

How should diverse stakeholder interests shape evaluations of complex water resources systems robustness when confronting deeply uncertain changes?

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Abstract

Robustness analysis can support long-term planning, design and operation of large-scale water infrastructure projects confronting deeply uncertain futures. Diverse actors, contextual specificities, sectoral interests, and risk attitudes make it difficult to identify an acceptable and appropriate robustness metric to rank decision alternatives under deep uncertainty. Here, we contribute an exploratory framework to demonstrate how methodological choices affect robustness evaluation. The framework is applied to a multi-actor, multi-sector Inchampalli-Nagarjuna Sagar (INS) water transfer megaproject in Southern India. We evaluate a suite of dynamic adaptive water transfer strategies discovered using evolutionary multi-objective direct policy search (EMODPS), a status quo strategy of no water transfer, and a strategy proposed by regional authorities. We evaluate robustness across wide-ranging scenarios that capture key uncertainties in potential future changes in reservoir inflows and water demands in the basins. Results show that the priorities of different actors, sectoral perspectives, and risk attitude significantly affect robustness rankings of strategies. We found that compromise strategies obtained from EMODPS are better able to balance the diverse performance requirements across various actors and sectors when compared to the no-transfer and proposed transfer. We reveal a key robustness tradeoff between the donor basin's ecological requirements and the recipient basin's socio-economic requirements. While robustness analysis is central to water infrastructure planning, we show why exploratory robustness analyses that engage with conflicting stakeholder objectives is vital for long-term sustainability. Furthermore, the selection of compromise solutions should be guided by an explicit understanding of how assumed risk attitudes shape stakeholders' understanding of consequential vulnerabilities.

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10 **Key Points:**

- 11 • Multi actor and sector robustness trade-offs are often not explored due to narrowly
12 defined robustness metrics
- 13 • Robustness rankings for decision alternatives across robustness metrics shed light
14 on trade-offs across actors, sectors, and risk attitudes
- 15 • Clarify donor-recipient conflicts between ecological needs and socio-economic de-
16 velopment for a proposed water transfer mega-project

Abstract

Robustness analysis can support long-term planning, design and operation of large-scale water infrastructure projects confronting deeply uncertain futures. Diverse actors, contextual specificities, sectoral interests, and risk attitudes make it difficult to identify an acceptable and appropriate robustness metric to rank decision alternatives under deep uncertainty. Here, we contribute an exploratory framework to demonstrate how methodological choices affect robustness evaluation. The framework is applied to a multi-actor, multi-sector Inchampalli-Nagarjuna Sagar (INS) water transfer megaproject in Southern India. We evaluate a suite of dynamic adaptive water transfer strategies discovered using evolutionary multi-objective direct policy search (EMODPS), a status quo strategy of no water transfer, and a strategy proposed by regional authorities. We evaluate robustness across wide-ranging scenarios that capture key uncertainties in potential future changes in reservoir inflows and water demands in the basins. Results show that the priorities of different actors, sectoral perspectives, and risk attitude significantly affect robustness rankings of strategies. We found that compromise strategies obtained from EMODPS are better able to balance the diverse performance requirements across various actors and sectors when compared to the *no-transfer* and *proposed* transfer. We reveal a key robustness tradeoff between the donor basin's ecological requirements and the recipient basin's socio-economic requirements. While robustness analysis is central to water infrastructure planning, we show why exploratory robustness analyses that engage with conflicting stakeholder objectives is vital for long-term sustainability. Furthermore, the selection of compromise solutions should be guided by an explicit understanding of how assumed risk attitudes shape stakeholders' understanding of consequential vulnerabilities.

1 Introduction

Water resources in many parts of the world face growing hydroclimatic and socio-economic pressures (Kummu et al., 2010; Mekonnen & Hoekstra, 2016; Bijl et al., 2018). Globally, water scarcity is projected to increase due to climate change impacts on mean temperature and precipitation variability, as well as increasingly extreme floods and droughts (Greve et al., 2018; Masson-Delmotte et al., 2021). The economic consequences of water scarcity are highly uncertain and sensitive to regions' capacities to adapt to these deeply uncertain hydro-climatic changes (Dolan et al., 2021). Large scale water infrastructure projects have a critical role in addressing these challenges (Grigg, 2019; Bhaduri et al., 2008; Gohari et al., 2013). Among them, inter-basin water transfer (IBWTs) megaprojects with investments of approximately \$2.7 trillion form a major global focus and pose severely challenging decision contexts (Shumilova et al., 2018).

IBWTs must balance irrigation needs, domestic water supply, hydro-electricity generation, and other uses across multiple participating river basins, requiring their design evaluation to consider the diverse interests of a broad array of sectors. Some IBWTs have been criticized for their ecological consequences and over-exploitation of donor basin's water resources, indicating that traditional evaluations are perhaps myopic about the long-term impacts on the impacted stakeholders (Wu et al., 2020; Gohari et al., 2013; Zhuang, 2016). These multi-decadal megaprojects require an understanding of the dynamic co-evolution of the coupled human-natural systems in which they are placed, especially in key drivers of climate and demands. Projections of these drivers are often deeply uncertain, challenging the traditional use of aggregated cost-benefit analysis to discover transfer policies. Deep uncertainty refers to conditions where parties to a decision lack a consensus on the likelihoods and/or distributional forms of key system inputs (Knight, 1921; Lempert, 2002; Lempert et al., 2006; Marchau et al., 2019). At the local scale, future runoff changes are deeply uncertain due to uncertainties associated with projections of potential future temperature and precipitation changes (Schewe et al., 2014; Bhave et al., 2018; Douville et al., 2021). Concurrent changes in socio-economic conditions are

69 also deeply uncertain, as they are a consequence of a multitude of factors pertaining to
70 the coupled human-natural system, changes in water demand priorities, and changing
71 policy landscapes (Quinn et al., 2018; Moallemi, Kwakkel, et al., 2020). Deep uncertainty
72 compounds existing challenges to traditional design approaches for IBWTs. For exam-
73 ple, a recent ex post evaluation of traditional design approaches for IBWTs have shown
74 that they often systematically underestimate water scarcity in the donor basin and over-
75 estimate the demands within the recipient basin (Huang et al., 2021).

76 Exploratory modelling-based frameworks such as Robust Decision Making, Many-
77 Objective Robust Decision Making (MORDM), Information Gap theory and Decision
78 Scaling seek to discover *robust* alternatives that perform well across a range of deeply
79 uncertain futures (Lamontagne et al., 2018; Moallemi, Zare, et al., 2020; Gold et al., 2019;
80 Moallemi et al., 2021; Kwakkel & Haasnoot, 2019; Hadjimichael et al., 2020; Ben-Haim,
81 2006; Brown et al., 2012). Robustness evaluation of IBWTs requires the analyst to de-
82 cide how to represent the multiple stakeholders involved. Although challenging, robust-
83 ness definition(s) should be identified through co-production of knowledge that includes
84 all relevant stakeholders (Moallemi, Zare, et al., 2020; Wyborn et al., 2019; Eriksen et
85 al., 2021; Bhavé et al., 2022). This would be best achieved by stakeholder workshops,
86 an iterative process that results in co-production of knowledge (Voinov et al., 2018). This
87 remains highly challenging for large-scale infrastructure projects as by their very nature,
88 they involve multiple actors spread across spatio-temporal and socio-economic gradients.
89 There may also be socio-political limitations in engaging a diverse group of stakehold-
90 ers due to differences in ideologies and varying degrees of understanding of the decision
91 process (Eriksen et al., 2021).

92 The emerging field of Decision Making Under Deep Uncertainty (DMDU) provides
93 a starting point to frame robustness definitions for the design and evaluation of IBWTs
94 (Marchau et al., 2019). Recent literature highlights a rapid proliferation of robustness
95 metrics and their impact on the preferential rank ordering of proposed alternative de-
96 signs and/or operational strategies (Borgomeo et al., 2018; Herman et al., 2015; Kwakkel,
97 Eker, & Pruyt, 2016; McPhail et al., 2018; Bartholomew & Kwakkel, 2020). In general,
98 robustness quantification requires the specification of methods for generating deeply un-
99 certain futures and aggregating evaluations of strategy performance across these futures
100 (Herman et al., 2015; McPhail et al., 2021). Generating deeply uncertain futures requires
101 an understanding and careful exploration of important system drivers as well as their
102 feasible ranges and plausible statistical properties (Quinn et al., 2018, 2020; McPhail
103 et al., 2020). The aggregate rank evaluations of robustness require an explicit consid-
104 eration of risk attitudes. Aggregation of robustness performance across sampled scenar-
105 ios for the future can be based on expected value analysis (Wald, 1950); higher-order
106 moments (Kwakkel, Haasnoot, & Walker, 2016); regret (Savage, 1951) or satisficing cri-
107 teria (Simon, 1956). Building on the general framework proposed by Herman et al. (2015),
108 McPhail et al. (2020, 2018) show that several underlying methodological choices tacit
109 to measuring robustness can substantially influence robustness-based rankings of deci-
110 sion alternatives. For example, performance aggregation across scenarios embeds assump-
111 tions regarding levels of risk aversion of stakeholders. Measuring robustness using tra-
112 ditional expected value focused metrics tacitly assumes risk neutrality, while minimax
113 or worst-case performance across scenarios represents high levels of risk aversion. Thus,
114 robustness criteria require a careful elicitation of requirements (or performance accept-
115 ability limits) from stakeholders (Herman et al., 2015; Kwakkel, Eker, & Pruyt, 2016).

116 In this study, we propose a framework to address the principal challenge of cap-
117 turing diverse stakeholder views in robustness assessments for large multi-actor infras-
118 tructure projects, a central concern when seeking to support co-production processes.
119 Our framework contextualizes how exploratory analysis of multiple robustness metrics
120 can better contextualize the implications of a broad range candidate robustness fram-
121 ings in capturing diverse stakeholder preferences and shaping performance evaluations.

