 culations from the numerical implementation. Such an approach was implemented in the 245 SUMMA model (Clark et al., 2015a,b), and we adopted it here to the development of 246 OpenWQ's core structure to improve scalability. This approach addresses a major problem 247 with many hydrological, hydrodynamic, and water quality models where the specification of 248 the model equations is intertwined with their numerical solution (Clark & Kavetski, 2010). 249 This specification complicates the selection and assessment of different model representa-250 tions (hypotheses) and makes introducing and evaluating alternative numerical methods 251 challenging. As such, the state variables of OpenWQ are only updated inside OpenWQ's 252 own numerical solver based on rates of chemical mass changes (time and space derivatives) 253 caused by different physical and biogeochemical processes. These processes and "rates of 254 change" are computed separately, in process-specific routines, and then passed into the nu-255 merical solver for a controlled and contained update of OpenWQ's state variables, which is 256 currently performed as a finite volume problem solved via a simple forward Euler method. 257 The separation of the numerical solver from the physics-biogeochemistry calculations will 258 enable improving the robustness of the numerical implementation in a contained and con-259 trolled manner in the future. 260

The modules in OpenWQ are divided into four groups. Each of these groups is respon-261 sible for the calculation of a chemical mass change (time derivative) driven by a particular 262 process (or phenomenon) or group of processes (or phenomena), specifically, (1) initial con-263 ditions (dm_ic) , (2) sinks and sources of chemical load (i.e., chemical mass entering or exiting 264 the model domain) (dm_ss) , (3) biogeochemical processes $(dm_dt_chemistry)$, and (4) phys-265 ical transport of chemical constituents with water flow $(dm_{-}dt_{-}transport)$ (Fig. 6). Each of 266 the modules, in turn, enables a series of modeling options that can be explored, but only 267 one module can be activated for each of the four spatiotemporal-derivative calculations. In 268 other words, the first layer of decision in representing a process pertains to the selection of 269 the key modules that will be responsible for computing each of these four derivatives. 270

CLASS OpenWQ_couplercall

OpenWQ_couplercalk

CALL

START

CLASS OpenWQ_couplerc

CLASS OpenWQ

CALL

START

Before space loop

SPACE

CLASS OpenWC

(BR

Figure 6: General modular architecture of OpenWQ that separates the different modules and methods available

3.2.4 State variables & Model-adaptive spatial and temporal discretization

271

OpenWQ was designed to automatically adapt to the hydro-model spatial and tempo-272 ral discretization structure, which can include Hydrological Response Units (HRUs), 1D–3D 273 spatial distributed meshes, structured or unstructured meshes, and multiple domains or hy-274 drological compartments (e.g., snow, soil, groundwater). OpenWQ's state variables (column 275 1 in Fig. 7), such as solute mass, are stored in a hierarchical data structure organized with 276 the following nested fields: (1) domain that refers to hydrological compartments (column 2) 277 in Fig. 7), (2) chemical species (column 3 in Fig. 7), and (3) sub-domain that refers to the 278 internal 1D–3D spatial discretization of the domains or hydrological compartments (column 279 4 in Fig. 7). The hierarchical data structures are built via the Armadillo C++ library for 280 linear algebra and scientific computing (Sanderson & Curtin, 2016, 2018). 281

The model interface and coupler are responsible for passing information about the 282 spatiotemporal configuration of the host hydro-model into OpenWQ. Such domain config-283 uration options are specific to each host hydro-model and depend on its particular domain 284 discretization scheme and model decisions specified by the users of the hydro-model. Once 285 that information is digested by the couplers and passed on to OpenWQ, a series of API 286 calls dynamically create the corresponding hierarchical data structures and sub-structures 287 that match those in the host model. Each of the data structures stores information about 288 a state variable or a supporting non-state variable (e.g., water fluxes, time derivatives). 289 The state-variable data structures record the spatiotemporal evolution of the mass of the 290 different chemical species tracked in each model domain or hydrological compartment (e.g., 291 snow, soil, lake). 292

STATE VARIABLES

arma::field(arma:field(arma::cube))

Figure 7: State variables are stored in OpenWQ via dynamic hierarchical data structures

293

3.3 Model setup and configuration: Inputs and Outputs

The configuration of OpenWQ is provided via JavaScript Object Notation (JSON) files. JSON is an ideal format for large data inputs that have a hierarchically structured relationship. JSON files are composed of key/value pairs, e.g., "RUN_MODE_DEBUG: TRUE". OpenWQ requires four JSON files to run. Each of these files deals with a particular aspect of the model setup and is given one of the following designations (1) "runManagement", (2) "Biogeochemistry", (3) "Configuration", and (4) "Source/Sink".

