Projection of future fire emissions over the contiguous US using explainable artificial intelligence and CMIP6 models

Sally S.-C. Wang¹, L Ruby Leung¹, and Yun Qian¹

¹Atmospheric Sciences and Global Change Division, Pacific Northwest National Laboratory

February 1, 2023

Abstract

Increasing temperature and water cycle changes due to warming climate may increase wildfire activities. Reliable projections of fire emissions are critical for informing fire management to address fire impacts on societies and ecosystems. Here, we construct a neural network (NN) model explained by the Shapley Additive explanation (SHAP) to predict fire $PM_{2.5}$ emissions change and understand their drivers over the contiguous US (CONUS) in the mid-21st century under a high greenhouse gas emissions scenario (SSP5-8.5). Using future meteorology and leaf area index (LAI) simulated by eight global climate models from the Coupled Model Intercomparison Project Phase 6 (CMIP6), future population density, and present-day land use and land cover (LULC) as input to the NN model, the total fire $PM_{2.5}$ emissions over CONUS are projected to increase by 4-75% (model spread). Among different regions, fire emissions in the western US are projected to increase more significantly in June-July-August (JJA) than in other seasons and regions, with the median ratios of future to present-day fire emissions ranging from 1.67 to 2.86. The increases in fire emissions are mainly driven by increasing normalized temperature (23-29%) and decreasing soil moisture (2-10%) in the future. When future LULC change is considered, the projected fire emissions further increase by 58%-83% over the western US compared to projections without LULC change because of future increases in vegetation fraction. The results highlight the important role of warmer temperature, decreasing soil moisture, and LULC change in increasing fire emissions in the future.

Projection of future fire emissions over the contiguous US using explainable artificial intelligence and CMIP6 models

3 Sally S.-C. Wang¹, L. Ruby Leung¹, and Yun Qian¹

4 ¹ Atmospheric Sciences and Global Change Division, Pacific Northwest National Laboratory,

5 Richland, Washington, USA

6

7	Corresponding author: Sally SC. Wang (sing-chun.wang@pnnl.gov) and L. Ruby Leung
8	(<u>Ruby.Leung@pnnl.gov</u>)

9 Key Points:

- Using explainable AI and CMIP6 outputs, total fire PM_{2.5} emissions over the US are projected to increase by 4-75% by mid-century.
- Fire emissions will increase by 67-186% in the western US in summer, due to reduced soil moisture and warmer temperature in the future.
- Land use and land cover change adds an extra ~50% increase of future fire emissions,
 driven by increased vegetation in the western US.

16

17 Abstract

- 18 Increasing temperature and water cycle changes due to warming climate may increase wildfire
- 19 activities. Reliable projections of fire emissions are critical for informing fire management to
- 20 address fire impacts on societies and ecosystems. Here, we construct a neural network (NN)
- 21 model explained by the Shapley Additive explanation (SHAP) to predict fire PM_{2.5} emissions
- change and understand their drivers over the contiguous US (CONUS) in the mid-21st century
- under a high greenhouse gas emissions scenario (SSP5-8.5). Using future meteorology and leaf
- 24 area index (LAI) simulated by eight global climate models from the Coupled Model
- Intercomparison Project Phase 6 (CMIP6), future population density, and present-day land use and land cover (LULC) as input to the NN model, the total fire PM_{2.5} emissions over CONUS are
- 27 projected to increase by 4-75% (model spread). Among different regions, fire emissions in the
- western US are projected to increase more significantly in June-July-August (JJA) than in other
- seasons and regions, with the median ratios of future to present-day fire emissions ranging from
- 30 1.67 to 2.86. The increases in fire emissions are mainly driven by increasing normalized
- temperature (23-29%) and decreasing soil moisture (2-10%) in the future. When future LULC
- 32 change is considered, the projected fire emissions further increase by 58%-83% over the western
- 33 US compared to projections without LULC change because of future increases in vegetation
- 34 fraction. The results highlight the important role of warmer temperature, decreasing soil
- 35 moisture, and LULC change in increasing fire emissions in the future.
- 36

37 Plain Language Summary

Climate change has increased fire frequency and size in the United States, particularly in the western US, causing property damage, threats to human life, and degraded air quality. Frequent

- 40 drought, enhanced fuel aridity, and increased fuel accumulation may cause larger and more
- frequent fires and emissions in the coming decades. Using explainable artificial intelligence
- 42 (XAI) and outputs from global climate models, we show that total fire PM_{2.5} emissions over the
- 43 contiguous US will be around 1.38 times the present-day emissions by the mid-21st century
- 44 under a high greenhouse gas emissions scenario. The fire PM_{2.5} emissions will double in the
- 45 western US during summer, mainly driven by drying trends in the soil along with increasing
- temperatures. When the future change of land use and land cover is also considered, the fire
- emissions in the western US will increase further by 50%, due to the increased vegetation
- 48 fractions in a warmer climate.

49 **1 Introduction**

Wildfires have become larger and more frequent across the United States over the past 50 two decades, with more notable changes in the western US. Several studies have suggested that 51 the increased wildfire burned areas are driven by increased fuel aridity, warming temperature, 52 53 and fuel management (Abatzoglou & Williams, 2016; Burke et al., 2021; Marlon et al., 2012; Westerling et al., 2006). With the increased fire activities in the recent five years, the burned area 54 has increased by ~ 3 times and costs in fire suppression have increased by ~ 8 times in the recent 55 decades compared to the 1980s (NIFC, 2022). The fires emitted large amounts of fine 56 particulates with a diameter of 2.5 µm or less (PM_{2.5}), leading to degraded air quality and 57 increased exposure to smoke from fires (Jaffe et al., 2020; Kaulfus et al., 2017; Liu et al., 2018; 58 59 O'Dell et al., 2019). Understanding how fire emissions will change under future climate and the

- 60 key factors controlling the changes in future fire emissions is important for fire management and
- 61 planning to reduce the impacts on human health and air quality in the coming decades.

Several studies have projected increased fire emissions over the contiguous US (CONUS) 62 in the future, mainly the western US, using different approaches (Ford et al., 2018; Liu et al., 63 2021; Neumann et al., 2021; Val Martin et al., 2015; Xie et al., 2022; Yue et al., 2013; Table 1). 64 One commonly used approach is the process-based fire model embedded in the Dynamic Global 65 Vegetation Model (DGVM), which uses empirical functions of fire activity and climate, 66 vegetation, and socioeconomic variables to estimate burned area and fire emissions (Pechony & 67 Shindell, 2009; Thonicke et al., 2010). For example, Ford et al. (2018) projected 54% and 50% 68 increases in organic carbon (OC) and black carbon (BC) over CONUS in mid-century under the 69 70 Representative Concentration Pathway (RCP) 8.5 scenario, using a process-based fire model implemented in Community Land Model (CLM) driven by meteorological fields simulated by 71 Community Earth System Model (CESM). Using the Sixth Coupled Model Intercomparison 72 73 Project (CMIP6) multimodel and multiensemble simulations, Xie et al. (2022) projected 130-260% increases in fire CO₂ emissions over the western US under the Shared Socioeconomic 74 Pathways (SSP) 5-8.5 scenario in late 21st century. In addition to DGVM, another widely used 75 approach is the statistical regression model, which is based on the statistical relationship between 76 fires and the associated predictors to predict burned area and fire emissions. Yue et al. (2013) 77 used regression and parameterization driven by climate simulations and projected increases of 78 80-170% in OC and BC over the western US in the mid-21st century. Recently, Liu et al. (2021) 79 used an empirical fire model with DGVM and projected 50% increases in fire PM_{2.5} emissions 80 over the western US in the mid-21st century, considering the changes in both climate and fuel. 81

The approaches mentioned above have their strengths and weaknesses. Process-based fire 82 models with DGVMs can simulate the non-linear relationships between fires and predictors and 83 capture the feedback among climate, vegetation, and fires. However, they often assume the same 84 relationships across the globe or regions and apply universal functions or the same set of 85 86 parameters to all the grid cells in a model (Pechony & Shindell, 2009; Thonicke et al., 2010). In addition, process-based models with DGVMs are more computationally expensive than 87 regression models, which are computationally efficient and can achieve promising performance. 88 89 However, the good performance of regression models is limited by the assumption of linear relationships between fires and predictors, which work well at large spatial scales (e.g., 90 ecoregions) but may not be applicable for projecting fire changes considering the non-linear 91 92 relationships among multiple factors at finer spatial scales. Combining the advantages of the two approaches, machine learning (ML) provides a solution to model the non-linear relationships 93 between fires and predictors at fine spatial scales efficiently. Prior studies have used various ML 94 approaches to predict fire occurrence, burned area, and fire emissions at different temporal and 95 spatial scales (Birch et al., 2015; Coffield et al., 2019; Cortez & Morais, 2007; Dillon et al., 96 2011; Kane et al., 2015; Wang et al., 2021; Wang & Wang, 2020). Recently, to fill the gaps of 97 interpretability of ML, explainable artificial intelligence (XAI) has been developed and applied 98 to understand the relationships learned by the ML and identify the key contributing predictors 99 (Adadi & Berrada, 2018; Arrieta et al., 2020). 100

Here, we leverage the power of XAI and the atmospheric forcing and vegetation from the
 CMIP6 multimodel simulations to project fire emissions over CONUS in the mid-21st century
 (2050-2065) under the SSP5-8.5 scenario. We first construct an artificial neural network (ANN)

- 104 to predict monthly fire PM_{2.5} emissions using predictors of local and large-scale meteorology,
- 105 land surface characteristics, and population density from observations and reanalysis data. We
- then evaluate the performance of the ANN model driven by the historical simulations of eight 107
- global climate models (GCMs) and identify the GCM biases that influence the fire emissions
 predicted by the ANN model. Two experiments are conducted for future projection. For the first
- predicted by the ANN model. Two experiments are conducted for future projection. For the first experiment, future fire emissions are projected by driving the ANN model with the future
- climate projections and LAI from the eight GCMs, future population density, and present-day
- 111 land use and land cover (LULC) from satellite observations. For the second experiment, we
- select four of the eight GCMs that provide future LULC outputs and drive the ANN with future
- climate projection, LAI, population density, vegetation distributions, and LULC. We compare
- the future projections with and without LULC change for the same four GCMs to isolate the
- 115 effects of LULC change on future fire emissions.