Our proposed exploratory robustness assessment provides a mechanism for formally broadening dialogue and the inclusion of diverse and potentially under-represented stakeholders. We apply this framework to the proposed Inchampalli- Nagarjuna Sagar (INS) IBWT in Southern India, which aims to transfer water from the Godavari (donor) to the Krishna (recipient) river basin with significant implications for millions of farmers as well as the pharmaceutical and software hub of Hyderabad, India. We extensively assess potential impacts on the participating basins and their water related sectors considering deeply uncertain changes in precipitation patterns and river flows due to uncertain potential future changes in Indian Summer Monsoon and anthropogenic water demands.

2 The Decision Context of the INS IBWT Megaproject

India faces a daunting challenge of ensuring water, food and energy security in a changing climate and rapidly evolving socio-economic conditions. India's National River Linking Project (NRLP) aims to improve water and food security via expansion of irrigated area by $\approx 350,000 \text{ km}^2$ using 30 water transfer projects totaling in length of $\approx 14,900 \text{ km}$ and a network of 3000 storage structures (Joshi, 2013; Bagla, 2014). If implemented fully, the NRLP will incur massive water infrastructure investment of $> \$2$ trillion, greater than 60% of the Indian economy of $\$3.17$ trillion. Within NRLP, the INS IBWT proposes to transfer water from the Godavari (donor) to the Krishna (recipient) basin, the two largest river basins of Southern India (Figure 1). The INS IBWT by itself has been justified due to a growing disparity between demand and supply between its participating basins. With a proposed $16,000 \text{ Mm}^3$ annual water transfer over 299 km classified the INS IBWT as a megaproject (NWDA, 2021; Veena et al., 2021; Shumilova et al., 2018). The water transfer is a major socio-economic development intervention for the Nagarjuna Sagar reservoir, which is stressed due to increasing agricultural and urban (primarily Hyderabad city) water demand, as well as demands from another regional political capital, Vijayawada. The INS IBWT is also going to impact the aquatic ecosystems downstream of the donor and local tribal populations that rely on the maintenance of minimum environmental flows.

Given the high stakes, deep uncertainty, and multi-stakeholder context, the INS IBWT requires a comprehensive evaluation to avoid potential decision lock-ins (Moallemi, Zare, et al., 2020). Average Godavari annual inflows at Perur gage station ($77,017 \text{ Mm}^3$) are more than double those at Nagarjuna Sagar on the river Krishna ($29,625 \text{ Mm}^3$) (Figure 1b), while their respective command area water demands are $\approx 603 \text{ Mm}^3$ and $\approx 8,535 \text{ Mm}^3$ (Figure 1c) (Veena et al., 2021). Mean annual precipitation (temperature) is projected to increase by 20-50% ($1^\circ - 5^\circ \text{C}$) in both basins by the end of century (Mishra & Lihare, 2016), but future water availability and demand dynamics will evolve in complex ways with changes in population as well as the efficiency of the multisectoral water dependent systems that evolve to meet the concomitant increasing human demands (Singh & Kumar, 2019), leading to deep uncertainty.

In this study, we employ the systems model and cooperative adaptive strategies contributed by Veena et al. (2021). Their original analysis focused on the stationary historical uncertainties affecting Godavari and Krishna inflows, exploiting a water balance model to track reservoir related fluxes, and assessed water transfer strategies against different priorities for environmental flows, domestic water supply and irrigation (please see Veena et al. (2021) for further details). The study formulated cooperative state-aware water transfer strategies where water transfers are decided based on the storage states of both the donor and the recipient reservoirs. Both the donor and the recipient reservoirs are also committed to transfer water to other reservoirs, which impose additional demands on the INS IBWT. These transfers are termed as 'predefined transfers' (PT). In this study, we further evaluate the cooperative adaptive INS IBWT operational strategies under deeply uncertain futures and contribute an exploratory framework to guide assessments of their robustness.

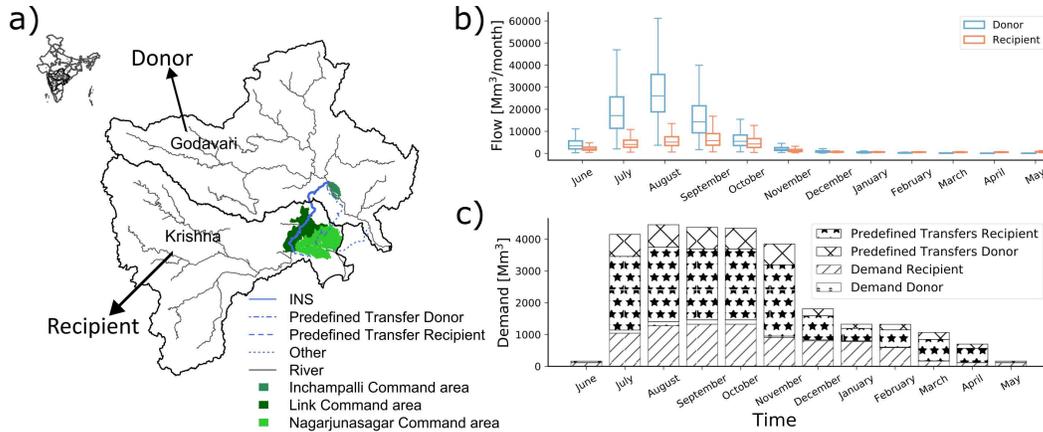


Figure 1. (a) Location of the Inchampalli - Nagarjunsagar (INS) water transfer project connecting the donor (Godavari) and recipient (Krishna) basins. The irrigated command areas for each basin is represented by shades of green. The predefined transfer from donor and recipient basins are also shown by dot-dashed and dashed line respectively. (b) Monthly stochastic inflows in donor basin (blue) and recipient basin (orange). (c) Monthly demands and predefined transfer for both basins.

174 Large scale water infrastructure projects such as the INS IBWT involve a number
 175 of actors and sectors, each with their own preferences and risk attitudes. Thus, multi-
 176 ple world views are invariably involved in its decision context. Prior literature has explored
 177 the consequences of multiple world views using multiple problem framings (Quinn
 178 et al., 2017; Kasprzyk et al., 2013; Singh et al., 2015; Lempert & Turner, 2021). Here,
 179 we propose a framework to support diverse stakeholders in exploring how they may define
 180 the robustness of an operational strategy. This framework can be used for deliberative
 181 analysis of candidate stakeholder preferences and/or as an exploratory modeling
 182 strategy for discovering the conflicts between stakeholders. The main actors involved in
 183 the INS IBWT are the donor (Godavari) basin, the recipient (Krishna) basin, and other
 184 basins dependent on water transfers from either of these (i.e., predefined transfers, PT).
 185 We also define a baseline system level actor that captures a risk neutral rational social
 186 planner acting on the expected value of performance objectives averaged over donor and
 187 recipient outcomes, following a standard assumption in simulation-optimization literature
 188 (Giuliani & Castelletti, 2016; McPhail et al., 2018; Loucks & Van Beek, 2017). Similarly,
 189 requirements of all other basins that depend upon the donor (Godavari) and recipient
 190 (Krishna) are represented by a system level PT actor.

191 The different sectors impacted by the INS IBWT are domestic, industrial, agricultural,
 192 and ecological. Domestic, industrial and agricultural sectors together constitute
 193 the water supply sector. Ecology is affected in two ways. First, minimum environmental
 194 flows (MEF) downstream of both reservoirs are dependent upon the transfer and reservoir
 195 operation rules. MEF has direct consequence on tribal communities downstream of
 196 the donor (Godavari) basin that depend upon fishing, thus it is also included here to represent
 197 the interests of the marginalized communities (Eriksen et al., 2021). Second, the
 198 volume of water transferred (transferred volume, TV) is also considered as a proxy of
 199 ecological impact. The lower the amount of water transferred, the lower the potential
 200 ecological impact of mixing waters of diverse quality and aquatic compositions. Using
 201 this rationale, we constructed two ecology related sectors: ecology-TV, and ecology-MEF.
 202 Thus, we envisage 12 actor-sector combinations that may emerge in the decision context

Table 1. Multiple actor-sector combinations explored for the INS IBWT. In each row, the X's identify which actor-sector combinations are used in robustness calculations. PT: predefined transfers for other reservoirs, TV: transfer volume, MEF: minimum environmental flows.

Combination of Actor-Sector	Actor				Sector		
	Donor	Recipient	System	PT System	Water Supply	Ecology-TV	Ecology-MEF
1	X				X		
2		X			X		
3	X						X
4		X					X
5			X		X		
6			X				X
7				X	X		
8	X	X			X		
9	X	X					X
10			X			X	
11			X		X	X	
12	X	X	X	X	X	X	X

203 of the INS IBWT (Table 1). The performance requirements for these are quantified using
 204 definitions discussed in the methods section below.

205 3 Methodology

206 Our main contribution is a formal exploratory modeling framework for better un-
 207 derstanding and transparently mapping the consequences of diverse actor and sector pref-
 208 erences as well as risk attitudes when defining robustness metrics within the MORDM
 209 framework (highlighted boxes in Figure 2). As is typical for the MORDM framework (Kasprzyk
 210 et al., 2013), our exploration of the INS IBWT begins with the identification of the de-
 211 cision context, alternative candidate problem formulations and generation of alternatives
 212 using many-objective optimization considering historical well-characterized uncertain-
 213 ties (WCU) (Section 3.1). Deeply uncertain factors that shape the performance of the
 214 alternative operational designs of the transfer are then identified and sampled in Step
 215 II (Section 3.2). We then explore tradeoffs across potential combinations of stakeholder
 216 preferences across multiple actors and sectors involved in or affected by the decision pro-
 217 cess (Section 3.3). These preference combinations together with risk attitude specifica-
 218 tion are used to re-evaluate the Pareto approximate operational transfer design strate-
 219 gies from Step I across scenarios identified in Step II (Section 3.4). In addition to eval-
 220 uating robustness under deep uncertainties (DU), we also analyze robustness under the
 221 internal hydroclimatic variability in the stochastic WCU baseline. The multivariate ro-
 222 bustness estimates thus obtained are further analyzed to identify key actor/sector trade-
 223 offs with a specific focus on the stability of alternatives ranking (Section 3.5). Finally,
 224 we identify the main drivers of system failure from uncertainties explored and clarifying
 225 how choice of robustness definitions affect inferences related to consequential trade-
 226 offs/vulnerabilities across stakeholder interests (Section 3.6).