The "RunManagement" file is the entry point to OpenWQ. It provides the basic instructions, simulation and model output decisions, and the full or relative paths to the other JSON files needed to run a model. Currently, OpenWQ supports HD5F and CSV output file formats. The "Biogeochemistry" file is where the chemical species and biogeochemical cycling frameworks are created and characterized. It uses a standard template structure for the characterization transformations and the chemical species involved. Different "Biogeochemistry" files can be prepared to simulate different pollution problems, which can be readily loaded simultaneously (or used separately or swapped as needed) into a simulation

via the "RunManagement" file. The "Biogeochemistry" file is also where the contaminant 308 species available for the simulation are defined and given a "tag" that can be invoked in 309 other input files. All cycling frameworks and their respective transformations are also pro-310 vided with a unique "tag" that can be used to flexibly set up the model as desired. These 311 cycling framework tags are particularly important in the "Configuration" file to assign the 312 desired transformations to each domain or hydrological compartment of the model (e.g., 313 snow, soil, groundwater). The "Configuration" file also includes information about the ini-314 tial conditions. A "Sink/Source" file provides information about the mass exchange with 315 regions outside the model domain. This information can include external chemical mass 316 loading into (source) or loss from (sink), the model domain that can be associated with 317 particular external water fluxes, as in the case of precipitation, or mass inputs that can 318 be independent of the hydrological cycle. Typical examples of such mass inputs can in-319 clude fertilizer application, which is a relatively localized source, or atmospheric deposition, 320 which is a relatively distributed source. There can be multiple "Sink/Source" files in a given 321 OpenWQ configuration. 322

323

3.4 Native Modules and Process representation

OpenWQ contains four main groups of modules, each producing rates of chemical mass change (time derivatives) associated with specific processes, phenomena, and model aspects, which are passed into the numerical solver for updating the state-variables throughout the simulation (see Fig. 6). The modules deal with (1) initial conditions, (2) sinks and sources, (3) transport with water, and (4) biogeochemistry.

The first two groups of modules deal with initial conditions and sinks and external 329 sources with a focus on translating user inputs into the simulation. For example, the sink 330 and source module applies chemical load as prescribed by the users that can include (1) 331 continuous load arising from sources like atmospheric deposition or the outlet of a wastewater 332 treatment plant or (2) episodic/instantaneous loads arising from sources such as fertilizer 333 and chemical spills. The remaining two groups of modules focus respectively on (1) the 334 physical transport of chemical constituents as water moves through the system and (2) 335 biogeochemical transformations. 336

337

3.4.1 Transport Module

Currently, there are two options available for computing the physical transport of dis-338 solved solutes and fully suspended sediments. The first option accounts for both advection 339 and dispersion and solves the hyperbolic-parabolic advection-diffusion partial differential 340 equation (PDE) in up to 3 dimensions depending on the host model spatial discretiza-341 tion scheme (Equation 1); the second option only accounts for advection (Equation 2). In 342 both cases, the PDEs are solved inside OpenWQ's numerical solver as described in Sec-343 tion 3.2.3 using the specific hydro-fluxes and corresponding source and recipient domain 344 cells prescribed by the host models. This means that for each water flux computed by the 345 host-model (at each grid cell and timestep), OpenWQ calculates the corresponding solute 346 mass transported, which will be intimately linked to the internal spatio-temporal water-flux 347 representations of the host-model (i.e., if the host model represents snow as a 1 vertical layer 348 domain (lumped), OpenWQ will compute solute concentrations at that spatial resolution; 349 if the host model runs at daily time steps, OpenWQ will calculate solute concentrations at 350 that temporal resolution). The physical transport can be between computation elements 351 (e.g., HRUs, grid cells) and across domains or hydrological compartments (e.g., runoff, soil 352 saturated, soil unsaturated, canopy), as prescribed by the host model. 353

$$\frac{\partial (Vc_s)}{\partial t} + \nabla \cdot (V\vec{u}c_s) = \nabla \cdot (V\mathbf{E} \cdot \nabla c_s) + S, \tag{1}$$

$$\frac{\partial (Vc_s)}{\partial t} + \nabla \cdot (V\vec{v}c_s) = S, \tag{2}$$

where V is the volume of the computational element/cell $[L^3]$, c_s is the concentration of a given dissolved substance $[ML^{-3}]$; $\vec{u} = (u_x, u_y, u_z)$ is the velocity in the x, y, and z directions $[LT^{-1}]$; **E** is the (diagonal) diffusivity tensor $[L^2T^{-1}]$; and S is a source term $[MT^{-1}]$ that is linked to the chemistry modules. The **E** term accounts for the combined effect of different mixing phenomena:

$$\mathbf{E} = \mathbf{E}^{turb} + \mathbf{E}^{tsd} + \mathbf{E}^{sgt} + \mathbf{E}^{d},\tag{3}$$

where \mathbf{E}^{turb} is the turbulent or eddy diffusivity, which is a complex phenomenon with multi-359 fractal behaviour dominated by friction forces, \mathbf{E}^{tsd} is the so-called Taylor shear dispersion 360 that arises from the unresolved vertical variation of the horizontal flow, \mathbf{E}^{sgt} accounts for 361 the sub-grid eddy viscosity and diffusivity arising from unresolved mixing occurring at sub-362 grid scale, and \mathbf{E}^d is the background molecular diffusion resulting from the probabilistic 363 Brownian motion concept occurring at particle scales. E is approximated to the dominant 364 turbulent dispersion based on the eddy viscosity concept: $\mathbf{E}^{turb} = \boldsymbol{\nu}_t / \sigma$, where σ is the 365 Prandtl-Schmidt number. Turbulent viscosity (ν_t) depends on shear velocity (\vec{u}^*) and on 366 a turbulent length scale $(l_t = 0.07h)$. In this model, this relationship is approximated by 367 an algebraic expression: $\nu_t \sim k \vec{u}^* l_t$, where k is a user-defined scaling factor to account for 368 sub-grid scale eddies (Costa et al., 2016). 369

