A Century Of Reforestation Reduced Anthropogenic Warming in the Eastern United States

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Abstract

Restoring and preventing losses of the world's forests are promising natural pathways to mitigate climate change. In addition to regulating atmospheric carbon dioxide concentrations, forests modify surface and near-surface air temperatures through biophysical processes. In the eastern United States (EUS), widespread reforestation during the 20th century coincided with an anomalous lack of warming, raising the question of whether reforestation contributed to biophysical cooling and slowed local climate change. Using new cross-scale approaches and multiple independent sources of data, our analysis uncovered links between reforestation and the response of both surface and air temperature in the EUS. Ground- and satellite-based observations showed that EUS forests cool the land surface by 1-2 °C annually, with the strongest cooling effect during midday in the growing season, when cooling is 2 to 5 °C. Young forests reducing midday air temperature by up to 1 °C. Our analyses of historical land cover and air temperature trends showed that the cooling benefits of reforestation extend across the landscape. Locations predominantly surrounded by reforestation were up to 1 °C cooler than neighboring locations that did not undergo land cover change, and areas dominated by regrowing forests were associated with cooling temperature trends in much of the EUS. Our work indicates that reforestation contributed to the historically slow pace of warming in the EUS, highlighting the potential for reforestation to provide local climate adaptation benefits in temperate regions worldwide.

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- 23 Key Points:
- Reforestation in the eastern United States (EUS) contributes to cooling the land surface
 and near-surface air temperature.
- The biophysical impacts of reforestation help explain the anomalous lack of 20th-century warming in the EUS
- Reforestation in temperate regions has the potential to provide biophysical climate
 adaptation benefits by cooling surface and air temperatures.
- 30

31 Abstract

32 Restoring and preventing losses of the world's forests are promising natural pathways to mitigate 33 climate change. In addition to regulating atmospheric carbon dioxide concentrations, forests 34 modify surface and near-surface air temperatures through biophysical processes. In the eastern 35 United States (EUS), widespread reforestation during the 20th century coincided with an 36 anomalous lack of warming, raising the question of whether reforestation contributed to 37 biophysical cooling and slowed local climate change. Using new cross-scale approaches and 38 multiple independent sources of data, our analysis uncovered links between reforestation and the 39 response of both surface and air temperature in the EUS. Ground- and satellite-based 40 observations showed that EUS forests cool the land surface by 1-2 °C annually, with the 41 strongest cooling effect during midday in the growing season, when cooling is 2 to 5 °C. Young 42 forests aged 25-50 years have the strongest cooling effect on surface temperature, which extends 43 to the near-surface air, with forests reducing midday air temperature by up to 1 °C. Our analyses 44 of historical land cover and air temperature trends showed that the cooling benefits of 45 reforestation extend across the landscape. Locations predominantly surrounded by reforestation 46 were up to 1 °C cooler than neighboring locations that did not undergo land cover change, and 47 areas dominated by regrowing forests were associated with cooling temperature trends in much 48 of the EUS. Our work indicates that reforestation contributed to the historically slow pace of 49 warming in the EUS, highlighting the potential for reforestation to provide local climate 50 adaptation benefits in temperate regions worldwide.

51 Plain Language Summary

52 A century of eastern US reforestation has had a cooling effect that helps to explain a lack of 53 regional warming in the 20th century, which stands in contrast to warming trends across the rest 54 of North America during the same period. Our study shows that forests across much of the 55 eastern United States have a substantial adaptive cooling benefit for surface temperature, and for 56 the first time, we demonstrate that this benefit also extends to near-surface air temperature. 57 Therefore, reforestation in temperate zones could provide a complementary set of benefits: 58 mitigating climate change by removing carbon dioxide from the atmosphere, while also helping 59 us adapt to rising temperatures by cooling surface and air temperatures over large areas.

60 **1 Introduction**

61 Drastic reductions in anthropogenic greenhouse gas emissions are necessary to address climate change. Nature-based Climate Solutions (NbCS), such as reforestation, have the potential to 62 63 provide additional mitigation through atmospheric CO₂ removal (Nolan et al., 2021; Novick et 64 al., 2022a,b; Seddon et al., 2020), but will only be effective if they are accompanied by economy-wide decarbonization. Changes to land cover and management, which are central to the 65 66 implementation of NbCS, can alter local temperature through changes to the surface energy 67 balance (Anderson et al., 2011). If these biophysical impacts are beneficial, then some NbCS 68 could serve as a tool for local adaptation in addition to global-scale climate mitigation.

- 69 Reforestation the NbCS with the highest CO₂ mitigation potential (Griscom et al., 2017) can
- 70 increase or decrease local surface temperature (T_s) depending on the balance of competing
- 71 mechanisms. In the tropics, forests evaporate substantially more water than grasslands, which
- promotes cooling by using energy that would otherwise heat the surface (Anderson et al., 2011;

73 Williams et al., 2021). Conversely, reforestation in boreal climates tends to warm the surface 74 through reductions in albedo (Lee et al., 2011). In the temperate zone, surface cooling from 75 increased evaporative and sensible heat fluxes usually outweighs albedo-driven warming, such 76 that temperate forests have lower T_s compared to non-forested ecosystems (Anderson et al., 77 2011; Bright et al., 2017; Windisch et al., 2021; Zhang et al., 2020).

78 Although T_s is relevant to many ecological processes (Farella et al., 2022), the near-surface air 79 temperature (T_a) is an equally important target for climate adaptation (Novick & Katul, 2020; 80 Winckler et al., 2019) because changes in T_a can have far-reaching effects, as biophysical 81 impacts on air temperature can be advected across the landscape (Winckler et al., 2019). While land-cover change affects T_s and T_a differently (Baldocchi & Ma, 2013; Helbig et al., 2021; 82 83 Novick & Katul, 2020; Winckler et al., 2019), quantifying the impacts of land-cover change on $T_{\rm a}$ has historically been challenging. Near-surface air temperature cannot be sensed remotely, 84 85 and its variation with height makes it hard to measure and interpret over forests (Novick & 86 Katul, 2020; Winckler et al., 2019). Most data-driven studies investigating biophysical impacts 87 on both T_s and T_a typically compare locations with similar macroclimates but different land cover, which captures the direct, local effects of land-cover change on T_s and T_a (e.g., Baldocchi 88 89 & Ma, 2013; Bright et al., 2017; Juang et al. 2007; Winckler et al., 2019; Windisch et al., 2021; 90 Zhang et al., 2020). However, this approach can overlook indirect, non-local effects (Baldocchi 91 & Ma, 2013) related to advection or changes in downstream cloud cover due to increased 92 upstream evapotranspiration. This study adopts a novel combination of approaches to evaluate the impacts of reforestation on both T_s and T_a , exploring both local and larger-scale effects across 93 94 various spatial and temporal scales.

