Long-term trends in productivity across Intermountain West lakes provide no evidence of widespread eutrophication

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Abstract

Eutrophication represents a major threat to freshwater systems and climate change is expected to drive further increases in freshwater primary productivity. However, long-term in-situ data is available for very few lakes and makes identifying trends and drivers of eutrophication challenging. Using remote sensing data, we conducted a retrospective analysis of long-term trends in trophic status across the Intermountain West, a region with understudied water quality trends and limited long-term datasets. We found that most lakes (55%) were not exhibiting shifts in trophic status from 1984-2019. Our results also show that increases in eutrophication were rare (3% of lakes) during this period, and that lakes exhibiting negative trends in trophic status were more common (17% of lakes). Lakes that were not trending occupied a wide range of lake and landscape characteristics, whereas lakes that were becoming less eutrophic tended to be in more heavily developed catchments. Our results highlight that while there are well-established narratives that climate change can lead to more eutrophication of lakes, this is not broadly observed in our dataset, with more lakes becoming more oligotrophic than lakes becoming eutrophic.

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1 2 3 4	Long-term trends in productivity across Intermountain West lakes provide no evidence of widespread eutrophication
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11	
12	Key Points:
13 14	• Remote sensing imagery captures long-term trends in lake productivity across the Intermountain West
15 16	• The majority of lakes observed in this dataset were not exhibiting shifts in trophic status from 1984-2019
17 18	• The incorporation of fine-scale lake climate data from new deep learning datasets results in substantial improvement to model accuracy
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32 Abstract

Eutrophication represents a major threat to freshwater systems and climate change is expected to 33 drive further increases in freshwater primary productivity. However, long-term in-situ data is 34 35 available for very few lakes and makes identifying trends and drivers of eutrophication challenging. Using remote sensing data, we conducted a retrospective analysis of long-term 36 trends in trophic status across the Intermountain West, a region with understudied water quality 37 trends and limited long-term datasets. We found that most lakes (55%) were not exhibiting shifts 38 in trophic status from 1984-2019. Our results also show that increases in eutrophication were 39 rare (3% of lakes) during this period, and that lakes exhibiting negative trends in trophic status 40 were more common (17% of lakes). Lakes that were not trending occupied a wide range of lake 41 and landscape characteristics, whereas lakes that were becoming less eutrophic tended to be in 42 43 more heavily developed catchments. Our results highlight that while there are well-established narratives that climate change can lead to more eutrophication of lakes, this is not broadly 44 observed in our dataset, with more lakes becoming more oligotrophic than lakes becoming 45 eutrophic. 46

47 Plain Language Summary

Lakes are often classified by their productivity. Low productive lakes generally represent deep 48 lakes with low amounts of algae. Whereas lakes with high levels of productivity support more 49 50 plant growth and have higher amounts of algae. The accumulation of nutrients in freshwater systems often results in increases in productivity and can lead to the development of algal 51 blooms. Algal blooms are a major concern due to their threat to ecosystem health, recreation, and 52 53 drinking water sources. Yet the lack of long-term field data across large scales has resulted in a limited understanding of 1) what factors are driving productivity trends and the development of 54 algal blooms across regions, and 2) are increasing trends representative of widespread 55

intensification or an increase in awareness and reporting. Therefore, there is a pressing need to
effectively monitor and understand these trends in order to inform management actions that
address their frequency and intensity. Here, we use data obtained from satellite imagery from
1984 - 2019 to document lake productivity trends in 1,169 lakes across the Intermountain West.
We show that substantial increases in productivity were rare, and that the majority of lakes have
not undergone widespread change.

62 **1 Introduction**

Widespread eutrophication is a global phenomenon that threatens water quality, 63 recreational industries, and ecosystem function (Paerl et al., 2001; Gatz, 2020; Amorim and 64 Moura, 2021). A common outcome of eutrophication is an increase in the biomass of 65 phytoplankton, both algae and cyanobacteria, in freshwater, transitional, and ocean environments 66 (Anderson et al., 2008; Hudnell, 2010; Wurtsbaugh et al, 2019). In many cases, this rapid and 67 excessive growth can become severe and lead to the development of Harmful Algal Blooms 68 (HABs) (Smith, 2003; Heisler et al. 2008). HABs are of particular concern due to the threats they 69 70 pose to human health and drinking water sources (Fleming et al., 2002; Falconer and Humpage et al., 2005; Christensen and Khan, 2020). Thus, the wide-ranging effects that eutrophication and 71 HABs have on aquatic systems and their threat to human health have highlighted the need to 72 understand the factors which drive them. 73

Generally, eutrophication and algal blooms are attributed to excessive loading of
nitrogen (N) and phosphorus (P) as well as high water temperatures (Rejmankova and
Komarkova, 2005; Paerl and Paul, 2012; Gobbler et al. 2016; Beaver et al. 2018). However, in
shallow lakes, warmer temperatures and higher light absorption have been found to be more
significant drivers of productivity (Kosten et al., 2012). In other words, the combination of

factors that drive rapid increases in lake productivity may differ between individual water bodies or geographic regions, hence smaller and more focused state and regional scale studies may be more useful in describing changes in lake productivity dynamics (Oleksy et al., 2022).