Fire model	Region	Scenario	Perio d	# GCM s	Projected changes	Note	Ref
Regression Parameterizatio n	WUS, 1980-2004	CMIP3 A1B scenario	2046- 2065	15	increase~80 % for OC and BC ~170% and 150% increase for OC and BC	Climate factor only; assume fuel consumption remains constant	Yue et al. (2013)
DVGM (CESM fully coupled)	WUS,200 0	RCP 4.5 RCP 8.5	2050	1	60% increase for OC decrease of 0.3% for OC	Online computed meteorology and prescribed sea-surface and sea-ice distributions	Val Martin et al. (2015)
DVGM (CLM)	CONUS, 2001-2010	RCP4.5 RCP8.5	2040- 2050 2090- 2100 2040- 2050 2090- 2100	1	~125% increase for OC and BC 54% and 50% increase for OC and BC 150% and 141% increase for OC and BC	Future simulations were driven by meteorology from archived CESM1 simulation	Ford et al. (2018)
Statistical model from Prestemon et al. (2016)	SEUS, 1992-2010	Scenario A2	2011- 2060	1	13-62% lower for PM 2.5	They include climate and socioeconomi c variables	Shankar et al. (2018)
Statistical model (empirical) and DGVM	WUS, 2001-2010	CMIP5 RCP 8.5 (meteorology for number/burne d area) +	2050- 2059	1	50% increase for PM2.5	They used two climate datasets (with different scenarios) for projection	Liu et al. (2021)

116 **Table 1.** Prior studies projecting future fire emissions over CONUS

		CMIP3 SRES- A2 (fuel loading and moisture)					
Regression (from Yue et al. (2013))	WUS, 1996-2005	RCP 4.5	2046- 2055	5	Average 8% increase for OC and BC (GISS-E2-R; small)	Climate factor only	Neuman n et al. (2021)
		RCP 8.5	2086- 2095		Average 17% increase for BC and OC (GISS-E2-R; small)		
DVGM (fully- coupled)	WUS, 2000-2014	CMIP6 SSP1- 2.6 CMIP6 SSP2- 4.5 CMIP6 SSP5-	2080- 2100	3	60-110% increase for CO2 (Aug- Sep) 100-150% increase for CO2 (Aug- Sep) 130-260%		Xie et al. (2022)
		8.5			increase for CO2		

118 **2 Data**

119 2.1 Fire-induced PM_{2.5} emission data

Monthly fire PM_{2.5} emission data for training and evaluating XAI is obtained from the 120 Global Fire Emissions Database (GFED) version 4, with a spatial resolution of 0.25° available 121 from 1997 to present. GFED obtains burned area from MODIS (MCD64A1) and provides fire 122 PM_{2.5} emissions estimated by the burned area, the emission factors from Akagi et al. (2011), and 123 the fuel loads and combustion completeness from the Carnegie-Ames-Stanford Approach 124 (CASA) biogeochemical model (van der Werf et al., 2017). In this study, the GFED fire PM_{2.5} 125 emission is the predictand of the neural network model. To match the spatial resolutions of the 126 CMIP6 models, we regrid the fire emissions to $1^{\circ} \times 1^{\circ}$ by area-weighted averaging. The target 127 grid is computed as a weighted mean of all the grids from the source grids that intersect with the 128 target grid and the weighting factor is the area of the intersection with each of the source grids. 129

130 2.2 Predictor data from observations and reanalysis for model training

131 We develop a neural network at $1^{\circ} \times 1^{\circ}$ grid resolution driven by predictor variables at a 132 monthly scale from 2000 to 2020. The predictor variables are regridded to $1^{\circ} \times 1^{\circ}$ by area-133 weighted averaging. Table 2 shows the predictors in the model with their sources and original

- 134 spatial and temporal resolutions. Compared to the variables used in Wang et al. (2022), this study
- used a subset of the variables that are accessible in the CMIP6 outputs.
- 136

137 **Table 2.** Predictor variables used in the NN model

Variables	Abbreviation	Categories	Temporal resolution	Spatial resolution	Data Source	References
Monthly mean relative humidity	RH	Local meteorology	monthly	32 km	North American Reanalysis (NARR)	Mesinger et al. (2006)
Monthly mean of daily precipitation	арср	Local meteorology	monthly	32 km	North American Reanalysis (NARR)	Mesinger et al. (2006)
Monthly mean wind speed	wndsp	Local meteorology	monthly	32 km	North American Reanalysis (NARR)	Mesinger et al. (2006)
Monthly Standardized Precipitation Evapotranspiration Index	SPEI	Local meteorology	monthly	0.5°×0.5°	SPEI	Vicente- Serrano et al. (2010)
Monthly mean vapor pressure deficit	VPD	Local meteorology	daily	4 km	gridMET	Abatzoglou (2013)
Monthly normalized meteorology (normalized temperature, RH, apcp, and VPD)	Nor.temp, nor.rhum, nor.apcp, nor.VPD	Local meteorology	monthly	32 km/ 4 km	North American Reanalysis (NARR) /gridMET	Mesinger et al. (2006); Abatzoglou (2013)
Monthly standard deviation of daily SVDs for northern California	SVD1_NCA and SVD2_NCA	Large-scale meteorological patterns	monthly	Regional	North American Reanalysis (NARR)	Wang et al. (2021)
Monthly standard deviation of daily SVDs for southern Rocky Mountain	SVD1_SRM and SVD2_SRM	Large-scale meteorological patterns	monthly	Regional	North American Reanalysis (NARR)	Wang et al. (2021)
Monthly standard deviation of daily SVDs for southeastern US (with 2-month lag)	SVD1_SElag2 and SVD2_SElag2	Large-scale meteorological patterns	monthly	Regional	North American Reanalysis (NARR)	Wang et al. (2021)
Monthly mean surface soil moisture	soilm	Land-surface properties	monthly	0.25°×0.25°	Global Land Data Assimilation System (GLDAS-2)	Xia et al. (2012)
Monthly mean vegetation fraction	Veg_frac	Land-surface properties	monthly	0.25°×0.25°	Global Land Data Assimilation System (GLDAS-2)	Xia et al. (2012)

Monthly mean Leaf Area Index	LAI	Land-surface properties	8 days	500 m	MODerate resolution Imaging Spectroradiometer (MODIS); LAI classification scheme	Myneni et al. (2015)
Land cover fraction	Grass, tree, shrub, crop, nonveg, resid	Land-surface properties	Yearly	0.05°×0.05°	MODerate resolution Imaging Spectroradiometer (MODIS); LAI classification scheme	Friedl (2015)
Population density	Рор	Socioeconomic and coordinate variables	Yearly	30 arc	Gridded Population of the World data collection (GPW v4)	CIESIN- Columbia University (2017)

139 2.2.1 Local meteorology

140	Local meteorology includes monthly mean relative humidity (RH) at 2m, total
141	precipitation, and wind speed at 10 m from the North American Regional Reanalysis (NARR
142	(Mesinger et al., 2006) and vapor pressure deficit (VPD) from the gridMET dataset (Abatzoglou
143	& Kolden, 2013). As drought dries the fuels and facilitates fire development, we also include the
144	monthly Standardized Precipitation Evapotranspiration Index (SPEI) (Vicente-Serrano et al.,
145	2010), which measures the drought severity based on climatic data as predictors.

To provide more information on the temporal variability of meteorology, we also include normalized meteorology, which is calculated by subtracting the long-term mean from 2000 to 2020 and dividing it by the standard deviation. We include the normalized monthly mean surface temperature, relative humidity (RH) at 2m, total precipitation, and VPD.

150 2.2.2 Large-scale meteorological patterns

Large-scale meteorological patterns influence wildfire occurrence and intensity by 151 modulating local meteorological conditions (Crimmins, 2006; L. Dong et al., 2021; Trouet et al., 152 2009; Zhong et al., 2020). Wang et al. (2021) showed that including predictors of large-scale 153 meteorological patterns associated with wildfires improves burned area prediction over CONUS. 154 Therefore, we include predictors representing the synoptic patterns driving fire emission 155 variability, constructed using the singular value decomposition (SVD) method (Wang et al., 156 2022). The leading nodes of SVDs were identified for the three regions where large fires 157 periodically occur, including northern California, southern Rocky Mountains, and the 158 southeastern US, as defined in Wang et al. (2021). For each region, we calculate the daily mean 159 fire PM2.5 emissions over the region and compute the day-to-day correlations between the 160 regional mean fire PM_{2.5} emissions and the five gridded daily meteorological variables (surface 161 temperature, 2-meter RH, U-wind and V-wind at surface, and geopotential height at 500 hPa 162 from NARR) for all $1^{\circ} \times 1^{\circ}$ grid cells within the large-scale domain, resulting in a correlation 163 map for each meteorological variable. The correlation maps are then used to derive the SVD 164

- 165 modes representing the large-scale meteorological patterns related to daily fire emissions.
- 166 Finally, we compute the monthly standard deviation of the daily SVD time series for the first two
- 167 SVD modes, representing the month-to-month variations of synoptic fluctuations and
- atmospheric instability. Note that we use U- and V-wind at the surface rather than at 850 hPa
- used in Wang et al. (2022) due to data availability in the CMIP6 archive. The detailed methods
- and discussions about the SVDs are provided in Wang et al. (2021) and Wang et al. (2022).