227 Building on and extending McPhail et al. (2018), Figure 3 elaborates key steps
 228 in the exploratory evaluation of robustness considering candidate choices associated with
 229 stakeholder preferences, their risk attitudes and scenario generation methods. Robust-
 230 ness calculations require specification of deeply uncertain factors and their sampling strate-
 231 gies (ψ , purple boxes). Each deeply uncertain world will be characterized by stochastic-
 232 ity (s , green boxes). Each decision alternative, θ , is re-evaluated using the systems model
 233 to quantify values of multiple performance objectives (f , dark green boxes) representing
 234 preferences of various actors and sectors. The vectors of performance objectives can be
 235 combined in different ways to represent combinations of stakeholder preferences (m_1, m_2, \dots, m_n ,
 236 yellow boxes). Finally, alternative representations of risk-attitudes in candidate robust-

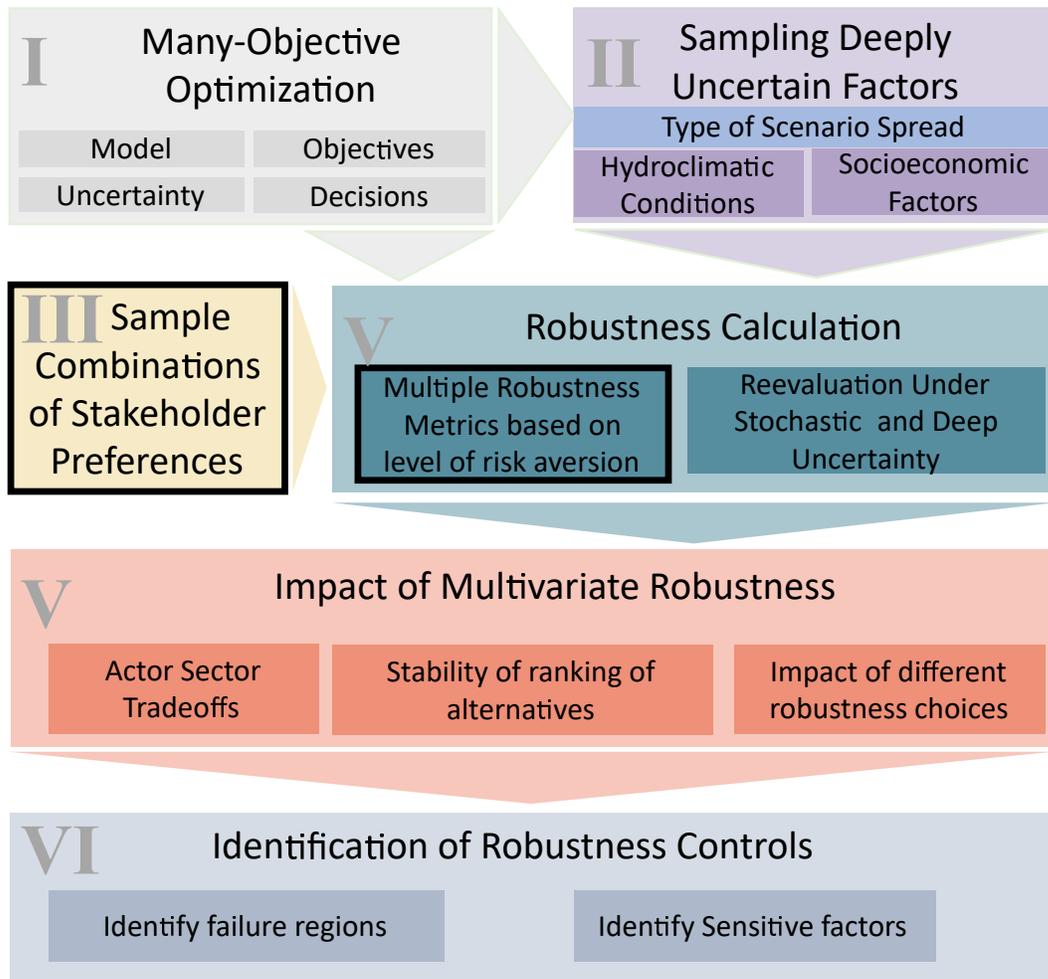


Figure 2. The six main stages in applying the MORDM framework to a decision problem. Black outlines highlight stages that include stakeholder preferences and their risk attitudes in the robustness assessment. This figure illustrates extension of MORDM framework adapted from the taxonomy of robustness frameworks presented in Herman et al. (2015).

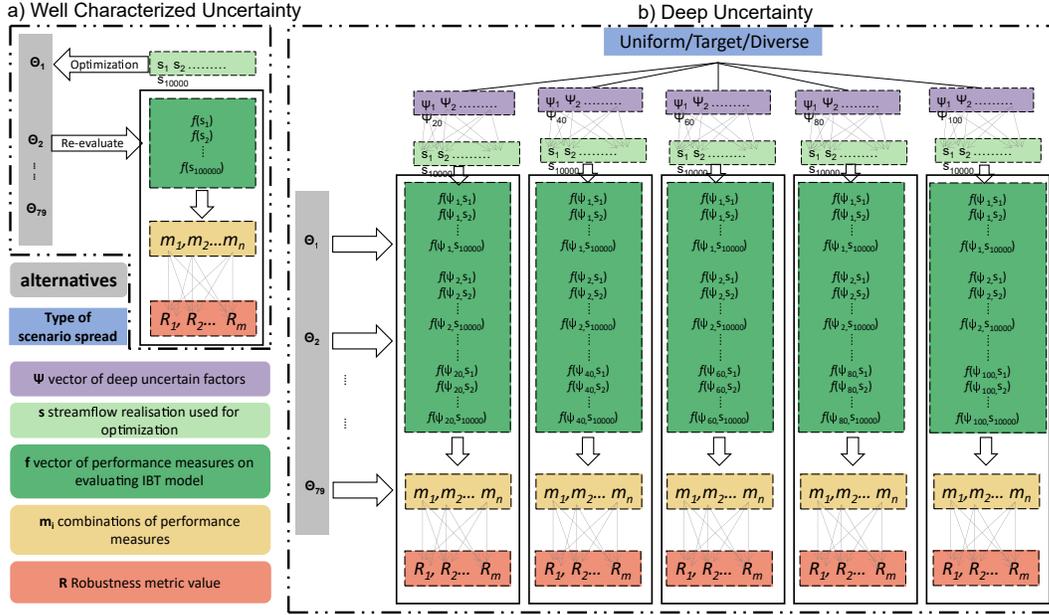


Figure 3. Evaluating the impact of metric definitions representing risk attitudes (orange), performance objectives (dark green) and their combinations (yellow) representing different stakeholders, and sampling strategies for stochastic (green) and deep (blue, purple) uncertainties on resultant robustness values. Shown are steps to quantify robustness under a) well-characterized, and b) deep uncertainties. Pareto-approximate alternatives (grey) are generated by many-objective optimization using stochastic streamflow realizations in (a). Each alternative is re-evaluated for a vector of performance objectives across a much larger stochastic set in (a). Deeply uncertain SOWs cover the multi-dimensional factor space using uniform, target, and diverse scenario spread types (blue box in b).

237 ness metrics are explored in terms of how they aggregate the performance of a decision
 238 alternative across sampled deeply uncertain states-of-the-world (SOWs, R_1, R_2, \dots, R_m ,
 239 orange box). In this way, we explore the influence of the choice of actor and sector com-
 240 binations, decision alternatives, robustness metrics, number of scenarios, and type of spread
 241 of scenarios on robustness estimates. As noted by Hadjimichael et al. (2020), it is dif-
 242 ficult in institutionally complex large-scale water resources systems for stakeholders to
 243 define and understand the implications of the alternative framings of robustness that could
 244 be considered. This study addresses this challenge by providing an exploratory frame-
 245 work that can broaden the representation of concerns while clarifying the consequences
 246 of incorporating them into alternative metrics of robustness. The following sections de-
 247 tail each of the key steps used to compute robustness.

248 3.1 Many-Objective Optimization

249 Veena et al. (2021) explored four problem formulations for the INS IBWT that
 250 quantified the tradeoffs across five system level objectives. The term ‘system-level’ refers
 251 to the fact that the performance objectives were regionally averaged across the partic-
 252 ipating basins. The objectives included reliability, resilience, and vulnerability of water
 253 demand satisfaction, reliability of maintaining minimum environmental flows, and reli-
 254 ability of avoiding high flow exceedances. The formulations compare dynamic and adap-
 255 tive rule-based operational designs against the status quo of no water transfer and a re-
 256 gional operational rule that has been suggested by the regional authorities, referred to

as the proposed rule throughout the paper. To better understand the value of information coordination across the donor and recipient basins, two types of dynamic rules were formulated by Veena et al. (2021): noncooperative that only condition the transfer decisions on the states of the donor reservoir and cooperative that condition them on the states of both the donor and recipient reservoirs. Pareto approximate strategies were generated using evolutionary multi-objective direct policy search (EMODPS) considering stochastic uncertainty (or WCU) of inflows. Stochasticity is represented using 10,000 realizations of synthetic inflows ($s_1, s_2, \dots, s_{10000}$) generated from historical inflows (1967-2012) (Veena et al., 2021; Kirsch et al., 2013; Herman et al., 2015) (Supplementary material S4). The procedure uses Cholesky decomposition to preserve the autocorrelation of inflows between the donor and recipient sites. *Cooperative* adaptive strategies outperformed all others indicating the importance of coordination between donor and recipient basins for managing water transfers and are, therefore, used in this study (79 in number) (Veena et al., 2021). Thus, we considered 81 INS IBWT operational design alternatives including the *proposed* and the status quo of *no-transfer*. These strategies are decisions (Step I in Figure 2, θ in Figure 3) to transfer water used for re-evaluating their performance under changing climates and demands to understand the long-term consequences of the INS IBWT for all the stakeholders involved. A brief overview of the model, objective functions, constraints and optimization procedure is included in Supplementary Material S1 to S3 and Table S1.