Large scale studies have highlighted that water quality trends are context dependent and 82 vary across regions (Beaver et al., 2018). However, some regions with unique landscape features 83 remain understudied regarding lake productivity trends. For example, the Intermountain West 84 region (including the US states Colorado, Idaho, Montana, Utah, and Wyoming) has very 85 different hydrological dynamics and landscape features compared with other regions, yet water 86 quality trends remain mostly undocumented. The region undergoes quick wet-dry seasonal 87 88 transitions, with most of the streamflow generated by snowmelt (Bales et al., 2006). Higher 89 gradients in temperature and precipitation with elevation make hydrologic processes significantly different compared with low-elevation regions (Bales et al., 2006). Land use in this region also 90 differs, with substantial amounts of grassland pasture and range contributing to increased organic 91 nutrient loading to streams and rivers (Agouridis et al., 2005). 92

93 An increase in awareness and reporting of HABs in the Intermountain West suggests that 94 lakes in the region may be becoming more eutrophic, yet our understanding of lake productivity 95 trends is very limited. As nation-wide research and understanding of HABs has grown, so have management and sampling plans, educational materials, and overall public awareness (Hudnell et 96 97 al. 2010). However, this increase in awareness and reporting has the potential to create a 98 perception that blooms are already increasing in intensity and frequency (Hallegraeff et al., 2021). Recent work in the region highlights that lakes are experiencing roughly equal trends of 99 changing from blue to green or changing green to blue, indicating there is not overwhelming 100 101 evidence that they are getting more eutrophic, where eutrophic lakes are generally more green

102 (Oleksy et al., 2022). It remains unclear whether this is a result of representative increases in

103 intensity or a result of heightened monitoring. Therefore, retrospective data analyses and long-

104 term monitoring are needed to identify consistent productivity trends (Hudnell, 2008),

105 particularly in understudied regions like the Western US.

Remote sensing and long-term satellite imagery create opportunities to address key 106 research gaps surrounding what factors are driving freshwater productivity across regions. In-situ 107 sampling methods are often limited by resources such as time and funding. Therefore, in-situ 108 water quality data tends to be focused on relatively large lakes (> 20 ha) and long-term records 109 tend to be rare (Stanley et al. 2019). Importantly, leveraging remote sensing data can address 110 111 water quality dynamics over large spatial and temporal scales where in situ data is lacking (Topp et al. 2020). Remote sensing data with high spatial and temporal coverage are also useful to 112 understand how global change is affecting productivity and bloom dynamics (Harvey et al. 2015; 113 Ho et al., 2017; Seegers et al. 2021). These tools can be used to determine water quality 114 parameters in freshwater systems such as chlorophyll-a (Boucher et al., 2018; Kuhn et al., 2019; 115 Papenfus et al., 2020), suspended sediments (Pavelsky and Smith, 2009), and organic matter 116 (Kutser et al., 2005; Slonecker et al., 2016). 117

In this study, we address two gaps in our understanding of lake productivity dynamics in the Intermountain West. Specifically, we aimed to identify 1) the historical prevalence of eutrophic lakes and whether this is an increasing trend of eutrophication, and 2) the drivers and spatial distribution of changes in trophic state. We use remote sensing imagery and in-situ chlorophyll-a data, covering 1984-2019, to predict chlorophyll-a and lake trophic state based solely on satellite imagery. This approach allowed us to document productivity trends in 1,169 lakes over 35 years. By increasing the level of understanding of historical trends in lake 125 productivity and their drivers in this region, our analysis can also shed light on the intensification

of algal blooms in lakes and provide important information for water quality management.

- 127 2 Materials and Methods
- 128 2.1 Data Sources and Processing
- 129

Our analysis used various remote sensing, water quality, lake and landscape features, and 130 climate datasets. We opted for a machine-learning approach that uses paired satellite reflectance 131 from Landsat observations and in-situ water quality data. We acquired Landsat data and in-situ 132 chlorophyll-a samples for model training from the AquaSat dataset (Ross et al., 2019). AquaSat 133 joins Landsat Tier 1 surface reflectance to water quality samples from the Water Quality Portal 134 135 (Read et al. 2017) and LAGOS-NE (Soranno et al. 2017) that occurred ± 1 day of a Landsat observation. We filtered AquaSat to only include observations over the Intermountain West 136 region and with Landsat scenes with less than 50% cloud cover. The resulting dataset included 137 138 1,340 observations across 249 lakes in the region. Reflectance values across the three different Landsat satellites used (5, 7, and 8) were standardized using the methodology outlined in 139 140 Gardner et al. (2021). We then identified various open-source datasets that captured 141 environmental drivers we hypothesized might be important for predicting chlorophyll-a. We merged Lake characteristics and catchment level metrics to our training dataset from the LakeCat 142 143 (Hill et al., 2018) and LAGOS-US (Cheruvelil et al., 2021), and HydroLAKES (Messager at al., 2016) datasets. Initially we joined lakes in the training set to corresponding lake polygons 144 145 included in NHDPlusV2. LakeCat, LAGOS-US, and HydroLAKES datasets were then added through common NHD identifiers. We selected metrics that were derived from these datasets 146 based on their potential to impact water quality (Table S1). 147

Daily surface water temperature and corresponding weather data (wind speed) were also included in our model development. We extracted daily water temperature from Willard et al. (2022), which includes estimated daily surface water temperature for 185,549 lakes across the US. In addition to daily surface temperature, we calculated prior 14-day mean temperatures for all 1,340 observations included in our training set. Then, we joined 14-day mean temperature and meridional wind speed to our training set using common NHD identifiers and the date of observation.