171 2.2.3 Land-surface properties and socioeconomic variables

We include predictors of monthly mean surface soil moisture, leaf area index (LAI), and 172 vegetation fraction from the North American Land Data Assimilation System (NLDAS-2) to 173 capture the effects of fuel availability and flammability (Xia et al., 2012). Land cover fraction of 174 175 the LAI classification scheme is obtained from the Terra and Aqua combined MODIS Land Cover Climate Modeling Grid (CMG) Version 6 data (Friedl & Sulla-Menashe, 2015). The LAI 176 classification scheme has 12 land types, including savannas, evergreen and deciduous broadleaf 177 forest, evergreen and deciduous needleleaf forest, grass, shrub, broadleaf crop, non-vegetated 178 179 land, urban, water bodies, and unclassified land. To match the land cover types in CMIP6, we combine savannas, evergreen and deciduous broadleaf forests as well as evergreen and 180 181 deciduous needleleaf forests in the MODIS data to one type, corresponding to the "tree" in the GCMs; we also combine urban, unclassified, and water bodies to a single type, corresponding to 182 183 "residual". The grass, shrub, broadleaf crop, and non-vegetated lands correspond to the grass, shrub, crop, and bare soil in the GCMs. This results in six different land cover types matching 184 185 the land types defined in CMIP6, including tree, grass, shrub, crop, bare soil, and residual. Note that the MODIS land cover data starts from 2001, so we use the data of 2001 for 2000. 186

Population density is included to represent human effects on wildfires. The population density data is obtained from the Gridded Population of the World (GPW) data collection for the years 2000, 2010, 2015, and 2020, with a spatial resolution of 30 arc-second (CIESIN-Columbia University, 2017). The populations in other years are linearly interpolated between the four years.

192 2.3 CMIP6 model data

We use the monthly output from the eight selected GCMs in the CMIP6 archive, 193 including the Australian Community Climate and Earth System Simulator Earth System Model 194 Version 1.5 (ACCESS-ESM 1.5), Community Earth System Model Version 2 (CESM2), Euro-195 196 Mediterranean Centre on Climate Change couped climate model standard configuration (CMCC-CM2-SR5), EC-Earth3 coupled Climate-Carbon Cycle (EC-Earth-CC), EC-Earth3 coupled with 197 the second-generation dynamic global vegetation model LPJ-GUESS (EC-Earth-Veg), EC-198 Earth-Veg with lower resolution (EC-Earth-VegLR), Geophysical Fluid Dynamics Laboratory 199 Earth System Model Version 4.1 (GFDL-ESM 4.1), and the Canadian Earth System Model 200 version 5 (CanESM 5) (Eyring et al., 2016; O'Neill et al., 2016). Model outputs include monthly 201 mean RH, surface wind speed, total precipitation, LAI, and surface soil moisture. SPEI and VPD 202 203 are computed based on monthly meteorology from CMIP6 data. SVDs are calculated from daily surface temperature, 2-meter RH, U-wind and V-wind at the surface, and geopotential height at 204 500 hPa. The CMIP6 model outputs are all regridded to $1^{\circ} \times 1^{\circ}$ by area-weighted averaging. 205 Outputs from historical simulations and SSP5-8.5 future scenario from variant r1i1p1f1 are used 206

for the present-day (2000-2014) simulation and future (2050-2065) projection. The SSP5-8.5

- 208 projects CO_2 emissions increasing strongly until 2080 and then slightly declining until the end of
- the century, with a radiative forcing peaking at 8.5 W/m² and a global mean temperature increase 140Gh = 2100
- of around 4°C by 2100 compared to the current era (Eyring et al., 2016; IPCC, 2022). The future population density data is obtained from the Global One-Eighth Degree Population Base Year
- and Projection Grids Based on the Shared Socioeconomic Pathways for 2010-2100, with a
- spatial resolution of 0.125° (Jones & O'Neill, 2020). We use the projected population density
- 214 under the SSP5 scenario.

As mentioned before, this study includes two projection experiments with four simulations, as shown in Table 3. The first experiment (land_fix) compares simulations 2 and 1, representing the impacts of future changes in meteorology, LAI, and population density on future fire emissions. The second experiment (land change) compares simulations 4 and 3, considering the effects of changes in meteorology, LAI, population density, as well as LULC on future fire emissions. The four GCMs are GFDL-ESM4, EC-Earth3-Veg, EC-Earth3-VegLR, EC-Earth-

221 CC, which provide future natural vegetation and land use distribution (Döscher et al., 2022;

222 Dunne et al., 2020; Hurtt et al., 2020; Lawrence et al., 2019; Song et al., 2021; Swart et al., 2019;

223 Ziehn et al., 2020). Accordingly, we use the outputs of vegetation distribution and LULC from 224 the four GCMs to project future fire emissions considering the effects of LULC change. The

the four GCMs to project future fire emissions considering the effects of LULC change. The outputs of vegetation distribution and LULC include monthly vegetation fraction and yearly land

cover fraction (tree, crop, grass, shrub, bare soil, and residual). Hereafter CMIP6 refers to the

227 CMIP6 models used in this study.

228	Table 3. Experiments and simulations with the sources and states of the variables used in this
229	study

Experiment	Simulation number	Meteorology and LAI	Population density	LULC
1. Land_fix (using eight	1	GCM present-day	GPW present-day	MODIS present- day
GCMs)	2	GCM future	GPW future	MODIS present- day
2. Land_change (using four	3	GCM present-day	GPW present-day	GCM present-day
GCMs)	4	GCM future	GPW future	GCM future

230

3 Method

232

233 **3.1 Artificial neural Network (ANN)**

ANN or neural network (NN) consists of several layers with interconnected neurons. The layers

include input layer, hidden layer, and output layer. Each neuron in the input layer represents one

- predictor (e.g., monthly RH, surface soil moisture etc.) and the neuron in the output layer is the
- target which is fire $PM_{2.5}$ emission (g m⁻²month⁻¹). During the training process, each neuron takes the weighted average from other neurons in the previous layer and transforms the average by the

the weighted average from other neurons in the previous layer and transforms the average by the nonlinear activation functions. The weights are randomly initialized for all the neurons at the

beginning of the training process. The same process applies to the neurons in the next layer until

the output layer is reached and the errors between the final output and the target are measured

using a loss function. Then the error is backpropagated from the last layer to the front layer to

calculate the contributions of the weights and update the weights to minimize the errors between

the final output and target using an optimization algorithm.

In this study, we construct an NN model with five layers, including one input layers with 245 25 neurons corresponding to the 25 predictors, three hidden layers with 200, 150, and 80 246 neurons, and one output layer corresponding to the predicted fire emission. The input variables 247 are normalized using min-max normalization; that is, we subtract the minimum value of the 248 predictor from each predictor value and divide the result by the difference between the maximum 249 and minimum value, resulting in values ranging between -1 to 1. The formula can be presented 250 as: $x' = \frac{x - \min(x)}{\max(x) - \min(x)}$, where x is the original value and x' is the normalized value. The 251 optimization algorithm is stochastic gradient descent (SGD), the learning rate is 0.01, and the 252 batch size is 64. The loss function is the mean square error (MSE), and the activation function is 253 254 the Rectified Linear Unit (ReLU).

255 **3.2 Model evaluation**

To evaluate the model, we apply the 10-fold cross validation (CV) technique. The whole dataset during 2000-2020 is randomly separated into ten equal-size splits. For each round of CV, the model is trained with nine splits of the data and the trained model is then used to predict burned area in the remaining one split. The final evaluation is based on root mean square error (RMSE). Besides RMSE, we also use the correlation coefficient and the index of agreement (IoA) between the observed and predicted fire emissions to evaluate our model performance. The formula of IoA can be expressed as:

263
$$IoA = 1 - \frac{\sum_{i=1}^{N} (y_i - \hat{y}_i)^2}{\sum_{i=1}^{N} (|\hat{y}_i - \bar{y}| + |y_i - \bar{y}|)^2}$$

where y_i is the observations, \hat{y}_i is prediction, and \bar{y} is the mean of the observations. The value of IoA ranges between 0 and 1, with values closer to 1 indicating a better fit.

266 **3.3 Shapley Additive explanation (SHAP)**

SHAP is an innovative approach that uses game theory to explain variable importance globally (i.e., the whole dataset) and locally (i.e., one sample) (Lundberg & Lee, 2017). Under the scope of SHAP, the predictors are the "players" in a co-operative game in which the goal is a prediction for a single target. Each predictor has its "playout" representing its contribution to the prediction, considering all possible combinations of the predictors. To calculate the predictor

- 272 contribution for predictor *i*, the SHAP value considers the differences in the model's predictions
- 273 f_x made by including and excluding the predictor *i* for all the combinations of predictors:
- 274

275
$$\phi_i = \sum_{S \subseteq F \setminus \{i\}} \frac{|S|! (F - |S| - 1)!}{F!} [f_x(S \cup \{i\}) - f_x(S)]$$

where ϕ_i is the weighted average of all marginal contribution of predictor *i*, *F* is the total number of features, *S* is the subset of predictors from all predictors except for predictor *i*, $\frac{|S|!(F-|S|-1)!}{F!}$ is the weighting factor counting the number of permutations of the subset *S*. $f_x(S)$ is the expected output given the predictors subset *S*. $[f_x(S \cup \{i\}) - f_x(S)]$ is the difference made by predictor *i*.