95 The Eastern United States (EUS) has undergone extensive reforestation over the last century 96 (Figs. 1A, B; Meehl et al., 2012), providing a unique opportunity to investigate the biophysical 97 impacts of large-scale land cover change. Additionally, the absence of warming over a large 98 portion of the EUS during this period (Fig. 1C; Ramankutty et al., 2010) raises the question of 99 whether reforestation has dampened the historic pace of warming in the region. To address this 100 question, we employ multiple independent data sources to evaluate both local and non-local 101 effects of reforestation on T_s and T_a . Our approach involves: 1) comparing locations with similar 102 climates but different land cover using both satellite and in situ observations to determine the local effects of reforestation on T_s and T_a , 2) exploring gradients in T_s across ecosystem 103 boundaries to uncover the potential local extent of such effects, and 3) analyzing historical 104 105 weather station, air temperature, and land-cover data to identify long-term links between T_a and 106 forest cover trends at landscape and regional scales. Through this comprehensive approach, we 107 aim to gain insight into the extent to which EUS reforestation has influenced historical rates of 108 regional warming and the potential of temperate zone reforestation for climate adaptation.

109 1.1 Historic land cover and climate trends in the region. Before European settlement, forests 110 occupied most of the land area in the EUS, with an uneven-aged stand structure sustained by 111 selective harvest and controlled burning by Indigenous tribes (Meehl et al., 2012). However, from the late 18th to early 20th century, forest cover in the EUS dramatically decreased due to 112 harvesting for timber and clearing for agriculture, resulting in forest losses exceeding 90% in 113 some locations (Carman, 2013; Hall et al., 2002; Houghton & Hackler, 2000). By 1930, 114 widespread land clearing had largely stopped, and forest cover began to increase with the 115 abandonment of marginal agricultural fields and active reforestation efforts (Carman, 2013; Hall 116

- et al., 2002; Houghton & Hackler, 2000; Meehl et al., 2012). Since 1900, millions of hectares of 117
- 118 forest have been added in the northeastern, southeastern, and midwestern US (Fig. 1B, and
- 119 Carman, 2013; Hall et al., 2002; Wear & Greis, 2012), mostly through the conversion of crop
- 120 and pastureland to deciduous forests, or to pine plantations providing softwood timber in some parts of the southeastern US (Mascioli et al., 2017). Many forests in the region are now 50-100
- 121
- 122 years old (Fig. 1C, Fig. S1), although frequent harvest of pine plantations suppresses stand age in
 - 123 the Southeast.
 - 124 This century of EUS reforestation coincides with an anomalous lack of regional warming, 125 sometimes referred to as a 'warming hole' (Fig. 1; Mascioli et al., 2017; Meehl et al., 2012; Z. 126 Pan et al., 2004; Partridge et al., 2018; Tosca et al., 2017). While most land areas worldwide 127 warmed during the twentieth century, much of the EUS experienced minor cooling, from -0.2 $^{\circ}$ C to -0.8 °C per 50 years (Fig. 1C). Proposed explanations for this cooling include internal 128 129 climate variability (Mascioli et al., 2017; Meehl et al., 2012), anthropogenic aerosols (Tosca et 130 al., 2017), agricultural intensification (Mueller et al., 2016), and increasing precipitation (Z. Pan 131 et al., 2004). However, mechanistic attribution remains elusive (Mascioli et al., 2017; Partridge 132 et al., 2018), and some of the mechanisms (e.g., agricultural intensification) are not relevant across the entire extent of the warming hole. Despite the established potential of reforestation to 133 134 affect local temperature, the biophysical impacts of regional reforestation over the past century 135 have not been thoroughly evaluated for their contribution to the 'warming hole' in the EUS.

136 2 Materials and Methods





Figure 1. The Southeastern United States' warming hole' and corresponding forest status. (A) Forest age estimates (1 km) as of 2019 calculated from forest age data from the North American Carbon Program. (B) Land conversion between agricultural land and forests from 1938 to 1992 calculated from 1 km FORE-SCE backcasting grids from the US Land Cover Trends project. The bounding box indicates the study area, and details on data sources are provided in the methods. (\mathbf{C}) Trend in temperature change from 1900 to 2010 (ΔT_a , °C/50 years) calculated using a season-trend model applied to 0.5° T_a grids from University of Delaware monthly climatologies provided by the NOAA. (D) Historical photo of 15-person planting crew on 05/09/1932 in Tucker County, West Virginia. Original photo at the Forest Service Office in Elinks, WV. Obtained from https://www.loc.gov/pictures/resource/hhh.wv0307.photos.041150p/

138 commonly used for assessing the biophysical impacts of land-cover change and is observable 139 from both flux towers and remote sensing platforms. In terrestrial ecosystems, T_s represents the 140 temperature of the uppermost layer of vegetation, reflecting the outcome of the interactive effects 141 of radiative transfer, leaf energy balance, eco-physiological controls on stomatal opening and 142 closure, and aerodynamics. The T_s is mechanistically connected to the near-surface T_a through 143 standard boundary layer theory for stratified flows. In vegetated systems, a thin layer of air 144 called the roughness sublayer lies between the vegetation and the surface layer. Near-surface $T_{\rm a}$ 145 is generally measured (and modeled) at a height of 2 m above the surface, but vertical profiles of $T_{\rm a}$ are influenced by canopy structural effects within the roughness sublayer. In short-stature 146 147 ecosystems such as grasslands and croplands, the roughness sublayer is typically 1-2 meters 148 thick, such that flux tower measurements of T_a are typically made in the surface layer. Above 149 forests with heights of 10 m or more, the roughness sublayer is much thicker (12-50 m), such 150 that tower $T_{\rm a}$ indicates the mean air temperature within the roughness sublayer, not the surface layer (Novick & Katul, 2020). Thus, comparing T_a observations across forested and non-forested 151 ecosystems is fraught with potential bias linked to canopy structural effects and measurement 152 153 height. To overcome this limitation, we focus on two proxies for T_a that are less sensitive 154 structure effects on near-surface T_a profiles, and that can be inferred from eddy covariance 155 sensible heat flux tower data using methodologies described elsewhere (Novick & Katul, 2020). 156 The first is the aerodynamic temperature (T_{aero}) , which represents the air temperature within the 157 upper reaches of the canopy. The second is the air temperature extrapolated aloft into the lower 158 reach of the surface layer (T_{extrap}) .