Using the same methods, we built our prediction dataset using LimnoSat-US (Topp et al., 155 156 2021). LimnoSat-US includes Landsat Collection 1, Tier 1 surface reflectance for lakes greater than 10 hectares in the U.S. spanning 1984 – 2020. Surface reflectance values represent the 157 median surface reflectance of a 120-meter buffer of the "deepest point" of a lake polygon. This 158 "deepest point" can be defined as the center of the largest circle that can fit within a lake 159 polygon. We joined the lake characteristics, catchment level metrics, and climate data described 160 above to our prediction dataset, resulting in 1,264,355 observations across 2,596 lakes in the 161 Intermountain West. 162

Lastly, we defined categories for three trophic states based on the following chlorophyll-a 163 thresholds: oligotrophic (0 - 2.6 ug/L), mesotrophic (2.7 - 7 ug/L), and eutrophic (> 7 ug/L). 164 These thresholds were inspired by the criteria outlined in the National Lakes Assessment (U.S. 165 Environmental Protection Agency, 2009). This categorical approach was taken because 166 167 predicting chlorophyll-a concentrations in freshwater systems with remote sensing has been notably challenging, particularly with Landsat imagery (Salem et al., 2017; Smith et al. 2021). 168 Landsat bands are relatively broad with a low signal-to-noise ratio, often resulting in predictions 169 170 of chlorophyll-a with high levels of uncertainty (Matthews, 2011). Furthermore, the accurate

171	prediction of chlorophyll-a is affected by complex optical conditions in various waterbodies with
172	higher levels of turbidity (Ruddick et al. 2001; Alvain et al. 2005). These challenges were
173	addressed by focusing on broad, trophic level predictions of chlorophyll-a.
174	2.2 Madel Development
174	2.2 Model Development
175	We developed an Extreme Gradient Boosting (XgBoost) model to classify categories of
176	
177	chlorophyll-a. These models build on machine learning concepts such as decision trees and
178	ensemble learning (Cheng and Guesterin, 2016). Decision trees represent a supervised learning
179	approach where training features are split into internal nodes and evaluated to form
180	homogeneous groups (terminal nodes) (Kotsiantis, 2013). Decision trees can comprise a single
181	univariate classifier or the combination of multiple classifiers, known as an ensemble classifier.
182	Gradient boosting is a method of ensemble learning where a series of models are built with
183	weights assigned to misclassified observations. Misclassified observations from the previous
184	model are used as training data for the next, and the result is an ensemble classifier that
185	represents an aggregation of individual classifiers and minimizes overall error (Pal, 2007).
186	We used a combination of optical and climatic variables to build a predictive model for
187	chlorophyll-a. Specifically, we calculated multiple band ratios that have been shown to explain
188	variation in phytoplankton blooms (Ho et al., 2017). We used a 14-day average of lake surface
189	temperature and daily meridional wind speed as additional predictor variables. We explored the
190	addition of static predictor variables (such as lake elevation or watershed land use) yet refrained
191	from including these in our final model because recent studies have shown that static predictor
192	variables can act as 'identifiers' and lead to overfitting and over-optimistic evaluation metrics

- 193 (Meyer et al., 2018). Thus, we selected only continuous predictor variables that we would not
- 194 expect to lead to substantial overfitting (Table 1).
- 195

196	Table 1	. Predictor	variables	used	for model	training.	
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Predictor variable	Description
Blue	Surface reflectance of blue band
Dwl	Dominant wavelength
Nir	Surface reflectance of Nir band
Swir2	Surface reflectance of Swir2 band
Red / Blue	Red / Blue
Red / Nir	Red / Nir
Nir / Red.	Nir / Red
Green / Blue	Green / Blue
Nir Sac	(Nir – 1.03) * Swir1
Nir – Red	Nir - Red
Red - Green	Red - Green
EVI	2.5*((Nir - Red)/(Nir + ((6*Red) - (7.5*Blue)) + 1))
GCI	Nir/(Green-1)
Mean 14-day Temp	14- day average surface water temperature (deg. C)
Wind	Meridional wind speed (m/s)

198	We partitioned our training set to reserve 20% for model testing and evaluation and 80%
199	for model training and parameter tuning. XgBoost models include a wide range of
200	hyperparameters and are one of the main tools used to reduce model variance. Hyperparameters
201	were tuned by first establishing a grid of conservative values (to prevent overfitting) and then
202	extracting the hyperparameters that resulted in the lowest validation loss. After training the final

model with these hyperparameters, model performance was evaluated through a confusion
 matrix which shows the relative accuracy of predictions across different categories.

