Prediction of future Alaskan lake methane emissions using a small-lake model coupled to a regional climate model

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Abstract

Methane emissions from lakes will increase with climate warming. However, CH4 these emissions are not presently in the surface schemes of Global Climate Models (GCMs). Because climate projections depend on future atmospheric CH4 concentrations, a positive feedback loop is not simulated. To address this issue, a one-dimensional model was developed to simulate future CH4 diffusive and ebullitive fluxes from four Alaskan lakes. The model was hindcast for validation (1976-2005) and forecast for prediction (2071-2100) with one-way coupling to raw meteorological data from the CanESM2 ensemble GCM. Three climate warming scenarios (RCPs 2.6, 4.5 and 8.5) simulated bottom water to warm by up to 2.24{degree sign}C, increasing the CH4 flux from the lakes by 38 - 129%. However, RCP 2.6 and 4.5 led to stabilized temperatures and CH4 emissions by 2100, at levels of 0.63 - 1.21{degree sign}C and 38 - 67%, respectively, above the 1976-2005 averages. The CH4 diffusion parameterization was transferable between the four lakes; however, different ebullition parameterizations were required for the two deeper lakes (~6-7 m mean depth) versus the two shallower lakes (~1-3 m mean depth). Relative to using observed meteorological forcing, which had a cold bias (-0.15 to -0.63 {degree sign}C) and RMSE of 0.38 to 0.90 {degree sign}C, the GCM-forced models had a warm bias (+0.96 to +3.13{degree sign}C) and marginally higher RMSE (1.03 to 3.50{degree sign}C) compared to observations. The results support continued efforts to couple CH4 lake-emission models to GCMs without downscaling meteorological data, allowing feedback between CH4 dynamics and future climates to be modelled.

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Prediction of future Alaskan lake methane emissions using a small-lake model coupled to a regional climate model

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12 Key Points:

- CH₄ emissions from lakes are not presently in the land surface schemes of Global
 Climate Models
- A one-dimensional lake model was developed to simulate future CH₄ diffusive and ebullitive fluxes from four Arctic lakes
- Three climate warming scenarios simulated the bottom water temperature to warm by up to 2.24°C, increasing the surface CH₄ flux from the four lakes by 38 129%.
- 19

Abstract 20

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from lakes are not presently in the land surface schemes of Global Climate Models (GCMs). 22

Modelled climate projections depend on future atmospheric CH₄ concentrations; therefore, a 23

positive feedback loop is not simulated. To address this issue, a one-dimensional lake model was 24

- 25 developed to simulate future CH₄ diffusive and ebullitive fluxes from four Arctic lakes. The
- model was hindcast for validation (1976-2005) and forecast for prediction (2071-2100) with one-26 way coupling to raw meteorological data from the CanESM2 ensemble GCM. Three climate 27
- warming scenarios (RCPs 2.6, 4.5 and 8.5) simulated the bottom water temperature to warm by 28
- up to 2.24°C, increasing the surface CH₄ flux from the four lakes by 38 129%. However, RCP 29
- 2.6 and 4.5 led to stabilized temperatures and CH₄ emissions by 2100, at levels of 0.63 1.21°C 30
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38 meteorological data, allowing feedback between CH₄ dynamics and future climates to be modelled.

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- 40

Plain Language Summary 41

Climate change in the Arctic is moving at a greater rate than in the rest of the world. The urgency 42

of characterizing greenhouse gas emissions from water bodies in high latitudes has become a 43

subject of intensive research during the last two decades. It is believed that methane (CH4) 44

emissions from freshwater systems are the most important source of uncertainty in the global 45 greenhouse gas budget, and their contribution has been excluded from earth systems models. 46

Previous models have been developed to determine these emissions; however, they often require 47

several inputs and lake characteristics that are not readily available. Considering this, we 48

developed a subroutine to determine CH4 fluxes and concentrations in the water column for four 49

lakes located in Arctic Alaska. Subsequently, we implemented meteorological data from a 50

climate model to predict future CH4 fluxes. The calculated increases in atmospheric fluxes were 51

significant; emissions from all the studied lakes are expected to increase at least 38% over the 52

next 80 years. Our study presents a simple formulation with limited constraints to estimate CH4 53

emissions. We expect that our subroutine could be embedded into climate models to predict 54

- 55 emissions from the Arctic and potentially from the rest of the globe.
- 56

58 1 Introduction

59

60 Atmospheric CH₄ emissions from lakes form a large portion of the global greenhouse gas (GHG)

budget (10-16%; Bastviken et al., 2011), with lake sediments being a reservoir for mineral and

62 organic carbon, which is released as CH₄ (as well as CO₂; Dean and Gorham, 1998; Zhang et al.,

63 2017). CH₄ emissions from lakes vary with seasonal changes in temperature and these

64 fluctuations are primary drivers of intra-annual variation in global CH₄ emissions (Greene et al.,

65 2014; Wik, Varner, and Walter Anthony, et al., 2016). Under the RCP 2.6 emission scenario

from 2080 - 2100, higher lake surface water temperatures and longer ice-free seasons are

expected globally (2.5°C and 15 days, respectively; Woolway and Merchant 2019); with
 warming 2-3 times more severe in the Arctic (Graversen et al., 2008; Belkin, 2009; Bintanja,

warming 2-3 times more severe in the Arctic (Graversen et al., 2008; Belkin, 2009; Bir
2018).

70 Rapid Arctic warming will increase CH₄ emissions by thawing permafrost, but will also increase

the rate of CH_4 release from lakes through enhanced microbial decomposition in the sediments

(Walter Anthony and Anthony, 2013). Higher temperatures will also decrease the duration of ice

rover, which is relevant as ice promotes dissolution of CH₄ bubbles into the water column, and

⁷⁴ subsequent accumulation under ice (Denfeld et al., 2018). A shorter duration of ice cover will

result in increased ebullition and diffusion; although increased oxidation is also expected,

renhanced CH₄ productivity may exceed this loss (Greene et al., 2014; Martinez-Cruz et al.,

77 2015).

78 The increased ebullitive and diffusive CH₄ fluxes, from temperature-driven enhancement of

79 productivity in anoxic sediments (Zeikus and Winfrey, 1976; Walter Anthony et al., 2006; Zhang

et al., 2017), suggests a positive feedback loop, because model projections of future climate

change depend strongly on the atmospheric concentration of GHGs (Harriss et al., 1993; Whalen,

2005). Small lakes are presently not resolved in climate models (MacKay et al., 2009), and so

83 there is a need to include this positive feedback from Arctic lake systems into climate model land

- 84 surface schemes.
- 85 Correspondingly, efforts have been made to incorporate emissions from freshwater systems into
- 86 regional and global GHG budgets. Observations from individual lakes have been up-scaled
- according to lake-size distributions (Bastviken, Ejlertsson and Tranvik, 2002; Bastviken et al.,
- 2004; Wik, Varner and Walter Anthony, et al., 2016; DelSontro, Beaulieu and Downing, 2018).
- 89 Process-based models have also been developed to simulate CH₄ dynamics in freshwater systems
- 90 (Tan, Zhuang and Walter Anthony, 2015; Stepanenko et al., 2016); however, they require lake
- 91 characteristics and calibration variables that are not typically measured.

92 Tan and Zhuang (2015) coupled a lake model to downscaled output from a climate model (one-

way coupling) to project present and future CH₄ emissions from Arctic lakes. Their results

showed that the emissions will roughly double by the end of the 21st century. However, because

of the bias-correction and interpolation applied in their downscaling, their models are not able to

96 be run as part of a climate model land surface scheme that would enable positive feedback

97 between CH₄ emissions and increased future temperatures.

- 98 The purpose of the present study is to develop a scalable CH₄ emission model (Hurtado Caicedo,
- 99 2019) that may be coupled to the one-dimensional (1D) Canadian Small Lake Model (CSLM;

- 100 MacKay 2012, MacKay et al., 2017). Although not part of the present study, these models can be
- 101 embedded in the land surface scheme (CLASS) of the Canadian Regional Climate Model
- 102 (CRCM; Verseghy and MacKay, 2017) and hence would be able to capture the feedback
- between CH_4 emissions and climate change. The specific objectives are to (1) assess the ability
- of the CSLM-CH₄ model to estimate present water temperatures and CH₄ emissions, from four
- Arctic lakes, without downscaling the CRCM output used as surface forcing; and (2) project
- 106 future CH₄ emissions under various GHG emission scenarios for the four Arctic lakes.

107 2 Methods

108 2.1 Study sites

- 109 Four Arctic kettle lakes near Toolik Field Station (68°38' N, 149°39' W) were modelled: Toolik
- Lake and Lakes E1, E5 and E6 (Figures 1, S1 and S2). The lakes are located in the continuous
- 111 permafrost zone on the North Slope of Alaska in the Brooks Range foothills region. These water
- bodies are part of the Arctic Long-Term Ecological Research (ARC LTER) program and have
- been widely studied in order to understand the impact of environmental change on their physical
- and chemical properties (O'Brien, 1992; MacIntyre et al., 2006; Walter Anthony et al., 2008;
- 115 Jorgenson et al., 2010; Hobbie and Kling, 2014).
- 116 Lakes characteristics (surface area, bathymetry and extinction coefficients) were available from
- 117 MacIntyre et al. (2018; Table 1). Given the predominantly flat-bottomed bowl shape of kettle
- 118 lakes (Fig. S1), simulations resolved from the free surface to the mean (Z_{mean}), as opposed to the
- 119 maximum (Z_{max}) lake depth (see MacKay, 2012).



120



124 **2.2 Forcing and validation data**

125 2.2.1 Observed meteorological forcing

126 Meteorological data was collected by Kling (2000) at the surface of Toolik Lake during the ice-

127 free seasons (June - September) of 2001-2005. Incident shortwave and longwave radiation were

- measured with a Kipp and Zonen CM3 radiation sensor at ~50 cm height above the lake surface.
- 129 Air temperature and relative humidity were measured with Vaisala HMP45C (2.5-3 m above the
- water surface), and wind speed was measured with a Met One 014A anemometer (3.8-5 m above
 the water surface, Table 2) and corrected to 10-m using a logarithmic model (Holmes, 2001).
- Given their proximity (Figure 1), the same meteorological forcing was applied to all four lakes.

133 2.2.2 Modelled meteorological forcing

- Baseline (1976-2005) and future (2071-2100) global climate model (GCM) forcing data were
- obtained from the simulations of the Canadian Earth System Model (CanESM2/CGCM4), which
- 136 was run as part of CMIP5 with a 50-km grid resolution. The emission scenarios included the
- 137 Representative Concentration Pathways 2.6 (RCP 2.6), 4.5 (RCP 4.5) and 8.5 (RCP 8.5). RCP
- 138 2.6 is a low greenhouse gas emission scenario in which changes to the radiative forcing would
- 139 lead to an increase in the global mean temperature of 1°C. Under RCP 2.6, global carbon
- emissions are projected to decrease to near zero by the end of 2100. Under the medium
- emissions scenario, RCP 4.5, changes to the radiative forcing cause a 2°C increase in air
- temperature under stabilized GHG emissions, and under the high emission scenario RCP 8.5,
- 143 changes to the radiative forcing result in an increase of 3.7°C in air temperature with a
- 144 continuous growth of GHG emissions (IPCC, 2013).

145 **2.2.3 Observed temperature time series**

- 146 Water temperature time-series at Toolik Lake and Lakes E5 and E6 were collected by S.
- 147 MacIntyre (ARC LTER; <u>https://arc-lter.ecosystems.mbl.edu</u>) during the summers of 2001-2005.
- 148 Water temperature was measured with moored, self-contained loggers (see Table S1 for logger
- 149 depths). In Toolik Lake RBR Ltd. TR-1050s (±0.002°C) and in Lakes E5 and E6, StowAway
- 150 Tidbit Loggers ($\pm 0.21^{\circ}$ C) were used.

151 2.2.4 Observed CH₄ fluxes

- 152 CH₄ ebullitive fluxes from Toolik Lake in 2004 were digitally retrieved from Tan et al., (2015;
- their Figure 10). They collected gas samples using submerged bubble traps over ebullition seeps
- along defined transects in the lake from April 28, 2003 to December 31, 2004. Subsequently,
- 155 lake-wide daily ebullition was determined as the sum of the fluxes from each seep type and
- 156 averaged over the lake surface area.
- 157
- 158 Continuous diffusion and ebullition data for lakes E1, E5 and E6 were not available; however,
- aggregate values of seasonal and mean annual fluxes were used as given in Sepulveda-Jauregui
- 160 et al. (2015). They determined the mean annual fluxes as the sum of different modes of emission
- 161 in the summer (ice-free season), winter and spring. Sepulveda-Jauregui et al. (2015) performed
- 162 measurements from June to July (2011–2012) and extrapolated the value to the entire ice-free
- 163 season. They calculated ebullitive fluxes by multiplying the average seep densities on each lake
- by the sum of daily ebullition from seeps of each type. We compared our simulated surface

diffusive fluxes from Toolik Lake during the summer seasons of 2010–2015 to measurements by

Eugster et al. (2020a), who deployed a three-dimensional ultrasonic anemometer-thermometer and a closed-path integrated off-axis cavity output spectrometer.

174	Table 1. Lake characteristics. Z _{max} and Z _{mean} are maximum and mean lake depths respectively, and k _d is
175	the extinction coefficient (MacIntyre, Cortés and Sadro, 2018).

	Latitude	Longitude	Area	Zmax	Zmean	k _d
	(°N)	(°W)	(km ²)	(m)	(m)	(m ⁻¹)
Toolik Lake	68.633	-149.607	1.49	26.0	7.4	0.6
Lake E1	68.626	-149.555	0.029	12.0	3.1	0.8
Lake E5	68.642	-149.458	0.109	12.9	6.4	1.1
Lake E6	68.643	-149.441	0.019	3.0	1.6	1.4

Table 2. Calibration parameters for each lake. The wind sensor height and hypolimnion turbulent

diffusivity were adjusted, on a lake-by-lake basis, to account for sub-grid-scale turbulent mixing and for

spatial variability in wind speed and surface drag (which is computed internally in the model). Clabile and

k_{ox}were adjusted to minimize CH₄ RMSE.

	Wind sensor height (m)	Clabile (mg m ⁻²)	Hypolimnion turbulent diffusivity, K _Z (m ² s ⁻¹)	CH4 oxidation k _{ox} (s ⁻¹)
	(A)	(B)	(C)	(D)
Range	3.8 - 10	90 - 280	$10^{-5} - 10^{-9}$	10 ⁻⁵ -10 ⁻⁸
Toolik Lake	3.8	280	10-7	1.74×10 ⁻⁵
Lake E1	5.0	280	10-7	1.74×10^{-5}
Lake E5	5.0	280	10-7	1.74×10^{-5}
Lake E6	3.8	280	10-7	1.74×10 ⁻⁵

187 2.2.5 Observed CH₄ concentrations

188 CH₄ concentration profiles were measured in Toolik Lake and lakes E1, E5 and E6 during the

ice-free seasons of 2013-2016 by MacIntyre and Cortés (2017). CH₄ was measured using the
 headspace equilibrium method and gas chromatography. A summary of the collected data and

191 sources specified in section 2.2 is presented in Table S2.

192 2.3 Modelling framework

193 Climate models are typically validated against long-term averages (e.g., climate normals);

194 however, they also show skill in reproducing characteristic large-scale distributions of air

195 temperature, precipitation, radiation and wind. For example, the dynamics of monsoon systems,

seasonal temperature changes and storm tracks. Climate models can even predict weather over

seasonal timescales and capture interannual variability (Randall et al. 2007). As a result, output
 from climate models is frequently applied to drive lake models to forecast the impacts of climate

change on lake hydrodynamics and water quality (e.g., Woolway et al. 2020; Bolkhari et al.

200 2022; Golub et al. 2022). When run as hindcasts, it is expected that these GCM-forced models

will reasonably capture observed variation in meteorological conditions over seasonal to inter-

202 annual timescales.

203 To evaluate forcing-bias when developing coupled lake-atmosphere models, it is a requirement

to test the accuracy of the lake model when forced directly with atmospheric (or climate) model

output, in comparison to being forced with observed meteorological forcing (e.g., Huang et al.

206 2010). Consequently, simulation dates were selected based on the availability of both

207 meteorological forcing data sets (observations and GCM model output) and water temperature

calibration data (ARC LTER). Three different calibration scenarios were run: (1) with observed

meteorological forcing for the ice-free season of individual years (Toolik Lake: 2001-2005; Lake
E5: 2002-2005; Lake E6: 2003-2005; no lake temperature data was available for Lake E1); (2)

with GCM-generated meteorological data for the ice-free season of individual years; and (3) with

GCM-generated meteorological data for the fee-free season of individual years, and (5) with GCM-generated meteorological data in a continuous 'baseline' run from 1976-2005. Simulations

(1) and (2) were to assess the error in applying GCM forcing without downscaling; simulation

214 (3) was to test for long-term model drift.