159 At regional scales, evaluating the impacts of land cover on T_s requires a strategy to control for 160 background variation in macroclimate. Here, the difference between MODIS T_s and T_a is evaluated, with the latter provided by the Daymet (Thornton et al., 2016) product ($T_{a,Daymet}$). 161 162 Remotely sensed surface temperature is frequently termed 'Land Surface Temperature' (LST), 163 but we use T_s here for consistency with the other temperature metrics. Daymet interpolates data 164 from meteorological weather stations, which are typically located over short grass surfaces, so it 165 is viewed here as a 'reference' T_a that is not influenced by variability in land cover. A summary 166 of the temperature metrics used here can be found in Table S1.

2.2 Forest age and land-cover change analyses. Gridded 0.5 $^{\circ}$ time series of monthly $T_{\rm a}$ were 167 168 obtained from the University of Delaware Air Temperature & Precipitation Dataset (Willmott & 169 Matsuura, n.d.) provided by the NOAA/OAR/ESRL PSD, Boulder, Colorado, USA, from 170 https://psl.noaa.gov. The gridded data are interpolated weather station data (Willmott & 171 Matsuura, 1995), primarily from GHCN2 (Global Historical Climatology Network) observations 172 and the GSOD (Global Surface Summary of Day) archive. Monthly mean air temperature (V4.01) was used to assess observed long-term changes in T_a across the continental United 173 174 States. Per-pixel temperature change in T_a was estimated using a season-trend model (function 175 'STM') in the 'greenbrown' package (Forkel & Wutzler, 2015) for R (R Core Team, 2020). The 176 season-trend model in 'greenbrown' is based on the additive decomposition model in refs 177 (Verbesselt et al., 2010) and (Verbesselt et al., 2012). Harmonic and linear terms are used to 178 model seasonal variation and trend, respectively, effectively 'detrending' the time series in a 179 single step (Verbesselt et al., 2012).

180 To assess forest age, we obtained continental forest age maps from the North American Carbon 181 Program (NACP), produced for the year 2006 (Y. Pan et al., 2012). Data were updated to give 182 accurate forest ages for 2019 by adding 13 years to all forest ages. This approach assumes that all 183 forests continued to regrow and were not cut between 2006 (when the product was produced) 184 and 2019. Although there are certainly locations for which this assumption will not hold, they 185 generally represent a small fraction of the land surface and are mostly contained in the 186 southeastern US where pine plantations are routinely harvested and replanted (Carman, 2013; 187 Fig. S1).

188 To assess change in forest status (Fig.1B), land-cover change between 1938 to 1992 was 189 calculated from 250 m FORE-SCE (FOREcasting SCEnarios of Land-use Change) backcasting grids from the US Land Cover Trends project (Sohl et al., 2007). Pixels were coded based on 190 191 their land cover (cropland or forest), and per-pixel change was calculated between 1938 and 192 1992. To exclude sites that had undergone multiple conversions in the 20th century, an 193 intermediate year, 1965, was also used. Pixels were classified into three categories: reforestation (change from cropland to forest), deforestation (forest to cropland), or 'no change' (land cover 194 195 type was consistent in 1938, 1965, and 1992).

196 **2.3 Differences in T_s - T_a by land cover.** To compare relative surface cooling between different 197 types of land cover on a regional scale, we used a remote sensing approach based on T_s retrievals 198 from the Moderate Imaging Resolution Spectroradiometer (MODIS; Wan et al., 2015) and T_a 199 estimates from the gridded 1 km Daymet product (Thornton et al., 2016). The MODIS Land 200 Surface Temperature/Emissivity Daily Product (MYD11A1v6.1) has a spatial resolution of 1 201 km. The timing of the Aqua MODIS overpass (~1:30 pm local time) generally corresponds to the 202 timing of daily maximum T_s estimated from tower measurements (Fig. S2). Thus, to evaluate 203 spatial variability in the difference between remotely sensed T_s , we corrected for macro-scale 204 climate variability by subtracting $T_{a,Daymet}$ from MODIS T_s :

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- 206 207

$$T_{\rm s} - T_{\rm a,Daymet} = \text{LST}_{1:30\text{pm, MYD11}} - T_{\rm a,Daymet}$$
(1)

208 If the near-surface air is warmer than the surface, $T_s - T_a$ is negative; if the air is cooler than the 209 surface, T_s - T_a is positive. Most often, especially at mid-day, the $T_s - T_a$ will be positive (Mildrexler et al. 2011, Novick & Barnes under review), but we expect that it will be more 210 211 negative (i.e., closer to zero) in forested versus non-forested sites. The monthly average T_s - $T_{a,Daymet}$ for each 1 km pixel in the study area was obtained for each month from 2002 to 2018. 212 213 The data presented in Figure 2 are the average monthly T_s - $T_{a,Daymet}$ for the full time-series. The 214 land cover type for each 1 km MODIS pixel was determined using the most recent 30 m National 215 Land Cover Database (NLCD) land cover dataset (Dewitz, 2019). Mixed pixels that contained 216 multiple land cover types at the 30 m level in each MODIS or Daymet 1 km pixel were classified 217 as the most common land cover type in the pixel, provided that the most common land cover 218 type comprised 70% or more of the pixel. There were relatively few pixels with grassland as the 219 dominant land cover type, so we combined croplands and grasslands for the spatial T_s - $T_{a,Davmet}$ 220 maps (Fig. 2A, Figs S3-S5).

221 **2.4 Paired site flux tower analyses.** To understand reforestation effects on T_s over the diurnal 222 cycle (e.g., Fig. 2A, B), we used a paired site approach (Stoy et al. 2023), relying on 223 observations from six forest-grassland site pairs in the study region. The site pairs are in 224 Arkansas, North Carolina (3 pairs), Indiana, and New Hampshire (Table S2), and each paired set 225 is separated by ~30 km or less. Site descriptions and details on eddy covariance data processing 226 are provided elsewhere (Zhang et al., 2020); Briefly, all data were quality-controlled and gap227 filled using community-accepted standards embedded in the REddyProc processing tool 228 (Wutzler et al., 2018). The T_s data were inferred from the outgoing longwave radiation using the 229 Stefan-Boltzmann law, with emissivity estimated as an empirical function of albedo following an 230 established approach (Juang et al., 2007). The attribution of changes in T_s to relevant mechanisms (i.e., variation in sensible versus latent heat flux) was accomplished through a 231 232 Taylor-series expansion of the site-level energy balance equation, as described in detail in 233 (Zhang et al., 2020), who presented results separately for each of the paired sites. In this study, 234 those results are then aggregated across the site pairs.