205 2.3 Data Analysis

206

To summarize lake trends and capture long-term changes in chlorophyll-a, we analyzed 207 208 the percent occurrence of trophic state observations. First, lakes included in our trend analysis 209 had to have at least two summer observations (June – September) for each year (1984-2019). 210 More conservative filtering criteria, such as at least 5 observations per year, was explored yet 211 had negligible effects on overall results and resulted in fewer lakes being included in our analysis. We specifically focused our analysis on summer observations to limit the effect that 212 213 snow and ice may have on our results. As a result, 1,169 lakes were included in our analysis 214 based on these criteria. For each summer, the percent occurrence of each trophic state observation was recorded. Then, the average percent occurrence for each trophic state was 215 recorded across two time periods: 1984 – 2004; and 2005 – 2019. Lastly, lakes were grouped 216 into the following categories based on the shift (if any) in trophic state during these two time 217 periods: 218 No trend: Change in % oligotrophic, % mesotrophic, and % eutrophic was less 219 1) than 10% across all three categories (Figure 1A) 220 2) Increasing in % Eutrophic: Number of eutrophic observations increased by 221

222 10% or more while the number of oligotrophic observations decreased by 10% or
223 more (Figure 1B)

3) Increasing in % Oligotrophic: Number of oligotrophic observations increased
by 10% or more while the number of eutrophic observations decreased by 10% or
more (Figure 1C).

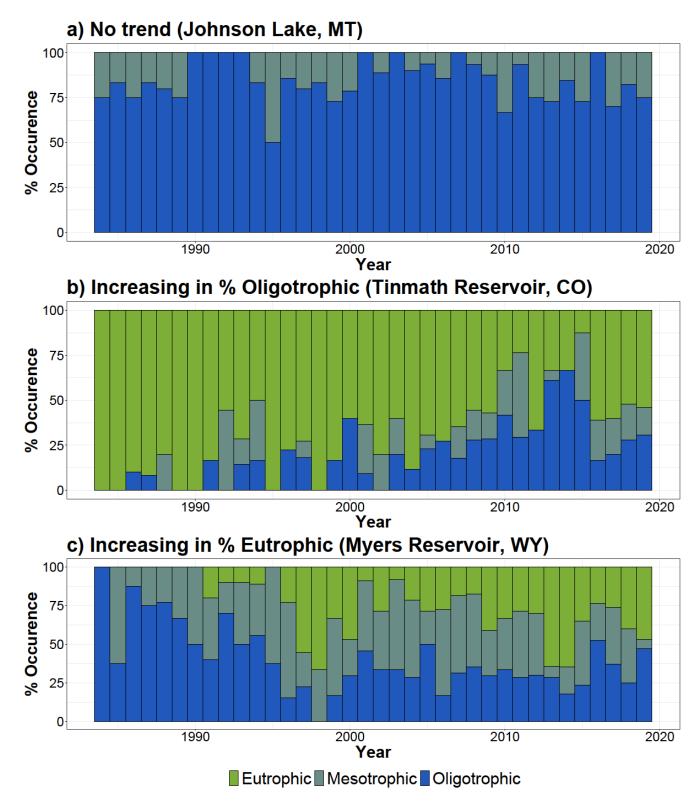


Figure 1. Examples of three possible trend categories based on the trends in % occurrence of

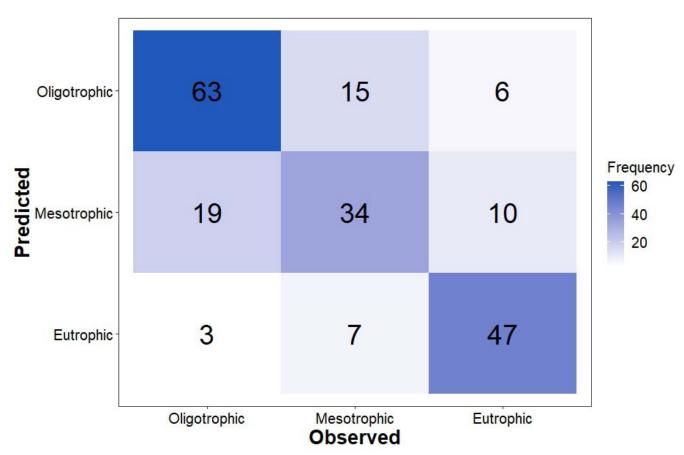
- 229 oligotrophic, mesotrophic, and eutrophic observations. Each panel included in this plot
- 230 represents the trends observed across three different lakes.