The GCM output is a statistical representation of the atmospheric conditions in each year, and

thus does not necessarily capture the exact timing of particular meteorological events (e.g., the

217 passage of a cold front; Figures S4-S6). Therefore, here we quantitatively compared mean

seasonal observed temperature profiles, as opposed to conventional contours of temperature

time-series with depth. The latter type of comparisons is given in Hurtado Caicedo (2019) and

220 Figure S7.

221 Temperature profiles for model initial conditions were not available from ARC LTER in 1976;

therefore, Toolik was initialized with historic data (Giblin and Kling, 2015), while E5 and E6

used initial profiles from 2002 and 2003, respectively. For simulations with future GCM data

224 (2071-2100), initial water temperatures were from observations in 2013. Initial CH₄ profiles

225 were as reported by MacIntyre and Cortés (2017). For long-term simulations, it is expected that

the model will lose knowledge of the initial conditions within the first year of simulation (1977

for the baseline run and 2072 for the future simulations; Figure S3), hence model results during

this time were disregarded.

229 2.3.1 Model description

- 230 The approach to determine CH₄ emissions consisted of two parts: (1) implementation of a pre-
- existing thermodynamic lake-tile model to calculate the surface mixed layer depth and unsteady
- temperature profile through the water column, and (2) the development of a new CH₄ subroutine
- to compute CH₄ ebullition, dissolved CH₄ flux from the sediments, diffusive CH₄ flux below the
- surface mixed layer and through the air-water interface.
- 235 *Temperature:* Temperature profiles were simulated using the 1D (vertical) thermodynamic
- 236 CSLM lake-tile model, which incorporates a bulk mixed-layer turbulent kinetic energy budget to
- simulate a 1 m² water column, characteristic of a lake with a specified surface area, mean depth
- and light extinction coefficient (Table 1; MacKay, 2012; MacKay et al., 2017). We acknowledge
- that a mean depth approach underestimates water temperature in the littoral zone and
- overestimates water temperature at depth. We justify this by noting that the resulting lower modelled fluxes in shallow zones and higher modelled fluxes at depth will partially cancel and
- that the objective here is to develop a feasible lake-tile CH_4 model and not to simulate CH_4
- fluxes as accurately as possible. The hypolimnion turbulent diffusivity K_z and wind sensor height
- were adjusted, on a lake-by-lake basis (Table 2), to account for variation in sub-grid-scale
- turbulent mixing and for spatial variability in wind speed and surface drag (which is computed
- internally in the model).
- 247 The model had $\Delta z = 0.5$ m vertical grid resolution and used 5 or 15-minute timesteps (Δt) when
- forced with observed and GCM-derived meteorological data, respectively. Ice cover was
- modelled according to the snowpack physics module of CLASS, which was shown to reproduce
- 250 ice-on/off dates to within 1 week of observations in Lake 239 (MacKay et al. 2017).
- 251 *CH₄ ebullition:* To model CH₄ emissions within a land-surface scheme, model parameterizations
- 252 must be a function of variables readily obtained from GCM output, which limits model
- complexity. We parameterize CH₄ ebullition for the deeper ($Z_{mean} = -6-7$ m; Toolik and E5) and
- shallower ($\mathbb{Z}_{\text{mean}} = \sim 1-3 \text{ m}; \text{ E1}, \text{ E6}$) lakes according to Eq. 1 and 2, respectively, which are

$$E = -0.00036T^3 + 0.16T^2 - 0.94T + 1.48$$
 (1)

$$E = 141 \times 1.19^{(T-20)}$$
(2)

257 The above expressions correlate the average ebullitive flux E (mg $CH_4 m^{-2} d^{-1}$) to the average

sediment surface temperature T (°C); assumed to be the modelled bottom water temperature.

259 CSLM is a water column model, that does not account for the distribution of sediment surface

area with lake depth. The model does not resolve littoral sediments, with higher temperatures

that contribute more ebullitive flux during summer (Bastviken et al., 2004; Wik, Varner, and

Walter Anthony, et al., 2016). To account for this, we use mean depth hypsometry and depth-

based parameterizations for the shallower (E1 and E6; Eq. 2) and deeper (Toolik and E5; Eq. 1)
lakes.

The parameterization in Eq. 1 was from 6806 ebullitive flux observations in three sub-Arctic lakes in northern Sweden (68°21′ N, 19°02′ E) during the ice-free seasons (June–September) of

- 267 2009-2014 (Wik et al., 2013, 2014; Wik, 2016). The parameterization in Eq. 2 (Fig. 1 and
- equation 2 in Aben et al., 2017) was developed for shallower Boreal ponds based on
- 269 measurements by DelSontro et al. (2016), including 77 observations during the ice-free season
- 270 (May–October of 2011, 2012 and 2014) from ten shallow ponds located in the Saguenay region
- of Quebec (48°23'N, 71°25' W), and 83 observations from three lakes in the Laurentian region
- 272 of Quebec (45°59' N 73°89' W).

Our attempts using Eq. 1 to determine ebullitive fluxes for the shallower lakes resulted in

- underestimation of emissions in comparison to observations; likely because we do not explicitly
- 275 resolve littoral sediments. These values were similar in order of magnitude to the results obtained
- for Toolik Lake. Shallower systems (ponds $< \sim 3$ m depth) have been shown to have CH₄ fluxes
- ~10 times greater than deeper systems (lakes ~3-30 m depth), supporting our usage of a different
 parameterization for ponds versus lakes (Wik, Varner and Walter Anthony, et al. 2016). For
- instance, ebullitive fluxes from the shallower systems can be up to ~9 times greater than those
- from the deeper systems (see Table S3). The other equations in Aben et al., (2017) were also
- evaluated (not shown) but performed worse than Eq. 1 and 2.
- 282 Dissolution of rising bubbles was neglected because of the shallow lake depths (Schmid et al.

283 2007). The present model also does not inhibit ebullition through ice during winter. Ebullition

- would be stored within the ice (Walter Anthony et al., 2008; Walter Anthony et al., 2006) and subsequently released during break up (Phalma, Poterson and Jaffriag, 1008a; Juntinger et al.
- subsequently released during break-up (Phelps, Peterson and Jeffries, 1998a; Juutinen et al.,
 2009; Karlsson et al., 2013). Consequently, methanotrophy (oxidation) of dissolved CH₄ from
- rising bubbles was also neglected, since previous work suggests that for systems < 20 m deep,
- less than 10% of the CH₄ is lost due to dissolution during rise (McGinnis et al., 2006; Schmid et
- al., 2007). These limitations will be addressed in future work.

290 Dissolved CH_4 and diffusive sediment flux of CH_4 : The CH_4 profile in the water column was

simulated by numerically solving the 1D diffusion equation (Eq. 3) with terms for turbulent

diffusivity, oxidation and production at the sediment-water interface (e.g., Schmid et al., 2007;

293 Jabbari et al., 2016):

$$\frac{\partial C_{CH_4}}{\partial t} = \frac{\partial}{\partial z} \left(K_z \frac{\partial C_{CH_4}}{\partial z} \right) - k_{Ox} (C_{CH_4}) + P$$
⁽³⁾

- Here, C_{CH_4} (mg L⁻¹) is the CH₄ concentration, K_Z (m² s⁻¹) is the vertical turbulent diffusivity 294 below the surface mixed layer (Table 2; $\sim 10^{-7}$ m² s⁻¹; Nakhaei et al. 2016), k_{Ox} (s⁻¹) is the first-295 order CH₄ oxidation rate coefficient (1.74×10^{-5} s⁻¹; Thottathil et al., 2019), P (mg CH₄ m⁻³ d⁻¹) is 296 the production term (Eq. 4), t is time and z is the vertical coordinate direction. We acknowledge 297 that K_Z, in these relatively shallow lakes, is likely orders of magnitude larger than this near-298 molecular value; however, the turbulent diffusivity and/or dissipation in a calibrated sub-grid-299 300 scale closure scheme is often not equal to observed values (Boegman et al., 2021; Lin et al., 2022). 301
- 302 Our first-order oxidation model was developed following Schmid et al. (2007). From a
- sensitivity analysis $(10^{-5} \text{ s}^{-1} < \text{k}_{\text{Ox}} < 10^{-8} \text{ s}^{-1})$, we selected the observed value $\text{k}_{\text{Ox}} \sim 1.74 \times 10^{-5} \text{ s}^{-1}$ by
- Thottathil et al. (2019) from 6 lakes in Quebec (Canadian Shield). For a CH₄ concentration of

- 305 $C_{CH_4} \sim 0.1 \ \mu \text{mol } \text{L}^{-1}$ (Figs S8-S11), the resulting oxidation was consistent with observed rates 306 $(\sim 10^{-6} \text{ to } \sim 10^{-5} \ \mu \text{mol } \text{L}^{-1} \text{ s}^{-1})$ from shallow Alaskan lakes (Lofton et al. 2014).
- 307 The oxidation term was applied throughout the water column and the turbulent diffusion term
- 308 was applied to drive CH₄ flux from the base of the surface mixed layer to the bottom cell. CH₄
- 309 production from sediments P (mg m⁻² s⁻¹) was applied as a flux into the cell above the sediment-
- 310 water interface (Tan et al. 2015):

$$P = R_C C_{labile} P Q_{10} \left(\frac{T - T_{pr}}{10}\right)$$
(4)

- 311 where R_{C} (0.02 s⁻¹) is the fraction of carbon converted per year, PQ_{10} (3.5) is the factor by which
- production increases with a 10°C rise in temperature, T_{pr} is the reference temperature for CH₄
- production, which is approximately equal to the yearly mean sediment temperature below the
- 314 water column (-3.0 °C) and T (°C) is the sediment surface temperature (Tan et al., 2015);
- assumed to equal the modelled bottom water temperature at the mean depth. Here, C_{labile} (mg m⁻
- ²) is the areal labile carbon density. To directly model the carbon pool requires soil incubation
- data and knowledge of the thickness of the talik layer (Tan et al., 2015). Unlike the lake-specific
- parameters in Table 1, these parameters are not known across the CRCM domain; therefore, we estimated $C_{labile} = 280 \text{ mg m}^{-2}$ (Table 2), by minimizing the difference between the observed and
- modelled CH₄ concentrations in the water column. Ideally, for the model to be scalable, C_{labile}
- will be the same for all lakes or a function of the lake-specific parameters in Table 1. This
- approach is not new, and it what is typically done to set the sediment oxygen demand in
- biogeochemical lake models (e.g., Scalo, Boegman and Piomelli, 2013).
- With Eq. 4, the model may account for both the ¹⁴C-enriched carbon pool from the upper
- 325 sediment layer constituted by newly settled organic matter, since both Toolik Lake and Lake E5
- are considered nonyedoma/nonthermokarst lakes (Stepanenko et al., 2011, 2016; Sepulveda-

Jauregui et al., 2015; Tan and Zhuang, 2015). According to Tan, Zhuang and Walter Anthony

- 328 (2015), CH₄ in nonyedoma lakes is mainly produced in surface sediments from newly deposited
- 329 ¹⁴C-enriched organic matter.
- 330 The constants R_{c} , PQ_{10} and T_{pr} were as specified by Zhuang et al. (2004) for the alpine tundra
- and polar desert ecosystem of the Toolik area. To account for productivity in the upper sediment
- layers (Peeters, Encinas Fernandez and Hofmann, 2019), and based on the shallow bowl-like
- bathymetry of the smaller lakes (Fig. S1 and S2), P was added fractionally at each depth of water
- 334 (Schwefel et al., 2018); this was accomplished by calculating P and dividing the result by the
- number of layers in the water column (14, 13, 6 and 4 layers for Toolik, E5, E1 and E6
- 336 respectively).

337 *Diffusive flux of CH*₄ *to the atmosphere.* The rate of CH₄ transfer across the water-air interface 338 was specified following Happell et al. (1995) and Schmid et al. (2007):

$$\mathbf{F}_{\mathrm{CH}_{4}} = \mathbf{f}(\mathbf{W})\mathbf{K}_{\mathrm{ex}}(\mathbf{C}_{\mathrm{W}} - \mathbf{C}_{\mathrm{A}}) \tag{5}$$

- 339 where K_{ex} (m s⁻¹) is the gas transfer velocity (O'Connor, 1983), f(W) is a coefficient that
- determines the influence of wind speed (equal to $1 + 0.058W^2$ for $W \le 5$ m s⁻¹ and $1 + 0.047W^2$
- for $W \ge 5 \text{ m s}^{-1}$; Schmid et al., 2007), C_w (mg L⁻¹) is the modelled CH₄ concentration at the water surface and C_A (mg L⁻¹) is the atmospheric equilibrium CH₄ concentration (Wiesenburg and
- Guinasso, 1979), which is a function of the surface temperature and the measured atmospheric
- CH₄ concentration (1700 ppb from 1976–2005 and projected to be on average 1300, 1600 and
- 345 3600 ppb under RCP scenarios 2.6, 4.5 and 8.5, respectively, from 2071–2100; Meinshausen et
- al., 2011). The diffusive CH₄ flux to the atmosphere was restricted during ice-cover period and
- 347 the accumulated gas was subsequently released into the atmosphere at ice-off.

348 **2.4 Flux aggregation**

The model computed CH₄ ebullitive and diffusive fluxes at 5 min intervals, when forced with

- observed meteorological data, and 15 min intervals, when forced with GCM-generated data.
- 351 Mean daily and mean annual fluxes were averages of these outputs over each day or year. Total
- surface fluxes refer to the sum of ebullitive and diffusive fluxes. These were compared to
- observed ebullitive and diffusive fluxes collected as described in section 2.2.4. We acknowledge that spatial and temporal extrapolation of the observations may introduce error (e.g., they neglect
- that spatial and temporal extrapolation of the observations may introduce error (e.g., they neglect background non-seep emissions); however, these were the best available data (Sepulveda-
- Jauregui et al. 2015) and followed well-established practices (Bastviken et al. 2011). For Toolik
- Lake, daily ebullitive fluxes from point source seeps in the lake centre, were digitally retrieved
- from (Tan et al., 2015), averaged over July 4 to August 18, 2004 and compared to the modelled
- 359 average over the same period.

360 **3 Results**

361 **3.1 Model calibration (observed meteorological forcing)**

- The coupled CSLM-CH₄ model (Hurtado Caicedo, 2019) was developed and calibrated for all
- four lakes over a 3 to 4-year duration simulation (Figures S8-S11) using the observed
- 364 meteorological data. Simulated temperature profiles had depth-dependent root-mean-square
- errors (RMSE) <3.2 °C, which were highest through the thermocline, where small differences in the simulated thermocline depth can lead to larger RMSE (e.g., Boegman and Sleep, 2012).
- the simulated thermocline depth can lead to larger RMSE (e.g., Boegman and Sleep, 2012).
 These were in the range of literature values, from large lake model applications (3 < RMSE < 7)
- ³⁶⁷ These were in the range of interature values, non-range rake moder appreations (3 < RWSE ³⁶⁸ °C; Huang et al., 2010; Paturi et al., 2012) and other pond simulations (RMSE < 3.03 °C;
- Nakhaei et al., 2018). The surface RMSE <1.5 °C was less than 1.96 °C for other small boreal
- 370 lake simulations (Stepanenko et al., 2016) and comparable to 1.5 °C from other Arctic lake
- 371 simulations (Tan et al., 2015).
- 372 The dissolved CH₄ concentrations were $< 0.6 \mu$ M (Figures S8-S10), which is reasonable in
- 373 comparison to errors from more complex CH_4 models (Stepanenko et al., 2011, 2016; Tan et al., 2015) The DMCF f_1 is the E1 (10.2 M) F_2 (10.2 M)
- 2015). The RMSE for Lakes E1 (< 0.2 μ M), E5 (< 0.05 μ M) and Toolik Lake (< 0.03 μ M) were
- consistent with the mean errors found by Tan et al., (2015) and Stepanenko et al. (2016), which ranged from 0.01 µM to 0.26 µM. However, the PMSE for Lake E6 were larger, ranging from
- ranged from 0.01 μ M to 0.26 μ M. However, the RMSE for Lake E6 were larger, ranging from 0.3 μ M to 0.6 μ M.
- $377 \quad 0.3 \mu M$ to $0.6 \mu M$.
- The simulated mean-daily summer ebullitive fluxes from Toolik Lake in 2004 were within 2% of measured values (7.04 vs. 6.91 g m⁻² d⁻¹, respectively: Table S3). Although ebullition

observations were not available for the other lakes (E1, E5 and E6) during the same years as

observed meteorological forcing (2013-16), observed ebullition data from 2011-12 were

included for comparison (Table S3). The simulated mean summer fluxes were 19% (9.53 mg CH₄ m⁻² d⁻¹), 21% (0.72 mg CH₄ m⁻² d⁻¹) and 30% (25.74 mg CH₄ m⁻² d⁻¹) lower than observed

 $_{383}$ CH4 II $^{-}$ d $^{-}$), 21% (0.72 IIg CH4 II $^{-}$ d $^{-}$) and 30% in E1, E5 and E6 over 2011-12.

The simulated mean-daily summer surface diffusive fluxes from Toolik Lake in 2015 and 2013 385 were within 20% of measured values by Eugster et al. (2020a; Table S4). As with the observed 386 387 ebullitive fluxes, observed diffusive fluxes were not available for lakes E1, E5 and E6 during 2014–2015, and so were compared to observed data from other years. For lakes E5 and E1, the 388 observed fluxes were larger than simulated value, whereas for E6 the simulated flux was larger 389 (Table S4). These data should be interpreted with caution, as the 2012 diffusive surface flux 390 observed by Sepulveda-Jauregui et al. (2015) from Toolik Lake is ~3 times larger than the value 391 observed by Eugster et al. (2020a). Therefore, values reported by Sepulveda-Jauregui et al. 392

393 (2015) for the shallower systems may also be overestimated.