- The approach to estimate T_{aero} and T_{extrap} is described elsewhere (Novick & Katul, 2020). Briefly,
- 236 T_{aero} is determined by first quantifying the mean ratio between T_{s} and the tower-measured $T_{\text{a}}(XT)$ when the measured sensible heat flux is near zero, implying that T_a and T_{aero} should be in 237 238 equilibrium. The determination of XT was performed for each hour of the day at each site 239 separately for the peak of the growing season (June-Sep.) and the dormant season (Nov.-Mar.) 240 and was then used to estimate T_{aero} from the observed T_s for every hourly or half-hourly 241 observation period. The T_{extrap} was then calculated assuming the logarithmic profiles from 242 Monin-Obukhov Similarity Theory (Monin & Obukhov, 1954), forced with the estimated T_{aero} and an estimate for the momentum roughness length for heat that varies as a function of 243 244 measured friction velocity as described in (Novick & Katul, 2020). Here, results are shown for 245 the temperature extrapolated into the first 10 meters of the surface layer after conceptually 246 'flattening' the ecosystems by replacing them with a rough surface characterized by two 247 roughness heights: one for momentum absorption and one for heat transfer (see Novick & Katul, 248 2020 for details and code). This manuscript extends the results presented in Novick & Katul 249 (2020) to the full set of paired sites in Table S3, noting that Novick & Katul only evaluate the 250 results from the Duke Forest sites (in North Carolina).
- The T_a data presented in Figure 2 are mean air temperature data measured from a T_a /RH probe (such as the HMP35 or HMP45c, Vaisala) above the canopy, usually at or near the height of the EC systems (see Table S2).
- 254 2.5 Broader synthesis of flux tower sites. The 58 flux towers included in the broader synthesis (Fig. 3A-D) include the paired tower sites, as well as datasets acquired from the AmeriFlux 255 256 network (Novick et al., 2018), and include towers installed and operated as part of the National 257 Science Foundation's National Ecological Observatory Network (Metzger et al., 2019). The 258 network data were limited to those that adopt a CC-BY-4.0 data use license. We first 259 downloaded all available site-years from the network, excluding wetlands and irrigated croplands 260 within the study region. Some sites and site-years were excluded due to missing information 261 about radiative fluxes that are required to derive or calculate the temperature metrics. The towers retained for this analysis are described in Table S3. The same approaches used to determine T_s , 262 T_{aero} , and T_{extrap} for the broader flux synthesis are the same as those described in section 2.4, with 263 264 the exception that the AmeriFlux data were not gap-filled; rather, data were filtered to exclude 265 those collected under very stable conditions in which turbulence generation is suppressed by 266 buoyancy forces, and to exclude excessively large anomalous observation of sensible heat flux 267 (i.e.,>1000 W/m²). For the purposes of this paper, croplands managed as corn/soy rotations were 268 treated as separate sites for corn and soy years. Estimates of canopy height and measurement 269 height are required to calculate T_{aero} and T_{extrap} . For most towers, they are available from the 270 AmeriFlux Biological, Ancillary Disturbance, and Metadata (BADM) database; In cases where 271 they were not available from the BADM, they were extracted from published studies. In rare

cases, a "best" guess was made based on known information about the canopy type or from sitephotos.

274

275 2.6 Land Surface Temperature across forest/agriculture boundaries. The NLCD was used to 276 identify areas with adjacent forest: agriculture boundaries with a continuous extent of land cover on either side of the boundary. Transects were created across boundaries in the east/west 277 278 direction at a ~90° angle. The transects were approximately 1 km long, with points every 10 m. 279 In total, 44 transects were created in the study area (Fig. S6). We then extracted T_s values from mid-summer clear-sky scenes (summer 2018) obtained from the Landsat Provisional Surface 280 281 Temperature Product (Cook, 2014; Cook et al., 2014). Only one T_s spatial profile was extracted 282 for each transect, even if multiple images were available. The spatial resolution of Landsat 283 imagery is 30 m, so the 10 m transects include several points within the same pixel. A smooth 284 transition suggests that biophysical feedbacks on T_s are linked to T_a through advection and the 285 formation of an internal boundary layer (Hsieh & Katul, 2009), since the surface itself is not 286 mixed. For the lines in Figure 3E, we used a hyperbolic tangent function fit to each side of the 287 transects (forest and agriculture). Distances were scaled from $-\pi$ to π for fitting, then back-288 transformed for plotting.

289

290 2.7 Impacts of Reforestation on Air Temperature: To investigate the impact of local land-

291 cover change on long-term air temperature trends, we used monthly air temperature data from

398 United States Historical Climate Network (USHCN) meteorological stations. Meteorological

stations, which were relatively evenly distributed across the study area (Fig. S8), allow us to associate locations more precisely with their land cover type compared to gridded temperature

associate locations more precisely with their land cover type compared to gridded temperature data. We evaluated the effects of land cover on annual temperature trends from 1900 to 2010 by

focusing on both maximum temperatures during the growing season (June, July, August) and

annual average temperatures. We did not conduct any interpolation of missing values.

298 We classified the weather stations based on the dominant land cover within a 500 m circular buffer surrounding each station, using 250 m FORE-SCE model backcasting grids. The 299 300 classification was based on three time points (1938, 1965, and 1992), as described in section 2.2. 301 Pixels were categorized as reforestation areas (change from agriculture to forest) if they were 302 agricultural in 1938, converted to forest cover by 1965, and maintained as forested land until 303 1992, or if they were agricultural in both 1938 and 1965 but became forested by 1992. The areas 304 predominantly surrounded by agriculture-to-forest conversion within 500 meters of the weather 305 stations were defined as 'reforest' sites. Out of the 398 USHCN stations, 153 were 306 predominantly surrounded by forest, 181 by agriculture, and 23 by reforestation. 41 of the sites underwent multiple land cover transitions or experienced deforestation in the 20th century. 307

308

309 To explore the impacts of reforestation on air temperature trends, we compared USHCN sites

310 predominantly surrounded by reforestation to sites where the land cover remained stable as either

agriculture or forest ('non-reforest') within a 50 km radius of the 'reforest' site. For each year,

312 we calculated the temperature difference between the reforest site and non-reforest site (or sites)

313 within the 50 km buffer. This approach builds upon the paired-site flux tower analyses described

314 in section 2.4 by extending our understanding of the effects of reforestation on surface

315 temperature to near-surface air temperature, covering a much longer period (back to the early

316 20th century), and including many more site 'pairs'. Our comparison of neighboring reforest and

317 non-reforest site pairs included analysis of both growing-season maximum temperatures and

318 average annual temperatures. We excluded sites that had more than 5 years of missing data from

1900 - 2010. We had a total of 22 reforest sites and 44 non-reforest sites within 50 km of the

reforest sites, resulting in 44 comparisons for annual average data and 42 comparisons for

321 growing season data. The lower number of growing season comparisons is due to missing data 322 for two sites. The reforest sites had a median of two non-reforest sites within a 50 km radius. We

based the 50 km radius on a previous study that showed changes in maximum air temperature up

to 50 km away from the site of land cover change (Cohn et al., 2019).