231	Lastly, trend-specific drivers were examined by determining how lake catchment,
232	hydrologic, and climate metrics explained differences across trends. We calculated variable
233	importance across trend categories by applying a random forest model using the randomForest
234	package in R (Liaw and Wiener, 2002). With this approach, we were able to classify the
235	reduction in accuracy that occurred across all three responses when certain variables were
236	excluded. All data processing, model development, statistical analysis, and visualizations were
237	done in Program R (R Core Team, 2022).
238	3 Results
239	3.1 Model Performance
240	
241	Model performance was evaluated through a confusion matrix as well as various
242	accuracy and error metrics (Table 2, Figure 2). In the range of oligotrophic values (0 - 2.6 ug/L),
243	observations had a balanced accuracy of 78% and only 7% of these observations were
244	misclassified as eutrophic (Table 2). Mesotrophic observations (2.7 - 7 ug/L) represented the
245	range of values with the lowest prediction accuracy. Our model reported a balanced accuracy of
246	69% for mesotrophic classifications (Table 2). The most common misclassification among
247	mesotrophic predictions was with observed classes that were oligotrophic (30%) (Figure 2).
248	Lastly, eutrophic observations (> 7 ug/L) represented the class with the highest prediction
249	accuracy (85%) (Table 2). In addition, there was relatively low prediction error with oligotrophic
250	classes (6%). Overall, our model reported a global accuracy of 70% with a 95% confidence
251	interval of between 63% and 76% (Table S2).
252	

253	Table 2. Model	evaluation	metrics for	each pro	edicted class.
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Statistic	Oligotrophic	Mesotrophic	Eutrophic
Sensitivity	0.7500	0.5397	0.8426
Specificity	0.8167	0.8440	0.8912
Neg Pred Value	0.8235	0.8041	0.9291
Pos Pred Value	0.7412	0.6071	0.7460
Prevalence	0.4118	0.3088	0.2794
Balanced Accuracy	78.33%	69.18%	85.79%

254



255

Figure 2. Confusion matrix illustrating the frequency and accuracy of predictions across all three

257 trophic states. The most common misclassification was among mesotrophic predictions that had

258 observed classes of oligotrophic (middle panel, far left). Overall, our model had a global

accuracy of 70% with a 95 % confidence interval of 63% - 76%.

The integration of fine-scale, daily temperature and climate features significantly improved our ability to predict across these trophic states. In terms of feature importance measured by model gain, mean 14 – day surface water temperature and meridional wind speed were the second and fourth most important predictor variables, behind the band ratio of blue to green and dominant wavelength (Figure 3). In addition, model scenarios without climate variables reported global accuracies of around 63%, with a 95% confidence interval of between 57 - 69%.

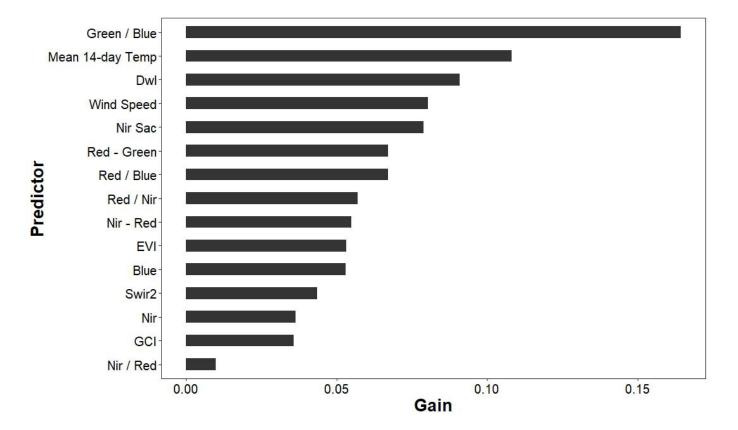
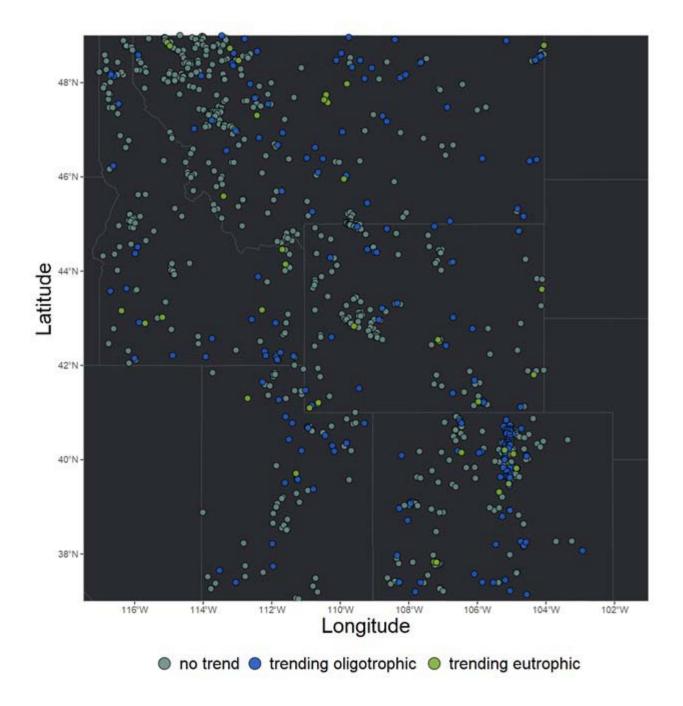


Figure 3. Feature importance, measured as model gain, for the predictor variables included in model development.