394 **3.2** Hindcast simulations (1976-2005 GCM forcing)

395 Simulated temperatures: Simulations forced with GCM data, initialized in individual years, were

better in comparison to observations, relative to the long-term GCM-forced simulations (Figure

2); except Toolik (in 2001) and E5 (in 2003 and 2004). The GCM-forced model had a systematic shift to higher temperatures in some years ($\leq \sim 4$ °C in Toolik, Fig. 2d; ~ 5 °C in E5), where the

seasonal thermocline intersected the lakebed. The shallower Lake E6, had more satisfactory

400 results (<2 °C bias). The temperature overestimation, simulated with the GCM, resulted from

401 higher incident longwave radiation in the GCM compared to the observed data (23 W m⁻² higher

402 on average; Table S5), warmer air temperatures (1 °C higher on average, with exception of 2005)

and higher wind speeds (1.5 m s^{-1} higher on average; Table S5). The GCM-forcing causes higher

404 temperatures and deepening of the thermocline relative to observations, and thus higher sediment

temperatures when the thermocline approaches the lakebed (mean depth).

In 2004, there was observed temperature data (Toolik, E5 and E6), as well as forcing for all three

simulations (forced with observed 2004 meteorology, forced with 2004 GCM meteorology and

forced with 1976-2005 GCM meteorology). Intercomparison of performance metrics (Table 3)

409 show RMSE (0.38 to 0.87 °C vs. 1.03 to 3.50 °C) and NRMSE (0.03 to 0.08 vs. 0.07 to 0.30) to

be marginally less for observed forcing, compared to GCM forcing. The observed and GCM

forced simulations had cold (+0.01 to -0.63 $^{\circ}$ C) and warm (+0.96 to +3.13 $^{\circ}$ C) biases,

412 respectively. The high R^2 (0.99 for observed forcing and 0.74-0.99 for GCM forcing) indicates

that the observed temperatures are explained by the models. In 2004, the GCM underestimated

observed shortwave radiation by 9.3%, overestimated longwave radiation by 6.4% and air

temperature by 4.6% (Table S5). However, it remains difficult to generalize this error as GCM

biases are different from year to year (Table S5). There was no discernable difference in metrics
for the 2004 GCM vs. 1976-2005 GCM forcing, indicating that long-term model drift was not an

418 obvious source of error.

419





Figure 2. Comparison of measured time-averaged water temperature with simulations. Black stars are measured temperature, blue lines show the simulations with observed meteorological data, red lines show simulations with GCM data and initial conditions at the beginning of the simulation period for each year, and green lines show simulations with GCM data starting in 1976. Temperature validation were from Jul. 16 to Aug. 11 in 2001, Jul. 2 to Aug. 10 in 2002 and Jun. 28 to Aug. 18 in 2004.

427

428 Simulated CH₄ fluxes: The baseline (1976-2005) average annual surface flux (sum of ebullition

- 429 and diffusion; Table 4A) was compared to the average total annual 2011-2012 emission
- 430 observations reported by Sepulveda-Jauregui et al. (2015). The simulated baseline fluxes for the
- 431 shallower lakes (Table 4A; from 8.43 ± 3.72 g CH₄ m⁻² yr⁻¹ in E1 to 8.65 ± 4.01 g CH₄ m⁻² yr⁻¹ in
- E6) were within 34% of the observations (Table 4B; from 9.4 g $CH_4 m^{-2} yr^{-1}$ in E1 to 13.3 g CH_4
- m^{-2} yr⁻¹ in E6). The observed total fluxes were higher in Lake E6 than in Lake E1, which was
- 434 captured by our model (Table 4). For the shallower lakes, the modelled surface fluxes were

- smaller than observed fluxes, which could result from Sepulveda-Jauregui et al. (2015)
- 436 extrapolating emissions from June-July observations, when sediment temperatures were much
- 437 higher, to the entire ice-free season (considered to be May-September) and therefore likely
- 438 overestimating fluxes (Section 3.1).
- The modelled baseline fluxes for the deeper lakes (Table 4A; from 3.79 ± 1.54 g CH₄ m⁻² yr⁻¹ in
- Toolik to 4.20 ± 1.52 g CH₄ m⁻² yr⁻¹ in E5) were larger than the observations from Sepulveda-
- 441 Jauregui et al. (2015), for 2011-2012 (Table 4B; from 2.0 g $CH_4 m^{-2} yr^{-1}$ in Toolik to 1.4 g CH_4
- 442 $m^{-2} yr^{-1}$ in E6). This discrepancy can be attributed to the warm bias of the GCM forced models,
- which was more prominent in the deeper systems (Table 3; Figure 2).

444 **3.3** Forecast simulations (2071-2100 GCM forcing)

- *Air temperature.* Under RCP scenarios 2.6, 4.5 and 8.5, the air temperature in the Toolik Lake
- 446 area was modelled to increase 3.6–7.3°C (Figure 3; Table 6) over the next 80 years. This is
- 447 consistent with the rates that have been reported for the entire Arctic over the past century
- 448 (Graversen et al., 2008; Belkin, 2009; Bintanja, 2018). RCP 8.5 shows a strong and continuous
- 449 increase in air temperature, which is more abrupt than under baseline conditions, RCP 4.5 has a
- 450 consistently increasing trend and RCP 2.6 shows a decline in air temperature over 2071-2100.



- Figure 3. Comparison of baseline and future simulations of average yearly air temperature. Baseline data
 correspond to GCM data in years 1976 2005 (black line) and future data correspond to GCM data in
 years 2071 2100 under RCPs, 2.6 (blue line), 4.5 (green line) and 8.5 (red line).
- 455 Bottom lake temperature. The simulated increases in bottom-water temperature (from 1976-2005
- to 2071–2100) were similar between the lakes (Table 6A; Figure 4 and Figure 5). For RCP 2.6,
- 457 4.5 and 8.5, the average bottom temperatures increased 0.61–0.84°C, 0.94–1.21°C and 1.82–
- 458 2.24°C, respectively, over 2071–2100 relative to the 1976–2005 baseline. Inter-annual variability
- in average bottom temperature was ~ 2 °C, with a more pronounced overall increase in bottom
- 460 lake temperature under RCP 8.5 (Figures 4 and 5).

Sediment CH₄ diffusion. Average diffusive fluxes from the lake sediments over 2071–2100 461

ranged between 40.24–72.03 mg CH₄ m⁻² d⁻¹ (under all RCPs). Lake E5 had the lowest fluxes of 462

40.24 - 61.28 mg CH₄ m⁻² d⁻¹ (Table 6 and Figure 4). The shallower lakes E1 and E6 had similar 463 increases in the sediment flux for all scenarios (increasing 8.25 - 27.94 mg CH₄ m⁻² d⁻¹ or 19–

464 63% from 1976-2005 to 2071-2100, Table 6B). These were larger than those of the deeper lakes 465

(increasing 5.62–21.03 mg CH₄ m⁻² d⁻¹ or 14–52%, Table 6B; Figures 4 and 5) and may be 466

associated with increased sediment warming (Table 6A). 467

CH₄ ebullitive flux. The simulated CH₄ ebullitive flux was most sensitive to the increases in lake 468 water temperature (Table 6C; Figures 4 and 5), with future fluxes more than doubling over the 469 baseline emissions under the high emission RCP 8.5 scenario in the deeper systems (increasing 470

3.57-3.67 mg CH₄ m⁻² d⁻¹ or 208–217%, Table 6B, Figure 4). In the shallower lakes, the higher 471

ebullitive fluxes had a smaller percentage change (115–118%), but larger absolute change 472

(12.59–13.12 mg CH₄ m⁻² d⁻¹, Table 6B), due to Eq. 2 being more sensitive to changes in 473 ebullition with temperature changes (Table 6A). This confirms the more significant impact of

474

climate change on increasing CH₄ ebullition (Aben et al., 2017; Walter et al., 2007; Wik, Varner 475 and Walter Anthony, et al., 2016) relative to increases in diffusion, particularly for shallow 476

systems. Ebullition comprised 15% to 47% of the total CH₄ flux from the four lakes (ebullition

477 plus diffusion under all scenarios), 68% of this total flux was produced by the two shallower 478

lakes (E1 and E6; Table 6C). 479

Ice-free days. The number of ice-free days increased from 103–111 (1976–2005) to 122–155 480

(2071–2100) (Table 6E). The number of ice-free days was estimated to be up to 19 days longer 481 under RCP 2.6 and up to 44 days longer under RCP 8.5. The baseline model (1976-2005)

482 estimates were consistent with the mean reported ice-free duration increase for glacial and post-483

glacial lakes of 20 days (Wik, Varner and Walter Anthony, et al., 2016). 484

Total surface CH4 flux. The total flux was modelled to increase by 4.15-4.57 mg CH4 m⁻² d⁻¹ 485

(40%; RCP 2.6), 6.73–7.74 mg CH₄ m⁻² d⁻¹ (65–67%; RCP 4.5) and 12.79–14.81 mg CH₄ m⁻² d⁻¹ 486

(123–129%; RCP 8.6) for the deeper lakes (Toolik and E5) and 8.66–8.94 mg CH₄ m⁻² d⁻¹ (38%; 487 RCP 2.6), 14.33–14.64 mg CH₄ m⁻² d⁻¹ (62%; RCP 4.5) and 29.24–30.29 CH₄ m⁻² d⁻¹ (127–

488 128%: RCP 8.6) for the shallower lakes (E1 and E6). Tan and Zhuang (2015) estimated similar 489

surface fluxes across the Arctic to increase 87% and 137%, under RCPs 2.6 and 8.5 respectively. 490

Trends over 2071-2100. The water temperature and CH₄ fluxes increased from the 1976-2005 491

492 average to the 2071-2100 average under all RCP scenarios, following the increases in GCM-

modelled air temperature. However, this did not always cause increasing trends in the simulated 493

494 temperature and CH₄ data during 2071–2100. For all lakes, during 2071–2100, RCP 8.5 showed

strong increases in bottom water temperature and CH₄ fluxes; RCP 4.5 showed minimal 495

increases in bottom water temperatures (<~1 °C) and flat trends in CH₄ production; and RCP 2.6 496

showed decreased bottom water temperatures (<~1 °C) and trends of decreased CH₄. This 497

498 suggests that RCP 2.6 and 4.5 will lead to stabilized Arctic lake temperatures and CH₄ emissions

by 2100, at levels 0.61–1.21 °C and 38–67%, respectively, above the 1976–2005 averages. 499

500





Figure 4. Comparison of baseline and future simulations for the deeper lakes (Toolik and E5; ~6-7 m
 mean depth), of yearly average A) bottom water temperature B) CH₄ diffusive fluxes at the sediment water interface C) CH₄ ebullition from sediments and D) total surface CH₄ fluxes.

- 506
- 507





Figure 5. Comparison of baseline and future simulations, for the shallower lakes (E1 and E6; \sim 1-3 m 513 mean depth), of yearly average A) bottom water temperature B) CH₄ diffusive fluxes at the sediment-514 water interface C) CH₄ ebullition from sediments and D) total surface CH₄ fluxes.

- 517 **Table 3.** Statistical metrics comparing (*i*) 2004 simulation forced with observed meteorology, (*ii*) 2004
- simulation forced with GCM meteorology and (*iii*) 1976-2005 simulation forced with GCM meteorology
- to (iv) observed temperatures in 2004. This was the only year with observation data and forcing data for all
- 520 3 simulations; observed data was not available for Lake E1 in 2004. The GCM data is a statistical
- 521 representation of the atmosphere and is not expected to reproduce observed meteorology on a day-to-day
- 522 basis; therefore, before computing the metrics, the higher spatial resolution field observations were 523 interpolated onto the 0.5 m depths of the model grids, then the observed and simulated temperature data
- were averaged over day-of-year 186 to day-of-year 230, resulting in a single representative mean
- temperature profile for cases (*i*) to (*iv*). RMSE = root-mean-square error of simulation vs. observation;
- 526 NRMSE = RMSE/mean(observation); Bias = sum(observation-simulation)/# observations; R^2 =
- 527 coefficient of determination.

	RMSE	NRMSE	Bias	\mathbf{R}^2		
	(°C)		(°C)			
	1	oolik Lake				
Observed 2004	0.90	0.07	-0.15	0.97		
GCM 2004	1.98	0.16	+1.86	0.76		
GCM 1976-2005	2.39	0.19	+2.28	0.80		
		Lake E5				
Observed 2004	0.87	0.08	-0.63	0.99		
GCM 2004	2.67	0.23	+2.38	0.81		
GCM 1976-2005	3.50	0.30	+3.13	0.77		
	Lake E6					
Observed 2004	0.38	0.03	-0.34	0.99		
GCM 2004	1.36	0.09	+1.31	0.77		
GCM 1976-2005	1.03	0.07	+0.96	0.74		

528 4 Discussion

- 529 We have modelled CH₄ emissions from Arctic lakes to increase as a function of increased future
- air temperature and corresponding increased simulated lake bottom water temperature. In this
- approach, the following aspects were not included in the model development and may have an
- impact on simulated future emissions: (1) release of carbon pools contained in permafrost soil as
- 533 GHGs due to thawing and overall lake/wetland area increase (13% larger, $\sim 0.72 \pm 0.19$ Mkm²
- under RCP 8.5; Zhang et al. 2017); (2) flux through ice cracks during ice-cover (likely
- insignificant; Phelps, Peterson, & Jeffries, 1998); (3) contribution of precipitation, which might
- affect hydrostatic pressure and soil moisture and promote organic matter decomposition by
- methanogens (Zhang et al., 2017); and (4) inhibition of surface CH₄ ebullition and subsequent
- storage during ice-cover (Phelps, Peterson and Jeffries, 1998b).
- 539 Comparison to previous models. Rather than attempting to improve model precision by
- downscaling climate model output to be consistent with the re-analysis of weather data (e.g., Tan
- and Zhuang 2015), we evaluated the implementation of a simple model that may, in theory, be
- scaled as part of a climate model land surface scheme. This was achieved by using data extracted
- 543 directly from the GCM, as would be done under two-way coupling.

Table 4. Summary of simulated mean annual surface fluxes forced with GCM data under baseline (1976-

548 2005) and future scenarios (2071-2100) and observed mean annual surface fluxes from 2011 and 2012

(Sepulveda-Jauregui et al., 2015). Percentage change of average value from the baseline value shown inparentheses.

		Observed total surface
	Total simulated surface flux	flux 2011-2012 from
	(g CH ₄ m ⁻² yr ⁻¹)	June-July extrapolation
		$(g CH_4 m^{-2} yr^{-1})$
	(A)	(B)
	Toolik	Lake
Baseline	3.79 ± 1.54	2.0
2.6	$5.30 \pm 2.16 \; (1.52/40\%)$	
4.5	$6.24 \pm 2.42 \ (2.45/65\%)$	
8.5	8.46 ± 4.23 (4.67/123%)	
	Lake	e E5
Baseline	4.20 ± 1.52	1.4
2.6	$5.87 \pm 1.74 \; (1.67/40\%)$	
4.5	$7.02 \pm 2.12 \; (2.83/67\%)$	
8.5	9.61 ± 4.30 (5.41/129%)	
	Lake	e El
Baseline	8.43 ± 3.72	9.4
2.6	11.59 ± 4.82 (3.16/38%)	
4.5	13.66 ± 5.51 (5.23/62%)	
8.5	$19.10 \pm 9.46 \ (10.67/127\%)$	
	Lake	e E6
Baseline	8.65 ± 4.01	13.3
2.6	11.91 ± 5.54 (3.26/38%)	
4.5	13.99 ± 6.01 (5.35/62%)	
8.5	19.70 ± 10.43 (11.05/128%)	

GCM scenario	Mean air temperature (°C)
Baseline (1976-2005)	-7.96
2.6 (2071-2100)	-4.37 (3.59°C)
4.5 (2071-2100)	-2.86 (5.10°C)
8.5 (2071-2100)	0.67 (8.63°C)

555	Table 5. Average air temperature from GCM data under baseline (1976-2005) and future scenarios
556	(2071-2100). Temperature changes from baseline are shown in parentheses.

The present study is distinct from Tan and Zhuang (2015) in that differing equations were

required to estimate the ebullitive flux from shallower ponds ($Z_{mean} = \sim 1-3 \text{ m}$; E1, E6) and deeper lakes ($Z_{mean} = \sim 6-7 \text{ m}$; Table 1). They followed a similar approach to this study (water

and sediment thermodynamics, gas and bubble transport and a sediment gas module with CH₄

562 dynamics) to simulate five Arctic lakes (thermokarst lakes, yedoma/non-yedoma in

563 continuous/discontinuous permafrost), including Toolik Lake (the methods are described in Tan

et al. (2015)). Their maximum lake depths and surface areas, were similar to ours, ranging

between 2.9 to 25 m and 1 to 149 ha; however, no validation of CH₄ concentrations in Toolik

Lake and Goldstream Lake (64.9°N 147.7°W; maximum depth 2.9 m, area 1.0 ha) was performed. Therefore, it was not possible to directly compare our model results with theirs.

568 Moreover, their C_{labile}, was calibrated against data from a single shallow lake (Suchi Lake, 69°N

569 161 °E, maximum depth 11 m, area 5.8 ha), which is likely more productive than the deeper

570 systems (Bastviken et al., 2004; Wik, Varner and Walter Anthony, et al., 2016) and could result

571 in higher CH₄ fluxes. In the present study, C_{labile} was not altered between the baseline and future

572 simulations, under the assumption that temperature was the sole driver of increased flux (Eq. 4).