3.4.2 Biogeochemistry module

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Biogeochemical cycling frameworks are characterized in the appropriate biogeochem-371 istry JSON file. Each cycling framework is given a "tag" that is used to load it into simu-372 lations. The characterization of each of these cycling frameworks involves the definition of 373 the associated biogeochemical transformations and the chemical species involved. In turn, 374 the characterization of each of these transformations requires information about the chem-375 ical species consumed and produced and the reaction kinetics. It allows for the creation of 376 reaction parameters that can be loaded into the kinetics solver, which relies on the com-377 prehensive C++ Mathematical Expression Toolkit Library (*ExprTk*) developed by Arash 378 Partow (1999–2020) (Partow, 1999). The implementation of ExpTk in OpenWQ is simple 379 to use and provides an efficient run-time mathematical expression parser and evaluation 380 engine. ExprTk supports numerous forms of functional, logical, and vector processing se-381 mantics and is easily extendible. The equations can be written with (1) multiple chemical 382 species, (2) user-defined parameters, and (3) built-in hydro-model dependencies. These 383 model dependencies are tailored to each hydro-model, but they usually include variables like soil moisture and air and soil temperature. 385

The biogeochemistry JSON file can be prepared manually following the appropriate 386 OpenWQ JSON structure (i.e., key-value pairs). Alternatively, the cycling frameworks can 387 be characterized via a diagram drawn using a *GraphML* editor. *GraphML* is an XML-based 388 file format for graphs, and there are several free GraphML editors, such as yEd. After 389 the diagrams have been drawn in GraphML format, they can be converted into OpenWQ's 390 JSON format using a Python script available on OpenWQ's GitHub repository and copied 391 locally during the coupling step defined in Section 3.2.1. This graphical option has been 392 developed to enable collaboration and co-creation of water quality models through a more 393 visual interaction, which can be particularly helpful for water quality modeling activities 394 that often involve interaction between different disciplines that may be less familiar with 395 JSON files. 396

The expressions used to represent the different reaction kinetics in the model via the biogeochemistry JSON files can take many forms and are deployed in OpenWQ via the ExprTk integration. However, these expressions often take the form of sequences of reaction networks involving single or multiple chemical species governed by first-, second-, or third-order kinetics (Eq. 4, Eq. 5, Eq. 6), respectively:

$$\frac{dc_A}{dt} = -k\lambda c_A,\tag{4}$$

$$\frac{dc_A}{dt} = -k\lambda c_A^2,\tag{5}$$

$$\frac{dc_A}{dt} = -k\lambda c_A^2 c_B,\tag{6}$$

where c_A and c_B [ML⁻³] are the concentrations of chemical species A and B, parameter/variable λ represents weather/hydrological dependencies (such as soil moisture and temperature), and k is the reaction rate [ML⁻³T⁻¹]. The reaction rate k can be provided as the reaction rate using standard maximum at a reference temperature (often 20°C) or using expressions that can include relationships with the hydrological/weather dependent variables/parameters.

3.5 Eco-modelling lab and Cross-model deplyoment: Benefits and Innova tion

OpenWQ's concept as a coupler and customizable biogeochemistry modeling frame-405 work allows for cross-model deployment of eco-lab, co-creation modeling capabilities, pro-406 viding a long-awaited transformative direction for innovation in water quality modeling that 407 tackles the inherent challenges of being at the intersection between several scientific fields, 408 including biogeochemistry, soil science, hydrology, hydrodynamics, and hydrogeology. This 100 deployment is an essential step to enhancing collaborative efforts and streamlining knowl-410 edge/innovation transfer between the different disciplines involved, ultimately benefiting 411 the entire environmental and ecological research and management community. The ability 412 to deploy OpenWQ across different hydro-models provides the following transformational 413 changes in the current paradigm of water quality modeling. 414

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 1. Optimization of investment because focus is given to a single biogeochemical tool,
 416 thus reducing code maintenance,
- Maximization of benefit because it allows (a) any existing hydro-tool to extend its capabilities to water quality and ecological studies, (b) progress achieved by the biogeochemistry and soil-science research community to be transferred into OpenWQ and automatically benefits multiple hydro-models via update installs, (c) OpenWQ's input files to be transferred across hydro-models, providing experience and knowledge transfer between environmental projects and research communities, and
 - 3. Reproducibility is reinforced because OpenWQ's input files are transferable across hydro-models, allowing for more rigorous cross-model comparisons.

⁴²⁵ Once different biogeochemical modeling hypotheses have been identified and set up ⁴²⁶ for testing in one hydro-model-OpenWQ coupled model, they can be easily transferred for ⁴²⁷ simulation in any other hydro-amodel that has been coupled to OpenWQ (Fig. 8).

$_{428}$ 4 Discussion

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4.1 Unifying Different Approaches to Water Quality Modelling

430 OpenWQ provides a unifying modeling framework to deploy different approaches to 431 physiochemically based soil and water quality modeling across existing hydrological, hydro-

Figure 8: Cross-model deployment of biogeochemistry modelling setup, hypothesis and scenários

dynamic, and hydrogeological models. Here, we emphasize how OpenWQ can help unify different modeling approaches to simulate different landscapes and pollution problems.