SHAP has been widely used to explain different machine learning models in many fields (Padarian et al., 2020; Stirnberg et al., 2021; Wang et al., 2021). It has been applied to deep neural network ("Deep SHAP") based on DeepLIFT method in Lundberg and Lee (2017), which

approximates the conditional expectations of SHAP values using a selection of background

samples.

285

286 4 Results

287 4.1 NN performance and variable importance

Figure 1 shows the maps of the observed and predicted fire emissions averaged over 288 2000–2020. The model reproduces the spatial patterns of fire emissions and captures the large 289 fire emissions over northern California. The spatial correlation between the observed and 290 291 predicted fire emissions is 0.79, showing a good agreement between the long-term observations and predictions. Additionally, the model can capture the interannual variability over CONUS, 292 with a correlation of 0.94 and RMSE of 4.84 g/m^2 . To assess model performance in different 293 regions, Figures S1a-d show the time series of observed and predicted fire PM_{2.5} emissions 294 averaged over the selected regions. The selected regions include the western forest region 295 (WFR), Mediterranean southern California (SCA), southwestern US (SWUS), and southeastern 296 297 US (SEUS) (color boxes in Figure 1b). These regions frequently burn and share similar fire regimes and vegetation types. The predicted interannual variability resembles the observed 298 interannual variability for the selected regions (IoA=0.31-0.95). However, the southwestern and 299 300 southeastern US have much smaller IoA and larger RMSE than the other two regions in the western US (Table 4). The model has larger biases over the regions with much lower fire 301 emissions (e.g., southeastern US) because the NN is dominated by grid cells that have large fire 302

303 emissions (i.e., western US) (Zhu et al., 2022).



Fig. 1. The burned area map averaged over 2000 to 2020 for (a) observation and (b) prediction; (c) Time series of observed (black) and predicted (red) monthly fire PM_{2.5} emissions averaged across CONUS. The color boxes in (b) denote four analysis regions: western forest area (WFR, red), Mediterranean southern California (SCA, blue), southwestern US (SWUS, dusty), and

309 southeastern US (SEUS, pink).

310	Table 4. Model performance aggregated over the regions, including the western forest region
311	(WFR), southern California (SCA), southwestern US (SWUS), and the southeastern US (SEUS),

	CONUS	WFR	SCA	SWUS	SEUS
IoA	0.94	0.88	0.95	0.31	0.44
RMSE (g m ⁻²)	4.84	5.02	0.73	0.37	0.94

312

To understand the leading variables in the NN model at different time scales, we analyze the absolute SHAP values at seasonal and interannual scales by averaging the SHAP values of all the grids over CONUS for each year and month and take the absolute values. Predictors with larger |SHAP| have greater contribution and therefore, more important to fire emission. Figure 2 shows the mean |SHAP| values at seasonal and interannual time scales for CONUS. At both time scales, local meteorology and land-surface variables are critical, as most of them are included in

- the top 10 variables. Specifically, normalized temperature, normalized RH, RH, SVD2_sElag2,
- 320 and soil moisture are the predominant variables controlling both interannual and seasonal
- variability of the fire emissions. Land-surface variables (LAI, vegetation fraction, and grass
- fraction) and VPD have larger contributions to fire emissions at the seasonal scale, as they have
- strong seasonality. Some variables, including SVD1_NCA, SVD1_SElag2, and SPEI, are more
- 324 important at the interannual time scale.



Fig. 2. Variable importance represented by the mean |SHAP| values at seasonal and interannual time scale.

328

329 **4.2 GCM-NN present-day performance**

Before utilizing the NN and GCM outputs to project future fire emissions, it is helpful to 330 331 evaluate the prediction of present-day fire emissions by the NN with GCM meteorology and LAI, GPW population density, and MODIS land cover from 2000 to 2014 as predictors. The NN 332 driven by GCM outputs is referred to as GCM-NN. Figure 3 shows the ratios of predicted to 333 observed total fire emissions for CONUS and the four regions for the eight GCMs. The GCM-334 NN reproduces the present-day fire emissions well for CONUS, with a median ratio of 1.27. 335 Focusing on the four regions where fires frequently occur, GCM-NN slightly overestimates fire 336 emissions over southwestern US and southeastern US where the median values of the ratios are 337

- 1.18 and 1.12, respectively. The fire emissions over the western forest region are underestimated
- by the GCM-NN (Table 5 and Figure 3). To better understand the model performance, we also
- evaluate the seasonal variability and spatial distributions of the GCM-NN predicted fire
 emissions. Figure 4 shows the seasonal distribution of the observed and predicted fire emissions
- driven by the observation and reanalysis meteorology (red line in Figure 4) and GCM
- meteorology (purple line in Figure 4). The GCM-NN predictions generally reproduce the
- observed seasonality of fire emissions and capture the peak in August. Although the seasonality
- is well reproduced, the predicted spatial distributions of GCM-NN are inconsistent with the
- 346 GFED fire emissions (Figure 5). For instance, the NN predictions driven by ACCESS, CESM,
- and CMCC outputs are overestimated over the Central Great Plains (Figures 5c-e). The NN
- driven by the outputs of CanESM and EC models overestimates fire emissions in the western
- forest region, mainly over the Rocky Mountains (Utah, Colorado, and Wyoming), and the NN
- 350 with GFDL outputs overestimates fire emissions over the southeastern US (Figures 5f-j).





354

Table 5. Observations (mean of GFED fire emissions) and multi-model projections of fire

emissions (Gg yr⁻¹) (median values of the 8 GCMs; $1Gg = 10^9$ g). The regions include western

357 forest region (WFR), southern California (SCA), southwestern US (SWUS), and southeastern US

358 (SEUS).

	Observed (2000- 2014)	Present-day (2000- 2014) NN model	Future (2050-2065) NN model
WFR	0.082 ± 1.006	0.061 ± 0.292	0.099 ± 0.336
SCA	0.051 ± 0.571	0.050 ± 0.097	0.070 ± 0.153
SWUS	0.029 ± 0.682	0.034 ± 0.063	0.069 ± 0.171
SEUS	0.044 ± 0.266	0.050 ± 0.245	0.055 ± 0.204
CONUS	0.041 ± 0.550	0.052 ± 0.266	0.062 ± 0.215



Fig. 4. Seasonal cycle of the averaged fire PM_{2.5} emission from GFED (black line), the NN model driven by reanalysis (red line), and the GCM-NN models (purple lines). The fire PM_{2.5}

364 emissions are averaged over CONUS during 2000-2014 by month.



365

Fig. 5. Spatial distributions of the mean fire $PM_{2.5}$ emissions (g m⁻² month⁻¹) averaged over 2000-

367 2014 from (a) GFED) and the NN model driven by observations (b) and different GCMs (c-j).

The inconsistency of the spatial patterns between GCM-NN and observations is 368 dominated by the spatial biases in summer (June-August), which is the peak season of fire 369 emissions (Figure S2). The spatial biases of the GCM-NN predicted fire emissions in summer 370 may be contributed by biases of the NN model, biases of the GCM outputs used as predictors for 371 the model, and interactions between these biases. To determine the relative contributions of the 372 NN model biases (reanalysis-NN minus GFED) and the GCM predictor biases (GCM outputs 373 minus reanalysis/observation), linear regression models are fitted to the model biases (GCM-NN 374 minus GFED) in June-July-August (JJA) for different regions using NN biases and predictor 375 biases at each grid point within the regions as predictors. Figure S3 shows the coefficient of the 376 predictors for each GCM and region and only the predictors with p<0.05 are shown. Predictors 377 with larger coefficients have larger contributions to the GCM-NN biases. For ACCESS, CESM, 378 and CMCC, the positive biases in the Central Great Plains are mainly contributed by the NN 379 380 biases (Figure S3e). The overestimations in the western forest region are due to biases in precipitation for CanESM while for the two EC models the fire emission biases are caused by 381 biases in normalized temperature and SPEI (Figure S3a). For the GFDL model, the 382 overestimations over southeastern US are mostly attributed to the NN biases (Figure S3d). 383

384 **4.3 Future fire emission projection**

We use the NN model with the future climate and LAI from the eight GCMs, future 385 population density from GPW, and present-day LULC from MODIS as predictors to project the 386 future fire emissions in 2050-2065 under the SSP5-8.5 scenario and compare with the GCM-NN 387 388 simulation driven with present-day GCM meteorology, GCM LAI, GPW population density, and MODIS LULC (i.e., land fix experiment in Table 3). This method assumes stationarity of the 389 relationships between fire emissions and their predictors learned by the NN model trained using 390 observations. The projections only consider the effects of changes in meteorology, LAI, and 391 population density and neglect the impacts of LULC changes. Figure 6 shows the spatial 392 distributions of the ratio of future to present-day fire emissions. All GCM-NN models project 393 394 increased fire emissions across the western US, with the largest enhancement over Pacific Northwest, northern California, and the southwestern US. For the southeastern US, five GCM-395

NN models project decreased fire emissions in the future, while CMCC, GFDL, and CanESM

397 project increased fire emissions.