573 Changes in lake productivity may occur, particularly as permafrost melts, however, these effects

are beyond the scope of the present work.

575 The CH₄ flux estimates simulated in this study can be considered conservative (Table 4). We

note that GHG emissions from water bodies are not implemented in CMIP5, which means the

577 contribution of these natural systems to climate change is underestimated (Zhang et al., 2017),

578 causing both future warming and consequent CH₄ emissions to be greater than modelled herein

579 with CMIP5 forcing.

580 *Comparison to regional and global estimates.* Many studies extrapolate measured CH₄ fluxes

from a limited number of water bodies to enable regional or global-scale emission estimates

582 (Bastviken et al., 2004; DelSontro et al., 2018; Wik, Varner and Walter Anthony, et al., 2016).

583 This approach often employs short-term sampling that neglects the temporal dynamics of

584 processes that govern CH₄ production, oxidation, ebullition, gas exchange at the water-air

585 interface and oxic CH₄ production (Wik, Thornton and Bastviken, et al., 2016). Consequently,

there are significant discrepancies in total CH₄ emissions estimates within the literature. For

example, the process-based model by Tan and Zhuang (2015) estimated present-day fluxes from

lakes north of 60°N to be 23.66 Tg CH₄ yr⁻¹ for a total lake area of 1.24×10^6 km², while

observation-based (22 freshwater ecosystems) extrapolation results from Bastviken et al., (2011)

590 gave 13.4 Tg CH₄ yr⁻¹ for lakes north of 54°N and a total area of 1.82×10^6 km². In contrast,

591 Wik, Varner and Walter Anthony, et al., (2016) estimated 16.5 ± 9.2 Tg CH₄ yr⁻¹ (total lake area

⁵⁵⁷

of 1.84×10^6 km²), using data from 733 northern lakes and updated lake-area estimates, including annual emissions of 8.3 ± 4.7 Tg CH₄ yr⁻¹ (total lake area of 1.45×10^6 km²) from all glacial/postglacial lakes.

Our study was limited to simple parameterizations of CH₄ fluxes from only four lakes; however, 595 596 we were able to compute CH₄ emissions year-round on sub-daily timescales, as they responded to changes in simulated sediment temperature. Extrapolating the mean flux from the four 597 simulated Alaskan lakes, using 1.45×10^6 km² of glacial/post-glacial lake area (Wik, Varner and 598 Walter Anthony, et al., 2016), gives total baseline (1976–2005) CH₄ emissions of 9.1 Tg CH₄ yr⁻ 599 ¹, which falls within 10% of the average updated estimate from Wik, Varner and Walter 600 Anthony, et al., (2016). While these comparisons must be interpreted with caution, given our 601 small sample size of 4 lakes, this result is interesting given that smaller lakes tend to have larger 602 areal CH₄ emissions (Wik et al., 2002) and our largest lake, (Toolik) is in the fourth smallest 603 lake-size class out of nine (Downing et al., 2006). Therefore, we likely overestimate the 604 extrapolated flux, which lies in the lower range of the estimate from Wik, Varner and Walter 605 Anthony, et al., (2016). This flux overestimation, based on lake size, may be compensated for by 606

607 our underestimation relative to observed fluxes (Table 4).

Our simulations show that the effect of climate change on areal (per m^2) CH₄ fluxes was more

609 pronounced in the shallower lakes (E1 and E6), in agreement with observations Wik, Varner and

610 Walter Anthony, et al., (2016); however, overall emissions were much higher from the deeper

lakes (Toolik and E5) due to their significantly larger surface area (e.g., Downing et al., 2006).

For instance, the annual lake-wide surface flux (ebullition and diffusion) from Toolik Lake and

Lake E5, respectively, were 5.64 and 0.46 Mg CH_4 yr⁻¹ during 1976-2005. For RCP 8.5, emissions increased ~2 times for both lakes (to 12.60 and 1.04 Mg CH_4 yr⁻¹, respectively: Table

emissions increased ~2 times for both lakes (to 12.60 and 1.04 Mg CH₄ yr⁻¹, respectively: Table 4). Fluxes from lakes E1 and E6, were 0.24 and 0.16 Mg CH₄ yr⁻¹, respectively, under baseline

conditions, and also increased ~ 2 times under RCP 8.5 (to 0.55 and 0.37 Mg CH₄ yr⁻¹,

617 respectively; Table 4). Notably, CH₄ emissions from these freshwater systems will remain above

the 1976-2005 averages, even with a conservative estimate of climate-driven warming (RCP

619 2.6), which indicates the urgency to include the influence of these natural systems in GHG

620 budgets and land surface schemes.

621 *Comparison of simulated CH*₄ *emissions with observations.* The differences between our

simulations and observations (Table 4), can be partially attributed to the methods used to

estimate seasonal and annual emissions. Sepulveda-Jauregui et al. (2015) note that there is

uncertainty in their measurements since it is assumed that single-day diffusion measurements are

representative of the entire ice-free season. However, they argue that this is the best (only)

available data, and these are well-established practices, as described in Bastviken et al., (2011).

627 This is to say that significant uncertainty amongst reported estimates of CH₄ emissions remains,

which does not allow for an accurate comparison. The methods used to quantify CH₄ fluxes

would ideally be near-continuous and evenly distributed in space. Certainly, attempts to improve

spatial and temporal sampling techniques in the field and laboratory will assist future studies to

obtain a more accurate representation of the overall CH₄ potential of lakes across the globe,
 especially regarding ebullitive fluxes, which are highly episodic.

633 *Errors from neglecting downscaling*. The GCM-forced model simulated lake temperature

profiles – at times – showed notable differences from the observations (Table 3), particularly at

the lake bottom (Figure 2), relative to when the model was forced with observed meteorological
data. To assess the impacts of the discrepancy in simulated bottom lake temperature, it is

- 637 important to compare the differences ebullitive and diffusive fluxes calculated under the different
- model forcing conditions. For 2004 (Jul. 5 to Aug. 19) in Lake E5, a difference of 5.31 °C was
- modelled between the average bottom temperature calculated with the observed meteorological
- 640 data and the long-term GCM-forced simulation (6.77°C and 12.09°C, respectively; Figure 2).
- For this period, the average ebullitive fluxes (Eq. 1) were 2.36 mg CH₄ m⁻² d⁻¹ (observed)
- 642 meteorological forcing) and 13.00 mg CH₄ m⁻² d⁻¹ (long-term GCM forcing). The seasonal
- average diffusive fluxes (Eq. 4) were 3.55 mg CH₄ m⁻² d⁻¹ (observed meteorological forcing) and $CH_4 = \frac{2}{3} \ln^2 (1 + 1) \ln^2 (1$
- 644 39.84 mg $CH_4 m^{-2} d^{-1}$ (long-term GCM forcing). This warm bias (Figure 2) led to an 645 overestimation of the surface CH_4 flux (1.40 g $CH_4 m^{-2} yr^{-1}$ observed in 2011–2012 vs. 4.20 g
- $CH_4 \text{ m}^{-2} \text{ yr}^{-1}$ simulated over 1976-2005). To minimize these effects on the interpretation of the
- results, we computed percentage changes between the GCM-forced baseline and GCM-forced
- 648 future projections (Table 6), considering that both simulations are subject to the bias and
- aggregation error. This approach is well established for assessing the simulated impacts of
- climate change (e.g., Plummer et al. 2006).

Model scalability and transferability. The temperature and CH₄ models were designed to have scalable parameterizations (which can be applied to many lakes based on a minimal number of readily available lake parameters, e.g., Table 1). However, transferability of the models to other lakes on the land surface, without the need for recalibration must be tested. Mackay (2012) found differences in lake surface temperatures to largely result from differences in surface area, depth,

- and transparency (all scalable model parameters; Table 1). Moreover, transferability of CSLM
 was shown where simulations of Canadian boreal, Alaskan Arctic and Swedish boreal lakes had
- some error but did not require parameter tuning or other model adjustments, which is essential
- for regional or global application. The present study was not designed to test model
- 660 transferability. We found that bottom temperatures were dependent on the surface wind drag (as
- 661 controlled by adjusting the wind sensor height). This may be a direct result of using a single set
- of observed meteorological data for all the lakes, which cover a ~5 km geographical region
- (Figure 1) and CRCM forcing data from a 50 km grid (e.g., variability in wind sheltering) or
- 664 poor specification of sediment heating (MacKay 2019). Therefore, application of the model to
- more lakes should be undertaken to better understand the need for individual lake calibration to
- 666 improve transferability in simulating sediment temperatures.
- The CH₄ model was more transferable than the temperature model. For the diffusive flux, the same C_{labile} , K_z and k_{ox} were used with all four lakes (Table 2). However, for the two larger and
- deeper lakes (Toolik and E5, $Z_{mean} \sim 6-7$ m and 10–150 ha), the ebullitive fluxes were estimated
- 670 from Arctic lake data using Eq. 1, which is a scalable parameterization based on the simulated
- sediment temperature (Table 1). However, this parameterization was inappropriate for the smaller shallower lakes (E1 and E6, $Z_{mean} \sim 1-3$ m and 2–3 ha), which required an ebullition
- equation developed for shallow boreal ponds. Observations show shallower systems (ponds $< \sim 3$
- 674 m depth) to have CH₄ fluxes ~ 10 times greater than deeper systems (lakes $\sim 3-30$ m depth),
- 675 supporting our usage of a different parameterization for ponds (Wik, Varner and Walter
- Anthony, et al. 2016). Further research is required to test if Eq. 1 is transferable to lakes larger
- than Toolik and E5. Ebullition flux equations should also be developed for shallower Arctic
- lakes like E1 and E6 (Aben et al., 2017).

679 **5** Conclusions

This study evaluated the ability of a computational one-dimensional CH4 model to estimate 680 historical, near present-day and future emissions from four Alaskan lakes, when forced with raw 681 (not downscaled) output data from a GCM. The GCM overestimated the bottom lake 682 temperature, which was subsequently used in the CH₄ sub-model. Simulated temperature error 683 metrics against observations were better when using observed meteorological forcing, compared 684 to raw GCM forcing (RMSE = 0.38 to 0.90 °C vs. 1.03 to 3.50 °C; $R^2 = 0.97$ to 0.99 vs. 0.74 to 685 0.81). There was no discernable difference in metrics for 2004 GCM vs. 1976-2005 GCM 686 forcing, indicating long term model drift was not an obvious source of error. Similar to previous 687 studies (Mackay, 2012), we found that CLASS simulated temperatures to be transferable (~1 °C 688 RMSE) between lakes without re-calibration; however, site specific adjustment of the wind 689 stress was required. This was also the case for the CH₄ fluxes from the deeper lakes (~6–7 m 690 691 mean depth); however, the shallower ($\sim 1-3$ m mean depth) required an ebullition equation derived from Boreal ponds. Future work should focus on improving simulation of bottom water 692 (sediment) temperatures, accounting for hypsometry that includes sediment area in littoral zones 693

694 where ebullition is maximal and developing transferable CH_4 flux parameterizations for shallow 695 Arctic lakes ($\sim 1-3$ m mean depth).

696 The three climate warming scenarios (RCPs 2.6, 4.5 and 8.5) all resulted in a significant increase

in total CH₄ emissions averaged over 2071-2100, relative to 1976-2005 (38-129%). However,

698 RCP 2.6 and 4.5 will lead to stabilized Arctic lake temperatures and CH₄ emissions by 2100, at

levels 0.61 - 1.21 °C and 38-67%, respectively, above the 1976–2005 averages. Overall

emissions from the larger two lakes (6–7 m mean depth and 10–150 ha) were modelled to be

higher in comparison to those from the smaller two systems (1-3 m mean depth and 2-3 ha) due

to their larger surface area; however, areal fluxes were larger from the smaller shallower lakes.
 Results from this work corroborate the urgent need to include the contributions of GHGs from

freshwater systems in regional and global climate models and associated positive feedback with

705 increased sediment temperatures.

Table 6. Summary of simulated mean ± standard deviation of model output forced with GCM data under baseline (1976-2005) and future

scenarios (2071-2100). Temperature and absolute/percentage change from baseline for CH₄ fluxes and ice-free days are shown in parentheses.

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	Bottom lake temperature (°C)	Sediment diffusive flux (mg CH4 m ⁻² d ⁻¹)	Sediment ebullitive flux (mg CH4 m ⁻² d ⁻¹)	Total surface flux (mg CH ₄ m ⁻² d ⁻¹)	Ice-free days
	(A)	(B)	(C)	(D)	(E)
	<i>Toolik Lake</i> $(k_{Ox} = 1$	$.74 \times 10^{-5} \mathrm{s}^{-1}$			
Baseline	3.61 ± 0.70	42.03 ± 3.88	1.65 ± 0.41	10.38 ± 1.54	110.86 ± 8.63
2.6	$4.24 \pm 0.78 \ (0.63)$	$47.86 \pm 4.76 \ (5.83/14\%)$	$2.68 \pm 0.60 \; (1.04/63\%)$	$14.53 \pm 2.16 \; (4.15/40\%)$	$128.83 \pm 6.70 \ (17.97/16\%)$
4.5	$4.55\pm 0.68\;(0.94)$	51.55 ± 3.92 (9.52/23%)	$3.39 \pm 0.66 \; (1.74/106\%)$	$17.10 \pm 2.42 \ (6.73/65\%)$	$134.97 \pm 9.39 \ (24.10/22\%)$
8.5	5.43 ± 0.66 (1.82)	$62.10\pm 6.36~(20.07/48\%)$	$5.21 \pm 0.96 \; (3.57/217\%)$	$23.17 \pm 4.23 \ (12.79/123\%)$	154.66 ± 10.98 (43.79/40%)
	<i>Lake E5</i> (k _{Ox} = 1.74	$\times 10^{-5} \mathrm{s}^{-1}$)			
Baseline	3.16 ± 0.47	40.24 ± 3.02	1.76 ± 0.50	11.50 ± 1.52	111.76 ± 8.65
2.6	3.77 ± 0.44 (0.61)	$45.87 \pm 3.37 \ (5.62/14\%)$	$2.77 \pm 0.55 \ (1.01/57\%)$	$16.08 \pm 1.74 \ (4.57/40\%)$	$129.83 \pm 6.57 \ (18.07/16\%)$
4.5	$4.28 \pm 0.51 \ (1.12)$	$50.68 \pm 3.74 \; (10.44/26\%)$	$3.51 \pm 0.64 \ (1.75/99\%)$	19.24 ± 2.12 (7.74/67%)	$135.34 \pm 8.90 \; (23.59/21\%)$
8.5	$5.12 \pm 0.57 (1.95)$	$61.28 \pm 6.25 \ (21.03/52\%)$	$5.43 \pm 1.04 \; (3.67/208\%)$	$26.32 \pm 4.30 \ (14.81/129\%)$	154.93 ± 11.26 (43.17/39%)
	<i>Lake E1</i> ($k_{Ox} = 1.74$	$\times 10^{-5} \mathrm{s}^{-1}$			
Baseline	3.41 ± 0.42	43.94 ± 4.12	10.92 ± 1.75	23.09 ± 3.72	105.55 ± 9.42
2.6	$4.25 \pm 0.42 \ (0.84)$	52.19 ± 5.73 (8.25/19%)	14.41 ± 2.83 (3.48/32%)	31.76 ± 4.82 (8.66/38%)	$123.24 \pm 6.15 \ (17.69/17\%)$
4.5	4.61 ± 0.51 (1.20)	$57.20 \pm 6.09 \ (13.27/30\%)$	$16.74 \pm 3.01 \ (5.82/53\%)$	37.42 ± 5.51 (14.33/62%)	$129.21 \pm 9.42 \ (23.66/22\%)$
8.5	5.55 ± 0.63 (2.14)	$71.00 \pm 9.42 \ (27.07/62\%)$	23.52 ± 5.12 (12.59/115%)	52.33 ± 9.46 (29.24/127%)	148.86 ± 11.57 (43.31/41%)
	<i>Lake E6</i> (k _{Ox} = 1.74	$\times 10^{-5} \mathrm{s}^{-1}$)			
Baseline	3.30 ± 0.37	44.09 ± 4.23	11.11 ± 1.86	23.69 ± 4.01	103.24 ± 9.57
2.6	$4.12 \pm 0.39 \ (0.82)$	$52.46 \pm 6.10 \ (8.36/19\%)$	14.71 ± 3.08 (3.59/32%)	$32.63 \pm 5.54 \ (8.94/38\%)$	$122.28 \pm 5.85 \ (19.03/18\%)$
4.5	4.51 ± 0.43 (1.21)	57.60 ± 6.31 (13.51/31%)	17.10 ± 3.23 (5.99/54%)	$38.34 \pm 6.01 \ (14.64/62\%)$	127.83 ± 9.31 (24.59/24%)
8.5	5.55 ± 0.58 (2.24)	$72.03 \pm 10.05 \ (27.94/63\%)$	24.23 ± 5.75 (13.12/118%)	53.98 ± 10.43 (30.29/128%)	147.41 ± 11.85 (44.17/43%)

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718 Data Availability Statement

- 719 Weather data, temperature time-series, CH₄ time-series were obtained from ARC LTER, NSF
- 720 Arctic Data Center site (<u>https://arcticdata.io/catalog/view/doi:10.18739/A2X54S</u>) and the
- 721 Environmental Data Center site of the Toolik Field Station in Alaska (https://arc-
- 122 <u>lter.ecosystems.mbl.edu/;</u> refer to reference list for specific yearly data). The second version of
- the CSLM source code is available at <u>https://github.com/MurrayMackay/CSLM</u>. All remaining
- relevant data supporting the conclusions of this study are included in the article and
- supplementary material. Further inquiries can be directed to the corresponding author.