270

272 3.2 Productivity Trends

274	Most lakes included in this study did not show trends in chlorophyll-a (Figure 4). Overall,
275	a total of 651 lakes (55%) did not meet our 10% thresholds for shifts across all three categories
276	More than half of the lakes that weren't changing from 1985-2019 were oligotrophic lakes with
277	most observations classified as oligotrophic. In contrast, 24% of lakes within this category were
278	eutrophic lakes. The remaining lakes (16%) in this trend category likely represent a more
279	complex, mesotrophic lake status.
280	The second most common trend we observed were lakes that had substantial shifts in
281	trophic status by becoming more oligotrophic. We found that 17% of lakes switched from
282	predominantly being classified as eutrophic to being classified primarily as oligotrophic. Most of
283	these lakes tended to be dominated by eutrophic observations, suggesting that they are eutrophic
284	lakes that are improving in water quality. Few lakes showed evidence of extreme (>30%) shifts
285	in oligotrophic observations. In other words, shifts in oligotrophic observations within this lake
286	trend was relatively moderate (10 - 30%, Figure S1).
287	Lastly, a minority (3%) of all lakes were shifting towards becoming more eutrophic.
288	Interestingly, these trends were equally distributed across lakes with high levels of eutrophic
289	observations and those with high levels of oligotrophic observations. In other words, lakes that
290	were predominately oligotrophic and were becoming more eutrophic were equally as common as
291	lakes that were eutrophic and were intensifying in this way. The magnitude of change was
292	similar to that of lakes that trended oligotrophic, with little evidence of extreme shifts in
293	eutrophic observations (Figure S1).
294	



295

Figure 4. Spatial distribution of trophic state trends across the five states included in this analysis.

299	The remaining lakes that were included in this analysis and did not fit into these rigid
300	categories reflect various levels of trophic state change. For example, 7% of lakes could be
301	described as becoming more oligotrophic and less mesotrophic by the same thresholds outlined
302	in Figure 1. In contrast, few lakes (1%) were found to be becoming more mesotrophic during this
303	time. The 12% of lakes that did not fit into these categories displayed slight trends in certain
304	categories (such as becoming more oligotrophic), but did not satisfy thresholds for trends in
305	other categories such that we would be confident of defining clear trends in productivity.
306	3.3 Drivers of Trends
307 308	Our random forest model was able to identify partially important variables for explaining
309	trends in productivity. Lake catchment data such as 30 year normal mean temperature, base flow
310	index, and mean runoff were more important in explaining overall lake trends (Figure 5).
311	Specifically, lakes becoming more oligotrophic tended to have longer residence times and were
312	located in catchments that were generally less forested and more developed (Figure 6). Whereas,
313	lakes that were becoming more eutrophic also tended to be less forested but were located in
314	smaller catchments and were shallower on average (4.13 m) compared with lakes that were not
315	trending (9.12 m). Lastly, a number of climate and landscape metrics displayed a high level of
316	variation across trophic state trends, however some of these metrics had significant cross
317	correlation with other variables (Figure S2).
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319	
320	
321	

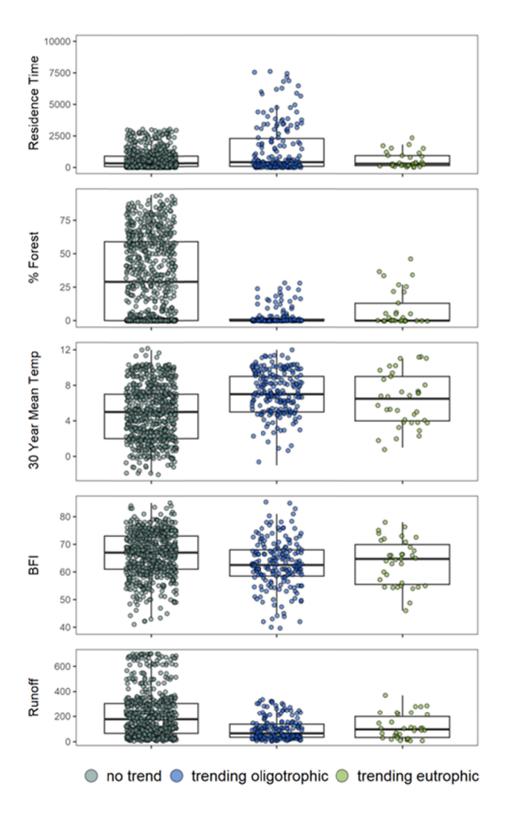




Figure 5. Boxplots across trend categories of the top five most important variables based on the decrease in accuracy from the overall (global) random forest model.