325

326 **2.8** Exploring the links between forest age and air temperature change within the Warming

Hole. To further investigate the influence of reforestation on the warming hole, we examined the relationship between forest age and recent growing season trends in T_a at the regional scale (Fig.

4D). To align our recent 'snapshot' of forest age with T_{a} , we compared the 1 km NACP-derived forest age maps with recent trends (1970-present) in T_{a} from the University of Delaware

forest age maps with recent trends (1970-present) in T_a from the University of Delaware climatologies (Willmott & Matsuura, n.d.). We used the season-trend approach described in

332 section 2.2 to estimate the slope of trends in annual T_a time-series from 1970-2017 (Fig. 4D).

The slope of the temperature change from 1970-present, ΔT_a , was calculated for each pixel, and forest age data were aggregated to the coarser scale of the T_a data (Fig. S1B) using a mode

function, "modal", in the 'raster' R package (Hijmans, n.d.), and then resampled using nearest neighbor interpolation. Then, we calculated focal correlations between aggregated forest age

337 (Fig. S1B) and temperature trends, specifically, a simple moving window correlation between

338 ΔT_a from 1970-present and forest age in a 5 x 5 window using the 'raster correlation' function in 339 the 'SpatialEco' R package (Evans, 2021).

339 the 'Spati340

We note that although gridded daily microscale (1 km) T_a products such as Daymet should be relatively insensitive to local impacts of land cover on T_a (Fig. S7), the fingerprint of reforestation may be detectable from coarser yearly mesoscale T_a estimates at 0.5-degree resolution. Observational and modeling studies support regional-scale impacts of land cover change on mesoscale T_a (Bonan, 2001; Mahmood et al., 2014), providing further justification for our approach.

347 3 Results

348 Across the study area, the difference between T_s and the $T_{a,Daymet}$ was more negative for forests

349 than grasslands and croplands most of the time (Fig. 2). Since the $T_{a,Daymet}$ normalizes for macro-

350 scale temperature variability, this result implies that forest surfaces are cooler than the surfaces

351 of nearby grasslands and croplands by the same amount. The effect was most pronounced during

352 the growing season, when forests were cooler than non-forests by an average of 0.5-2 $^{\circ}$ C (Fig.

353 2A, B), with smaller reductions observed in spring and fall.

Next, we leveraged rich surface energy balance information from eddy covariance (EC) flux towers, beginning with six co-located ('paired') forest and grassland sites in the study region (Table S2; Zhang et al., 2020). Across these paired sites, forest T_s was 4-5 °C cooler, on average, than nearby grasslands during midday periods (Fig. 2D, E), driven primarily by enhanced evapotranspiration in summer and enhanced sensible heat flux in winter that outweighed albedodriven warming effects in the darker forests (Zhang et al., 2020).



Figure 2. A forest surface cooling effect is evident in both MODIS and flux tower (A-B) observations (**C**, **D**). (**A**) Average difference between daily ~1:30 pm surface temperature (T_s) and daily maximum air temperature (T_a , D_{aymet}) for 2003-2018 for forests (top row) and combined grasslands and croplands ('other', second row). Negative values indicate cooler surface than air temperature (surface cooling) and positive values indicate warmer surface air temperature. (**B**) The seasonal cycle of $T_s - T_{a,Daymet}$ for forests (blue), grasslands (yellow), and croplands (orange) for the study region (bolded lines), with latitude ranges indicated by faint dashed lines. (**C**) Remotely sensed surface temperature (T_s), corrected by the T_a , D_{aymet} reference, shown as a function of forest age, using July data for the period 2003-2018. (**D**, **E**) Diurnal time series of the difference between forest and grassland T_s (blue), T_{aero} (black), T_{extrap} (gray), and tower-measured T_a (yellow) for six eddy covariance site pairs (Table S3) for the growing season (**D**) and the dormant season (**E**).

- 360 Although the paired-site approach represents a 'gold standard' for understanding the biophysical
- impacts of land-cover change (Zhang et al., 2020), only a handful of forest-grassland flux tower
- 362 pairs exist. To expand the scope of inference, we synthesized T_s observations from 58 Ameriflux
- 363 EC tower sites across the EUS, again correcting for macro-scale climate variability with $T_{a,Daymet}$ 364 (Table S3). The results revealed widespread daytime surface cooling in forests compared to non-
- forests (Fig. 3A). The difference between tower-derived T_s and the reference $T_{a,Daymet}$ was 4.5-5.5
- 366 ° C lower for forests than grasslands during the growing season (Fig. 3A). When comparing
- 367 forests to croplands, T_s was more similar than for grasslands, but tall forests were still relatively
- 368 cooler than croplands (Fig. 3A).

369 **3.1 Land cover effects on near-surface** T_a . For the surface cooling to extend beyond the stand-370 scale and thus contribute to the warming hole, changes in T_s must translate to changes in near-371 surface T_a .

- 372 We adopted several new or emerging approaches to quantify the extent to which the forest 373 cooling effects on T_s extend to T_a . First, we harnessed flux tower data to estimate metrics of T_a 374 that are less sensitive to canopy effects (11): T_{aero} and T_{extrap} (Table S1). In the paired sites and 375 across the regional network of flux towers, the midday growing season T_{aero} and T_{extrap} are cooler 376 for forests than for grasslands (Fig. 2D, E, and 3A-D). Consistent with expectations, forests have 377 the strongest cooling effect on surface temperatures (Fig. 3A) and smaller effects on air 378 temperatures (Fig. 3B-D). T_{aero} and T_{extrap} were similar between forests and croplands, although 379 T_{aero} was substantially lower for tall forest stands (Fig. 3B-D). Despite the confounding influence 380 of canopy effects on near-surface T_a profiles, tower-measured T_a also suggests a cooling effect of
- 381 forests, although differences across biomes were smaller (Fig. 3D).

Next, to provide an independent perspective on surface and air temperature coupling, we used high-resolution (30 m) Landsat T_s retrievals to evaluate the extent to which transitions in T_s at forest-cropland boundaries were smooth or abrupt (Fig. 3E; Fig. S6). A relatively smooth temperature transition from cooler forests to warmer croplands was observed (Fig. 3E) that extends over length scales of several hundred meters.