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Prediction of future Alaskan lake methane emissions using a small-lake model coupled to a regional climate model

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12 Key Points:

- CH₄ emissions from lakes are not presently in the land surface schemes of Global
 Climate Models
- A one-dimensional lake model was developed to simulate future CH₄ diffusive and ebullitive fluxes from four Arctic lakes
- Three climate warming scenarios simulated the bottom water temperature to warm by up to 2.24°C, increasing the surface CH₄ flux from the four lakes by 38 129%.
- 19

Abstract 20

Methane emissions from lakes will increase with climate warming. However, CH₄ emissions 21

from lakes are not presently in the land surface schemes of Global Climate Models (GCMs). 22

Modelled climate projections depend on future atmospheric CH₄ concentrations; therefore, a 23

positive feedback loop is not simulated. To address this issue, a one-dimensional lake model was 24

- 25 developed to simulate future CH₄ diffusive and ebullitive fluxes from four Arctic lakes. The
- model was hindcast for validation (1976-2005) and forecast for prediction (2071-2100) with one-26 way coupling to raw meteorological data from the CanESM2 ensemble GCM. Three climate 27
- warming scenarios (RCPs 2.6, 4.5 and 8.5) simulated the bottom water temperature to warm by 28
- up to 2.24°C, increasing the surface CH₄ flux from the four lakes by 38 129%. However, RCP 29
- 2.6 and 4.5 led to stabilized temperatures and CH₄ emissions by 2100, at levels of 0.63 1.21°C 30
- and 38 67%, respectively, above the 1976-2005 averages. The CH₄ diffusion parameterization 31
- was transferable between the four lakes; however, different ebullition parameterizations were 32
- required for the two deeper lakes (~6-7 m mean depth) versus the two shallower lakes (~1-3 m 33

mean depth). Relative to using observed meteorological forcing, which had a cold bias (-0.15 to -34

0.63 °C) and RMSE of 0.38 to 0.90 °C, the GCM-forced models had a warm bias (+0.96 to 35

+3.13°C) and marginally higher RMSE (1.03 to 3.50°C) compared to observations. The results 36

support continued efforts to couple CH₄ lake-emission models to GCMs without downscaling 37

38 meteorological data, allowing feedback between CH₄ dynamics and future climates to be modelled.

- 39
- 40

Plain Language Summary 41

Climate change in the Arctic is moving at a greater rate than in the rest of the world. The urgency 42

of characterizing greenhouse gas emissions from water bodies in high latitudes has become a 43

subject of intensive research during the last two decades. It is believed that methane (CH4) 44

emissions from freshwater systems are the most important source of uncertainty in the global 45 greenhouse gas budget, and their contribution has been excluded from earth systems models. 46

Previous models have been developed to determine these emissions; however, they often require 47

several inputs and lake characteristics that are not readily available. Considering this, we 48

developed a subroutine to determine CH4 fluxes and concentrations in the water column for four 49

lakes located in Arctic Alaska. Subsequently, we implemented meteorological data from a 50

climate model to predict future CH4 fluxes. The calculated increases in atmospheric fluxes were 51

significant; emissions from all the studied lakes are expected to increase at least 38% over the 52

next 80 years. Our study presents a simple formulation with limited constraints to estimate CH4 53

emissions. We expect that our subroutine could be embedded into climate models to predict 54

- 55 emissions from the Arctic and potentially from the rest of the globe.
- 56

58 1 Introduction

59

60 Atmospheric CH₄ emissions from lakes form a large portion of the global greenhouse gas (GHG)

budget (10-16%; Bastviken et al., 2011), with lake sediments being a reservoir for mineral and

62 organic carbon, which is released as CH₄ (as well as CO₂; Dean and Gorham, 1998; Zhang et al.,

63 2017). CH₄ emissions from lakes vary with seasonal changes in temperature and these

64 fluctuations are primary drivers of intra-annual variation in global CH₄ emissions (Greene et al.,

65 2014; Wik, Varner, and Walter Anthony, et al., 2016). Under the RCP 2.6 emission scenario

from 2080 - 2100, higher lake surface water temperatures and longer ice-free seasons are

expected globally (2.5°C and 15 days, respectively; Woolway and Merchant 2019); with
 warming 2-3 times more severe in the Arctic (Graversen et al., 2008; Belkin, 2009; Bintanja,

warming 2-3 times more severe in the Arctic (Graversen et al., 2008; Belkin, 2009; Bir
2018).

70 Rapid Arctic warming will increase CH₄ emissions by thawing permafrost, but will also increase

the rate of CH_4 release from lakes through enhanced microbial decomposition in the sediments

(Walter Anthony and Anthony, 2013). Higher temperatures will also decrease the duration of ice

rover, which is relevant as ice promotes dissolution of CH₄ bubbles into the water column, and

⁷⁴ subsequent accumulation under ice (Denfeld et al., 2018). A shorter duration of ice cover will

result in increased ebullition and diffusion; although increased oxidation is also expected,

renhanced CH₄ productivity may exceed this loss (Greene et al., 2014; Martinez-Cruz et al.,

77 2015).

78 The increased ebullitive and diffusive CH₄ fluxes, from temperature-driven enhancement of

79 productivity in anoxic sediments (Zeikus and Winfrey, 1976; Walter Anthony et al., 2006; Zhang

et al., 2017), suggests a positive feedback loop, because model projections of future climate

change depend strongly on the atmospheric concentration of GHGs (Harriss et al., 1993; Whalen,

2005). Small lakes are presently not resolved in climate models (MacKay et al., 2009), and so

83 there is a need to include this positive feedback from Arctic lake systems into climate model land

- 84 surface schemes.
- 85 Correspondingly, efforts have been made to incorporate emissions from freshwater systems into
- 86 regional and global GHG budgets. Observations from individual lakes have been up-scaled
- according to lake-size distributions (Bastviken, Ejlertsson and Tranvik, 2002; Bastviken et al.,
- 2004; Wik, Varner and Walter Anthony, et al., 2016; DelSontro, Beaulieu and Downing, 2018).
- 89 Process-based models have also been developed to simulate CH₄ dynamics in freshwater systems
- 90 (Tan, Zhuang and Walter Anthony, 2015; Stepanenko et al., 2016); however, they require lake
- 91 characteristics and calibration variables that are not typically measured.

92 Tan and Zhuang (2015) coupled a lake model to downscaled output from a climate model (one-

way coupling) to project present and future CH₄ emissions from Arctic lakes. Their results

showed that the emissions will roughly double by the end of the 21st century. However, because

of the bias-correction and interpolation applied in their downscaling, their models are not able to

96 be run as part of a climate model land surface scheme that would enable positive feedback

97 between CH₄ emissions and increased future temperatures.

- 98 The purpose of the present study is to develop a scalable CH₄ emission model (Hurtado Caicedo,
- 99 2019) that may be coupled to the one-dimensional (1D) Canadian Small Lake Model (CSLM;

- 100 MacKay 2012, MacKay et al., 2017). Although not part of the present study, these models can be
- 101 embedded in the land surface scheme (CLASS) of the Canadian Regional Climate Model
- 102 (CRCM; Verseghy and MacKay, 2017) and hence would be able to capture the feedback
- between CH_4 emissions and climate change. The specific objectives are to (1) assess the ability
- of the CSLM-CH₄ model to estimate present water temperatures and CH₄ emissions, from four
- Arctic lakes, without downscaling the CRCM output used as surface forcing; and (2) project
- 106 future CH₄ emissions under various GHG emission scenarios for the four Arctic lakes.

107 2 Methods

108 2.1 Study sites

- 109 Four Arctic kettle lakes near Toolik Field Station (68°38' N, 149°39' W) were modelled: Toolik
- Lake and Lakes E1, E5 and E6 (Figures 1, S1 and S2). The lakes are located in the continuous
- 111 permafrost zone on the North Slope of Alaska in the Brooks Range foothills region. These water
- bodies are part of the Arctic Long-Term Ecological Research (ARC LTER) program and have
- been widely studied in order to understand the impact of environmental change on their physical
- and chemical properties (O'Brien, 1992; MacIntyre et al., 2006; Walter Anthony et al., 2008;
- 115 Jorgenson et al., 2010; Hobbie and Kling, 2014).
- 116 Lakes characteristics (surface area, bathymetry and extinction coefficients) were available from
- 117 MacIntyre et al. (2018; Table 1). Given the predominantly flat-bottomed bowl shape of kettle
- 118 lakes (Fig. S1), simulations resolved from the free surface to the mean (Z_{mean}), as opposed to the
- 119 maximum (Z_{max}) lake depth (see MacKay, 2012).



120



124 **2.2 Forcing and validation data**

125 2.2.1 Observed meteorological forcing

126 Meteorological data was collected by Kling (2000) at the surface of Toolik Lake during the ice-

127 free seasons (June - September) of 2001-2005. Incident shortwave and longwave radiation were

- measured with a Kipp and Zonen CM3 radiation sensor at ~50 cm height above the lake surface.
- 129 Air temperature and relative humidity were measured with Vaisala HMP45C (2.5-3 m above the
- water surface), and wind speed was measured with a Met One 014A anemometer (3.8-5 m above
 the water surface, Table 2) and corrected to 10-m using a logarithmic model (Holmes, 2001).
- Given their proximity (Figure 1), the same meteorological forcing was applied to all four lakes.

133 2.2.2 Modelled meteorological forcing

- Baseline (1976-2005) and future (2071-2100) global climate model (GCM) forcing data were
- obtained from the simulations of the Canadian Earth System Model (CanESM2/CGCM4), which
- 136 was run as part of CMIP5 with a 50-km grid resolution. The emission scenarios included the
- 137 Representative Concentration Pathways 2.6 (RCP 2.6), 4.5 (RCP 4.5) and 8.5 (RCP 8.5). RCP
- 138 2.6 is a low greenhouse gas emission scenario in which changes to the radiative forcing would
- 139 lead to an increase in the global mean temperature of 1°C. Under RCP 2.6, global carbon
- emissions are projected to decrease to near zero by the end of 2100. Under the medium
- emissions scenario, RCP 4.5, changes to the radiative forcing cause a 2°C increase in air
- temperature under stabilized GHG emissions, and under the high emission scenario RCP 8.5,
- 143 changes to the radiative forcing result in an increase of 3.7°C in air temperature with a
- 144 continuous growth of GHG emissions (IPCC, 2013).

145 **2.2.3 Observed temperature time series**

- 146 Water temperature time-series at Toolik Lake and Lakes E5 and E6 were collected by S.
- 147 MacIntyre (ARC LTER; <u>https://arc-lter.ecosystems.mbl.edu</u>) during the summers of 2001-2005.
- 148 Water temperature was measured with moored, self-contained loggers (see Table S1 for logger
- 149 depths). In Toolik Lake RBR Ltd. TR-1050s (±0.002°C) and in Lakes E5 and E6, StowAway
- 150 Tidbit Loggers ($\pm 0.21^{\circ}$ C) were used.

151 2.2.4 Observed CH₄ fluxes

- 152 CH₄ ebullitive fluxes from Toolik Lake in 2004 were digitally retrieved from Tan et al., (2015;
- their Figure 10). They collected gas samples using submerged bubble traps over ebullition seeps
- along defined transects in the lake from April 28, 2003 to December 31, 2004. Subsequently,
- 155 lake-wide daily ebullition was determined as the sum of the fluxes from each seep type and
- 156 averaged over the lake surface area.
- 157
- 158 Continuous diffusion and ebullition data for lakes E1, E5 and E6 were not available; however,
- aggregate values of seasonal and mean annual fluxes were used as given in Sepulveda-Jauregui
- 160 et al. (2015). They determined the mean annual fluxes as the sum of different modes of emission
- 161 in the summer (ice-free season), winter and spring. Sepulveda-Jauregui et al. (2015) performed
- 162 measurements from June to July (2011–2012) and extrapolated the value to the entire ice-free
- 163 season. They calculated ebullitive fluxes by multiplying the average seep densities on each lake
- by the sum of daily ebullition from seeps of each type. We compared our simulated surface

diffusive fluxes from Toolik Lake during the summer seasons of 2010–2015 to measurements by

Eugster et al. (2020a), who deployed a three-dimensional ultrasonic anemometer-thermometer and a closed-path integrated off-axis cavity output spectrometer.

174	Table 1. Lake characteristics. Z _{max} and Z _{mean} are maximum and mean lake depths respectively, and k _d is
175	the extinction coefficient (MacIntyre, Cortés and Sadro, 2018).

	Latitude	Longitude	Area	Zmax	Zmean	k _d
	(°N)	(°W)	(km ²)	(m)	(m)	(m ⁻¹)
Toolik Lake	68.633	-149.607	1.49	26.0	7.4	0.6
Lake E1	68.626	-149.555	0.029	12.0	3.1	0.8
Lake E5	68.642	-149.458	0.109	12.9	6.4	1.1
Lake E6	68.643	-149.441	0.019	3.0	1.6	1.4

Table 2. Calibration parameters for each lake. The wind sensor height and hypolimnion turbulent

diffusivity were adjusted, on a lake-by-lake basis, to account for sub-grid-scale turbulent mixing and for

spatial variability in wind speed and surface drag (which is computed internally in the model). Clabile and

k_{ox}were adjusted to minimize CH₄ RMSE.

	Wind sensor height (m)	Clabile (mg m ⁻²)	Hypolimnion turbulent diffusivity, K _Z (m ² s ⁻¹)	CH4 oxidation k _{ox} (s ⁻¹)
	(A)	(B)	(C)	(D)
Range	3.8 - 10	90 - 280	$10^{-5} - 10^{-9}$	10 ⁻⁵ -10 ⁻⁸
Toolik Lake	3.8	280	10-7	1.74×10 ⁻⁵
Lake E1	5.0	280	10-7	1.74×10^{-5}
Lake E5	5.0	280	10-7	1.74×10^{-5}
Lake E6	3.8	280	10-7	1.74×10 ⁻⁵

187 2.2.5 Observed CH₄ concentrations

188 CH₄ concentration profiles were measured in Toolik Lake and lakes E1, E5 and E6 during the

ice-free seasons of 2013-2016 by MacIntyre and Cortés (2017). CH₄ was measured using the
 headspace equilibrium method and gas chromatography. A summary of the collected data and

191 sources specified in section 2.2 is presented in Table S2.

192 2.3 Modelling framework

193 Climate models are typically validated against long-term averages (e.g., climate normals);

194 however, they also show skill in reproducing characteristic large-scale distributions of air

195 temperature, precipitation, radiation and wind. For example, the dynamics of monsoon systems,

seasonal temperature changes and storm tracks. Climate models can even predict weather over

seasonal timescales and capture interannual variability (Randall et al. 2007). As a result, output
 from climate models is frequently applied to drive lake models to forecast the impacts of climate

change on lake hydrodynamics and water quality (e.g., Woolway et al. 2020; Bolkhari et al.

200 2022; Golub et al. 2022). When run as hindcasts, it is expected that these GCM-forced models

will reasonably capture observed variation in meteorological conditions over seasonal to inter-

202 annual timescales.

203 To evaluate forcing-bias when developing coupled lake-atmosphere models, it is a requirement

to test the accuracy of the lake model when forced directly with atmospheric (or climate) model

output, in comparison to being forced with observed meteorological forcing (e.g., Huang et al.

206 2010). Consequently, simulation dates were selected based on the availability of both

207 meteorological forcing data sets (observations and GCM model output) and water temperature

calibration data (ARC LTER). Three different calibration scenarios were run: (1) with observed

meteorological forcing for the ice-free season of individual years (Toolik Lake: 2001-2005; Lake
E5: 2002-2005; Lake E6: 2003-2005; no lake temperature data was available for Lake E1); (2)

with GCM-generated meteorological data for the ice-free season of individual years; and (3) with

GCM-generated meteorological data for the fee-free season of individual years, and (5) with GCM-generated meteorological data in a continuous 'baseline' run from 1976-2005. Simulations

(1) and (2) were to assess the error in applying GCM forcing without downscaling; simulation

214 (3) was to test for long-term model drift.

The GCM output is a statistical representation of the atmospheric conditions in each year, and

thus does not necessarily capture the exact timing of particular meteorological events (e.g., the

217 passage of a cold front; Figures S4-S6). Therefore, here we quantitatively compared mean

seasonal observed temperature profiles, as opposed to conventional contours of temperature

time-series with depth. The latter type of comparisons is given in Hurtado Caicedo (2019) and

220 Figure S7.

221 Temperature profiles for model initial conditions were not available from ARC LTER in 1976;

therefore, Toolik was initialized with historic data (Giblin and Kling, 2015), while E5 and E6

used initial profiles from 2002 and 2003, respectively. For simulations with future GCM data

224 (2071-2100), initial water temperatures were from observations in 2013. Initial CH₄ profiles

225 were as reported by MacIntyre and Cortés (2017). For long-term simulations, it is expected that

the model will lose knowledge of the initial conditions within the first year of simulation (1977

for the baseline run and 2072 for the future simulations; Figure S3), hence model results during

this time were disregarded.