3.2 Sampling of deeply uncertain factors

Here, we explore eight deeply uncertain factors (ψ , Figure 3) to capture potential impacts on river flows due to uncertain future changes in Indian Summer Monsoon precipitation patterns and demands; six related to inflows and two related to demands (Table 2, ψ in Figure 3). Demand factors are applied as multipliers to the historical demands to represent candidate increases in the future. Six factors are used to generate different monsoon dynamics in the inflows including changes in log-space annual mean, log-space standard deviation and interannual variability of inflows. The equations to generate inflows from monsoon factor ranges are adapted from Quinn et al. (2018). Each deeply uncertain inflow defined by a combination of six monsoon related factors is paired with 10,000 realizations of inflows that represent WCU. The generated inflows are evaluated using available climate projections for the study region from the Inter-Sectoral Impact Model Intercomparison Project. These span five GCMs and four representative concentration pathways (RCPs) (Warszawski et al., 2014; Singh & Kumar, 2019) (Figure S1).

Deeply uncertain futures are sampled from within the space of plausible ranges of uncertain factors. We explore alternative sampling approaches that vary in how they focus on specific regions of the space or cover the entire space following McPhail et al. (2020). Vectors of the eight factors listed in Table 2 are generated using three sampling strategies - diverse, target, and uniform. Diverse sampling identifies locations of interest within the feasible range of uncertain factors, then generate samples around those locations (Anghileri et al., 2018; Giuliani & Castelletti, 2016; Haasnoot et al., 2012; Huskova et al., 2016; McPhail et al., 2018). This represents the general scenario generation approach followed in climate change impact studies where, first specific carbon emissions trajectories are specified, followed by using multiple climate models to generate possible climates for each trajectory. On the other hand, the targeted approach samples the scenario space such that different uncertain factors increase or decrease together monotonically (Beh et al., 2015b, 2014, 2015a). It follows that targeted sampling is useful in contexts where, changes in uncertain factors are highly correlated and would cover a smaller region of the overall feasible space. Finally, uniform sampling explores the entire multi-dimensional scenario space by sampling points within this space using Latin hypercube sampling (Herman et al., 2015; Kasprzyk et al., 2013; Kwakkel, 2017; Kwakkel et al., 2015; McPhail et al., 2018; Singh et al., 2015; Quinn et al., 2018). Further details on the generation of samples are provided in Supplementary Figure S2. For each sampling scheme, 20, 40, 60, 80,

Table 2. List of deep uncertain factors used to generate scenarios with change in monsoonal dynamics and socio-economic changes.

Deeply uncertain factors	Lower bound	Upper bound	Remarks
Log-space mean multiplier, inflows	0.95	1.05	Annual increase or decrease in mean annual inflows
Log-space std multiplier, inflows	0.5	1.5	Change in interannual variability of inflows
Log-space C_1 multiplier, inflows	0.5	1.5	Change in amplitude of annual monsoon
Log-space C_2 multiplier, inflows	0.5	1.5	Change in amplitude of semiannual monsoon
Log-space ϕ_1 delta (radians), inflows	$-2\pi/12$	$+2\pi/12$	Shift of annual monsoon
Log-space ϕ_2 delta (radians), inflows	$-2\pi/12$	$+2\pi/12$	Shift of semiannual monsoon
Demand factor, donor basin	1	1.5	Relative increase in donor demand
Demand factor, recipient basin	1	1.5	Relative increase in recipient demand

310 and 100 samples of vectors are generated. The reader is encouraged to refer to McPhail
 311 et al. (2020) for more details on the distributional sampling of scenarios for targeted spread
 312 and diverse futures.

313 3.3 Sampling combinations of stakeholder preferences

314 As detailed in section 2, we explore 12 actor-sector combinations that represent the
 315 diverse stakeholders involved in the INS IBWT. To quantify the water supply related
 316 sectoral performances, the vulnerability measure (Vul) is used as follows,

$$Vul = \frac{\sum_{t=1}^T (ad_t - d_t)}{\sum_{t=1}^T ad_t} \times 100 \quad (1)$$

317 In equation 1, d_t is the demand satisfied, ad_t is the actual demand, for each time period
 318 t , and T is the total number of time periods. The vulnerability measure can also be ex-
 319 pressed in terms of average volumetric deficits by multiplying with the total demand.
 320 Preferences of the ecology-TV sector is quantified as the mean annual transfer volumes
 321 for a water transfer alternative. The performance for the Ecology-MEF (J_{EF}) sector is
 322 quantified using a reliability measure as,

$$J_{EF} = \left(1 - \frac{\sum_{t=1}^T EF_t}{T} \right) \quad (2)$$

$$EF_t = \begin{cases} 1 & \text{if } (ef_t < mef_t) \\ 0 & \text{else} \end{cases} \quad (3)$$

323 where ef_t is the flow released as environmental flow and mef_t is the MEF at time
 324 t . MEF to be released downstream are set at 30% of the mean historical flow following
 325 recommendation by Smakhtin (2006). These sectoral performances are evaluated at the
 326 donor, recipient, and system level.

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3.4 Robustness metrics

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Several robustness metrics have been developed and applied to analyze performance of complex water resources systems, each representing a unique way to attain aggregate performance rankings for alternative solution strategies across a large number of uncertain SOWs (McPhail et al., 2018, 2020; Herman et al., 2015; Giuliani & Castelletti, 2016; Kwakkel, Eker, & Pruyt, 2016). The means of computing these aggregations are important in how they tacitly indicate the risk attitude of the decision maker(s). Here, we illustrate four aggregation strategies for robustness metrics that have been commonly used in the literature and represent a range of risk-attitudes (in order of increasing risk aversion): the maximax, Laplace, minimax regret, and maximin metrics (Table 3). The maximax metric (i.e., ‘best’) represents a low inherent level of risk aversion, as its calculation is only based on the best performance over all the scenarios. In contrast, the maximin metric (i.e., ‘worst’) has a very high level of intrinsic risk aversion as it only considers the worst performance of all scenarios, leading to a very conservative solution (Bertsimas & Sim, 2004). Thus, across all decision alternatives, the alternative that has the maximum worst-off performance across all deeply uncertain scenarios is deemed to be most robust. Similarly, the minimax regret metric assumes that the selected decision alternative will minimize the largest regret possible, focusing again on the worst-case relative performance. Laplace’s principle of insufficient reason, referred to as Laplace from hereon, is representative of a risk neutral metric as its calculation is based on the mean performance over all the scenarios considered. For each performance objective, values are estimated and rescaled between 0 and 1 to allow a comparison between objectives in calculation of robustness metrics.

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When multiple actors and sectors are involved, the implications of performance aggregation across the actor-sector combinations as well as scenarios need to be explored carefully. Stakeholders and decision makers are not likely to know *a priori* the complex effects of aggregation or how to specify robustness metrics as noted in Hadjimichael et al. (2020). To better aid stakeholders in understanding the relative implications of alternative robustness metrics, we more carefully distinguish the conceptual definition of candidate metrics across how they are aggregated with respect to scenarios as well as performance objectives. For example, when applied to a single performance objective, the maximin metric would focus on the minimum (‘worst’) performance value across all scenarios. The multi-objective version of maximin selects the worst performing objective across all of the performance objectives as well as scenarios considered (Table 3). This version of the metric tracks maximal regret or loss across the four performance objectives across alternatives and scenarios. A total of 12 actor-sector combinations along with four levels of risk aversion result in 48 combinations of stakeholder interests and risk attitude assumptions.

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3.5 Impact of multivariate robustness

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A total of 432,000,000 robustness evaluations were carried out for each of the 81 alternatives. These result from a combination of 12 performance objectives, 4 robustness metrics, 300 (20+40+60+80+100) scenario sample sizes, 10000 stochastic realizations, and 3 scenario spread types (Figure 3). Rank stability of alternatives across the candidate specifications of robustness definitions is evaluated 720 times, representing 12 performance objectives, 4 robustness metrics, 5 scenario sample sizes, and 3 scenario spread types. An alternative is ranked 81 if it attains the highest robustness value and 1 for the least robustness value. We summarize the rankings via the median and the interquartile range (IQR) of the ranks under WCU and DU sampling cases. A strategy is defined as having a stable ranking if there is little or no change in median rank defined under WCU and DU. We classify a strategy as having an unstable ranking when the difference in median rank between WCU and DU is greater than 20 or has high (>60) IQR rank

Table 3. List of different robustness metrics considered in the analysis along with equations for aggregation. The multi-objective version of each metric is applied when multiple performance objectives are included. In the equations listed, n denotes the number of performance objectives considered, f denotes objective function performance, s denotes the set of SOWs in the analysis, a denotes the value to be evaluated across all alternatives and j denotes the j th SOW from the set of s . For example, considering the actor-sector combination in row 8 of Table 1, two performance objectives considered in robustness calculations would be the vulnerability of water supply of donor (Godavari) and recipient (Krishna) basins. When evaluated across deeply uncertain scenarios, the worst value across donor and recipient basins would be selected for each alternative. The alternative with the maximum-worst off case would then be identified as most robust.