Fig. 6. Spatial distributions of the ratio of future (2050-2065) to present-day (2000-2014) fire PM_{2.5} emissions. The ratio is shown in the base 10 logarithmic scale.

401

398

Over CONUS, the GCM-NN projects a median ratio of future to present-day fire 402 emissions of 1.38, showing increased future fire emissions (Figure 7a). Across the four regions, 403 the GCM-NN projects an increase in fire emissions of 13-115%, with the largest median ratio of 404 1.85 in the southwestern US (Figure 7a). Generally, more significant enhancement in fire 405 emissions is found over the western US, where the western forest region and southern California 406 have median ratios of 1.40 and 1.22, respectively. As for the southeastern US, different models 407 project increases and decreases in fire emissions, resulting in a relatively small median ratio of 408 1.11. We further analyze the median ratios of future to present-day fire emissions for each region 409 and season in Figure 7b. For CONUS, September-October-November (SON) and June-July-410 August (JJA) have the largest enhancement of fire emissions, with a median ratio of 1.7 and 411 1.58, respectively. Like CONUS, the western forest region is projected to have the largest 412 enhancement of fire emissions in SON (median ratio = 2.29) and the second largest enhancement 413 in JJA (median ratio = 1.67). For other regions, including southern California, southwestern US, 414

and the southeastern US, the GCM-NN projects the greatest enhancement in JJA, with a median

416 ratio of 1.82, 2.86, and 1.33, respectively (Figure 7b).

417



418

Fig. 7. (a) The ratio of future (2050-2065) to present-day (2000-2014) fire $PM_{2.5}$ emissions for

420 CONUS and the four regions; (b) the median values of the ratios from 8 GCMs in CONUS and

421 the four regions for each season.

Compared to the projections of fire carbon emissions by the GCMs (CESM2, EC3-Earth-422 CC, EC3-Earth-Veg, EC3-Earth-VegLR, and GFDL-ESM4.1) from CMIP6, the spatial patterns 423 424 predicted by GCM-NN are very similar for present-day and future (Figure S4). Although the GCM-NN projection of PM_{2.5} emissions is driven by the changes in meteorology, LAI, and 425 population density, while the GCM projection of carbon emissions considers the changes in all 426 427 factors including LULC changes, the predicted ratios of future to present-day fire emissions by the GCM-NN are within a similar range as the GCMs (Figures S5a and 5b). For the seasonality, 428 since only CESM2 and GFDL-ESM1.4 provide monthly output, we compare the mean values of 429 the two ratios for the four seasons over CONUS and the four regions. Both GCM and GCM-NN 430 431 project the largest enhancement in JJA over CONUS (Figures S5c and 5d). However, there are differences in the peak seasons of enhancement for the four regions. For example, GCM-NN 432 433 projects the largest enhancements of fire emissions in JJA, while GCM projects the largest enhancements in MAM for the western forest region and the southeastern US. The comparisons 434 demonstrate the GCM-NN projections are generally consistent with the process-based models 435 with DGVM embedded in GCMs despite the differences in the emitted species and peak fire 436 seasons. 437

Lastly, we utilize the four GCM-NN models, which provides outputs of future vegetation 438 distribution and LULC (GFDL-ESM4, EC-Earth3-CC, EC-Earth3-Veg, EC-Earth3-Veg-LR), to 439 project fire emissions considering the effects of land use and land cover change in 2050-2065. 440 These GCM-NN results (land change) are compared with those predicted using the same LULC 441 derived from MODIS for both the future and the present-day (land fix) (Table 3). Figure 8a 442 shows the median ratio of future to present-day fire emissions for CONUS and four regions 443 based on the four GCMs. The projected median ratios increase for CONUS and the western US 444 (WRF, SCA, and SWUS) by 58-83% when LULC changes are considered, which may be 445 attributed to increases in vegetation, tree, and grass fraction in the future (Figure 8b) and these 446

three land types are the predictors with larger contributions to the fire emissions (Figure 2). Note

that vegetation fraction is the sum of tree, grass, shrub, and crop fraction. The results indicate the

importance of land use and land cover in controlling future fire emissions. Meanwhile, large

uncertainties exist in the projected land use and land cover change (Prestele et al., 2016).





Fig. 8. (a) The median values of the ratio of future to present-day fire emissions from the four GCM-NN for CONUS and the four regions with and without including land use and land cover changes; (b) the percentage change of tree, grass, and vegetation fraction calculated as the differences in the average of the four models between 2050-2065 and 2000-2014 divided by the average of the four models over 2000-2014.

457

458 **4.4 Factors driving increased fire emissions in the future**

Figure 9 shows the differences in seasonal mean SHAP values between 2050-2065 and 459 2000-2014 from the eight GCM-NN for the four regions based on the projection excluding 460 effects of future LULC changes (land fix experiment). The SHAP values represent the 461 contributions of the predictors, and the positive difference in SHAP values indicates increasing 462 contribution to the future fire emissions. We show the changes in SHAP values for the summer 463 (JJA) and fall (SON), as the larger enhancement of fire emissions is projected to occur in the two 464 seasons for all the regions. For the western forest region, the increased fire emissions in JJA are 465 driven by increasing normalized temperature and LAI and decreasing normalized RH and soil 466 moisture (Figure 9a and Figure S6a). All of the GCMs show increased normalized temperature 467 and LAI over the western forest region, with a median change of 29% and 20%, while the 468 normalized RH and soil moisture only slightly decrease by 2.19% and 0.39% (Figure S6). 469 Similar to the western forest region, the growing fire emissions in JJA in Southern California 470 result from increasing normalized temperature and decreasing soil moisture, with a median 471 change of 29% and -1.8%, respectively (Figure 9b and Figure S6b). For the southwestern US, 472 decreased normalized RH (-4.6%) and soil moisture (-2.6%) as well as increased normalized 473 temperature (+24%) and VPD (+18%) contribute to the enhanced fire emissions in summer 474 475 (Figure 9c and Figure S6c). Lastly, increased normalized temperature and decreased normalized RH are the drivers of the increased fire emissions in JJA for the southeastern US (Figure 9d and 476 Figure S6d). In the mid-21st century, normalized temperature is projected to increase by 23% and 477

normalized RH is projected to decrease by 13% in JJA over the southeastern US. Interestingly,

the changes in VPD contribution to fire emissions are much smaller than the changes in

normalized temperature for the western US (Figures 9a-c). This can be explained by the fact that

the increase in future VPD over western US is mainly driven by increasing temperature (Figures

482 S6 and S7), so the NN model shows larger contributions from normalized temperature (Figure 2

483 and Figure 9).



484

Fig. 9. The distributions of the changes in seasonal mean SHAP values from the eight GCMs between 2050-2065 and 2000-2014 for JJA and SON for the (a) western forest region, (b)

487 Southern California, (c) southwestern US, and (d) southeastern US.

SON also shows larger fire emissions in the future, particularly over the western forest 488 region and southern California, which are mainly contributed by increased normalized 489 temperature and decreased soil moisture (Figures 9a and 9b). Normalized temperature is 490 projected to increase by 336% and 164% in SON for the western forest region and southern 491 California, respectively (Figure S8). The soil moisture in the two regions is projected to decrease 492 493 by 3.2%. In summary, reduced soil moisture is the key common factor driving the increased fire emissions in both summer and fall across the western US (the western forest region, Southern 494 California, and the southwestern US). Normalized temperature here represents the temporal 495 variation of temperature. When we look at the actual changes in temperature across the 8 GCMs, 496

the median temperature is projected to increase by 2.25~3.23 K in JJA and 2.60~2.80 K in SON
for the four regions.

499 **5 Discussions and conclusions**

This study constructs a NN model explained by the SHAP to predict fire emissions and to 500 understand factors driving the changes in fire emissions over CONUS in the mid-21st century 501 under a high greenhouse gas emissions scenario (SSP5-8.5). The NN model shows promising 502 results, with an RMSE of 0.085 g/m² and an IoA of 0.53 at 1° grid level, and it reproduces the 503 interannual variability of the fire emissions over CONUS and selected regions (IoA=0.31-0.95). 504 Although the NN model performance slightly degrades compared to the XGBoost model 505 developed in Wang et al. (2022) due to coarser resolution and differences in the ML approaches, 506 the NN model still outperforms other process-based models in simulating the spatial distribution 507 and temporal variability of fire emissions. This study uses NN instead of XGBoost because the 508 XGBoost cannot discriminate between the present-day and future meteorology simulated by the 509 GCMs as both of their probability distributions fall within the probability distributions of the 510 511 observations used for training, so the samples from the two periods would fall into the same node. On the other hand, NN can reflect the small changes in the inputs through the combination 512 of weights and activation function. 513

Driven by the GCM outputs (i.e., GCM-NN) for the present-day, the GCM-NN 514 predictions generally reproduce the observed seasonality of fire emissions, but the predicted 515 spatial distributions of GCM-NN deviate from the GFED fire emissions. When we consider the 516 contributions from NN biases and predictor biases, for ACCESS, CESM2, and CMCC, the 517 518 overestimation of fire emissions over the Central Great Plains can be attributed to the NN biases. The overestimation in western forest region is due to biases in precipitation for CanESM and 519 biases in normalized temperature and SPEI for the EC-Earth models, consistent with the warm 520 521 and dry biases identified in the GCMs in CMIP6 (J. Dong et al., 2022; Srivastava et al., 2020). The positive biases over the southeastern US in GFDL-ESM can be mainly attributed to the NN 522 biases. 523