229 2.3.1 Model description

- 230 The approach to determine CH₄ emissions consisted of two parts: (1) implementation of a pre-
- existing thermodynamic lake-tile model to calculate the surface mixed layer depth and unsteady
- temperature profile through the water column, and (2) the development of a new CH₄ subroutine
- to compute CH₄ ebullition, dissolved CH₄ flux from the sediments, diffusive CH₄ flux below the
- surface mixed layer and through the air-water interface.
- 235 *Temperature:* Temperature profiles were simulated using the 1D (vertical) thermodynamic
- 236 CSLM lake-tile model, which incorporates a bulk mixed-layer turbulent kinetic energy budget to
- simulate a 1 m² water column, characteristic of a lake with a specified surface area, mean depth
- and light extinction coefficient (Table 1; MacKay, 2012; MacKay et al., 2017). We acknowledge
- that a mean depth approach underestimates water temperature in the littoral zone and
- overestimates water temperature at depth. We justify this by noting that the resulting lower modelled fluxes in shallow zones and higher modelled fluxes at depth will partially cancel and
- that the objective here is to develop a feasible lake-tile CH_4 model and not to simulate CH_4
- fluxes as accurately as possible. The hypolimnion turbulent diffusivity K_z and wind sensor height
- were adjusted, on a lake-by-lake basis (Table 2), to account for variation in sub-grid-scale
- turbulent mixing and for spatial variability in wind speed and surface drag (which is computed
- internally in the model).
- 247 The model had $\Delta z = 0.5$ m vertical grid resolution and used 5 or 15-minute timesteps (Δt) when
- forced with observed and GCM-derived meteorological data, respectively. Ice cover was
- modelled according to the snowpack physics module of CLASS, which was shown to reproduce
- 250 ice-on/off dates to within 1 week of observations in Lake 239 (MacKay et al. 2017).
- 251 *CH₄ ebullition:* To model CH₄ emissions within a land-surface scheme, model parameterizations
- 252 must be a function of variables readily obtained from GCM output, which limits model
- complexity. We parameterize CH₄ ebullition for the deeper ($Z_{mean} = -6-7$ m; Toolik and E5) and
- shallower ($\mathbb{Z}_{\text{mean}} = \sim 1-3 \text{ m}; \text{ E1}, \text{ E6}$) lakes according to Eq. 1 and 2, respectively, which are

$$E = -0.00036T^3 + 0.16T^2 - 0.94T + 1.48$$
 (1)

$$E = 141 \times 1.19^{(T-20)}$$
(2)

257 The above expressions correlate the average ebullitive flux E (mg $CH_4 m^{-2} d^{-1}$) to the average

sediment surface temperature T (°C); assumed to be the modelled bottom water temperature.

259 CSLM is a water column model, that does not account for the distribution of sediment surface

area with lake depth. The model does not resolve littoral sediments, with higher temperatures

that contribute more ebullitive flux during summer (Bastviken et al., 2004; Wik, Varner, and

Walter Anthony, et al., 2016). To account for this, we use mean depth hypsometry and depth-

based parameterizations for the shallower (E1 and E6; Eq. 2) and deeper (Toolik and E5; Eq. 1)
lakes.

The parameterization in Eq. 1 was from 6806 ebullitive flux observations in three sub-Arctic lakes in northern Sweden (68°21′ N, 19°02′ E) during the ice-free seasons (June–September) of

- 267 2009-2014 (Wik et al., 2013, 2014; Wik, 2016). The parameterization in Eq. 2 (Fig. 1 and
- equation 2 in Aben et al., 2017) was developed for shallower Boreal ponds based on
- 269 measurements by DelSontro et al. (2016), including 77 observations during the ice-free season
- 270 (May–October of 2011, 2012 and 2014) from ten shallow ponds located in the Saguenay region
- of Quebec (48°23'N, 71°25' W), and 83 observations from three lakes in the Laurentian region
- 272 of Quebec (45°59' N 73°89' W).

Our attempts using Eq. 1 to determine ebullitive fluxes for the shallower lakes resulted in

- underestimation of emissions in comparison to observations; likely because we do not explicitly
- 275 resolve littoral sediments. These values were similar in order of magnitude to the results obtained
- for Toolik Lake. Shallower systems (ponds $< \sim 3$ m depth) have been shown to have CH₄ fluxes
- ~10 times greater than deeper systems (lakes ~3-30 m depth), supporting our usage of a different
 parameterization for ponds versus lakes (Wik, Varner and Walter Anthony, et al. 2016). For
- instance, ebullitive fluxes from the shallower systems can be up to ~ 9 times greater than those
- from the deeper systems (see Table S3). The other equations in Aben et al., (2017) were also
- evaluated (not shown) but performed worse than Eq. 1 and 2.
- 282 Dissolution of rising bubbles was neglected because of the shallow lake depths (Schmid et al.

283 2007). The present model also does not inhibit ebullition through ice during winter. Ebullition

- would be stored within the ice (Walter Anthony et al., 2008; Walter Anthony et al., 2006) and subsequently released during break up (Phalma, Poterson and Jaffriag, 1008a; Juntinger et al.
- subsequently released during break-up (Phelps, Peterson and Jeffries, 1998a; Juutinen et al.,
 2009; Karlsson et al., 2013). Consequently, methanotrophy (oxidation) of dissolved CH₄ from
- rising bubbles was also neglected, since previous work suggests that for systems < 20 m deep,
- less than 10% of the CH₄ is lost due to dissolution during rise (McGinnis et al., 2006; Schmid et
- al., 2007). These limitations will be addressed in future work.

290 Dissolved CH_4 and diffusive sediment flux of CH_4 : The CH_4 profile in the water column was

simulated by numerically solving the 1D diffusion equation (Eq. 3) with terms for turbulent

diffusivity, oxidation and production at the sediment-water interface (e.g., Schmid et al., 2007;

293 Jabbari et al., 2016):

$$\frac{\partial C_{CH_4}}{\partial t} = \frac{\partial}{\partial z} \left(K_z \frac{\partial C_{CH_4}}{\partial z} \right) - k_{Ox} (C_{CH_4}) + P$$
⁽³⁾

- Here, C_{CH_4} (mg L⁻¹) is the CH₄ concentration, K_Z (m² s⁻¹) is the vertical turbulent diffusivity 294 below the surface mixed layer (Table 2; $\sim 10^{-7}$ m² s⁻¹; Nakhaei et al. 2016), k_{Ox} (s⁻¹) is the first-295 order CH₄ oxidation rate coefficient (1.74×10^{-5} s⁻¹; Thottathil et al., 2019), P (mg CH₄ m⁻³ d⁻¹) is 296 the production term (Eq. 4), t is time and z is the vertical coordinate direction. We acknowledge 297 that K_Z, in these relatively shallow lakes, is likely orders of magnitude larger than this near-298 molecular value; however, the turbulent diffusivity and/or dissipation in a calibrated sub-grid-299 300 scale closure scheme is often not equal to observed values (Boegman et al., 2021; Lin et al., 2022). 301
- 302 Our first-order oxidation model was developed following Schmid et al. (2007). From a
- sensitivity analysis $(10^{-5} \text{ s}^{-1} < \text{k}_{\text{Ox}} < 10^{-8} \text{ s}^{-1})$, we selected the observed value $\text{k}_{\text{Ox}} \sim 1.74 \times 10^{-5} \text{ s}^{-1}$ by
- Thottathil et al. (2019) from 6 lakes in Quebec (Canadian Shield). For a CH₄ concentration of

- 305 $C_{CH_4} \sim 0.1 \ \mu \text{mol } \text{L}^{-1}$ (Figs S8-S11), the resulting oxidation was consistent with observed rates 306 $(\sim 10^{-6} \text{ to } \sim 10^{-5} \ \mu \text{mol } \text{L}^{-1} \text{ s}^{-1})$ from shallow Alaskan lakes (Lofton et al. 2014).
- 307 The oxidation term was applied throughout the water column and the turbulent diffusion term
- 308 was applied to drive CH₄ flux from the base of the surface mixed layer to the bottom cell. CH₄
- 309 production from sediments P (mg m⁻² s⁻¹) was applied as a flux into the cell above the sediment-
- 310 water interface (Tan et al. 2015):

$$P = R_C C_{labile} P Q_{10} \left(\frac{T - T_{pr}}{10}\right)$$
(4)

- 311 where R_{C} (0.02 s⁻¹) is the fraction of carbon converted per year, PQ_{10} (3.5) is the factor by which
- production increases with a 10°C rise in temperature, T_{pr} is the reference temperature for CH₄
- production, which is approximately equal to the yearly mean sediment temperature below the
- 314 water column (-3.0 °C) and T (°C) is the sediment surface temperature (Tan et al., 2015);
- assumed to equal the modelled bottom water temperature at the mean depth. Here, C_{labile} (mg m⁻
- ²) is the areal labile carbon density. To directly model the carbon pool requires soil incubation
- data and knowledge of the thickness of the talik layer (Tan et al., 2015). Unlike the lake-specific
- parameters in Table 1, these parameters are not known across the CRCM domain; therefore, we estimated $C_{labile} = 280 \text{ mg m}^{-2}$ (Table 2), by minimizing the difference between the observed and
- modelled CH₄ concentrations in the water column. Ideally, for the model to be scalable, C_{labile}
- will be the same for all lakes or a function of the lake-specific parameters in Table 1. This
- approach is not new, and it what is typically done to set the sediment oxygen demand in
- biogeochemical lake models (e.g., Scalo, Boegman and Piomelli, 2013).
- With Eq. 4, the model may account for both the ¹⁴C-enriched carbon pool from the upper
- 325 sediment layer constituted by newly settled organic matter, since both Toolik Lake and Lake E5
- are considered nonyedoma/nonthermokarst lakes (Stepanenko et al., 2011, 2016; Sepulveda-

Jauregui et al., 2015; Tan and Zhuang, 2015). According to Tan, Zhuang and Walter Anthony

- 328 (2015), CH₄ in nonyedoma lakes is mainly produced in surface sediments from newly deposited
- 329 ¹⁴C-enriched organic matter.
- 330 The constants R_{c} , PQ_{10} and T_{pr} were as specified by Zhuang et al. (2004) for the alpine tundra
- and polar desert ecosystem of the Toolik area. To account for productivity in the upper sediment
- layers (Peeters, Encinas Fernandez and Hofmann, 2019), and based on the shallow bowl-like
- bathymetry of the smaller lakes (Fig. S1 and S2), P was added fractionally at each depth of water
- 334 (Schwefel et al., 2018); this was accomplished by calculating P and dividing the result by the
- number of layers in the water column (14, 13, 6 and 4 layers for Toolik, E5, E1 and E6
- 336 respectively).

337 *Diffusive flux of CH*₄ *to the atmosphere.* The rate of CH₄ transfer across the water-air interface 338 was specified following Happell et al. (1995) and Schmid et al. (2007):

$$\mathbf{F}_{\mathrm{CH}_{4}} = \mathbf{f}(\mathbf{W})\mathbf{K}_{\mathrm{ex}}(\mathbf{C}_{\mathrm{W}} - \mathbf{C}_{\mathrm{A}}) \tag{5}$$

- 339 where K_{ex} (m s⁻¹) is the gas transfer velocity (O'Connor, 1983), f(W) is a coefficient that
- determines the influence of wind speed (equal to $1 + 0.058W^2$ for $W \le 5$ m s⁻¹ and $1 + 0.047W^2$
- for $W \ge 5 \text{ m s}^{-1}$; Schmid et al., 2007), C_w (mg L⁻¹) is the modelled CH₄ concentration at the water surface and C_A (mg L⁻¹) is the atmospheric equilibrium CH₄ concentration (Wiesenburg and
- Guinasso, 1979), which is a function of the surface temperature and the measured atmospheric
- CH₄ concentration (1700 ppb from 1976–2005 and projected to be on average 1300, 1600 and
- 345 3600 ppb under RCP scenarios 2.6, 4.5 and 8.5, respectively, from 2071–2100; Meinshausen et
- al., 2011). The diffusive CH₄ flux to the atmosphere was restricted during ice-cover period and
- 347 the accumulated gas was subsequently released into the atmosphere at ice-off.

348 **2.4 Flux aggregation**

The model computed CH₄ ebullitive and diffusive fluxes at 5 min intervals, when forced with

- observed meteorological data, and 15 min intervals, when forced with GCM-generated data.
- 351 Mean daily and mean annual fluxes were averages of these outputs over each day or year. Total
- surface fluxes refer to the sum of ebullitive and diffusive fluxes. These were compared to
- observed ebullitive and diffusive fluxes collected as described in section 2.2.4. We acknowledge that spatial and temporal extrapolation of the observations may introduce error (e.g., they neglect
- that spatial and temporal extrapolation of the observations may introduce error (e.g., they neglect background non-seep emissions); however, these were the best available data (Sepulveda-
- Jauregui et al. 2015) and followed well-established practices (Bastviken et al. 2011). For Toolik
- Lake, daily ebullitive fluxes from point source seeps in the lake centre, were digitally retrieved
- from (Tan et al., 2015), averaged over July 4 to August 18, 2004 and compared to the modelled
- 359 average over the same period.

360 **3 Results**

361 **3.1 Model calibration (observed meteorological forcing)**

- The coupled CSLM-CH₄ model (Hurtado Caicedo, 2019) was developed and calibrated for all
- four lakes over a 3 to 4-year duration simulation (Figures S8-S11) using the observed
- 364 meteorological data. Simulated temperature profiles had depth-dependent root-mean-square
- errors (RMSE) <3.2 °C, which were highest through the thermocline, where small differences in the simulated thermocline depth can lead to larger RMSE (e.g., Boegman and Sleep, 2012).
- the simulated thermocline depth can lead to larger RMSE (e.g., Boegman and Sleep, 2012).
 These were in the range of literature values, from large lake model applications (3 < RMSE < 7)
- ³⁶⁷ These were in the range of interature values, non-range rake moder appreations (3 < RWSE ³⁶⁸ °C; Huang et al., 2010; Paturi et al., 2012) and other pond simulations (RMSE < 3.03 °C;
- Nakhaei et al., 2018). The surface RMSE <1.5 °C was less than 1.96 °C for other small boreal
- 370 lake simulations (Stepanenko et al., 2016) and comparable to 1.5 °C from other Arctic lake
- 371 simulations (Tan et al., 2015).
- 372 The dissolved CH₄ concentrations were $< 0.6 \mu$ M (Figures S8-S10), which is reasonable in
- 373 comparison to errors from more complex CH_4 models (Stepanenko et al., 2011, 2016; Tan et al., 2015) The DMCF f_1 is the E1 (10.2 M) F_2 (10.2 M)
- 2015). The RMSE for Lakes E1 (< 0.2 μ M), E5 (< 0.05 μ M) and Toolik Lake (< 0.03 μ M) were
- consistent with the mean errors found by Tan et al., (2015) and Stepanenko et al. (2016), which ranged from 0.01 µM to 0.26 µM. However, the PMSE for Lake E6 were larger, ranging from
- ranged from 0.01 μ M to 0.26 μ M. However, the RMSE for Lake E6 were larger, ranging from 0.3 μ M to 0.6 μ M.
- $377 \quad 0.3 \mu M$ to $0.6 \mu M$.
- The simulated mean-daily summer ebullitive fluxes from Toolik Lake in 2004 were within 2% of measured values (7.04 vs. 6.91 g m⁻² d⁻¹, respectively: Table S3). Although ebullition

observations were not available for the other lakes (E1, E5 and E6) during the same years as

observed meteorological forcing (2013-16), observed ebullition data from 2011-12 were

included for comparison (Table S3). The simulated mean summer fluxes were 19% (9.53 mg CH₄ m⁻² d⁻¹), 21% (0.72 mg CH₄ m⁻² d⁻¹) and 30% (25.74 mg CH₄ m⁻² d⁻¹) lower than observed

 $_{383}$ CH4 II $^{-}$ d $^{-}$), 21% (0.72 IIg CH4 II $^{-}$ d $^{-}$) and 30% in E1, E5 and E6 over 2011-12.

The simulated mean-daily summer surface diffusive fluxes from Toolik Lake in 2015 and 2013 385 were within 20% of measured values by Eugster et al. (2020a; Table S4). As with the observed 386 387 ebullitive fluxes, observed diffusive fluxes were not available for lakes E1, E5 and E6 during 2014–2015, and so were compared to observed data from other years. For lakes E5 and E1, the 388 observed fluxes were larger than simulated value, whereas for E6 the simulated flux was larger 389 (Table S4). These data should be interpreted with caution, as the 2012 diffusive surface flux 390 observed by Sepulveda-Jauregui et al. (2015) from Toolik Lake is ~3 times larger than the value 391 observed by Eugster et al. (2020a). Therefore, values reported by Sepulveda-Jauregui et al. 392

393 (2015) for the shallower systems may also be overestimated.