Metri namec	Description	Method of com- bining multiple performance objectives (ag- gregation of "n" metrics)	Equation	Metric Choice
Maximin	Worst-case per- for- mance	worst case perfor- mance among 'n' objectives	$\min(\min f_1(s), \min f_2(s), \dots, \min f_n(s))$	Max
Maximax	Best-case per- for- mance	best case perfor- mance among 'n' objectives	$\max(\max f_1(s), \max f_2(s), \dots, \max f_n(s))$	Max
Laplace's principle of insuf- ficient reason	Mean per- for- mance	mean perfor- mance among 'n' objectives	$\text{mean}(\text{mean} f_1(s), \text{mean} f_2(s), \dots, \text{mean} f_n(s))$	Max
Minimax regret	The worst case of making a wrong decision in a given scenario	worst case cost of wrong decision in any given sce- nario among 'n' objectives	$\max(\max r_1(s), \max r_2(s), \dots, \max r_n(s))$ $r_i(s_j) = \max((f(s)) - f(s_j))$	Min

378 under DU. We also explore the impact of these choices on the inferred stability of a strat-
 379 egy.

380 Along with the rank stability of a strategy, the degree of change in the quantified
 381 robustness of a transfer strategy when moving from the internal variability focus of WCU
 382 sampling to broader DU sampling could also be of interest to stakeholders. We define
 383 this change in terms of median and IQR rank of strategies. We classify the strategy as
 384 “improving” for an increase in median rank or decrease in IQR rank, “deteriorating” for
 385 a decrease in median rank or increase in IQR rank, or “similar” for a difference in me-
 386 dian or IQR rank that falls within ± 2 ranked slots of original WCU value. We also as-
 387 sess the impact of using various actor-sector combinations on resultant robustness per-
 388 ception of strategies. For the transfer strategies identified, we perform a detailed assess-
 389 ment of robustness controls to identify which factors among the many considered are driv-
 390 ing robustness gradients across deeply uncertain scenarios (Step V, Figure 2).

391 3.6 Identification of robustness controls

392 This step identifies which deeply uncertain factors are most responsible for the fail-
 393 ure of alternatives to meet the performance requirements implied for each of the differ-
 394 ent robustness metrics (robustness controls). We use Classification and Regression Trees
 395 (CART) to identify the relative importance of different factors for meeting performance
 396 requirements specified across alternative robustness metrics across sampled scenarios.
 397 CART requires input of the uncertain factors of focus and their performance outcomes
 398 (success or failure) (Step VI, Figure 2). The method then recursively partitions the fac-
 399 tor space into subgroups to explain variation in failure or success outcomes (e.g., iden-
 400 tifying the combinations of uncertain factors as well as their specific values that result
 401 in performance failures). Given that CART identifies the region of factor space that leads
 402 to failures, it facilitates scenario discovery where decision makers can more carefully pin-
 403 point the most consequential scenarios of concern for a given INS IBWT operational de-
 404 sign alternative. This step was completed using the ‘rpart’ package to generate pruned
 405 trees and prevent overfitting using a ten-fold cross-validation process (Breiman et al.,
 406 1984; Therneau et al., 2010).

407 4 Results

408 4.1 Multi-sectoral performance of transfer strategies

409 We first analyze the multi-sector tradeoffs across the 81 water transfer strategies
 410 for the INS IBWT for the three sectors: ecology-TV, water supply, and ecology-MEF.
 411 Their performance is analyzed at the system level by estimating the average performances
 412 across both donor (Godavari) and recipient (Krishna) basins (Figure 4a, b). The sys-
 413 tem level performance of each strategy across all SOWs under WCU (DU) is plotted as
 414 a line crossing the three vertical axes, each representing a sectoral performance in Fig-
 415 ure 4a(b). Across the 79 Pareto-approximate strategies, the average volumetric deficits
 416 ranged from 222-348 Mm³ (2.4% - 3.8% of total demands) for the water supply sector
 417 under WCU (Figure 4a). For these strategies, the reliability of maintaining MEF ranged
 418 from 97-98% for the ecology-MEF sector, while mean annual transfer volumes ranged
 419 from 4985-7730 Mm³ for the ecology-TV sector, under WCU. Notable is the tradeoff be-
 420 tween the ecology-MEF and water supply sectors at the system level, a 1% increase in
 421 MEF reliability requires a concurrent increase of 118 Mm³ in average volumetric deficits.
 422 The *proposed* strategy results in the worst performance for the ecology-MEF (MEF re-
 423 liability of 96.3%) and ecology-TV (mean annual transfer volume of 13,437 Mm³) sec-
 424 tors. The *no-transfer* strategy results in the worst performance of the water supply sec-
 425 tor with an average volumetric deficit of 1547 Mm³ (17% of total demands), respectively,
 426 at the system level. We surmise that the transfer of water between the Godavari and Kr-

427 ishna basins is likely to force decision makers to consider the significant tradeoffs between
 428 the water supply and ecology sectors in both basins.

429 On further analyzing these strategies under deeply uncertain futures, we find a sub-
 430 stantial deterioration in the performance of the water supply and ecology-MEF sectors
 431 when compared to the narrower evaluation of performance under WCU (DU, Figure 4b).
 432 The average volumetric deficits across the Pareto-approximate strategies increase from
 433 222 Mm³- 348 Mm³ to 1,593 Mm³-1,820 Mm³ as we transition from an emphasis on hy-
 434 droclimatic internal variability in the WCU evaluations to the broader uncertainties posed
 435 by climate and demand changes. Similarly, the reliability of maintaining MEFs reduces
 436 from 97-98% under WCU to 90- 91% under DU. The mean annual transfer volume re-
 437 duces from 13,437 Mm³ under WCU to 8302 Mm³ under DU for the proposed strategy.
 438 However, the annual volumetric transfers do not change substantially for the 79 dynamic
 439 state-aware solutions as they adapt to changing inflow and demand conditions under the
 440 DU SOWs. The *proposed* strategy attains a 90% reliability of maintaining MEF, the worst
 441 performance for the ecology-MEF sector under the DU SOWs across all strategies. The
 442 *no-transfer* strategy attains the highest performance for the ecology-MEF sector under
 443 DU futures but still results in the lowest performance for the water supply sector. Thus,
 444 even under the more challenging DU SOWs, the Pareto approximate strategies deteri-
 445 orate less than the *proposed* and *no-transfer* strategies.

446 We further identify four strategies that represent different possible compromises
 447 between the three sectors at the system level. The *Best Water Supply strategy* attains
 448 the highest performance in the water supply sector from the system perspective under
 449 WCU (red line, Figure 4). This strategy is likely to be of high interest to all water users
 450 including farmers and urban centers as well as regional water planners who typically pri-
 451 oritize augmentation of freshwater supply. The *Best Ecology-MEF* strategy attains the
 452 highest performance for the ecology-MEF sector at the system level under both the WCU
 453 and DU SOWs (purple line, Figure 4). Considering the ecological services provided by
 454 the Godavari River downstream of the proposed Inchampalli dam site, these strategies
 455 would be of interest to ecologists and dependent downstream water users. The *Best Ecology-*
 456 *TV* strategy results in the lowest annual volumetric transfers from the Godavari to the
 457 Krishna river under both the WCU and DU SOWs (yellow line, Figure 4). This strat-
 458 egy would be of interest to stakeholders who would be concerned about the potential im-
 459 plications of mixing the waters of the Godavari with the Krishna, resulting in the intro-
 460 duction of new aquatic species in the Krishna River. It will also be of interest to stake-
 461 holders concerned with the cost of constructing and maintaining of the INS IBWT it-
 462 self. The *Compromise* strategy represents the willingness of stakeholders to negotiate
 463 across sectors under both the WCU and DU SOWs (blue line, Figure 4). Together, these
 464 four strategies along with the *proposed* and *no-transfer* strategies, represent a range of
 465 stakeholders' interests including regional planning authorities, environmentalists, ecol-
 466 ogists, water users, tribal populations dependent on MEFs, etc. We further examine these
 467 in more detail w.r.t to implied actor-sector tradeoffs as well as implications of robust-
 468 ness definitions.

4.1.1 Key Actor-Sector Tradeoffs under WCU and DU

470 We now examine the tradeoffs between the three sectors for each actor perspective
 471 (donor-Godavari, recipient-Krishna, and system) associated with the INS IBWT to fur-
 472 ther understand the compromises faced by the participating basins (Figure 5). The av-
 473 erage demand deficits for the water supply sector under WCU ranged from 24-33 Mm³,
 474 415-672 Mm³, and 222-348 Mm³ for the donor, recipient and system, respectively. The
 475 reliability of maintaining MEF, representing the ecology-MEF sector, ranges from 94-
 476 97%, 99-99%, and 97-98% under WCU for the donor, recipient, and system, respectively.
 477 A key tradeoff emerges between the ecology-MEF and water supply sectors of the donor
 478 basin where increasing demand satisfaction by 9 Mm³ is attained at the cost of 2% re-

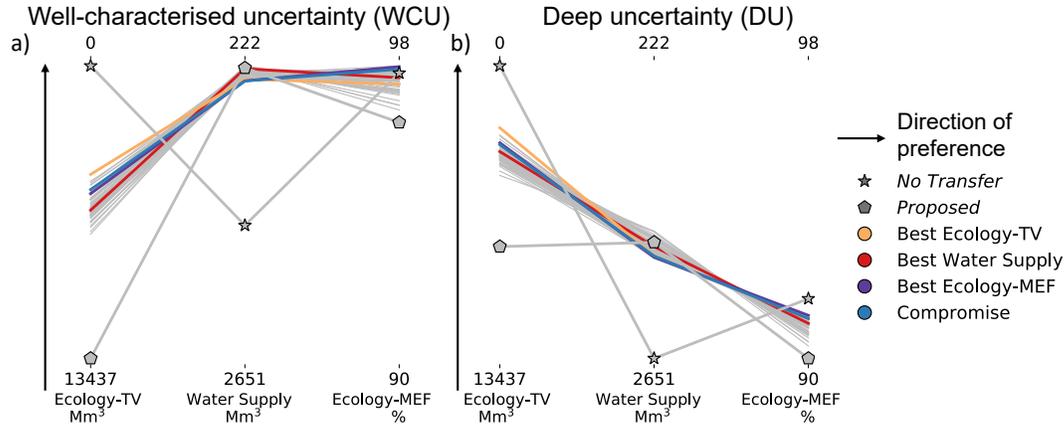


Figure 4. (a) Parallel coordinate plots showing performance of each sector for the system actor for all strategies under well-characterized uncertainty (WCU). Each vertical axis represents sectoral performance ranging from lowest (bottom) to highest (top) performance. Each strategy is represented by a line crossing the three axes. (b) Same as (a) but for all strategies reevaluated under deeply uncertain (DU) futures.