The GCM-NN models project that fire emissions will increase by 38% (median value) in 524 the mid-century under the SSP5-8.5 scenario, with a larger increase over the southwestern US 525 (85%; median value) and western forest region (40%; median value). The largest enhancements 526 are projected to occur in JJA for most selected regions, including southern California, 527 528 southwestern US, and the southeastern US, with increases of 33-186%. As for the western forest region, fire emissions are projected to increase most in SON (129%). The projected increasing 529 trends are consistent with the projections from prior studies (Table 1). For instance, Xie et al. 530 (2022) showed that the fire CO₂ emissions are projected to increase by 130 to 260% over western 531 North America under the SSP5-8.5 scenario using three CMIP6 earth system models. Although 532 the emission species are different, our GCM-NN projects smaller increases in fire PM2.5 533 534 emissions in summer over the western US (33-186%), as the GCM-NN does not consider the changes in LULC as well as the feedback between fire and climate. More recently, Liu et al. 535 (2021) showed that the fire PM_{2.5} emissions over the western US are projected to increase by 536 537 50% in the mid-21st century under the RCP8.5 scenario, which is close to the median value of the enhancement over the western US by our GCM-NN (~50%). Their projection considering future 538 changes in fuel is roughly equal to the GCM-NN projection while our GCM-NN projection does 539

540 not include future changes in fuel load. The smaller projected enhancements considering the changes in fuels may be due to fire model biases and smaller projected fuel changes (Liu et al., 541 2021). Considering the LULC change simulated by the four GCM models, the projected fire 542 543 emissions by GCM-NN increase by 58-83% over the western US, compared to the future projection without LULC change. The projection considering future LULC change is close to the 544 projection estimated by Xie et al. (2022), while the projected LULC change contains certain 545 uncertainties (Prestele et al., 2016). Recent studies suggest that the climate sensitivity is 546 substantially higher in CMIP6 than in CMIP5, showing more than one-quarter of models 547 projecting warming larger than 4.7°C when doubling atmospheric CO2 concentrations from pre-548 industrial levels (Ribes et al., 2021; Zelinka et al., 2020). The eight GCMs used in this study 549 except for GFDL-ESM have higher equilibrium climate sensitivity (ECS) larger than the 550 multimodel mean shown in Zelinka et al. (2020). Therefore, our projection using CMIP6 climate 551 output may be subject to the "hotter" CMIP6 projection, but we focus on the median value of the 552 multimodel projections, which would not be influenced by extreme values (i.e., the model with 553

554 very high sensitivity).

Using SHAP to explain the NN model, we identify the crucial factors driving the future 555 enhanced fire emissions over CONUS. Increased normalized temperature and decreased soil 556 557 moisture and normalized RH are the common and key drivers leading to the enhanced fire emissions in JJA and SON across the western US (Figure 9). Soil moisture and RH is projected 558 to decrease by 0.36-2.6% and 2.2-4.6%, respectively; the normalized temperature is projected to 559 increase by 24-366% in the future, which is consistent with prior studies showing decreases in 560 moisture and increases temperature will enhance future fire risk given abundant fuels (Jain et al., 561 2022). As shown in Figure 8, increases in fire emissions considering future LULC change may 562 be attributed to increases in the grass, tree, and vegetation fraction ($\sim 15\%$ across the western 563 US). The fire emission projection generally agrees with the projections in prior studies and 564 reveals the importance of the drying trend and LULC change in controlling future fire emissions 565 (Brey et al., 2018; Yu et al., 2022). 566

This study utilizes XAI and CMIP6 GCM outputs to project future fire PM_{2.5} emissions. 567 This approach assumes the relationships between predictors and fire emissions remain the same 568 in the future as at present, while the relationships may change. Uncertainties may exist in NN's 569 extrapolation ability and behavior, which would be a concern for making future projections. For 570 instance, prior studies have shown that the NN's structure and activation function and whether 571 the training data is diverse have substantial impacts on extrapolation behavior (Xu et al., 2021). 572 Hernanz et al. (2022) evaluated different MLs' extrapolation behavior in predicting surface 573 temperature. Their results show that the extrapolation capability is better when training and 574 testing data overlap at a certain level. In this study, the data space of the observation/reanalysis 575 data and historical and future simulation of GCMs overlaps, suggesting minor extrapolated errors 576 in our NN model projection (Figure S9). One notable limitation of the NN model is its lack of 577 feedback between fire emissions and climate, which might underestimate the projection of future 578 fire emissions (Zou et al., 2020). Lastly, large uncertainties also exist in the projection of future 579 LULC; future work is needed to improve the future projection of LULC to better project fire 580 emissions, as climate change projection can also be influenced by LULC change (Bukovsky et 581 al., 2021). Overall, based on the relationships at present-day learned by NN, rising temperature, 582 583 decreasing moisture, and changes in LULC (i.e., increases in vegetation) are key factors

- contributing to increasing future fire emissions over the western US, where the intensity and
- frequency of large fires have significantly risen in recent decades.
- 586

587 Acknowledgments

- 588 This research was supported by the U.S. Department of Energy Office of Science Biological and
- 589 Environmental Research through the Regional and Global Model Analysis and Multisector
- 590 Dynamics program areas. PNNL is operated by Battelle Memorial Institute for the U.S.
- 591 Department of Energy under contract DE-AC06-76RLO-1830. This research was performed at
- 592 PNNL and also partially funded under Assistance Agreement No. RD835871 by the U.S.
- 593 Environmental Protection Agency to Yale University through the SEARCH (Solutions for
- 594 Energy, AiR, Climate, and Health) Center. It has not been formally reviewed by EPA. The views
- 595 expressed in this document are solely those of the SEARCH Center and do not necessarily reflect
- those of the Agency. EPA does not endorse any products or commercial services mentioned in
- 597 this publication.
- 598

599 **Open Research**

- 600 The datasets and model output are publicly accessible online at
- 601 <u>https://zenodo.org/record/7094709#.Yyj5y-zMLlw</u>.

602 **References**

- Abatzoglou, J. T. (2013). Development of gridded surface meteorological data for ecological
 applications and modelling. *International Journal of Climatology*, 33(1), 121–131.
 https://doi.org/10.1002/joc.3413
- Abatzoglou, J. T., & Kolden, C. A. (2013). Relationships between climate and macroscale area
 burned in the western United States. *International Journal of Wildland Fire*, 22(7), 1003–
 1020. https://doi.org/10.1071/WF13019
- Abatzoglou, J. T., & Williams, A. P. (2016). Impact of anthropogenic climate change on wildfire
 across western US forests. *Proceedings of the National Academy of Sciences*, *113*(42),
 11770–11775. https://doi.org/10.1073/pnas.1607171113
- Adadi, A., & Berrada, M. (2018). Peeking Inside the Black-Box: A Survey on Explainable
 Artificial Intelligence (XAI). *IEEE Access*, 6, 52138–52160.
 https://doi.org/10.1109/ACCESS.2018.2870052
- Akagi, S. K., Yokelson, R. J., Wiedinmyer, C., Alvarado, M. J., Reid, J. S., Karl, T., et al.
 (2011). Emission factors for open and domestic biomass burning for use in atmospheric models. *Atmospheric Chemistry and Physics*, 11(9), 4039–4072.
- 618 https://doi.org/10.5194/acp-11-4039-2011
- 619 Arrieta, A. B., Díaz-Rodríguez, N., Del Ser, J., Bennetot, A., Tabik, S., Barbado, A., et al.
- 620 (2020). Explainable Artificial Intelligence (XAI): Concepts, taxonomies, opportunities
- and challenges toward responsible AI. *Information Fusion*, *58*, 82–115.
- 622 https://doi.org/10.1016/j.inffus.2019.12.012