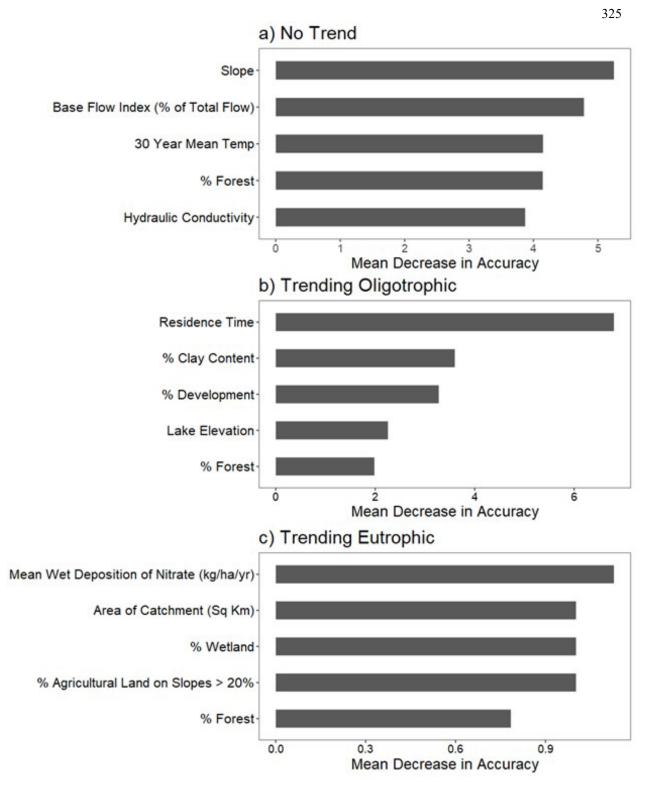


Figure 6. Mean decrease in accuracy of the top five variables used to explain each trend category in the random forest model. The mean decrease in accuracy describes variable

importance by quantifying how much accuracy is lost by excluding that particular variable.

329 **4 Discussion**

Eutrophication and the development of algal blooms are global phenomena that threaten aquatic systems. Given the effects of global change and the expected increasing intensity of these disturbances, there has been a substantial level of interest in investigating recent productivity trends in lakes and reservoirs. Our analysis found that most lakes in the Intermountain West region have remained relatively static in terms of their productivity over the last 35 years. In addition, we found that a greater percentage of lakes were improving with regards to productivity, as opposed to becoming more eutrophic.

337 4.1 Productivity Trends

338

The majority of lakes included in this analysis showed no evidence of substantial changes 339 in trophic state and supplement other regional-scale analyses of in-situ chlorophyll-a data. This is 340 341 consistent with previous analyses demonstrating that magnitude, severity, and duration of algal blooms are not intensifying in US lakes (Wilkinson et al. 2022). Similarly, long-term trends of 342 Florida lakes have indicated that a majority (73%) have not shown evidence of changes in 343 chlorophyll-a and trophic state (Canfield et al., 2018). While there is a growing concern of 344 eutrophication and HABs becoming pervasive in the Intermountain West, our results build on 345 recent studies that suggest no indication of widespread intensification in algal blooms. Rather, 346 the large percentage of lakes not trending combined with the presence of algal blooms across the 347 region suggest a historical baseline of eutrophication and that blooms could have predated the 348 1980s. 349

Our analysis revealed that, in fact, the smallest percentage (3%) of lakes were trending eutrophic. Global analyses of long-term phytoplankton blooms have shown a substantial (68 %) number of lakes to be increasing in bloom intensity (Ho et al., 2019). However, only 5% of U.S.

lakes have been shown to be increasing in the same metric over the past 40 years (Wilkinson et al., 2022). In addition, a minority of lakes (13%) in the Rocky Mountain region have shown to be shifting from blue to greener wavelengths during this time (Oleksy et al., 2022). With our analysis, we show that concerns regarding the widespread intensification of algal blooms are not captured in our analysis of chlorophyll-a and trophic state.

Our analysis of lakes that were trending eutrophic revealed several important hydrologic 358 and climate factors associated with eutrophication. Specifically, 30-year normal mean 359 temperatures tended to be higher among lakes trending eutrophic and an important variable for 360 361 explaining overall trends. In addition, hydrologic variables such as lake depth and lake area revealed that lakes trending eutrophic tended to be smaller and shallower than other lakes. Small, 362 shallow lakes are often more productive than deeper lakes because of the effects that lake 363 morphology can have on ecosystem structure (Richardson et al., 2022; Henderson et al., 2021). 364 Shallow lakes have also been shown to be more sensitive to climate conditions (Mooij et al., 365 2007) and could explain the interaction between climate and depth driving these trends. 366 In contrast, 19 % of study lakes were found to be improving by simultaneously becoming 367 less eutrophic and more oligotrophic. Lake-specific characteristics reveal that lakes improving in 368 water quality were in more developed and less forested catchments, as well as at lower 369 elevations. These results are consistent with studies on water clarity (Topp et al., 2021), lake 370 color (Oleksy et al., 2022), and chlorophyll-a (Wilkinson et al., 2022), that highlight 371 372 improvements in water quality metrics over the same time period. These trends have been hypothesized to be the result of management actions or restoration projects (Wilkinson et al., 373 2022), although we lacked the information to make conclusions about the mechanisms of these 374 375 trends. However, concentrations of nutrients across urban watersheds have significantly