479 reduction in MEF requirements under WCU. Notably, the proposed strategy attains the
 480 highest performance (99.4%) in the ecology-MEF sector for the recipient-Krishna basin,
 481 but it does so by incurring a concurrent loss of MEF reliability in the donor-Godavari
 482 basin (93%). This results in the proposed strategy performing the worst for the ecology-
 483 MEF sector at the system level (96.3%). Thus, gains by sharing water between the Go-
 484 davari and Krishna basins will entail a tradeoff between the water supply sector of the
 485 recipient-Krishna basin and ecology-MEF sector of the donor-Godavari basin, even when
 486 considering historical hydroclimatic variability.

487 The ecology-MEF sector witnesses a substantial system level performance reduction
 488 under DU futures, which is primarily due to the deteriorating MEF reliability of the
 489 donor-Godavari basin. Under DU futures, we observe a small reduction in MEF reli-
 490 ability for the recipient-Krishna basin despite an overall reduction in mean annual wa-
 491 ter transferred. This suggests that water transfers may alleviate some MEF shortages
 492 in the recipient basin. We also find a reduction in system level water supply performance
 493 under DU futures, driven primarily by substantial reduction in for the recipient-Krishna
 494 basin. Historically, the donor-Godavari basin has had lower demand and hence the im-
 495 pact on water supply performance is limited. Importantly, for all strategies, including
 496 *proposed* and *no-transfer*, a reduced performance for water supply and ecology-MEF sec-
 497 tors for all actors, and an increased performance for ecology-TV sector, is observed un-
 498 der DU futures compared to WCU. Reduced transfer volumes under DU compared to
 499 WCU is due to change in water availability and increased demands in both the basins

500 4.2 Rank stability of strategies

501 Decision analysis frameworks should provide insights for how problem framing in-
 502 fluences the preferential ordering of suggested actions across the diverse actors and sec-
 503 tors that have stakes. In our study, different robustness metrics represent alternative world
 504 views by exploring candidate performance goals across actor-sector combinations and
 505 their risk attitudes. It further follows that each robustness metric is likely to result in
 506 a different rank ordering of decision alternatives. The rank stability of the decision al-
 507 ternatives may thus be an additional feature of interest to planners, especially in deci-
 508 sion contexts where it is conceptually challenging to define the appropriate robustness

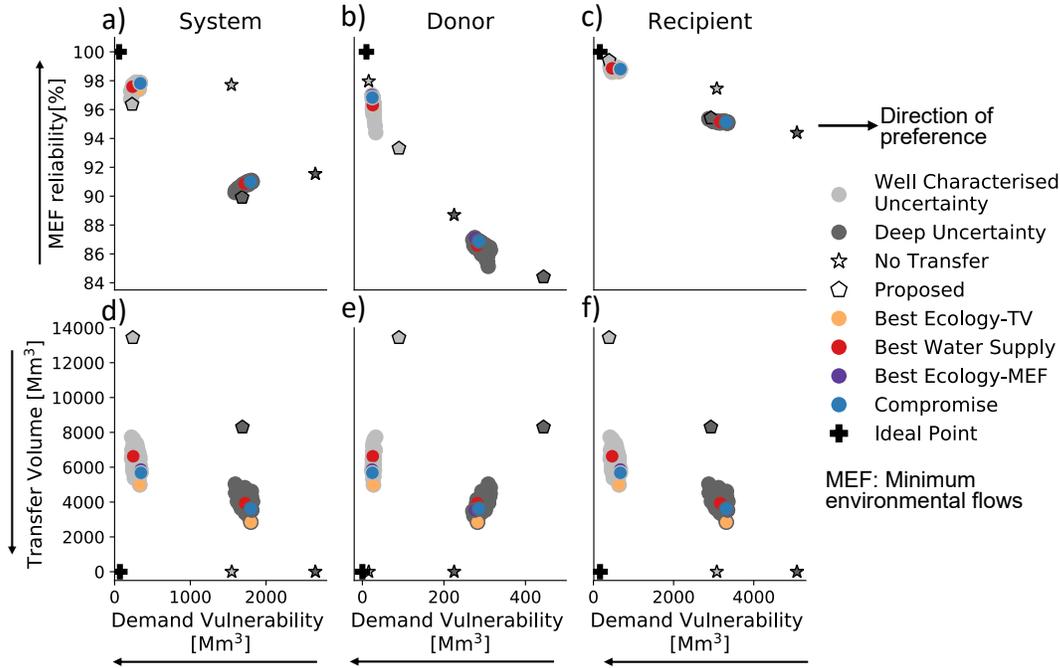


Figure 5. Trade-off between (a-c) vulnerability of water supply and reliability of maintaining MEF; (d-f) vulnerability of water supply and mean annual transfer volumes for (b, e) donor, (c, f) recipient and (a, d) system. The Pareto-approximate strategies are highlighted by circles. Performance under well-characterized uncertainties is shown by light grey circles while deep uncertainties in dark grey. MEF: minimum environmental flows.

509 metrics such as the INS IBWT. To investigate this, we plot the median and inter-quartile
 510 range (IQR) of the rank obtained by a strategy across all 720 robustness metric defini-
 511 tions under both WCU and DU (Figure 6a, b). A strategy with highest median rank and
 512 lowest IQR of rank indicates a high robustness irrespective of the choice of robustness
 513 definitions. The plausible highest rank in this study is 81 as there are 81 strategies and
 514 lowest is rank 1. Note that a strategy with high rank under WCU may not maintain its
 515 rank under DU. This can occur when a strategy is overly trained on historical data and
 516 exhibits a high-performance deterioration when exposed to DU futures. We further de-
 517 fine a strategy as stable when the difference in median rank of WCU and DU is less than
 518 20 or IQR rank of strategy is smaller than 60 under DU (shaded regions in Figure 6a,
 519 b). This choice of thresholds was determined after investigating the impact of different
 520 thresholds on resultant inferences of solution stability (Figure S3).

521 We find that the ranking of strategies is quite stable across the WCU and DU SOWs
 522 indicating that strategies tend to maintain similar relative performance under both cases
 523 (see also supplementary Figure S3). The stability of a strategy implies that the alter-
 524 native robustness-based preferential ordering of that strategy is largely consistent across
 525 multiple worldviews. The *proposed* strategy attains low median rank and high IQR of
 526 rank suggesting an overall low robustness with high variability across robustness defi-
 527 nitions. The *no-transfer* strategy attains the highest median rank across all robustness
 528 definitions under both WCU and DU SOWs but also exhibits a greater instability in rank-
 529 ing as indicated by its highest IQR in both cases. Table 4 summarizes the median rank,
 530 IQR of rank, as well as the stability ranking outcomes for the selected water transfer strate-
 531 gies. The Pareto-approximate strategies attain lower median ranks (i.e., higher median
 532 rank is preferred over lower ranks) when compared to the *no-transfer* strategy. They also

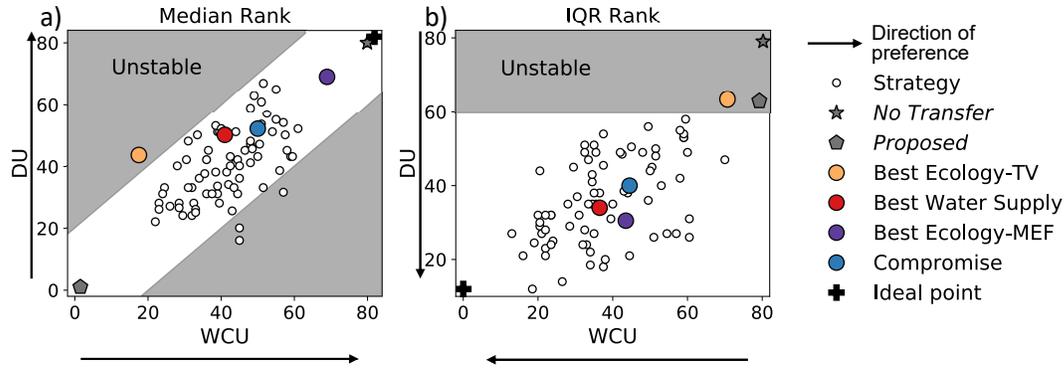


Figure 6. The (a) median and (b) interquartile range (IQR) of rank for a strategy under WCU (x-axis) and DU (y-axis). A total of 81 strategies are ranked using 720 robustness metrics under both WCU and DU. The ideal point with highest median rank and lowest interquartile range is highlighted by a plus symbol in both panels. Grey shading represents regions of instability w.r.t strategy ranking. See text for more details.