623 Birch, D. S., Morgan, P., Kolden, C. A., Abatzoglou, J. T., Dillon, G. K., Hudak, A. T., & Smith, 624 A. M. S. (2015). Vegetation, topography and daily weather influenced burn severity in central Idaho and western Montana forests. *Ecosphere*, 6(1), art17. 625 https://doi.org/10.1890/ES14-00213.1 626 Brey, S. J., Barnes, E. A., Pierce, J. R., Wiedinmyer, C., & Fischer, E. V. (2018). Environmental 627 Conditions, Ignition Type, and Air Quality Impacts of Wildfires in the Southeastern and 628 Western United States. Earth's Future, 6(10), 1442–1456. 629 https://doi.org/10.1029/2018EF000972 630 Bukovsky, M. S., Gao, J., Mearns, L. O., & O'Neill, B. C. (2021). SSP-Based Land-Use Change 631 Scenarios: A Critical Uncertainty in Future Regional Climate Change Projections. 632 Earth's Future, 9(3), e2020EF001782. https://doi.org/10.1029/2020EF001782 633 Burke, M., Driscoll, A., Heft-Neal, S., Xue, J., Burney, J., & Wara, M. (2021). The changing risk 634 and burden of wildfire in the United States. Proceedings of the National Academy of 635 Sciences, 118(2). https://doi.org/10.1073/pnas.2011048118 636 CIESIN-Columbia University. (2017). Gridded Population of the World, Version 4 (GPWv4): 637 Population Density, Revision 11 [Data set]. Palisades, NY: Socioeconomic Data and 638 Applications Center (SEDAC). https://doi.org/10.7927/H49C6VHW 639 Coffield, S. R., Graff, C. A., Chen, Y., Smyth, P., Foufoula-Georgiou, E., & Randerson, J. T. 640 (2019). Machine learning to predict final fire size at the time of ignition. International 641 642 Journal of Wildland Fire, 28, 861–873. Cortez, P., & Morais, A. (2007). A Data Mining Approach to Predict Forest Fires using 643 Meteorological Data (pp. 512–523). Portugal. 644 645 Crimmins, M. A. (2006). Synoptic climatology of extreme fire-weather conditions across the southwest United States. International Journal of Climatology, 26(8), 1001–1016. 646 https://doi.org/10.1002/joc.1300 647 Dillon, G. K., Holden, Z. A., Morgan, P., Crimmins, M. A., Heyerdahl, E. K., & Luce, C. H. 648 (2011). Both topography and climate affected forest and woodland burn severity in two 649 regions of the western US, 1984 to 2006. Ecosphere, 2(12), art130. 650 https://doi.org/10.1890/ES11-00271.1 651 Dong, J., Lei, F., & Crow, W. T. (2022). Land transpiration-evaporation partitioning errors 652 responsible for modeled summertime warm bias in the central United States. Nature 653 654 Communications, 13(1), 336. https://doi.org/10.1038/s41467-021-27938-6 Dong, L., Leung, L. R., Qian, Y., Zou, Y., Song, F., & Chen, X. (2021). Meteorological 655 Environments Associated With California Wildfires and Their Potential Roles in Wildfire 656 Changes During 1984–2017. Journal of Geophysical Research: Atmospheres, 126(5), 657 e2020JD033180. https://doi.org/10.1029/2020JD033180 658 Döscher, R., Acosta, M., Alessandri, A., Anthoni, P., Arsouze, T., Bergman, T., et al. (2022). 659 The EC-Earth3 Earth system model for the Coupled Model Intercomparison Project 6. 660 661 Geoscientific Model Development, 15(7), 2973–3020. https://doi.org/10.5194/gmd-15-2973-2022 662 Dunne, J. P., Horowitz, L. W., Adcroft, A. J., Ginoux, P., Held, I. M., John, J. G., et al. (2020). 663 The GFDL Earth System Model Version 4.1 (GFDL-ESM 4.1): Overall Coupled Model 664 Description and Simulation Characteristics. Journal of Advances in Modeling Earth 665 Systems, 12(11), e2019MS002015. https://doi.org/10.1029/2019MS002015 666 Eyring, V., Bony, S., Meehl, G. A., Senior, C. A., Stevens, B., Stouffer, R. J., & Taylor, K. E. 667 (2016). Overview of the Coupled Model Intercomparison Project Phase 6 (CMIP6) 668

669	experimental design and organization. Geoscientific Model Development, 9(5), 1937-
670	1958. https://doi.org/10.5194/gmd-9-1937-2016
671	Ford, B., Martin, M. V., Zelasky, S. E., Fischer, E. V., Anenberg, S. C., Heald, C. L., & Pierce, J.
672	R. (2018). Future Fire Impacts on Smoke Concentrations, Visibility, and Health in the
673	Contiguous United States. GeoHealth, 2(8), 229–247.
674	https://doi.org/10.1029/2018GH000144
675	Friedl, M., & Sulla-Menashe, D. (2015). MCD12C1 MODIS/Terra+Aqua Land Cover Type
676	Yearly L3 Global 0.05Deg CMG V006 [Data set]. NASA EOSDIS Land Processes
677	DAAC. https://doi.org/10.5067/MODIS/MCD12C1.006
678	Hernanz, A., García-Valero, J. A., Domínguez, M., & Rodríguez-Camino, E. (2022). A critical
679	view on the suitability of machine learning techniques to downscale climate change
680	projections: Illustration for temperature with a toy experiment. Atmospheric Science
681	Letters, 23(6), e1087. https://doi.org/10.1002/asl.1087
682	Hurtt, G. C., Chini, L., Sahajpal, R., Frolking, S., Bodirsky, B. L., Calvin, K., et al. (2020).
683	Harmonization of global land use change and management for the period 850–2100
684	(LUH2) for CMIP6. Geoscientific Model Development, 13(11), 5425–5464.
685	https://doi.org/10.5194/gmd-13-5425-2020
686	IPCC. (2022). Impacts, Adaptation, and Vulnerability. Contribution of Working Group II to the
687	Sixth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge
688	University Press.
689	Jaffe, D. A., O'Neill, S. M., Larkin, N. K., Holder, A. L., Peterson, D. L., Halofsky, J. E., &
690	Rappold, A. G. (2020). Wildfire and prescribed burning impacts on air quality in the
691	United States. Journal of the Air & Waste Management Association, 70(6), 583–615.
692	https://doi.org/10.1080/10962247.2020.1749731
693	Jain, P., Castellanos-Acuna, D., Coogan, S. C. P., Abatzoglou, J. T., & Flannigan, M. D. (2022).
694	Observed increases in extreme fire weather driven by atmospheric humidity and
695	temperature. Nature Climate Change, 12(1), 63-70. https://doi.org/10.1038/s41558-021-
696	01224-1
697	Jones, B., & O'Neill, B. C. (2020). Global One-Eighth Degree Population Base Year and
698	Projection Grids Based on the Shared Socioeconomic Pathways. Palisades, New York:
699	NASA Socioeconomic Data and Applications Center (SEDAC). Retrieved from
700	https://doi.org/10.7927/m30p-j498
701	Kane, V. R., Cansler, C. A., Povak, N. A., Kane, J. T., McGaughey, R. J., Lutz, J. A., et al.
702	(2015). Mixed severity fire effects within the Rim fire: Relative importance of local
703	climate, fire weather, topography, and forest structure. Forest Ecology and Management,
704	358, 62–79. https://doi.org/10.1016/j.foreco.2015.09.001
705	Kaulfus, A. S., Nair, U., Jaffe, D., Christopher, S. A., & Goodrick, S. (2017). Biomass Burning
706	Smoke Climatology of the United States: Implications for Particulate Matter Air Quality.
707	Environmental Science & Technology, 51(20), 11731–11741.
708	https://doi.org/10.1021/acs.est.7b03292
709	Lawrence, D. M., Fisher, R. A., Koven, C. D., Oleson, K. W., Swenson, S. C., Bonan, G., et al.
710	(2019). The Community Land Model Version 5: Description of New Features,
711	Benchmarking, and Impact of Forcing Uncertainty. Journal of Advances in Modeling
712	Earth Systems, 11(12), 4245-4287. https://doi.org/10.1029/2018MS001583
713	Liu, Yawen, Zhang, K., Qian, Y., Wang, Y., Zou, Y., Song, Y., et al. (2018). Investigation of
714	short-term effective radiative forcing of fire aerosols over North America using nudged

715	hindcast ensembles. Atmospheric Chemistry and Physics, 18(1), 31–47.
716	https://doi.org/10.5194/acp-18-31-2018
717	Liu, Yongqiang, Liu, Y., Fu, J., Yang, CE., Dong, X., Tian, H., et al. (2021). Projection of
/18	future wildlife emissions in western USA under chimate change: contributions from
719 720	<i>Fire</i> , 31(1), 1–13. https://doi.org/10.1071/WF20190
721	Lundberg, S., & Lee, S. (2017). A Unified Approach to Interpreting Model Predictions. Long
722	Beach, CA, USA.
723	Marlon, J. R., Bartlein, P. J., Gavin, D. G., Long, C. J., Anderson, R. S., Briles, C. E., et al.
724	(2012). Long-term perspective on wildfires in the western USA. Proceedings of the
725	National Academy of Sciences, 109(9), E535–E543.
726	https://doi.org/10.1073/pnas.1112839109
727	McKinnon, K. A., Poppick, A., & Simpson, I. R. (2021). Hot extremes have become drier in the
728	United States Southwest, Nature Climate Change, 11(7), 598–604.
729	https://doi.org/10.1038/s41558-021-01076-9
730	Mesinger, F., DiMego, G., Kalnay, E., Mitchell, K., Shafran, P. C., Ebisuzaki, W., et al. (2006).
731	North American Regional Reanalysis. Bulletin of the American Meteorological Society.
732	87(3), 343–360. https://doi.org/10.1175/BAMS-87-3-343
733	Myneni, R., Knyazikhin, Y., & Park, T. (2015). MCD15A2H MODIS/Terra+Agua Leaf Area
734	Index/FPAR 8-day L4 Global 500m SIN Grid V0006. NASA EOSDIS Land Processes
735	DAAC. https://doi.org/10.5067/MODIS/MCD15A2H.006
736	Neumann, J. E., Amend, M., Anenberg, S., Kinney, P. L., Sarofim, M., Martinich, J., et al.
737	(2021). Estimating PM2.5-related premature mortality and morbidity associated with
738	future wildfire emissions in the western US. Environmental Research Letters, 16(3),
739	035019. https://doi.org/10.1088/1748-9326/abe82b
740	NIFC. (2022). Suppression Costs. National Interagency Fire Center. Retrieved from
741	https://www.nifc.gov/fire-information/statistics/suppression-costs
742	O'Dell, K., Ford, B., Fischer, E. V., & Pierce, J. R. (2019). Contribution of Wildland-Fire Smoke
743	to US PM2.5 and Its Influence on Recent Trends. Environmental Science & Technology,
744	53(4), 1797–1804. https://doi.org/10.1021/acs.est.8b05430
745	O'Neill, B. C., Tebaldi, C., van Vuuren, D. P., Eyring, V., Friedlingstein, P., Hurtt, G., et al.
746	(2016). The Scenario Model Intercomparison Project (ScenarioMIP) for CMIP6.
747	Geoscientific Model Development, 9(9), 3461–3482. https://doi.org/10.5194/gmd-9-3461-
748	2016
749	Padarian, J., McBratney, A. B., & Minasny, B. (2020). Game theory interpretation of digital soil
750	mapping convolutional neural networks. SOIL, 6(2), 389–397.
751	https://doi.org/10.5194/soil-6-389-2020
752	Pechony, O., & Shindell, D. T. (2009). Fire parameterization on a global scale. Journal of
753	Geophysical Research: Atmospheres, 114(D16). https://doi.org/10.1029/2009JD011927
754	Prestele, R., Alexander, P., Rounsevell, M. D. A., Arneth, A., Calvin, K., Doelman, J., et al.
755	(2016). Hotspots of uncertainty in land-use and land-cover change projections: a global-
756	scale model comparison. Global Change Biology, 22(12), 3967–3983.
757	https://doi.org/10.1111/gcb.13337
758	Ribes, A., Qasmi, S., & Gillett, N. P. (2021). Making climate projections conditional on
759	historical observations. Science Advances, 7(4), eabc0671.
760	https://doi.org/10.1126/sciadv.abc0671