1. Sub-catchment versus catchment-scale modeling. OpenWQ provides a unique oppor-434 tunity to harmonize spatial scales in water quality modeling. Water quality models 435 usually belong to one of two scales regarding the spatial domain, (a) sub-catchment 436 and (b) catchment. Sub-catchment water quality models, including river-reach mod-437 els such as WASP (Wool et al., 2020; Di Toro et al., 1983) and QUAL2E (Brown 438 & Barnwell, 1987), lake models such as MyLake (Saloranta & Andersen, 2007) and 439 Delft3D (Lesser et al., 2004), and aquifer models such as MODFLOW-MT3D (Har-440 baugh, 2005; Bedekar et al., 2016) and FEFLOW (Trefry & Muffels, 2007), require 441 the characterization of both horizontal and vertical boundary conditions because they 442 do not represent entire closed systems such a river basin. Instead, they focus on sub-443 regions within such closed systems. On the other hand, catchment models such as 444 INCA (Jackson-Blake et al., 2016; A. J. Wade et al., 2002; Whitehead et al., 1998) 445 and SWAT (Arnold et al., 2012) only require vertical boundary conditions (e.g., pre-446 cipitation) because they focus on the larger closed system that contains the entire 447 basin area. Accordingly, these two scales of models tend to focus on different pol-448 lution problems. Sub-catchment models tend to focus more on point sources, such 449 as wastewater discharge, and catchment models often look at diffuse pollution (e.g., 450 agriculture nutrients and fertilizer use). These differences also result in different 451 chemical species and biogeochemical cycles of focus, e.g., river models often address 452 problems related to low Dissolved Oxygen (DO) levels caused by biomass decomposi-453 tion and BOD (Biological Oxygen Demand) arising from wastewater discharges and 454 other point sources discharging directly in river and lakes. 455

OpenWQ enables multi-scale chemistry simulations that can be integrated into 456 hydro-models of sub-catchment and catchment scales; thus, it can help bridge 457 the gap between these two approaches, as well as allow for better integration of 458 cross-dependent biogeochemical cycles. For example, whereas the DO and nitrogen 459 cycles are often simulated in river-reach models, catchment models tend to focus 460 more on the nitrogen cycle and assume that there is an unlimited amount of DO avail-461 able. Although this assumption could be considered valid in many cases for surface 462 runoff, it is certainly erroneous for water bodies such as rivers, lakes, and groundwater. 463

- 2. Simple versus detailed biogeochemistry representation. Water quality models have 465 been developed with varying degrees of detail in the representation of biogeochemical 466 processes (Costa, Baulch, et al., 2020a). For example, while INCA and HYPE 467 (Lindström et al., 2010) provide simpler biogeochemical modeling approaches for 468 the nitrogen and phosphorous cycles, requiring a smaller number of reaction-kinetic 469 parameters to calibrate them may be more suitable for data-scarce regions; more 470 complex biogeochemical models like HSPF ("Hydrological Simulation Program-471 Fortran, User's manual for version 11: U.S. Environmental Protection Agency" 472 1997) arguably provide higher model fidelity but may only be applicable in data-rich 473 environments. However, when selecting a model for a particular region, modelers 474 often have to make compromises with process representation because some models 475 have more detailed physics-based coupled water-energy balance computations such 476 as AnnAGNPS (Bosch et al., 1998) but may offer more limited biogeochemical 477 capabilities (Costa, Baulch, et al., 2020a). OpenWQ enables addressing this issue 478 through its flexible eco-modelling lab, which allows testing different biogeochemical 479 conceptual models and modelling hypotheses, from simple biogeochemical cycles and 480 transformations to more complex and intertwined reaction networks involving dozens 481 or hundreds of chemical species. 482
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3. Unconnected versus interconnected biogeochemical cycling representation. Most catchment models simulate biogeochemical cycles in isolation. For example, popular

models such as SWAT and HYPE simulate the phosphorous and nitrogen cycles 486 without integrating them with the DO-BOD cycle, despite them being strongly 487 interrelated because DO is used in both cycles. In the nitrogen cycle, DO is used 488 in nitrification, where ammonia is oxidated into nitrite (NO_3) and nitrate (NO_2) . 489 In the DO-BOD cycle, DO is used in the biological, aerobic decomposition of 490 organic matter. OpenWQ aims to allow for a biogeochemistry representation that 491 is less compartmentalized, making interactions between cycling frameworks more 492 fluid and closer to reality (i.e., model fidelity). In reality, DO dynamics affect the 493 cycling of many chemical constituents, from nutrients to heavy metals. OpenWQ's 494 eco-modelling lab addresses this issue through its open reaction-network solver, 495 which provides the flexibility to deploy any number of biogeochemical cycling 496 representations of any number of chemical species simultaneously, which can be 497 connected or unconnected. 498

- 4. Background chemical transport driver. Typically, hydro-models are developed and 500 maintained by research communities with specific research motivations driven by a 501 regional context. These modeling tools often become highly specialized for particular 502 environments and applications, and modelers commonly find it hard to find a water 503 quality modeling tool that integrates such regionally important hydro-transport mod-504 eling capabilities. There is also in-house expertise in such communities and research 505 groups that is passed on over the years between elements of the research community, 506 which can make them reluctant to switch to other modeling tools because it may 507 involve steep learning curves. OpenWQ aims to address this issue by enabling exten-508 sion to water quality modeling capabilities directly on those models so that modelers 509 can continue using the hydro-modeling tool that they consider more suitable to the 510 environments on which they focus. 511
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4.2 Advancing Current Model Development Paradigms