Figure 3: Extension of surface cooling to the near-surface air. (**A-D**) *Difference in tower-measured or* derived temperature metrics and reference $T_{a, Daymet}$. Each boxplot shows the monthly midday growing season site-level means, with forested ecosystems split by canopy height ($\leq 20 \text{ m or } > 20 \text{ m}$). Horizontal lines indicate the median, boxes indicate the interquartile range, whiskers indicate the data range (=1.5 times the interquartile range), and the symbols' +' indicate outliers. Letters indicate significant differences between groups evaluated using a two-sample t-test at a significance level of p = 0.05. Note the y-axis range varies between panels. Each panel is informed by data from 13 grass towers, 19 cropland towers, 10 towers in forests < 20 m tall, and 16 towers in forests > 20 m tall. (**E**) Horizontal profiles ("transects") of T_s across the forest:cropland boundary. Values of ΔT_s (sampled every 10 m) are relative to the temperature at the forest:cropland boundary. Negative values of ΔT_s indicate lower temperatures than at the boundary, and positive values of ΔT_s indicate higher temperatures than at the boundary. Points left of the zerodistance line indicate forest (blue), and points right of the zero-distance line indicate agricultural croplands (orange). The blue and orange lines indicate a hyperbolic tangent function fit to the pooled transect data. 388 3.2 Evaluating the signatures of land-cover change in long-term meteorological observations. To evaluate the signatures of land-cover change in long-term meteorological 389 390 observations, we compared long-term records of air temperature from weather station sites, 391 estimates of forest age, and gridded air temperature data. First, we examined the long-term 392 temperature trends from 398 USHCN sites in the study region. Our analysis showed that 393 temperatures remained relatively stable throughout the 20th century, with a decline in average 394 temperatures in the late 1950s (Fig. 4A). Next, we investigated the cooling effect of regrowing forests by analyzing the difference between T_s and the reference $T_{a,Daymet}$ as a function of forest 395 396 age. Results indicated that the cooling effect was strongest for forests around 25 years old, as the 397 difference reached its lowest point at that age (Fig. 4B).

398 To further explore the cooling effect of reforestation, we compared 10-year rolling means of T_a 399 from historical USHCN stations in sites predominantly surrounded by reforestation ('reforest') 400 and neighboring sites that did not undergo land cover conversion ('non-reforest') within 50 km. 401 In terms of average annual temperatures, reforested sites were consistently cooler than their non-402 reforesting counterparts throughout the 20th century (Fig. 4B, blue line). The results were more 403 nuanced for maximum growing season temperatures. Sites with nearby reforestation tended to be 404 warmer in the early half of the 20th century and cooler in the latter half, and the magnitude of 405 this cooling effect increased throughout the study period. By the end of the 20th century, 406 reforesting sites were up to 1 °C cooler than their non-reforesting neighbors in terms of 407 maximum growing season air temperatures (Fig. 4B, red line).

Finally, to explore the relationship between forest age and temperature trends, we explored 408 409 spatial variation in the correlation between decadal-scale growing season T_a trends (from University of Delaware 0.5° monthly climatologies) and forest age across the continental US 410 411 (Fig. 4D). A positive correlation coefficient indicates that pixels dominated by younger forests 412 (i.e., < 100 years as of 2019) experienced less warming during this period than pixels dominated 413 by more mature stands (see inset to Fig. 4D). This correlation coefficient was positive across 414 much of the study region on the annual timescale (63% of pixels; Fig. 4D), and when considering 415 growing season temperature trends alone (Fig. S9B), consistent with the expectation that 416 reforesting areas experience less warming. However, there is substantial variability across the 417 study region, with negative relations between forest age and ΔT_a observed, including in the 418 central Appalachians and Ozarks.



Figure 4. Relationships between land cover, long-term temperature trends, and forest age. (A) Average long-term trend in pooled annual T_a anomalies for 398 USHCN sites in the EUS. The blue line represents the 10-year moving average. (B) Difference in air temperature between USCHN sites predominantly surrounded by reforestation and non-reforesting sites located within a 50 km radius, presented as 10-year moving averages. The red line shows the difference in summer maximum temperatures, and the blue line shows the difference in average annual temperatures. Negative values indicate that the reforesting site was cooler than its neighboring non-reforesting site(s). (C) Remotely sensed July surface temperature (T_s) corrected by the T_{a} , D_{aymet} reference displayed as a function of forest age for the period 2003-2018. (D) Spatial moving window correlation (5 x 5 window) between forest age and recent long-term T_a trends (1970-present). A negative correlation indicates older forests are associated with greater cooling, and a positive correlation indicates

younger forests are associated with greater cooling (inset).

419

420 4 Discussion and Conclusions

421 Our study provides strong evidence that reforestation has biophysical climate benefits in the EUS 422 and demonstrates a clear relationship between observed land-cover change and observed 423 temperature changes. Taken together, our findings indicate reforestation has a cooling effect on surface and near-surface air temperature in the EUS, and likely contributed to the slower pace of 424 425 warming in the region. Both ground- and satellite-based observations indicate that EUS forests 426 cool the land surface by 1-2 °C annually (Fig. 2) compared to nearby surfaces with short-stature 427 vegetation. During midday in the growing season, surface cooling is 2 to 5 °C (Fig. 2C, Fig. 3A), 428 and forests aged 25-50 years exhibit the strongest cooling effect (Fig. 4C). The surface cooling 429 extends to the near-surface air, with forests reducing midday growing-season air temperature by 430 up to 1 °C (Fig. 3D, Fig. 4B). Historical weather station data and regional-scale gridded long-431 term air temperature data analyses establish a link between reforestation and the observed lack of 432 warming in the EUS (Figure 4). In particular, weather stations located near reforesting areas 433 recorded temperatures that were 0.5 - 1.0 °C cooler than stations surrounded by land that 434 experienced little change in forest cover. Overall, these results highlight the substantial 435 adaptation potential of reforestation as a natural climate solution.

436

437 4.1 Assessing land cover impacts on surface and near-surface air temperature. Assessing 438 the direct consequence of land-cover change on surface temperature is relatively straightforward, 439 given the abundance of satellite observations of T_s and complementary measurements of surface 440 energy fluxes from ground-based towers. We have known for some time that reforestation tends 441 to increase T_s in the boreal zone (due to large reductions in albedo, Bright et al., 2017; Swann et 442 al., 2010) and decreases T_s in the tropics (due to enhanced latent heat fluxes, Bonan, 2008). In 443 this study of temperate ecosystems, satellite and flux tower data are combined to demonstrate 444 that, in mid-latitude forests of the eastern US, forest-driven reductions to T_s are: (1) 2-5 °C at 445 midday during the growing season (Fig. 2, Fig. 3A), (2) achievable within ~25 years of forest 446 regeneration (Fig. 4C), and (3) stable across mature forest classes (Fig. 2C). These results, 447 combined with other recent work on the topic, contribute to a growing consensus that 448 reforestation tends to have a direct cooling effect on T_s across much of the temperate zone.