394 **3.2** Hindcast simulations (1976-2005 GCM forcing)

395 Simulated temperatures: Simulations forced with GCM data, initialized in individual years, were

better in comparison to observations, relative to the long-term GCM-forced simulations (Figure

2); except Toolik (in 2001) and E5 (in 2003 and 2004). The GCM-forced model had a systematic shift to higher temperatures in some years ($\leq \sim 4$ °C in Toolik, Fig. 2d; ~ 5 °C in E5), where the

seasonal thermocline intersected the lakebed. The shallower Lake E6, had more satisfactory

400 results (<2 °C bias). The temperature overestimation, simulated with the GCM, resulted from

401 higher incident longwave radiation in the GCM compared to the observed data (23 W m⁻² higher

402 on average; Table S5), warmer air temperatures (1 °C higher on average, with exception of 2005)

and higher wind speeds (1.5 m s^{-1} higher on average; Table S5). The GCM-forcing causes higher

404 temperatures and deepening of the thermocline relative to observations, and thus higher sediment

temperatures when the thermocline approaches the lakebed (mean depth).

In 2004, there was observed temperature data (Toolik, E5 and E6), as well as forcing for all three

simulations (forced with observed 2004 meteorology, forced with 2004 GCM meteorology and

forced with 1976-2005 GCM meteorology). Intercomparison of performance metrics (Table 3)

409 show RMSE (0.38 to 0.87 °C vs. 1.03 to 3.50 °C) and NRMSE (0.03 to 0.08 vs. 0.07 to 0.30) to

be marginally less for observed forcing, compared to GCM forcing. The observed and GCM

forced simulations had cold (+0.01 to -0.63 $^{\circ}$ C) and warm (+0.96 to +3.13 $^{\circ}$ C) biases,

412 respectively. The high R^2 (0.99 for observed forcing and 0.74-0.99 for GCM forcing) indicates

that the observed temperatures are explained by the models. In 2004, the GCM underestimated

observed shortwave radiation by 9.3%, overestimated longwave radiation by 6.4% and air

temperature by 4.6% (Table S5). However, it remains difficult to generalize this error as GCM

biases are different from year to year (Table S5). There was no discernable difference in metrics
for the 2004 GCM vs. 1976-2005 GCM forcing, indicating that long-term model drift was not an

418 obvious source of error.

419





Figure 2. Comparison of measured time-averaged water temperature with simulations. Black stars are measured temperature, blue lines show the simulations with observed meteorological data, red lines show simulations with GCM data and initial conditions at the beginning of the simulation period for each year, and green lines show simulations with GCM data starting in 1976. Temperature validation were from Jul. 16 to Aug. 11 in 2001, Jul. 2 to Aug. 10 in 2002 and Jun. 28 to Aug. 18 in 2004.

427

428 Simulated CH₄ fluxes: The baseline (1976-2005) average annual surface flux (sum of ebullition

- 429 and diffusion; Table 4A) was compared to the average total annual 2011-2012 emission
- 430 observations reported by Sepulveda-Jauregui et al. (2015). The simulated baseline fluxes for the
- 431 shallower lakes (Table 4A; from 8.43 ± 3.72 g CH₄ m⁻² yr⁻¹ in E1 to 8.65 ± 4.01 g CH₄ m⁻² yr⁻¹ in
- E6) were within 34% of the observations (Table 4B; from 9.4 g $CH_4 m^{-2} yr^{-1}$ in E1 to 13.3 g CH_4
- m^{-2} yr⁻¹ in E6). The observed total fluxes were higher in Lake E6 than in Lake E1, which was
- 434 captured by our model (Table 4). For the shallower lakes, the modelled surface fluxes were

- smaller than observed fluxes, which could result from Sepulveda-Jauregui et al. (2015)
- 436 extrapolating emissions from June-July observations, when sediment temperatures were much
- 437 higher, to the entire ice-free season (considered to be May-September) and therefore likely
- 438 overestimating fluxes (Section 3.1).
- The modelled baseline fluxes for the deeper lakes (Table 4A; from 3.79 ± 1.54 g CH₄ m⁻² yr⁻¹ in
- Toolik to 4.20 ± 1.52 g CH₄ m⁻² yr⁻¹ in E5) were larger than the observations from Sepulveda-
- 441 Jauregui et al. (2015), for 2011-2012 (Table 4B; from 2.0 g $CH_4 m^{-2} yr^{-1}$ in Toolik to 1.4 g CH_4
- 442 $m^{-2} yr^{-1}$ in E6). This discrepancy can be attributed to the warm bias of the GCM forced models,
- which was more prominent in the deeper systems (Table 3; Figure 2).

444 **3.3** Forecast simulations (2071-2100 GCM forcing)

- *Air temperature.* Under RCP scenarios 2.6, 4.5 and 8.5, the air temperature in the Toolik Lake
- 446 area was modelled to increase 3.6–7.3°C (Figure 3; Table 6) over the next 80 years. This is
- 447 consistent with the rates that have been reported for the entire Arctic over the past century
- 448 (Graversen et al., 2008; Belkin, 2009; Bintanja, 2018). RCP 8.5 shows a strong and continuous
- 449 increase in air temperature, which is more abrupt than under baseline conditions, RCP 4.5 has a
- 450 consistently increasing trend and RCP 2.6 shows a decline in air temperature over 2071-2100.



- Figure 3. Comparison of baseline and future simulations of average yearly air temperature. Baseline data
 correspond to GCM data in years 1976 2005 (black line) and future data correspond to GCM data in
 years 2071 2100 under RCPs, 2.6 (blue line), 4.5 (green line) and 8.5 (red line).
- 455 Bottom lake temperature. The simulated increases in bottom-water temperature (from 1976-2005
- to 2071–2100) were similar between the lakes (Table 6A; Figure 4 and Figure 5). For RCP 2.6,
- 457 4.5 and 8.5, the average bottom temperatures increased 0.61–0.84°C, 0.94–1.21°C and 1.82–
- 458 2.24°C, respectively, over 2071–2100 relative to the 1976–2005 baseline. Inter-annual variability
- in average bottom temperature was ~ 2 °C, with a more pronounced overall increase in bottom
- 460 lake temperature under RCP 8.5 (Figures 4 and 5).

Sediment CH₄ diffusion. Average diffusive fluxes from the lake sediments over 2071–2100 461

ranged between 40.24–72.03 mg CH₄ m⁻² d⁻¹ (under all RCPs). Lake E5 had the lowest fluxes of 462

40.24 - 61.28 mg CH₄ m⁻² d⁻¹ (Table 6 and Figure 4). The shallower lakes E1 and E6 had similar 463 increases in the sediment flux for all scenarios (increasing 8.25 - 27.94 mg CH₄ m⁻² d⁻¹ or 19–

464 63% from 1976-2005 to 2071-2100, Table 6B). These were larger than those of the deeper lakes 465

(increasing 5.62–21.03 mg CH₄ m⁻² d⁻¹ or 14–52%, Table 6B; Figures 4 and 5) and may be 466

associated with increased sediment warming (Table 6A). 467

CH₄ ebullitive flux. The simulated CH₄ ebullitive flux was most sensitive to the increases in lake 468 water temperature (Table 6C; Figures 4 and 5), with future fluxes more than doubling over the 469 baseline emissions under the high emission RCP 8.5 scenario in the deeper systems (increasing 470

3.57-3.67 mg CH₄ m⁻² d⁻¹ or 208–217%, Table 6B, Figure 4). In the shallower lakes, the higher 471

ebullitive fluxes had a smaller percentage change (115–118%), but larger absolute change 472

(12.59–13.12 mg CH₄ m⁻² d⁻¹, Table 6B), due to Eq. 2 being more sensitive to changes in 473 ebullition with temperature changes (Table 6A). This confirms the more significant impact of

474

climate change on increasing CH₄ ebullition (Aben et al., 2017; Walter et al., 2007; Wik, Varner 475 and Walter Anthony, et al., 2016) relative to increases in diffusion, particularly for shallow 476

systems. Ebullition comprised 15% to 47% of the total CH₄ flux from the four lakes (ebullition

477 plus diffusion under all scenarios), 68% of this total flux was produced by the two shallower 478

lakes (E1 and E6; Table 6C). 479

Ice-free days. The number of ice-free days increased from 103–111 (1976–2005) to 122–155 480

(2071–2100) (Table 6E). The number of ice-free days was estimated to be up to 19 days longer 481 under RCP 2.6 and up to 44 days longer under RCP 8.5. The baseline model (1976-2005)

482 estimates were consistent with the mean reported ice-free duration increase for glacial and post-483

glacial lakes of 20 days (Wik, Varner and Walter Anthony, et al., 2016). 484

Total surface CH4 flux. The total flux was modelled to increase by 4.15-4.57 mg CH4 m⁻² d⁻¹ 485

(40%; RCP 2.6), 6.73–7.74 mg CH₄ m⁻² d⁻¹ (65–67%; RCP 4.5) and 12.79–14.81 mg CH₄ m⁻² d⁻¹ 486

(123–129%; RCP 8.6) for the deeper lakes (Toolik and E5) and 8.66–8.94 mg CH₄ m⁻² d⁻¹ (38%; 487 RCP 2.6), 14.33–14.64 mg CH₄ m⁻² d⁻¹ (62%; RCP 4.5) and 29.24–30.29 CH₄ m⁻² d⁻¹ (127–

488 128%: RCP 8.6) for the shallower lakes (E1 and E6). Tan and Zhuang (2015) estimated similar 489

surface fluxes across the Arctic to increase 87% and 137%, under RCPs 2.6 and 8.5 respectively. 490

Trends over 2071-2100. The water temperature and CH₄ fluxes increased from the 1976-2005 491

492 average to the 2071-2100 average under all RCP scenarios, following the increases in GCM-

modelled air temperature. However, this did not always cause increasing trends in the simulated 493

494 temperature and CH₄ data during 2071–2100. For all lakes, during 2071–2100, RCP 8.5 showed

strong increases in bottom water temperature and CH₄ fluxes; RCP 4.5 showed minimal 495

increases in bottom water temperatures (<~1 °C) and flat trends in CH₄ production; and RCP 2.6 496

showed decreased bottom water temperatures (<~1 °C) and trends of decreased CH₄. This 497

498 suggests that RCP 2.6 and 4.5 will lead to stabilized Arctic lake temperatures and CH₄ emissions

by 2100, at levels 0.61–1.21 °C and 38–67%, respectively, above the 1976–2005 averages. 499

500





Figure 4. Comparison of baseline and future simulations for the deeper lakes (Toolik and E5; ~6-7 m
 mean depth), of yearly average A) bottom water temperature B) CH₄ diffusive fluxes at the sediment water interface C) CH₄ ebullition from sediments and D) total surface CH₄ fluxes.

- 506
- 507





Figure 5. Comparison of baseline and future simulations, for the shallower lakes (E1 and E6; \sim 1-3 m 513 mean depth), of yearly average A) bottom water temperature B) CH₄ diffusive fluxes at the sediment-514 water interface C) CH₄ ebullition from sediments and D) total surface CH₄ fluxes.

- 517 **Table 3.** Statistical metrics comparing (*i*) 2004 simulation forced with observed meteorology, (*ii*) 2004
- simulation forced with GCM meteorology and (*iii*) 1976-2005 simulation forced with GCM meteorology
- to (iv) observed temperatures in 2004. This was the only year with observation data and forcing data for all
- 520 3 simulations; observed data was not available for Lake E1 in 2004. The GCM data is a statistical
- 521 representation of the atmosphere and is not expected to reproduce observed meteorology on a day-to-day
- 522 basis; therefore, before computing the metrics, the higher spatial resolution field observations were 523 interpolated onto the 0.5 m depths of the model grids, then the observed and simulated temperature data
- were averaged over day-of-year 186 to day-of-year 230, resulting in a single representative mean
- temperature profile for cases (*i*) to (*iv*). RMSE = root-mean-square error of simulation vs. observation;
- 526 NRMSE = RMSE/mean(observation); Bias = sum(observation-simulation)/# observations; R^2 =
- 527 coefficient of determination.

	RMSE	NRMSE	Bias	\mathbf{R}^2		
	(°C)		(°C)			
	1	oolik Lake				
Observed 2004	0.90	0.07	-0.15	0.97		
GCM 2004	1.98	0.16	+1.86	0.76		
GCM 1976-2005	2.39	0.19	+2.28	0.80		
		Lake E5				
Observed 2004	0.87	0.08	-0.63	0.99		
GCM 2004	2.67	0.23	+2.38	0.81		
GCM 1976-2005	3.50	0.30	+3.13	0.77		
	Lake E6					
Observed 2004	0.38	0.03	-0.34	0.99		
GCM 2004	1.36	0.09	+1.31	0.77		
GCM 1976-2005	1.03	0.07	+0.96	0.74		

528 4 Discussion

- 529 We have modelled CH₄ emissions from Arctic lakes to increase as a function of increased future
- air temperature and corresponding increased simulated lake bottom water temperature. In this
- approach, the following aspects were not included in the model development and may have an
- impact on simulated future emissions: (1) release of carbon pools contained in permafrost soil as
- 533 GHGs due to thawing and overall lake/wetland area increase (13% larger, $\sim 0.72 \pm 0.19$ Mkm²
- under RCP 8.5; Zhang et al. 2017); (2) flux through ice cracks during ice-cover (likely
- insignificant; Phelps, Peterson, & Jeffries, 1998); (3) contribution of precipitation, which might
- affect hydrostatic pressure and soil moisture and promote organic matter decomposition by
- methanogens (Zhang et al., 2017); and (4) inhibition of surface CH₄ ebullition and subsequent
- storage during ice-cover (Phelps, Peterson and Jeffries, 1998b).
- 539 Comparison to previous models. Rather than attempting to improve model precision by
- downscaling climate model output to be consistent with the re-analysis of weather data (e.g., Tan
- and Zhuang 2015), we evaluated the implementation of a simple model that may, in theory, be
- scaled as part of a climate model land surface scheme. This was achieved by using data extracted
- 543 directly from the GCM, as would be done under two-way coupling.

Table 4. Summary of simulated mean annual surface fluxes forced with GCM data under baseline (1976-

548 2005) and future scenarios (2071-2100) and observed mean annual surface fluxes from 2011 and 2012

(Sepulveda-Jauregui et al., 2015). Percentage change of average value from the baseline value shown inparentheses.

		Observed total surface			
	Total simulated surface flux	flux 2011-2012 from June-July extrapolation			
	(g CH ₄ m ⁻² yr ⁻¹)				
		$(g CH_4 m^{-2} yr^{-1})$			
	(A)	(B)			
	Toolik Lake				
Baseline	3.79 ± 1.54	2.0			
2.6	$5.30 \pm 2.16 \; (1.52/40\%)$				
4.5	$6.24 \pm 2.42 \ (2.45/65\%)$				
8.5	8.46 ± 4.23 (4.67/123%)				
	Lake E5				
Baseline	4.20 ± 1.52	1.4			
2.6	$5.87 \pm 1.74 \; (1.67/40\%)$				
4.5	$7.02 \pm 2.12 \; (2.83/67\%)$				
8.5	9.61 ± 4.30 (5.41/129%)				
	Lake E1				
Baseline	8.43 ± 3.72	9.4			
2.6	11.59 ± 4.82 (3.16/38%)				
4.5	13.66 ± 5.51 (5.23/62%)				
8.5	$19.10 \pm 9.46 \ (10.67/127\%)$				
	Lake E6				
Baseline	8.65 ± 4.01	13.3			
2.6	11.91 ± 5.54 (3.26/38%)				
4.5	$13.99 \pm 6.01 \ (5.35/62\%)$				
8.5	$19.70 \pm 10.43 \; (11.05/128\%)$				

GCM scenario	Mean air temperature (°C)
Baseline (1976-2005)	-7.96
2.6 (2071-2100)	-4.37 (3.59°C)
4.5 (2071-2100)	-2.86 (5.10°C)
8.5 (2071-2100)	0.67 (8.63°C)

555	Table 5. Average air temperature from GCM data under baseline (1976-2005) and future scenarios
556	(2071-2100). Temperature changes from baseline are shown in parentheses.

The present study is distinct from Tan and Zhuang (2015) in that differing equations were

required to estimate the ebullitive flux from shallower ponds ($Z_{mean} = \sim 1-3 \text{ m}$; E1, E6) and deeper lakes ($Z_{mean} = \sim 6-7 \text{ m}$; Table 1). They followed a similar approach to this study (water

and sediment thermodynamics, gas and bubble transport and a sediment gas module with CH₄

562 dynamics) to simulate five Arctic lakes (thermokarst lakes, yedoma/non-yedoma in

563 continuous/discontinuous permafrost), including Toolik Lake (the methods are described in Tan

tet al. (2015)). Their maximum lake depths and surface areas, were similar to ours, ranging

between 2.9 to 25 m and 1 to 149 ha; however, no validation of CH₄ concentrations in Toolik

Lake and Goldstream Lake (64.9°N 147.7°W; maximum depth 2.9 m, area 1.0 ha) was performed. Therefore, it was not possible to directly compare our model results with theirs.

568 Moreover, their C_{labile}, was calibrated against data from a single shallow lake (Suchi Lake, 69°N

569 161 °E, maximum depth 11 m, area 5.8 ha), which is likely more productive than the deeper

570 systems (Bastviken et al., 2004; Wik, Varner and Walter Anthony, et al., 2016) and could result

571 in higher CH₄ fluxes. In the present study, C_{labile} was not altered between the baseline and future

572 simulations, under the assumption that temperature was the sole driver of increased flux (Eq. 4).

573 Changes in lake productivity may occur, particularly as permafrost melts, however, these effects

are beyond the scope of the present work.

575 The CH₄ flux estimates simulated in this study can be considered conservative (Table 4). We

note that GHG emissions from water bodies are not implemented in CMIP5, which means the

577 contribution of these natural systems to climate change is underestimated (Zhang et al., 2017),

578 causing both future warming and consequent CH₄ emissions to be greater than modelled herein

579 with CMIP5 forcing.