Our focus on providing a flexible biogeochemical model representation framework that 513 is deployable on any process-based hydro-models and offers a unique opportunity to trans-514 form the way environmental studies involving water quality modeling are carried out. It 515 also provides a concrete way for streamlining collaboration, co-creation, and knowledge 516 and innovation transfer between the hydrological and biogeochemistry/soil science com-517 518 munities. Such streamlining is materialized by OpenWQ's coupler interface, which links developments/progress in the hydro-modeling tool (hydro-modeling community) with de-519 velopments/progress in OpenwQ (biogeochemistry communities). Another major issue with the current water quality modeling paradigm is that most models offer little flexibility to 521 adapt the representation of biogeochemistry processes to (1) local/region context, (2) data 522 availability, and (3) application(s) of interest, resulting in the need to make difficult choices 523 when selecting a modeling tool because compromises between the representation of hydro-524 fluxes and biogeochemistry are often required (Costa, Baulch, et al., 2020a). 525

526 4.3 Limitations

In OpenWQ, the physical transport of chemical constituents relies on the quality of the water fluxes passed by the host hydro-model. Although this may be an advantage because it allows the user to choose the hydro-model that best suits the modeling needs, it can become an issue if the calculated fluxes in the hydro-model are not accurate. Modelers should select the hydro-model coupled to OpenWQ that is the most suitable for the application at hand or pursue the coupling of OpenWQ to another hydro-model.

The native biogeochemistry module of OpenWQ that provides water quality-lab capabilities assumes that biogeochemical cycling can be represented via a series of sequential and parallel reaction networks. Although this representation is true for most chemical constituents, pollution problems, and environmental studies, some biogeochemistry may involve

formulations that rely on variable dependencies that may not be available in a particular cou-537 pled model. For example, simulating microbiological pollution in lakes, rivers, and beaches, 538 such as contamination with fecal coliforms originating from leaking septic tanks and wastew-539 ater discharges, requires the simulation of die-off rates due to exposure to solar radiation. 540 In the case of hydrodynamic models, such information may not be available. In the case of 541 hydrological models, which typically deal with such data, that variable dependency may not 542 have been passed into OpenWQ during the development of the coupler interface, so updates 543 to the coupler may be needed. OpenWQ's coupler functions have been designed to make 544 the addition of new dependency variables straightforward. 545

The portability of OpenWQ is materialized through a series of coupler functions and 546 wrapper interface functions that enable its coupling to hydro-models written in C++ or 547 Fortran. However, the use of a more standardized model interface framework, such as 548 the Basic Model Interface (BMI Peckham et al., 2013; E. W. Hutton et al., 2020), could 549 help streamline further the coupling process through the use of standard control and query 550 functions. This could make that model both easier to learn and easier to couple with other 551 software elements. BMI, for example, currently supports five languages: C, C++, Fortran, 552 Java, and Python. 553

554 5 Conclusions

This paper describes a unified framework for enabling multi-biogeochemical modeling capabilities in existing hydro-models. The framework, which we call Open Water Quality (OpenWQ), was designed with both coupler and water quality-lab modeling capabilities to enable (1) flexible co-creation and testing of biogeochemistry modeling representations, (2) systematic implementation and evaluation of alternative modeling approaches for process representation, and (3) identification of specific causes of model weaknesses.

This work arises from the recognition that the hydrological transport of contaminants 561 in the environment strongly affects their concentrations in aquatic ecosystems, but the inter-562 action between the biogeochemical and hydrological-hydrodynamic-hydrogeological commu-563 nities lacks the appropriate mechanisms for an efficient transfer of knowledge and innovation 564 between the two communities. Many hydro-models are developed and maintained by spe-565 cific research communities addressing particular regions and climate zones. The investment 566 in such tools creates invaluable in-house expertise, making them less likely to switch to 567 other modeling tools and hindering their expansion to water quality and ecological studies. 568 OpenWQ aims to address this problem and provide a concrete direction for innovation in 569 570 connecting communities through an optimized plug-in-like water quality model that can be coupled to existing hydro-modeling tools, extending the value of these tools for water 571 quality and ecological studies in their region of focus. Different process representations and 572 different spatial configurations can be integrated into the structural model core, which en-573 ables users to decompose the modeling problem into the individual decisions made as part 574 of model development and evaluate different "fine-grained" model development decisions in 575 a systematic and controlled way. 576

OpenWQ can provide the necessary model flexibility to progress toward answering the 577 following fundamental modeling questions and challenges: (1) which hydro-biogeochemical 578 processes should be represented explicitly in different environmental settings, and, corre-579 spondingly, which processes can be ignored or greatly simplified; (2) what modeling ap-580 proaches should be used to represent the dominant biogeochemistry at the spatial scale 581 of the model discretization; (3) how should heterogeneity in pollution and biogeochemical 582 processes be represented across spatial scales, including the complexity of transport across 583 landscapes; and (4) how can we provide insights into the sources of model uncertainty. The 584 companion paper describes the integration of OpenWQ into two hydro-models, SUMMA and 585 CRHM, describing how coupling interfaces between the two models have not only enabled 586 water quality simulation capacities in these host hydro-models but, even more importantly, 587

established a direct and permanent link for the transfer of innovation between the associated modeling communities, promoting cooperation and co-creation. Example applications of pollution studies enabled by the coupling of the tools are also provided to begin to address some of these fundamental modeling challenges.