449 Assessing the impact of land cover change on near-surface air temperature is more challenging 450 due to technical and methodological limitations. Yet, understanding the potential of reforestation 451 and other natural climate solutions to confer climate adaptation benefits requires that we quantify 452 land cover impacts on near-surface air temperature. Furthermore, for the surface cooling to 453 extend beyond the stand scale and thus contribute to the warming hole, changes in T_s must 454 translate to changes in $T_{\rm a}$. In turn, evaluating the impacts of land cover on near-surface air 455 temperatures requires overcoming technical and methodological challenges, including data 456 scarcity and canopy structural effects that prevent straightforward comparisons of air 457 temperature measured above different ecosystems. We adopted several new or emerging 458 approaches to overcome the technical and methodological challenges associated with evaluating 459 T_{a} .

460 First, we harnessed flux tower data to estimate metrics of T_a that are less sensitive to canopy 461 effects. By relying on proxies for near-surface air temperature that are relatively insensitive to 462 canopy structural effects (e.g., T_{acro} and T_{extrap}), we demonstrate that, across much of the study 463 region, the cooling effect of forests on T_s extends to the near-surface air temperature above the 464 canopy (Fig. 2E, F; Figs. 3A-D). Forest cover impacts on T_a are smaller than impacts on T_s (~ 1 465 °C for midday growing season periods, Fig. 2D, Fig. 3) but still consequential, particularly 466 compared to the magnitude of historic and predicted changes in T_a due to climate change.

467 Next, to provide an independent perspective on surface and air temperature coupling, we used 468 high-resolution (30 m) Landsat T_s retrievals to evaluate the extent to which transitions in T_s at forest-cropland boundaries were smooth or abrupt. A smoother transition suggests that 469 470 biophysical feedbacks on T_s are linked to T_a through the formation of an internal boundary layer, 471 assuming the thermal inertia of leaves is small. A relatively smooth temperature transition from 472 cooler forests to warmer croplands was observed (Fig. 3E) that extends over length scales typical 473 of adjustment distances needed for the equilibration of the air temperature internal boundary layer. Furthermore, the smooth transition of T_s across forest-grassland boundaries implies 474 475 coupling between T_s and T_a that extends at least several hundred meters beyond the ecosystem 476 boundary. This result helps to confirm previous theoretical modeling (Hsieh & Katul, 2009; Li & 477 Wang, 2019) and augments a similar exercise in the tropics (Cohn et al., 2019) by documenting 478 these processes at a large spatial scale over much of the EUS. Importantly, it is over this distance that biophysical impacts on T_a are substantial enough to feedback on T_s . It is likely that 479 480 reforestation impacts on T_a extend further, but beyond a few hundred meters are not great enough 481 to drive significant changes in grassland $T_{\rm s}$.

482 Our analyses have shown that forests in the EUS exert a cooling effect on both surface and air 483 temperatures, and that this effect can extend across ecosystem boundaries. Importantly, this 484 cooling benefit is most pronounced during midday summer periods (Fig. 2C, D; Fig. 4B), which 485 are typically associated with high heat stress and extreme events. These findings suggest that, in 486 temperate zones, reforestation may provide the greatest climate adaptation benefit precisely 487 when it is most needed.

488 4.2 Linking reforestation to the observed lack of warming in the Eastern US. To establish a 489 link between reforestation and the observed lack of warming in the EUS, we analyzed historical 490 weather station data to associate near-surface air temperature records with land cover changes 491 during the 20th century. Additionally, we conducted a regional-scale analysis to investigate the 492 relationship between spatial patterns of reforestation and temperature patterns in the EUS.

Overall, our analysis of long-term T_a records from 398 USHCN weather stations is consistent 493 with the general understanding of 20th temperature trends in the EUS as it reveals no overall 494 495 warming trend (Fig. 1B; Fig. 4A). We observed a sharp decrease in average temperatures in the 496 1950s, which corresponds with previous studies that found an abrupt climactic regime shift in 497 1957-1958 in the EUS (Partridge et al., 2018; Rogers, 2013). The causes of this abrupt, uniform 498 cooling are likely multifaceted. The abrupt cooling could be related to changes in jet stream 499 position, specifically sharp decreases in the Meridional Circulation Index (MCI; Tosca et al., 2017). However, the impact of decreases in the MCI is greater on winter temperatures, which 500 decreased more uniformly during the twentieth century than summer temperatures (Fig. S10). 501 502 Therefore, focusing on summer temperature time-series in addition to annual temperature time-503 series can help to distinguish the influence of the MCI from other mechanisms, including 504 reforestation.

505 We investigated the signatures of reforestation in long-term climate data by analyzing the 506 difference in long-term T_a records between weather stations predominantly surrounded by 507 reforestation and nearby (within 50 km) weather stations that did not undergo land cover change 508 (remained agriculture or forest) throughout the 20th century. In the early 20th century, we found 509 that sites with nearby reforestation tended to be warmer than sites without reforestation in terms 510 of maximum daily growing season temperatures (Fig. 4B), which is consistent with expectations 511 if these areas were sparsely vegetated at the time (Fig. 1). However, as the 20th century 512 progressed, sites predominantly surrounded by reforestation became increasingly cool relative to 513 their non-reforesting neighbors within 50 km. By the end of the 20th century, reforesting sites 514 were up to 1 °C cooler than their non-reforesting neighbors in terms of maximum growing 515 season air temperatures (Fig. 4B). The magnitude of the cooling effect is consistent with the 516 results from the tower air temperature comparisons (Fig. 2D, Fig. 3), providing independent 517 evidence of the impact of reforestation on near-surface air temperature.

- 518 Analysis of the forest cooling by age indicated that the cooling effect of regrowing forests takes
- about 25 years to fully develop, with forests around 25 years old exhibiting the strongest cooling f_{22}
- 520 effect (Fig. 4C). This time frame is consistent with the time it takes for regenerating forests in the 521 region to achieve levels of hydrological function comparable to mature forests (Ford et al.,
- 521 region to achieve revers of hydrological function comparable to mature forests (Ford et al., 522 2011). During the growing season, when the cooling impact of regrowing forests is expected to
- 523 be strongest, reforesting sites were consistently cooler than nearby non-reforesting sites by the
- 100 strongest, reforesting sites were consistently cooler than nearby non reforesting sites by the 100 strongest, reforesting sites were consistently cooler than nearby non reforesting sites by the 100 strongest, reforesting sites were consistently cooler than nearby non reforesting sites by the 100 strongest, reforesting sites were consistently cooler than nearby non reforesting sites by the 100 strongest, reforesting sites were consistently cooler than nearby non reforesting sites by the
- 525 widespread agricultural abandonment and federal reforestation efforts in the 1930s as age zero
- 526 (e.g., Fig. 4D). The trends in maximum growing season temperature increased throughout the
- 20^{th} century, supporting the idea that as forests grew, the summer cooling effect increased (Fig.
- 528 4B).