580 *Comparison to regional and global estimates.* Many studies extrapolate measured CH₄ fluxes

from a limited number of water bodies to enable regional or global-scale emission estimates

582 (Bastviken et al., 2004; DelSontro et al., 2018; Wik, Varner and Walter Anthony, et al., 2016).

583 This approach often employs short-term sampling that neglects the temporal dynamics of

584 processes that govern CH₄ production, oxidation, ebullition, gas exchange at the water-air

585 interface and oxic CH₄ production (Wik, Thornton and Bastviken, et al., 2016). Consequently,

there are significant discrepancies in total CH₄ emissions estimates within the literature. For

example, the process-based model by Tan and Zhuang (2015) estimated present-day fluxes from

lakes north of 60°N to be 23.66 Tg CH₄ yr⁻¹ for a total lake area of 1.24×10^6 km², while

observation-based (22 freshwater ecosystems) extrapolation results from Bastviken et al., (2011)

590 gave 13.4 Tg CH₄ yr⁻¹ for lakes north of 54°N and a total area of 1.82×10^6 km². In contrast,

591 Wik, Varner and Walter Anthony, et al., (2016) estimated 16.5 ± 9.2 Tg CH₄ yr⁻¹ (total lake area

⁵⁵⁷

of 1.84×10^6 km²), using data from 733 northern lakes and updated lake-area estimates, including annual emissions of 8.3 ± 4.7 Tg CH₄ yr⁻¹ (total lake area of 1.45×10^6 km²) from all glacial/postglacial lakes.

Our study was limited to simple parameterizations of CH₄ fluxes from only four lakes; however, 595 596 we were able to compute CH₄ emissions year-round on sub-daily timescales, as they responded to changes in simulated sediment temperature. Extrapolating the mean flux from the four 597 simulated Alaskan lakes, using 1.45×10^6 km² of glacial/post-glacial lake area (Wik, Varner and 598 Walter Anthony, et al., 2016), gives total baseline (1976–2005) CH₄ emissions of 9.1 Tg CH₄ yr⁻ 599 ¹, which falls within 10% of the average updated estimate from Wik, Varner and Walter 600 Anthony, et al., (2016). While these comparisons must be interpreted with caution, given our 601 small sample size of 4 lakes, this result is interesting given that smaller lakes tend to have larger 602 areal CH₄ emissions (Wik et al., 2002) and our largest lake, (Toolik) is in the fourth smallest 603 lake-size class out of nine (Downing et al., 2006). Therefore, we likely overestimate the 604 extrapolated flux, which lies in the lower range of the estimate from Wik, Varner and Walter 605 Anthony, et al., (2016). This flux overestimation, based on lake size, may be compensated for by 606

607 our underestimation relative to observed fluxes (Table 4).

Our simulations show that the effect of climate change on areal (per m^2) CH₄ fluxes was more

609 pronounced in the shallower lakes (E1 and E6), in agreement with observations Wik, Varner and

610 Walter Anthony, et al., (2016); however, overall emissions were much higher from the deeper

lakes (Toolik and E5) due to their significantly larger surface area (e.g., Downing et al., 2006).

For instance, the annual lake-wide surface flux (ebullition and diffusion) from Toolik Lake and

Lake E5, respectively, were 5.64 and 0.46 Mg CH_4 yr⁻¹ during 1976-2005. For RCP 8.5, emissions increased ~2 times for both lakes (to 12.60 and 1.04 Mg CH_4 yr⁻¹, respectively: Table

emissions increased ~2 times for both lakes (to 12.60 and 1.04 Mg CH₄ yr⁻¹, respectively: Table 4). Fluxes from lakes E1 and E6, were 0.24 and 0.16 Mg CH₄ yr⁻¹, respectively, under baseline

conditions, and also increased ~ 2 times under RCP 8.5 (to 0.55 and 0.37 Mg CH₄ yr⁻¹,

617 respectively; Table 4). Notably, CH₄ emissions from these freshwater systems will remain above

the 1976-2005 averages, even with a conservative estimate of climate-driven warming (RCP

619 2.6), which indicates the urgency to include the influence of these natural systems in GHG

620 budgets and land surface schemes.

621 *Comparison of simulated CH*₄ *emissions with observations.* The differences between our

simulations and observations (Table 4), can be partially attributed to the methods used to

estimate seasonal and annual emissions. Sepulveda-Jauregui et al. (2015) note that there is

uncertainty in their measurements since it is assumed that single-day diffusion measurements are

representative of the entire ice-free season. However, they argue that this is the best (only)

available data, and these are well-established practices, as described in Bastviken et al., (2011).

627 This is to say that significant uncertainty amongst reported estimates of CH₄ emissions remains,

which does not allow for an accurate comparison. The methods used to quantify CH₄ fluxes

would ideally be near-continuous and evenly distributed in space. Certainly, attempts to improve

spatial and temporal sampling techniques in the field and laboratory will assist future studies to

obtain a more accurate representation of the overall CH₄ potential of lakes across the globe,
 especially regarding ebullitive fluxes, which are highly episodic.

633 *Errors from neglecting downscaling*. The GCM-forced model simulated lake temperature

profiles – at times – showed notable differences from the observations (Table 3), particularly at

the lake bottom (Figure 2), relative to when the model was forced with observed meteorological
 data. To assess the impacts of the discrepancy in simulated bottom lake temperature, it is

- 637 important to compare the differences ebullitive and diffusive fluxes calculated under the different
- model forcing conditions. For 2004 (Jul. 5 to Aug. 19) in Lake E5, a difference of 5.31 °C was
- modelled between the average bottom temperature calculated with the observed meteorological
- 640 data and the long-term GCM-forced simulation (6.77°C and 12.09°C, respectively; Figure 2).
- For this period, the average ebullitive fluxes (Eq. 1) were 2.36 mg CH₄ m⁻² d⁻¹ (observed)
- 642 meteorological forcing) and 13.00 mg CH₄ m⁻² d⁻¹ (long-term GCM forcing). The seasonal
- average diffusive fluxes (Eq. 4) were 3.55 mg CH₄ m⁻² d⁻¹ (observed meteorological forcing) and $CH_4 = \frac{2}{3} \ln^2 (1 + 1) \ln^2 (1$
- 644 39.84 mg $CH_4 m^{-2} d^{-1}$ (long-term GCM forcing). This warm bias (Figure 2) led to an 645 overestimation of the surface CH_4 flux (1.40 g $CH_4 m^{-2} yr^{-1}$ observed in 2011–2012 vs. 4.20 g
- $CH_4 \text{ m}^{-2} \text{ yr}^{-1}$ simulated over 1976-2005). To minimize these effects on the interpretation of the
- results, we computed percentage changes between the GCM-forced baseline and GCM-forced
- 648 future projections (Table 6), considering that both simulations are subject to the bias and
- aggregation error. This approach is well established for assessing the simulated impacts of
- climate change (e.g., Plummer et al. 2006).

Model scalability and transferability. The temperature and CH₄ models were designed to have scalable parameterizations (which can be applied to many lakes based on a minimal number of readily available lake parameters, e.g., Table 1). However, transferability of the models to other lakes on the land surface, without the need for recalibration must be tested. Mackay (2012) found differences in lake surface temperatures to largely result from differences in surface area, depth,

- and transparency (all scalable model parameters; Table 1). Moreover, transferability of CSLM
 was shown where simulations of Canadian boreal, Alaskan Arctic and Swedish boreal lakes had
- some error but did not require parameter tuning or other model adjustments, which is essential
- for regional or global application. The present study was not designed to test model
- 660 transferability. We found that bottom temperatures were dependent on the surface wind drag (as
- 661 controlled by adjusting the wind sensor height). This may be a direct result of using a single set
- of observed meteorological data for all the lakes, which cover a ~5 km geographical region
- (Figure 1) and CRCM forcing data from a 50 km grid (e.g., variability in wind sheltering) or
- 664 poor specification of sediment heating (MacKay 2019). Therefore, application of the model to
- more lakes should be undertaken to better understand the need for individual lake calibration to
- 666 improve transferability in simulating sediment temperatures.
- The CH₄ model was more transferable than the temperature model. For the diffusive flux, the same C_{labile} , K_z and k_{ox} were used with all four lakes (Table 2). However, for the two larger and
- deeper lakes (Toolik and E5, $Z_{mean} \sim 6-7$ m and 10–150 ha), the ebullitive fluxes were estimated
- 670 from Arctic lake data using Eq. 1, which is a scalable parameterization based on the simulated
- sediment temperature (Table 1). However, this parameterization was inappropriate for the smaller shallower lakes (E1 and E6, $Z_{mean} \sim 1-3$ m and 2–3 ha), which required an ebullition
- equation developed for shallow boreal ponds. Observations show shallower systems (ponds $< \sim 3$
- 674 m depth) to have CH₄ fluxes ~ 10 times greater than deeper systems (lakes $\sim 3-30$ m depth),
- 675 supporting our usage of a different parameterization for ponds (Wik, Varner and Walter
- Anthony, et al. 2016). Further research is required to test if Eq. 1 is transferable to lakes larger
- than Toolik and E5. Ebullition flux equations should also be developed for shallower Arctic
- lakes like E1 and E6 (Aben et al., 2017).

679 **5** Conclusions

This study evaluated the ability of a computational one-dimensional CH4 model to estimate 680 historical, near present-day and future emissions from four Alaskan lakes, when forced with raw 681 (not downscaled) output data from a GCM. The GCM overestimated the bottom lake 682 temperature, which was subsequently used in the CH₄ sub-model. Simulated temperature error 683 metrics against observations were better when using observed meteorological forcing, compared 684 to raw GCM forcing (RMSE = 0.38 to 0.90 °C vs. 1.03 to 3.50 °C; $R^2 = 0.97$ to 0.99 vs. 0.74 to 685 0.81). There was no discernable difference in metrics for 2004 GCM vs. 1976-2005 GCM 686 forcing, indicating long term model drift was not an obvious source of error. Similar to previous 687 studies (Mackay, 2012), we found that CLASS simulated temperatures to be transferable (~1 °C 688 RMSE) between lakes without re-calibration; however, site specific adjustment of the wind 689 stress was required. This was also the case for the CH₄ fluxes from the deeper lakes (~6–7 m 690 691 mean depth); however, the shallower ($\sim 1-3$ m mean depth) required an ebullition equation derived from Boreal ponds. Future work should focus on improving simulation of bottom water 692 (sediment) temperatures, accounting for hypsometry that includes sediment area in littoral zones 693

694 where ebullition is maximal and developing transferable CH_4 flux parameterizations for shallow 695 Arctic lakes ($\sim 1-3$ m mean depth).

696 The three climate warming scenarios (RCPs 2.6, 4.5 and 8.5) all resulted in a significant increase

in total CH₄ emissions averaged over 2071-2100, relative to 1976-2005 (38-129%). However,

698 RCP 2.6 and 4.5 will lead to stabilized Arctic lake temperatures and CH₄ emissions by 2100, at

levels 0.61 - 1.21 °C and 38-67%, respectively, above the 1976–2005 averages. Overall

emissions from the larger two lakes (6–7 m mean depth and 10–150 ha) were modelled to be

higher in comparison to those from the smaller two systems (1-3 m mean depth and 2-3 ha) due

to their larger surface area; however, areal fluxes were larger from the smaller shallower lakes.
 Results from this work corroborate the urgent need to include the contributions of GHGs from

freshwater systems in regional and global climate models and associated positive feedback with

705 increased sediment temperatures.

Table 6. Summary of simulated mean ± standard deviation of model output forced with GCM data under baseline (1976-2005) and future

scenarios (2071-2100). Temperature and absolute/percentage change from baseline for CH₄ fluxes and ice-free days are shown in parentheses.

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	Bottom lake temperature (°C)	Sediment diffusive flux (mg CH4 m ⁻² d ⁻¹)	Sediment ebullitive flux (mg CH4 m ⁻² d ⁻¹)	Total surface flux (mg CH4 m ⁻² d ⁻¹)	Ice-free days		
	(A)	(B)	(C)	(D)	(E)		
	Toolik Lake $(k_{Ox} = 1.74 \times 10^{-5} \text{ s}^{-1})$						
Baseline	3.61 ± 0.70	42.03 ± 3.88	1.65 ± 0.41	10.38 ± 1.54	110.86 ± 8.63		
2.6	$4.24 \pm 0.78 \ (0.63)$	$47.86 \pm 4.76 \ (5.83/14\%)$	$2.68 \pm 0.60 \; (1.04/63\%)$	$14.53 \pm 2.16 \ (4.15/40\%)$	$128.83 \pm 6.70 \ (17.97/16\%)$		
4.5	$4.55\pm 0.68\;(0.94)$	51.55 ± 3.92 (9.52/23%)	$3.39 \pm 0.66 \; (1.74/106\%)$	$17.10 \pm 2.42 \ (6.73/65\%)$	$134.97 \pm 9.39 \ (24.10/22\%)$		
8.5	5.43 ± 0.66 (1.82)	$62.10\pm 6.36~(20.07/48\%)$	$5.21 \pm 0.96 \; (3.57/217\%)$	$23.17 \pm 4.23 \ (12.79/123\%)$	154.66 ± 10.98 (43.79/40%)		
	<i>Lake E5</i> ($k_{Ox} = 1.74 \times 10^{-5} \text{s}^{-1}$)						
Baseline	3.16 ± 0.47	40.24 ± 3.02	1.76 ± 0.50	11.50 ± 1.52	111.76 ± 8.65		
2.6	3.77 ± 0.44 (0.61)	$45.87 \pm 3.37 \ (5.62/14\%)$	$2.77 \pm 0.55 \ (1.01/57\%)$	$16.08 \pm 1.74 \ (4.57/40\%)$	$129.83 \pm 6.57 \ (18.07/16\%)$		
4.5	$4.28 \pm 0.51 \ (1.12)$	$50.68 \pm 3.74 \; (10.44/26\%)$	$3.51 \pm 0.64 \ (1.75/99\%)$	19.24 ± 2.12 (7.74/67%)	$135.34 \pm 8.90~(23.59/21\%)$		
8.5	$5.12 \pm 0.57 (1.95)$	$61.28 \pm 6.25 \ (21.03/52\%)$	$5.43 \pm 1.04 \; (3.67/208\%)$	$26.32 \pm 4.30 \ (14.81/129\%)$	154.93 ± 11.26 (43.17/39%)		
	Lake E1 ($k_{Ox} = 1.74 \times 10^{-5} \text{s}^{-1}$)						
Baseline	3.41 ± 0.42	43.94 ± 4.12	10.92 ± 1.75	23.09 ± 3.72	105.55 ± 9.42		
2.6	$4.25 \pm 0.42 \ (0.84)$	52.19 ± 5.73 (8.25/19%)	14.41 ± 2.83 (3.48/32%)	31.76 ± 4.82 (8.66/38%)	$123.24 \pm 6.15 \ (17.69/17\%)$		
4.5	4.61 ± 0.51 (1.20)	$57.20 \pm 6.09 \ (13.27/30\%)$	$16.74 \pm 3.01 \ (5.82/53\%)$	37.42 ± 5.51 (14.33/62%)	$129.21 \pm 9.42 \ (23.66/22\%)$		
8.5	5.55 ± 0.63 (2.14)	$71.00 \pm 9.42 \ (27.07/62\%)$	23.52 ± 5.12 (12.59/115%)	52.33 ± 9.46 (29.24/127%)	148.86 ± 11.57 (43.31/41%)		
	<i>Lake E6</i> ($k_{Ox} = 1.74 \times 10^{-5} \text{ s}^{-1}$)						
Baseline	3.30 ± 0.37	44.09 ± 4.23	11.11 ± 1.86	23.69 ± 4.01	103.24 ± 9.57		
2.6	$4.12 \pm 0.39 \ (0.82)$	$52.46 \pm 6.10 \ (8.36/19\%)$	14.71 ± 3.08 (3.59/32%)	$32.63 \pm 5.54 \ (8.94/38\%)$	$122.28 \pm 5.85 \ (19.03/18\%)$		
4.5	4.51 ± 0.43 (1.21)	57.60 ± 6.31 (13.51/31%)	17.10 ± 3.23 (5.99/54%)	$38.34 \pm 6.01 \ (14.64/62\%)$	$127.83 \pm 9.31 \ (24.59/24\%)$		
8.5	5.55 ± 0.58 (2.24)	$72.03 \pm 10.05 \ (27.94/63\%)$	24.23 ± 5.75 (13.12/118%)	53.98 ± 10.43 (30.29/128%)	147.41 ± 11.85 (44.17/43%)		

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717

718 Data Availability Statement

- 719 Weather data, temperature time-series, CH₄ time-series were obtained from ARC LTER, NSF
- 720 Arctic Data Center site (<u>https://arcticdata.io/catalog/view/doi:10.18739/A2X54S</u>) and the
- 721 Environmental Data Center site of the Toolik Field Station in Alaska (https://arc-
- 122 <u>lter.ecosystems.mbl.edu/;</u> refer to reference list for specific yearly data). The second version of
- the CSLM source code is available at <u>https://github.com/MurrayMackay/CSLM</u>. All remaining
- relevant data supporting the conclusions of this study are included in the article and
- supplementary material. Further inquiries can be directed to the corresponding author.

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