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598 **References**

- Arheimer, B., Dahné, J., Donnelly, C., Lindström, G., & Strömqvist, J. (2012). Water
 and nutrient simulations using the HYPE model for Sweden vs. the Baltic Sea basin
 influence of input-data quality and scale. *Hydrology Research*, 43(4), 315-329. doi:
 10.2166/nh.2012.010
- Arnold, J. G., Moriasi, D. N., Gassman, P. W., Abbaspour, K. C., White, M. J., Srinivasan,
 R., ... Jha, M. K. (2012). SWAT: Model Use, Calibration, and Validation. *Trans.*ASABE, 55(4), 1491–1508.
- Arnold, J. G., Srinivasan, R., Muttiah, R. S., & Williams, J. R. (1998). Large area hydrologic
 modeling and assessment Part I: Model Development. JAWRA Journal of the American
 Water Resources Association, 34(1), 73-89. doi: https://doi.org/10.1111/j.1752-1688
 .1998.tb05961.x
- Beck, M. B. (1987). Water quality modeling: A review of the analysis of uncertainty. *Water Resources Research*, 23(8), 1393–1442. doi: 10.1029/WR023i008p01393
- Bedekar, V., Morway, E. D., Langevin, C. D., & Tonkin, M. J. (2016). MT3D-USGS version
 1: A U.S. Geological Survey release of MT3DMS updated with new and expanded transport capabilities for use with MODFLOW (Tech. Rep.). Reston, VA. doi: 10.3133/tm6A53
- Bosch, D., Theurer, F., Bingner, R., & Felton, G. (1998). Evaluation of the AnnAGNPS
 water quality model. Agricultural Non-Point Source Water Quality Models: Their Use
 and Application; John, EP, Daniel, LT, Rodney, LH, Eds, 45–54.
- Brown, L., & Barnwell, T. (1987). The enhanced stream water quality models qual2e and qual2e-uncas: documentation and user manual. *Environmental Protection Agency*.
- Clark, M. P., & Kavetski, D. (2010). Ancient numerical daemons of conceptual hydrological
 modeling: 1. fidelity and efficiency of time stepping schemes. Water Resources Research,
 46(10). doi: https://doi.org/10.1029/2009WR008894
- Clark, M. P., Kavetski, D., & Fenicia, F. (2011). Pursuing the method of multiple working
 hypotheses for hydrological modeling. *Water Resources Research*, 47(9). doi: https://
 doi.org/10.1029/2010WR009827
- Clark, M. P., Nijssen, B., Lundquist, J. D., Kavetski, D., Rupp, D. E., Woods, R. A., ...
 Rasmussen, R. M. (2015a). A unified approach for process-based hydrologic modeling: 1.
 modeling concept. *Water Resources Research*, 51 (4), 2498-2514. Retrieved from https://
 agupubs.onlinelibrary.wiley.com/doi/abs/10.1002/2015WR017198
 doi.org/10.1002/2015WR017198
- Clark, M. P., Nijssen, B., Lundquist, J. D., Kavetski, D., Rupp, D. E., Woods, R. A.,
 Marks, D. G. (2015b). A unified approach for process-based hydrologic model ing: 2. model implementation and case studies. Water Resources Research, 51(4), 2515 2542. Retrieved from https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1002/
 2015WR017200 doi: https://doi.org/10.1002/2015WR017200
- Costa, D., Baulch, H., Elliott, J., Pomeroy, J., & Wheater, H. (2020a). Modelling nutrient
 dynamics in cold agricultural catchments: A review. *Environmental Modelling & Software*,

- ⁶³⁸ *124*, 104586. doi: https://doi.org/10.1016/j.envsoft.2019.104586
- Costa, D., Baulch, H., Elliott, J., Pomeroy, J., & Wheater, H. (2020b). Modelling nutrient
 dynamics in cold agricultural catchments: A review. *Environmental Modelling & Software*,