533 maintain higher rank stability as exhibited by their low IQR (i.e., low IQR is preferred)
 534 as well as consistency of ranking between the WCU and DU SOWs. The *Best Ecology-*
 535 *MEF* strategy attains the highest median rank among the Pareto approximate strate-
 536 gies and has low IQR. The *Compromise* strategy has a relatively high median ranking,
 537 as well as lower IQR of rank under both WCU and DU SOWs. The *Best Ecology-TV*
 538 strategy is found to be unstable based on the criteria discussed above, which is mainly
 539 attributed to the poor performance of this strategy for the water supply sector. Over-
 540 all, the selected strategies display advantages over one another either w.r.t individual sec-
 541 toral performance or in rank stability across robustness choices. Ideally, a strategy with
 542 the highest median rank and lowest IQR of rank across the robustness definitions should
 543 be preferred. Such a strategy would maintain performance irrespective of the choice of
 544 actor-sector combinations and risk attitudes. However, we find that the median rank and
 545 IQR of rank have a trade-off across the set of strategies analyzed here. This indicates
 546 that strategies that attain a high rank across robustness metrics also display greater vari-
 547 ability of ranking, resulting in lower performances in certain actor-sector combinations.
 548 Thus, choosing an appropriate water transfer strategy for the INS IBWT would be dif-
 549 ficult and require careful consideration of involved tradeoffs under deeply uncertain fu-
 550 tures.

551 4.3 Impact of stakeholder(s) interests and risk-attitudes on perceived 552 robustness

553 A key objective of this study is to demonstrate how decision makers may explore
 554 different risk attitudes or stakeholders' interests in the evaluation of design alternatives
 555 robustness using the complex context of the INS IBWT. The exploratory evaluation of
 556 the consequences of the different risk attitudes across candidate robustness metrics can
 557 provide a broader context for how outcomes may be classified as being consequential across
 558 the range from full optimism to extreme pessimism. We visualize the variation of robust-
 559 ness values across actor-sector combinations, and risk attitudes for six selected strate-
 560 gies under DU SOWs as bar plots in Figure 7. We reiterate that across the candidate
 561 operational strategies for the INS IBWT, the preferred robustness for the Maximax, Laplace,
 562 and maximin metrics assumes maximization. Similarly, to choose the best robustness
 563 value for the minimax regret metric, the robustness values are subtracted from a value
 564 of 1 for consistency as this regret measure is minimized. Across all robustness metrics,

Table 4. The median and IQR of rank for selected strategies under WCU and DU.

Strategy Name	Selection criteria	Median Rank		IQR rank		Comment on Stability		Whether median (IQR rank) improves from WCU to DU
		WCU	DU	WCU	DU	Difference in median rank of WCU and DU	Based on IQR rank of strategy	
Proposed	Baseline strategy	1.5	1	79	63	Stable	Instable	Similar (Improving)
No-transfer	Status quo	80	80	80	79	Stable	Instable	Same (Similar)
Best Ecology-TV	Strategy with minimum transfer volume under WCU and DU	17.5	43.5	70.5	63.5	Instable	Instable	Improving (Improving)
Best Ecology-MEF	Best performance for ecology-MEF under DU	69	69	43.5	30.5	Stable	Stable	Same (Improving)
Best Water Supply	Best performance for water supply under WCU and DU	41	50	36.5	34	Stable	Stable	Improving (Improving)
Compromise	Strategy with compromise performance across sectors	50	52	44.5	40	Stable	Stable	Similar (Improving)

565 the highest robustness value is attained by a variety of strategies depending upon the
 566 choice of actor-sector combination is emphasized. This shows that a single robust INS
 567 IBWT operational strategy cannot easily be identified without a deeper engagement with
 568 the trade-offs between different risk attitudes and carefully evaluating the choice of which
 569 actor-sectors that have a central role in decision making.

570 Figure 7 shows that the *no-transfer* strategy attains the highest robustness value
 571 compared to the other strategies across all levels of risk aversion for actor-sector com-
 572 binations of donor water supply, donor ecology-MEF and system ecology-TV. It is ex-
 573 pected that the *no-transfer* strategy results as being robust for the donor (Inchampalli)
 574 water supply and donor ecology-MEF combination as it avoids conflicts in resource shar-
 575 ing with the recipient basin. The *proposed* strategy is found to be robust for the recip-
 576 ient (Nagarjuna Sagar dam) water supply across all metrics and recipient ecology-MEF
 577 except for minimax regret. In summary, for donor related combinations, the *no-transfer*
 578 strategy is robust, while for recipient related combinations the highest metric value is
 579 attained by the *proposed* strategy. Not opting for the water transfer would be in the best
 580 interest of donor's water supply and ecology goals, while the *proposed* strategy entails
 581 the highest possible value of annual volumetric transfers. Similarly, for system ecology-
 582 TV which focuses on minimizing the transfer volume, the *no-transfer* strategy attains
 583 the highest robustness as the transfer volume is set to the minimum value of zero. Al-
 584 ternatively, system level actors for the INS IBWT are mainly decision makers focused
 585 on the overall average benefits across both the Inchampalli and Nagarjuna Sagar com-
 586 mand areas.

587 As expected, the INS IBWT increases the robustness of water supply at the sys-
 588 tem level. Across all levels of risk aversion, the Pareto optimal strategies display greater
 589 robustness when compared to the *no-transfer* strategy for the water supply sector at the
 590 system level. Note also that at the system level, the robustness of strategies for Ecology-
 591 MEF sector is markedly different than for the water supply sector suggesting that stake-
 592 holders with a high preference towards the water supply sector may select strategies that
 593 pose higher risks for violating MEFs. The *no-transfer* strategy attains greater robust-
 594 ness compared to other strategies for the Laplace and maximin metrics at the system
 595 level for the ecology-MEF sector as well as across all actors and sectors ('All' in Figure
 596 7). The Laplace metric captures risk-neutral mean performance across scenarios while
 597 the maximin metric captures risk averse performance. Among the optimal strategies, the
 598 *Best Ecology MEF* strategy attains high robustness for the maximin metric. The *Best*
 599 *Ecology TV* strategy attains the highest robustness when considering the minimax re-
 600 gret metric across all actors and sectors. Recall that this metric emphasizes alternative
 601 INS IBWT operational strategies that have minimal deterioration in their performance
 602 from an optimal baseline.

603 Metric combination number 12 (Table 4) represented as "All" in Figure 7 consid-
 604 ers all actors and sectors related to the INS IBWT. This robustness assessment metric
 605 is more stringent and difficult to attain high levels of performance compared to other actor-
 606 sector combinations. However, it does identify INS IBWT operational strategies that are
 607 consistently classified as robust across the different levels of risk aversion. This consis-
 608 tency is partially an artifact of the compensatory effects of combining actors and sec-
 609 tors in the measure of robustness. For example, the water supply sector may fail in cer-
 610 tain scenarios, but those failures are in aggregate countered by increasing levels of suc-
 611 cess for the ecology-MEF sector. Overall, we find that assumed levels of risk aversion
 612 has a far more dominant effect on candidate robustness measures than the number of
 613 samples and type of sampling strategy (Figure S4). In summary, we find that the *no-*
 614 *transfer* strategy remains robust when considering donor water supply, donor ecology-
 615 MEF and system ecology-TV actor-sector combinations, across all deeply uncertain fu-
 616 tures. On the other hand, the best water supply strategy performs the best for system
 617 water supply and donor, recipient water supply actor-sector combinations (Figure 7). Fur-

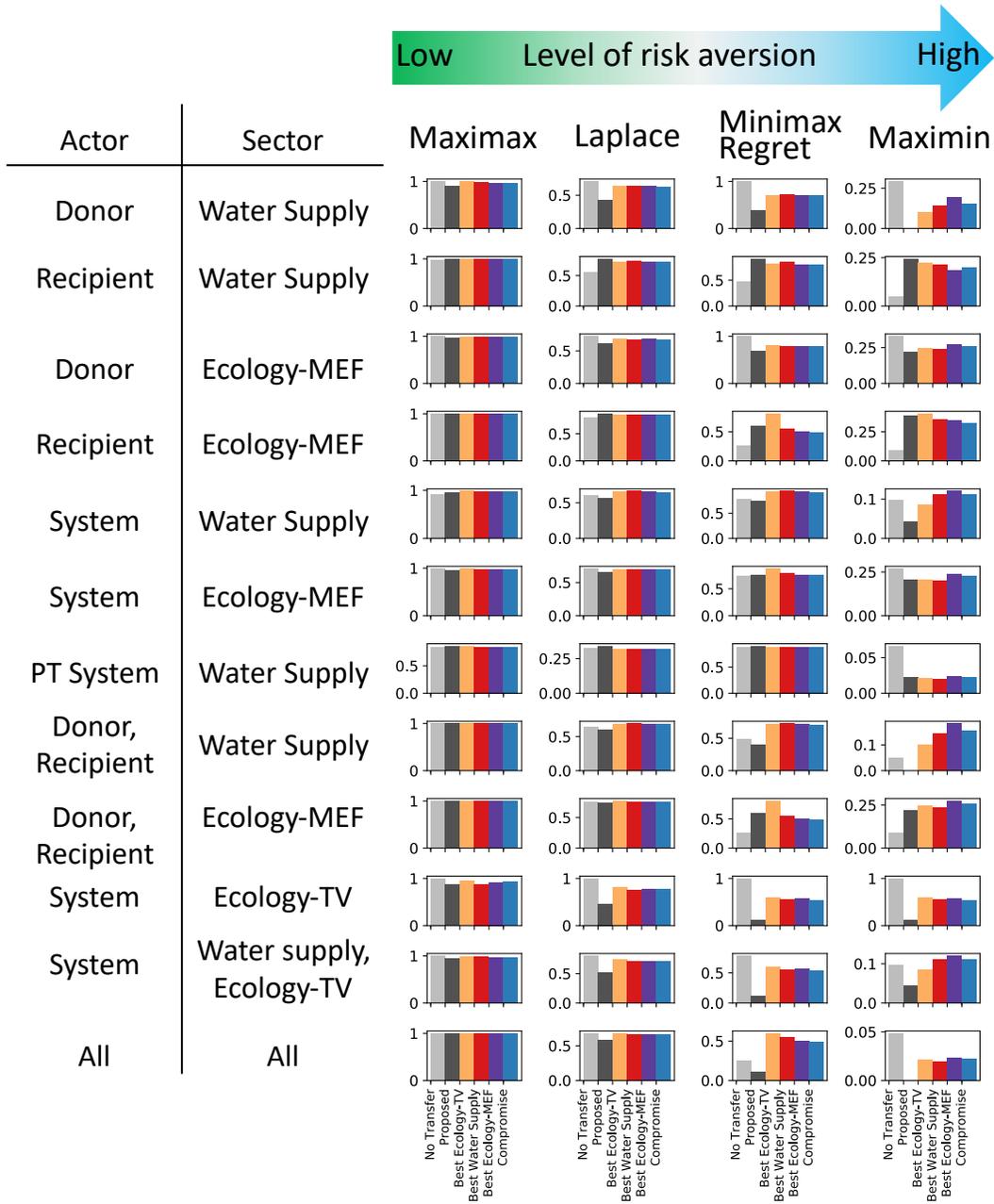


Figure 7. Robustness of selected strategies (from Table 1) for each combination of actor-sector and varying levels of risk aversion for uniform type sampling of scenarios. The arrow represents the increasing level of risk aversion with Maximax as least risk averse and maximin as highest risk averse.

thermore, when considered the most risk averse metric, the *no-transfer* strategy emerges as the most robust as it balances the deterioration in recipient water supply actor-sector against improvement in donor ecology-MEF and system ecology-TV actor-sector.