529 Regional-scale analysis of gridded long-term air temperature data provided an independent, 530 complementary approach to the weather station analyses. The analysis showed that younger 531 forests were associated with lower historic rates of warming across most of the study region (Fig. 532 4D). This finding, combined with the weather station analyses, suggests that reforestation had a 533 cooling effect on near-surface air temperature across a wide swath of the EUS. Parsing the trends 534 in long-term T_a time series as a function of land cover and historic land-cover change (e.g., the 535 results in Figure 4) captures both the influence of regional-scale non-local effects and the finer-536 scale direct effects. While uncertainties remain regarding the characterization of forest age and 537 historical land cover, our results support the expectation that 20th-century EUS reforestation had 538 a net cooling effect that extended well beyond the surface and local stand-scale. For weather 539 stations located near to, but outside of, reforested areas, the effect amounts to a suppression of T_a on the order of 0.5 - 1 °C for the latter half of the 20th century. 540

541 **4.3 Non-local effects of land-use change.** A more holistic perspective on the efficacy of nature-542 based climate solutions accounts for the possibility that local (e.g., ecosystem-scale) changes in 543 land cover can initiate non-local impacts over much broader scales (Swann et al., 2012; Williams 544 et al., 2021). For example, reforestation can lead to increased evapotranspiration, resulting in 545 increased cloud cover and precipitation (Cerasoli et al., 2021; Manoli et al., 2016) that extend 546 across the landscape, which would tend to amplify local cooling effects, particularly during the 547 daytime. Localized land-cover changes can also cause shifts in atmospheric circulation, which 548 can have continental or even global-scale consequences for temperature, precipitation,

549 cloudiness, and other meteorological drivers (Pongratz et al., 2010; Swann et al., 2012; Winckler 550 et al., 2019). Earth system models have been the primary tool for exploring these 551 teleconnections, with many studies suggesting a dramatic role for non-local processes to amplify 552 or counteract the local impacts of land-cover change on surface and near-surface temperature. 553 However, results from these modeling studies are sometimes contradictory, and the difference 554 between model predictions and observations can be large (Bonan, 2008; De Hertog et al., 2022).

555 This study employs observational approaches to indirectly assess the non-local effects of land-

556 cover change. We investigate how surface temperature effects can extend to the air and be

transported across the landscape, providing a generalizable method to investigate these connections. Our approach includes exploring gradients in surface temperature (T_s) across

econvectoris. Our approach includes exploring gradients in surface temperature (1_s) across ecosystem boundaries, as shown in Figure 3, to uncover the potential local extent of such effects.

560 This connection is crucial because long-term records of air temperature, which are used to

561 delineate the warming hole, are made in open clearings. By uncovering how surface temperature

562 effects extend to the air, we can gain insight into how local changes in land cover may initiate

563 non-local impacts across much broader scales. It is important to note that these observational

approaches have limitations, as they cannot fully account for the possibility that non-local effects

565 of land-use change within and outside the study region are influencing long-term temperature

- trends in the EUS.
- 567

568 For example, changes in agricultural management, such as agricultural intensification and

569 increased irrigation use, are known to have a local cooling effect (Mueller et al., 2016), and

570 many areas of the EUS have experienced these management shifts (Spangler et al., 2020).

571 Changes in agricultural management that promote cooling could obscure the influence of

572 reforestation in the comparison of long-term trends from forested and cropland ecosystems (e.g.,

573 Figure 4). Thus, the results presented here can be considered conservative. It is also possible that

574 land-use changes occurring well outside of the study region may be driving widespread trends in

575 long-term temperature trends in the EUS via teleconnections (Swann et al., 2018), or that certain

576 non-local temperature effects of reforestation or deforestation may be opposite in sign to the 577 local effects (De Hertog et al., 2022; Pongratz et al., 2021). These non-local teleconnections are

578 currently only possible to explore with modeling studies that tend to rely on idealized

579 experiments that force instantaneous and extreme changes in land cover across broad regions.

580 Evaluating all modes of historical land-use change that might have affected climate in the EUS

581 would be computationally expensive (if not impossible) and would still be unlikely to resolve,

582 with precision, the relatively small changes in T_s and T_a that are revealed by our observation-

583 driven approach.

584

585 4.4 Conclusions. Various hypotheses have been proposed to explain the observed lack of 20th-586 century warming in the eastern United States (e.g., Meehl et al., 2012; Z. Pan et al., 2004; 587 Partridge et al., 2018; Tosca et al., 2017). Our work does not identify widespread reforestation as 588 the sole factor causing the EUS warming hole or its trend, but multiple independent data sources 589 suggest it may have been an important contributor to this lack of historic regional warming. 590 Beyond that, our study provides robust evidence of local biophysical climate benefits of 591 reforestation in the EUS. The strong and persistent increase in forest cover throughout the region in the 20th century contributed to cooling, which is consistent with observed temperature 592

593 changes. In addition, our findings demonstrate that reforestation has a consistent cooling effect 594 on both surface and air temperatures, especially during midsummer periods when high 595 temperatures can be most harmful. These findings emphasize the potential for reforestation to 596 provide local climate adaptation benefits in temperate regions such as the EUS, highlighting the 597 importance of biophysical co-benefits of nature-based climate solutions.

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613

614 **Open Research**

615 The data used in this study are freely accessible. The AmeriFlux tower data are available from

- 616 the AmeriFlux data portal (<u>https://ameriflux.lbl.gov/</u>), and the various remote sensing and
- 617 meteorological network data are available from links and repositories described in the methods
- 618 section. To ensure reproducibility, we will make intermediate data products, code, and
- 619 transformed data products (i.e., those derived from raw publicly available data) publicly
- available on a Dryad repository prior to publication of this manuscript. The Dryad repository will
- have a persistent DOI identifier to ensure accessibility and citation. If site data are not yet
- available for download via AmeriFlux at the time of publication, we will include the raw data inour Dryad repository.
- 624
- 625 *Data are available for peer-reviewers at the following Dryad repository:
- 626 <u>https://datadryad.org/stash/share/Xc8kjs7DMNAeTdHFfTLWC2gTXV3zip8ZMxTN3N-6mdQ</u> 627
- 628 To facilitate access to the data, we will provide direct links to the data or detailed instructions on
- 629 how to access the data efficiently. We will also follow any acknowledgments or citation
- 630 guidelines provided by the repository or data/software source when creating our final
- 631 Availability Statement and referencing the data in our References section.
- 632

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