124, 104586. doi: https://doi.org/10.1016/j.envsoft.2019.104586

641

- Costa, D., Burlando, P., & Liong, S.-Y. (2016). Coupling spatially distributed river and
 groundwater transport models to investigate contaminant dynamics at river corridor
 scales. *Environmental Modelling & Software*, 86, 91-110. doi: https://doi.org/10.1016/
 j.envsoft.2016.09.009
- Costa, D., Pomeroy, J., & Wheater, H. (2018). A numerical model for the simulation of snowpack solute dynamics to capture runoff ionic pulses during snowmelt: The
 pulse model. Advances in Water Resources, 122, 37-48. doi: https://doi.org/10.1016/
 j.advwatres.2018.09.008
- Costa, D., & Pomeroy, J. W. (2019). Preferential meltwater flowpaths as a driver of prefer ential elution of chemicals from melting snowpacks. Science of The Total Environment,
 662, 110–120. doi: 10.1016/J.SCITOTENV.2019.01.091
- Costa, D., Pomeroy, J. W., Brown, T., Baulch, H., Elliott, J., & Macrae, M. (2021).
 Advances in the simulation of nutrient dynamics in cold climate agricultural basins:
 Developing new nitrogen and phosphorus modules for the Cold Regions Hydrological
 Modelling Platform. Journal of Hydrology, 603, 126901. doi: https://doi.org/10.1016/
 j.jhydrol.2021.126901
- ⁶⁵⁸ Costa, D., Roste, J., Pomeroy, J., Baulch, H., Elliott, J., Wheater, H., & Westbrook, C.
 (2017). A modelling framework to simulate field-scale nitrate response and transport during snowmelt: The wintra model. *Hydrological Processes*, 31(24), 4250-4268. doi: https://doi.org/10.1002/hyp.11346
- Costa, D., Shook, K., Spence, C., Elliott, J., Baulch, H., Wilson, H., & Pomeroy, J. W.
 (2020). Predicting variable contributing areas, hydrological connectivity, and solute
 transport pathways for a canadian prairie basin. Water Resources Research, 56(12),
 e2020WR027984. doi: https://doi.org/10.1029/2020WR027984
- Davies, T. D., Brimblecombe, P., Tranter, M., Tsiouris, S., Vincent, C. E., Abrahams,
 P., & Blackwood, I. L. (1987). *The Removal of Soluble Ions from Melting Snowpacks*(H. G. Jones & W. J. Orville-Thomas, Eds.). Dordrecht: Springer Netherlands. doi:
 10.1007/978-94-009-3947-9_20
- Di Toro, D. M., Fitzpatrick, J. J., Thomann, R. V., & Hydroscience, I. (1983). Documentation For Water Quality Analysis Simulation Program (WASP) And Model Verification Program (MVP) (Tech. Rep.).
- Frank, E. G., & Schlenker, W. (2016). Balancing economic and ecological goals. *Science*, 353(6300), 651-652. doi: 10.1126/science.aaf9697
- Fu, B., Merritt, W. S., Croke, B. F., Weber, T. R., & Jakeman, A. J. (2019). A review of catchment-scale water quality and erosion models and a synthesis of future prospects. *Environmental Modelling & Software*, 114, 75–97. doi: 10.1016/J.ENVSOFT.2018.12
 .008
- Harbaugh, A. W. (2005). MODFLOW-2005, the US Geological Survey modular groundwater model: the ground-water flow process (Vol. 6). US Department of the Interior, US
 Geological Survey Reston, VA, USA.
- Hutton, C., Wagener, T., Freer, J., Han, D., Duffy, C., & Arheimer, B. (2016). Most
 computational hydrology is not reproducible, so is it really science? Water Resources *Research*, 52(10), 7548–7555. doi: https://doi.org/10.1002/2016WR019285
- Hutton, E. W., Piper, M. D., & Tucker, G. E. (2020). The basic model interface 2.0: A
 standard interface for coupling numerical models in the geosciences. Journal of Open
 Source Software, 5(51), 2317. doi: 10.21105/joss.02317
- Hydrological Simulation Program-Fortran, User's manual for version 11: U.S. Environmental Protection Agency. (1997). National Exposure Research Laboratory, Athens, Ga., EPA/600/R-97/080, 755 p.
- Jackson-Blake, L. A., Wade, A. J., Futter, M. N., Butterfield, D., Couture, R.-M., Cox,

B. A., ... Whitehead, P. G. (2016). The INtegrated CAtchment model of phosphorus dynamics (INCA-P): Description and demonstration of new model structure
and equations. *Environmental Modelling & Software*, 83, 356–386. doi: 10.1016/
J.ENVSOFT.2016.05.022