376 decreased over the past 20 years and have been directly attributed to the Clean Water Act (Stets et al., 2020). Given the greater variable importance of developed land use across lakes becoming 377 more oligotrophic (3.9 compared to 1.6 among no trend lakes), it is possible that water quality 378 implementation projects have had a positive effect on mitigating eutrophication in the region. 379 Despite the 35-year study period and wide range of lakes involved, the remote sensing 380 381 data used in this study may not capture various spatial and temporal characteristics of eutrophication or algal blooms. Algal blooms tend to have high temporal and spatial variance in 382 the short term, as wind dynamics drive the spatial distribution of phytoplankton blooms (Bosse et 383 384 al., 2019). Therefore, the 16-day return period for Landsat observations may not capture shortterm peaks in chlorophyll-a. Furthermore, some images can be unusable due to extensive cloud 385 cover and may extend the period between observations up to months at a time. However, given 386 that our analysis includes 35 years of data across 1,169 lakes, we would expect to capture 387 widespread eutrophication and the spatial clustering of eutrophication trends if it were present. 388 Additionally, Landsat's long-term record restricted us to coarse analyses of chlorophyll-a 389 and trophic state. Our analysis does not capture cyanobacteria dynamics or those of cyanotoxins 390 directly. Satellites with spectral resolution to capture cyanobacteria abundance, such as MERIS 391 392 and Sentinel-3, have lacked the data availability for long-term, retrospective analyses (Coffer et al., 2020). However, future studies that are able to capture trends in cyanobacteria blooms 393 specifically will help provide further context regarding the concerns of bloom intensification. 394

395

4.2 Modeling Approach

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Our research focused on leveraging long-term remote sensing and environmental datasets
 that would supplement the ongoing debate regarding recent trends in phytoplankton blooms.

399 While the application of remote sensing for inland water quality monitoring has grown over the past decade (Topp et al., 2020), the retrieval of certain optical properties such as chlorophyll-a 400 has remained a challenge (Matthews, 2011). However, by incorporating daily surface 401 temperature and meridional wind speed from datasets leveraging modern deep learning 402 techniques we were able to show substantial improvements in model accuracy. The incorporation 403 of fine-scale lake climate data over the 35-year time span of this study was instrumental to our 404 ability to document trophic state changes and add evidence to the ongoing debate regarding the 405 recent trends in increasing eutrophication and HABs. 406

407 Most notably, surface water temperature was the second most important predictor variable of our trophic state model and could be important for a wide range of remote sensing 408 based water quality models. Water temperature has proven to be an important predictor of 409 chlorophyll-a across inland lakes (Liu et al. 2019; Karcher et al. 2020) as well as oceans 410 (Dunstan et al. 2018). However, applied remote sensing models that predict chlorophyll-a are 411 often limited to strictly optical predictors such as band-ratio (blue-green) models. These models 412 work well in waterbodies where other parameters such as colored dissolved organic matter co-413 vary with chlorophyll-a (O'Reilly et al., 1998). However, in optically complex waterbodies with 414 higher levels of turbidity and dissolved organic matter band-ratio models struggle to accurately 415 retrieve chlorophyll-a concentrations (Tzortziou et al., 2007; Zheng and DiGiacomo, 2007; 416 Witter et al., 2009). Thus, relying on surface reflectance for predictive models has resulted in a 417 418 lack of generalizability across a wide range of waterbodies. However, the incorporation of surface water temperature seems to have supplemented existing band-ratio features to better 419 420 predict across a wide range of lake types.

Wind speed was another climate predictor variable that was substantially important in predicting trophic state. Correlations between wind speed and chlorophyll have been shown using remote sensing at global scales (Kahru et al., 2010). In addition, wind speed has been documented as an important driver of cyanobacterial bloom development with blooms favoring warm, calm weather (Kanoshina et al. 2003). Overall, the integration of daily, fine-scale weather data greatly improved our ability to predict trophic state and is likely to have a positive impact on similar approaches that leverage remote sensing data.

428 **5** Conclusions

With increases in global lake temperatures (Maberly et al., 2020), lakes globally are 429 expected to become more eutrophic as a response to climate change (Yang et al., 2020). Yet, 430 there have been conflicting results thus far regarding intensifying eutrophication and algal 431 blooms in U.S. and global lakes (Ho et al., 2019, Wilkinson et al., 2022, Topp et al., 2021). 432 While increasing eutrophication is a major threat to freshwaters, our analysis found that lakes in 433 the Intermountain West region have not undergone widespread change. Rather, we found that 434 most lakes were not changing, and a substantial number of lakes were becoming less eutrophic 435 and more oligotrophic over this time period. In addition, the number of eutrophic lakes that have 436 not undergone substantial change over this time period suggests algal blooms have been present 437 in the region since at least the early 1980s. These results highlight the complex nature of 438 observing changes in freshwater lakes across large scales. However, our results suggest that 439 despite the processes that drive eutrophication (warmer temperatures, nutrient accumulation, 440 etc.) which have increased over the past several decades, we haven't yet observed a concurrent 441 increase in eutrophication from a large, unbiased sample of 1,169 lakes in the Intermountain 442

443	West. This suggested suggesting controls on eutrophication in this region are complex and need
444	further additional study.
445	
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451	
452	Open Research
453	
454	The data used for this paper (Hydrolakes, LakeCat, AquaSat, LimnoSat, and LAGOS) are all
455	freely available to download in online repositories (Messager et al., 2016; Hill et al., 2018; Ross
456	et al., 2019; Topp et al., 2021; Cheruvelil et al., 2021). Links to the where this data can be
457	downloaded can be found in the code for this analysis. The code used for this analysis can be
458	found at https://github.com/SamSillen/ProductivityTrends.
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467 **References**

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