761	Shankar, U., Prestemon, J. P., McKenzie, D., Talgo, K., Xiu, A., Omary, M., et al. (2018).
762	Projecting wildfire emissions over the south-eastern United States to mid-century.
763	International Journal of Wildland Fire, 27(5), 313–328.
764	https://doi.org/10.1071/WF17116
765	Song, X., Wang, DY., Li, F., & Zeng, XD. (2021). Evaluating the performance of CMIP6
766	Earth system models in simulating global vegetation structure and distribution. <i>Advances</i>
767	in Climate Change Research, 12(4), 584–595.
768	https://doi.org/10.1016/i.accre.2021.06.008
769	Srivastava, A., Grotiahn, R., & Ullrich, P. A. (2020). Evaluation of historical CMIP6 model
770	simulations of extreme precipitation over contiguous US regions. <i>Weather and Climate</i>
771	<i>Extremes</i> , 29, 100268, https://doi.org/10.1016/i.wace.2020.100268
772	Stirnberg, R., Cermak, J., Kotthaus, S., Haeffelin, M., Andersen, H., Fuchs, J., et al. (2021).
773	Meteorology-driven variability of air pollution (PM_1) revealed with explainable machine
774	learning. Atmospheric Chemistry and Physics, 21(5), 3919–3948.
775	https://doi.org/10.5194/acp-21-3919-2021
776	Swart, N. C., Cole, J. N. S., Kharin, V. V., Lazare, M., Scinocca, J. F., Gillett, N. P., et al.
777	(2019). The Canadian Earth System Model version 5 (CanESM5.0.3). Geoscientific
778	<i>Model Development</i> , 12(11), 4823–4873. https://doi.org/10.5194/gmd-12-4823-2019
779	Thonicke, K., Spessa, A., Prentice, I. C., Harrison, S. P., Dong, L., & Carmona-Moreno, C.
780	(2010). The influence of vegetation, fire spread and fire behaviour on biomass burning
781	and trace gas emissions: results from a process-based model. <i>Biogeosciences</i> , 7(6), 1991–
782	2011. https://doi.org/10.5194/bg-7-1991-2010
783	Trouet, V., Taylor, A. H., Carleton, A. M., & Skinner, C. N. (2009). Interannual variations in fire
784	weather, fire extent, and synoptic-scale circulation patterns in northern California and
785	Oregon. Theoretical and Applied Climatology 95: 349-360, 95, 349–360.
786	https://doi.org/10.1007/s00704-008-0012-x
787	Val Martin, M., Heald, C. L., Lamarque, JF., Tilmes, S., Emmons, L. K., & Schichtel, B. A.
788	(2015). How emissions, climate, and land use change will impact mid-century air quality
789	over the United States: a focus on effects at national parks. Atmospheric Chemistry and
790	Physics, 15(5), 2805–2823. https://doi.org/10.5194/acp-15-2805-2015
791	Vicente-Serrano, S. M., Beguería, S., & López-Moreno, J. I. (2010). A Multiscalar Drought
792	Index Sensitive to Global Warming: The Standardized Precipitation Evapotranspiration
793	Index. Journal of Climate, 23(7), 1696–1718. https://doi.org/10.1175/2009JCLI2909.1
794	Wang, S. SC., & Wang, Y. (2020). Quantifying the effects of environmental factors on wildfire
795	burned area in the south central US using integrated machine learning techniques.
796	Atmospheric Chemistry and Physics, 20(18), 11065–11087. https://doi.org/10.5194/acp-
797	20-11065-2020
798	Wang, S. SC., Qian, Y., Leung, L. R., & Zhang, Y. (2021). Identifying key drivers of wildfires
799	in the contiguous US using machine learning and game theory interpretation. Earth's
800	Future, 9(6), e2020EF001910. https://doi.org/10.1029/2020EF001910
801	Wang, S. SC., Qian, Y., Leung, L. R., & Zhang, Y. (2022). Interpreting machine learning
802	prediction of fire emissions and comparison with FireMIP process-based models.
803	Atmospheric Chemistry and Physics, 22(5), 3445-3468. https://doi.org/10.5194/acp-22-
804	3445-2022

805	van der Werf, G. R., Randerson, J. T., Giglio, L., van Leeuwen, T. T., Chen, Y., Rogers, B. M.,
806	et al. (2017). Global fire emissions estimates during 1997–2016. Earth System Science
807	Data, 9(2), 697–720. https://doi.org/10.5194/essd-9-697-2017
808	Westerling, A. L., Hidalgo, H. G., Cayan, D. R., & Swetnam, T. W. (2006). Warming and Earlier
809	Spring Increase Western U.S. Forest Wildfire Activity. Science, 313(5789), 940–943.
810	https://doi.org/10.1126/science.1128834
811	Xia, Y., Mitchell, K., Ek, M., Sheffield, J., Cosgrove, B., Wood, E., et al. (2012). Continental-
812	scale water and energy flux analysis and validation for the North American Land Data
813	Assimilation System project phase 2 (NLDAS-2): 1. Intercomparison and application of
814	model products. Journal of Geophysical Research: Atmospheres, 117(D3).
815	https://doi.org/10.1029/2011JD016048
816	Xie, Y., Lin, M., Decharme, B., Delire, C., Horowitz, L. W., Lawrence, D. M., et al. (2022).
817	Tripling of western US particulate pollution from wildfires in a warming climate.
818	Proceedings of the National Academy of Sciences, 119(14), e2111372119.
819	https://doi.org/10.1073/pnas.2111372119
820	Xu, K., Zhang, M., Li, J., Du, S. S., Kawarabayashi, K., & Jegelka, S. (2021, March 2). How
821	Neural Networks Extrapolate: From Feedforward to Graph Neural Networks. arXiv.
822	https://doi.org/10.48550/arXiv.2009.11848
823	Yu, Y., Mao, J., Wullschleger, S. D., Chen, A., Shi, X., Wang, Y., et al. (2022). Machine
824	learning-based observation-constrained projections reveal elevated global socioeconomic
825	risks from wildfire. Nature Communications, 13(1), 1250.
826	https://doi.org/10.1038/s41467-022-28853-0
827	Yue, X., Mickley, L. J., Logan, J. A., & Kaplan, J. O. (2013). Ensemble projections of wildfire
828	activity and carbonaceous aerosol concentrations over the western United States in the
829	mid-21st century. Atmospheric Environment, 77, 767–780.
830	https://doi.org/10.1016/j.atmosenv.2013.06.003
831	Zelinka, M. D., Myers, T. A., McCoy, D. T., Po-Chedley, S., Caldwell, P. M., Ceppi, P., et al.
832	(2020). Causes of Higher Climate Sensitivity in CMIP6 Models. Geophysical Research
833	Letters, 47(1), e2019GL085782. https://doi.org/10.1029/2019GL085782
834	Zhong, S., Yu, L., Heilman, W. E., Bian, X., & Fromm, H. (2020). Synoptic weather patterns for
835	large wildfires in the northwestern United States—a climatological analysis using three
836	classification methods. Theoretical and Applied Climatology, 76.
837	https://doi.org/10.1007/s00704-020-03235-y
838	Zhu, Q., Li, F., Riley, W. J., Xu, L., Zhao, L., Yuan, K., et al. (2022). Building a machine
839	learning surrogate model for wildfire activities within a global Earth system model.
840	Geoscientific Model Development, 15(5), 1899–1911. https://doi.org/10.5194/gmd-15-
841	1899-2022
842	Ziehn, T., Chamberlain, M. A., Law, R. M., Lenton, A., Bodman, R. W., Dix, M., et al. (2020).
843	The Australian Earth System Model: ACCESS-ESM1.5. Journal of Southern Hemisphere
844	Earth Systems Science, 70(1), 193–214. https://doi.org/10.1071/ES19035
845	Zou, Y., Wang, Y., Qian, Y., Tian, H., Yang, J., & Alvarado, E. (2020). Using CESM-RESFire
846	to understand climate-fire-ecosystem interactions and the implications for decadal
847	climate variability. Atmospheric Chemistry and Physics, 20(2), 995–1020.
848	https://doi.org/10.5194/acp-20-995-2020
849	
850	

manuscript submitted to Journal of Geophysical Research: Atmospheres