- ⁶⁹⁶ Jones, H. G. (1999). The ecology of snow-covered systems: a brief overview of nutrient ⁶⁹⁷ cycling and life in the cold. *Hydrological Processes*, 13(14-15), 2135–2147. doi: 10.1002/ ⁶⁹⁸ (SICI)1099-1085(199910)13:14/15(2135::AID-HYP862)3.0.CO;2-Y
- Keith, D. A., Ferrer-Paris, J., Nicholson, E., Bishop, M. J., Polidoro, B. A., Ramirez-Llodra,
 E., ... Kingsford, R. T. (2022). A function-based typology for earth's ecosystems. *Nature*,
 610(7932), 513–518. doi: 10.1038/s41586-022-05318-4
- ⁷⁰² Kitware. (2022). *CMake: Reference documentation*.
- Lesser, G., Roelvink, J., van Kester, J., & Stelling, G. (2004). Development and validation of
 a three-dimensional morphological model. *Coastal Engineering*, 51(8), 883-915. (Coastal
 Morphodynamic Modeling) doi: https://doi.org/10.1016/j.coastaleng.2004.07.014
- Lilbaek, G., & Pomeroy, J. W. (2007). Modelling enhanced infiltration of snowmelt ions into frozen soil. *Hydrological Processes*, 21(19), 2641–2649. doi: 10.1002/hyp.6788
- Lindström, G., Pers, C., Rosberg, J., Strömqvist, J., & Arheimer, B. (2010). Development and testing of the HYPE (Hydrological Predictions for the Environment) water quality model for different spatial scales. *Hydrology Research*, 41 (3-4), 295. doi: 10.2166/nh.2010
 .007
- Marsh, P., & Pomeroy, J. W. (1999). Spatial and temporal variations in snowmelt runoff
 chemistry, Northwest Territories, Canada. Water Resources Research, 35(5), 1559–1567.
 doi: 10.1029/1998WR900109
- Mekonnen, B. A. (2016). Modeling and management of water quantity and quality
 in cold-climate Prairie Watersheds (Unpublished doctoral dissertation). University of
 Saskatchewan.
- Moore, D. S., Bingner, R. L., & Theurer, F. D. (2006). ANNAGNPS: Accounting for
 Snowpack, Snowmelt, and Soil Freeze-Thaw. In *Eighth federal interagency sedimentation conference (8thfisc)*, (pp. 475–482). Reno, NV, USA.
- Partow, A. (1999). ExprTk C++ Mathematical Expression Library.
- Peckham, S. D., Hutton, E. W., & Norris, B. (2013). A component-based approach to
 integrated modeling in the geosciences: The design of csdms. *Computers & Geosciences*,
 53, 3–12.
- Pomeroy, J., Brown, T., Fang, X., Shook, K., Pradhananga, D., Armstrong, R., ... Lopez
 Moreno, J. (2022). The cold regions hydrological modelling platform for hydrological diagnosis and prediction based on process understanding. *Journal of Hydrology*, 615, 128711. doi: https://doi.org/10.1016/j.jhydrol.2022.128711
- Pomeroy, J., Davies, T., Jones, H., & Marsh, P. (1999). Transformations of snow chemistry in the boreal forest: accumulation and volatilization. *Hydrological Processes*, 13(1415), 2257–2273.
- Pomeroy, J. W., Davies, T. D., & Tranter, M. (1991). The Impact of Blowing Snow on
 Snow Chemistry. In Seasonal snowpacks: Processes of compositional change (pp. 71–113).
 Springer, Berlin, Heidelberg. doi: 10.1007/978-3-642-75112-7_4
- Pomeroy, J. W., Gray, D. M., Brown, T., Hedstrom, N. R., Quinton, W. L., Granger, R. J.,
 & Carey, S. K. (2007). The cold regions hydrological model: A platform for basing process
 representation and model structure on physical evidence. *Hydrological Processes*, 21(19),
 2650–2667. doi: {10.1002/hyp.6787}
- Refsgaard, J., & Storm, B. (1995). Mike she.[in:] singh vp (ed.), computer models of
 watershed hydrology. *Water Resources Publication, Colorado*, 809–847.
- Richmond, B. (2003). STELLA: An Introduction to Systems Thinking.
- Saloranta, T. M., & Andersen, T. (2007). Mylake—a multi-year lake simulation model code
- suitable for uncertainty and sensitivity analysis simulations. *Ecological Modelling*, 207(1),
 45-60. (Uncertainty in Ecological Models) doi: https://doi.org/10.1016/j.ecolmodel.2007
- .03.018

- Sanderson, C., & Curtin, R. (2016). Armadillo: a template-based c++ library for linear
 algebra. Journal of Open Source Software, 1(2), 26. doi: 10.21105/joss.00026
- Sanderson, C., & Curtin, R. (2018). A user-friendly hybrid sparse matrix class in c++. , 422–430. doi: $10.1007/978-3-319-96418-8_50$
- Shoemaker, L. (1997). Compendium of tools for watershed assessment and tmdl development.
- Shook, K., Pomeroy, J. W., Spence, C., & Boychuk, L. (2013). Storage dynamics simulations in prairie wetland hydrology models: evaluation and parameterization. *Hydrological Processes*, 27(13), 1875–1889. doi: 10.1002/hyp.9867
- Trefry, M. G., & Muffels, C. (2007). Feflow: A finite-element ground water flow and transport modeling tool. *Groundwater*, 45(5), 525-528. doi: https://doi.org/10.1111/ j.1745-6584.2007.00358.x
- Wade, A., Jackson, B., & Butterfield, D. (2008). Over-parameterised, uncertain 'mathematical marionettes' How can we best use catchment water quality models? An example of
 an 80-year catchment-scale nutrient balance. Science of The Total Environment, 400(1-3),
 52–74. doi: 10.1016/j.scitotenv.2008.04.030
- Wade, A. J., Durand, P., Beaujouan, V., Wessel, W. W., Raat, K. J., Whitehead, P. G.,
 ... Lepisto, A. (2002). A nitrogen model for european catchments: Inca, new model
 structure and equations. *Hydrology and Earth System Sciences*, 6(3), 559–582. doi: 10.5194/hess-6-559-2002
- Wellen, C., Kamran-Disfani, A. R., & Arhonditsis, G. B. (2015). Evaluation of the current state of distributed watershed nutrient water quality modeling. *Environmental Science and Technology*, 49(6), 3278–3290. doi: 10.1021/es5049557
- Whitehead, P., Wilson, E., Butterfield, D., & Seed, K. (1998). A semi-distributed integrated flow and nitrogen model for multiple source assessment in catchments (inca): Part ii application to large river basins in south wales and eastern england. Science of The Total Environment, 210-211, 559-583. doi: https://doi.org/10.1016/S0048-9697(98)00038-2
- Wool, T., Ambrose, R. B., Martin, J. L., & Comer, A. (2020). Wasp 8: The next generation
 in the 50-year evolution of usepa's water quality model. *Water*, 12(5). doi: 10.3390/
 w12051398
- Yang, X., Tetzlaff, D., Soulsby, C., & Borchardt, D. (2022). Hiwaq v1.0: A flexible catchment water quality assessment tool with compatibility for multiple hydrological model structures. *Geoscientific Model Development Discussions*, 2022, 1–26. doi: 10.5194/gmd-2022-239

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