4.4 Influence of deeply uncertain factors

Scenario discovery helps identify deeply uncertain factors, which drive the performance deterioration of objective functions and potential strategy failure. Here, we identify which uncertain factors control the robustness of transfer alternatives to DU SOWs using CART to perform scenario discovery for each of the 79 Pareto-approximate strategies (Step VI of Figure 2). As an example, we perform this analysis for the system level water supply metric, the actor sector combination 5 for a uniform sampling of scenarios (Figure 8). Notably, the order of influence of the deeply uncertain factors on strategy failure is found to be the same across all strategies: amplitude of inflows to the donor and recipient basins, mean inflows to the donor and recipient basins, standard deviation of inflows to the donor and recipient basins and demands in the donor basin. Climate models struggle to reproduce the complex spatio-temporal patterns of the Indian Summer Monsoon (Kodra et al., 2012; Konduru & Takahashi, 2020; Saha et al., 2021), but understanding potential future river flows is crucial to understanding potential strategy failure. Our analysis suggests an urgent need to focus on understanding the potential temporal dynamics of future hydro-climatology of this region given its significantly important role in influencing strategy failure.

5 Conclusion

We apply an innovative framework to a major water transfer project in India, to illustrate how the role of different sectoral priorities, stakeholder preferences, policy options, uncertainties and robustness metrics, affect robustness assessments. This study contributes a proof-of-concept to demonstrate how evolving analytical frameworks can support infrastructure planning and decision making under uncertainty. Our results reveal how tacit assumptions within robustness metrics could influence the perceived robustness of INS IBWT policies. We find stronger variation in robustness values across different risk-attitudes and actor-sector combinations compared to sampling choices. Different actor-sector combinations may yield different robustness values of selected strategies. For example, when risk averse measures of robustness are applied to donor favoured measures of system performance, we find that the *no-transfer* strategy has the highest robustness. Alternatively, the *proposed* transfer is also identified as the highest rank for a selected stakeholder preferences which are recipient centred. Our analysis suggests that while the high-cost INS IBWT infrastructure investment may be considered feasible under historically observed stationary climatic conditions, that future climate change effects have the potential to strongly degrade its robustness performance across all of the operational strategies and actor-sector concerned assessed. In assessing the robustness of the INS IBWT, the distribution of scenarios has a greater impact on the inferred robustness values versus the number of scenarios considered, in agreement with prior analysis by McPhail et al. (2020). Overall, this study highlights the importance of an exploratory evaluation of the robustness of mega-investments projects.

The choice of robustness metric presents a daunting challenge for the complex decision context of the INS IBWT. It follows that an easy to attain performance goal will lead to high robustness values while a stricter performance requirement that maintain key system performance goals may result in lower robustness values. The ranking across robustness metrics therefore does not distinguish the relative value or importance of the underlying metrics to real operations, but rather the consequences of risk attitudes and stakeholder preferences. This could be altered in future studies with stakeholder elicitation to discover acceptable and stricter performance values. For example, the *proposed*

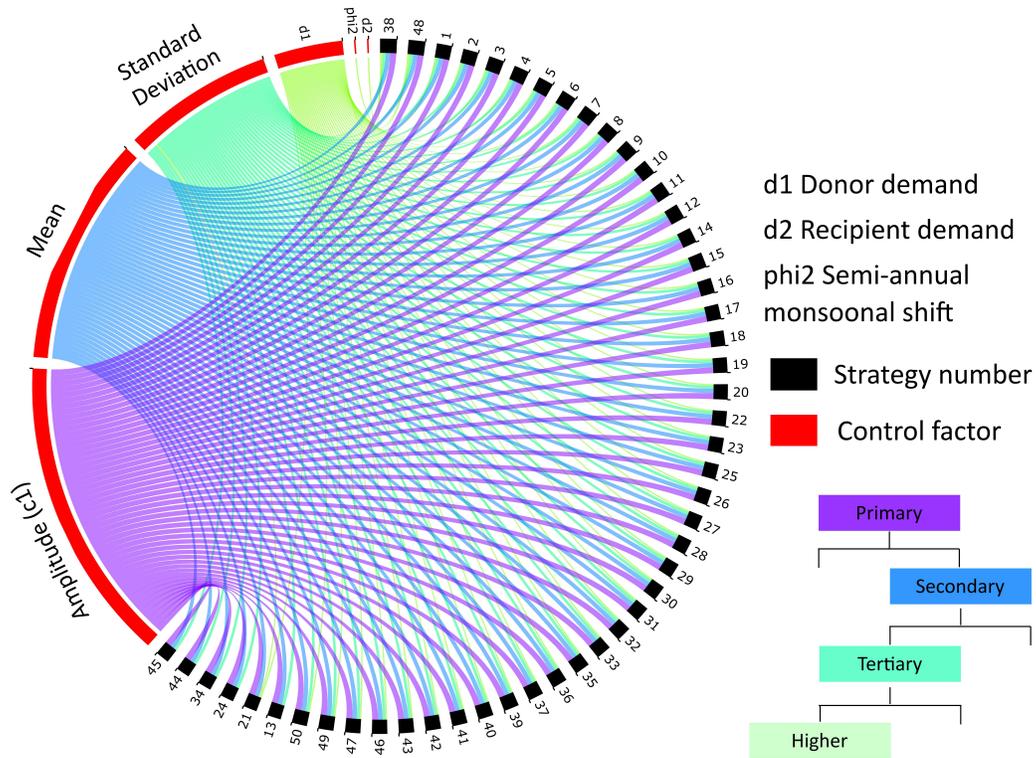


Figure 8. Understanding the importance of different uncertain factors on performance of optimized strategies using classification and regression trees (CART). Shown are the ranking of deeply uncertain factors: changes in amplitude, standard deviation, demands in recipient, mean, demand donor and phase shift in determining robustness of strategies. Purple, blue, green and lime green colors represent primary, secondary, tertiary and higher factor ranking. The CIRCOS plot displays the uncertain factors as the circles outer edge in red and each optimized strategy is shown on the circle's outer edge in black. A purple line connecting a strategy to a factor indicates that factor being the primary control on strategy failure under deeply uncertain futures.

668 strategy attains 90% reliability of maintaining MEF under the DU SOWs which is the
 669 worst performance compared to other strategies. Here, the contention between differ-
 670 ent decision makers on the acceptable level of risk emphasizes that future work would
 671 need to clarify the accepted value of reliability or other performance requirements. In
 672 other words, 90% reliability may be seen as a failure or sufficient across diverse decision
 673 makers.

674 In this study, we constructed combinations of actor-sector preferences based on an
 675 understanding of the stakeholders involved in the INS IBWT. The exploratory robust-
 676 ness assessment framework contributed here has significant potential to provide a quan-
 677 titative basis for stakeholder elicitation using a participatory modeling framework and
 678 aid in building a shared understanding of potential irreversible decision lock-ins. Such
 679 participatory approaches require inclusive thinking to account for different worldviews,
 680 priorities and preferences of marginalized communities and avoiding the monopolization
 681 of project benefits (Eriksen et al., 2021). While we know that the ‘planners’ associated
 682 with this project want to minimize system level deficits and that all stakeholders are nei-
 683 ther well represented nor consulted, there are issues regarding their understanding of de-
 684 cision analysis terminology and techniques. So, to facilitate an appropriate uptake of such
 685 approaches, it will require investments in building capacity and understanding of robust-
 686 ness, uncertainty, risks, and participatory stakeholder engagement. Additionally, research
 687 on the applicability and usefulness of approaches such as dynamic planning, will help
 688 improve the design and management of institutionally complex water resources systems
 689 balancing conflicting demands and complex interdependent risks.

690 Recent research has highlighted the complex nature of IBWTs and their multi-faceted
 691 challenges. We contribute to this growing body of literature by highlighting the type of
 692 information that advanced decision support can provide for better engaging a variety of
 693 stakeholders. This framework could also be extended to other robustness metrics such
 694 as satisficing criteria and higher-order moments. Analyzing the robustness of alterna-
 695 tives against different thresholds using the satisficing criteria, usefully indicates their sta-
 696 bility and is worth exploring, especially during participatory engagement. Stakeholders
 697 may implicitly favor one actor-sector over others because of hidden assumptions within
 698 their robustness analysis. The framework in this paper offers a means of revealing those
 699 hidden assumptions and making the decision process transparent. This has benefits of
 700 1) ensure stakeholders are not blind to potential risks and trade-offs and 2) aid the co-
 701 production process by providing insight into the implications for all actors-sectors.

702 **6 Open Research**

703 All code for replicating the analysis and figure generation can be found at [https://](https://doi.org/10.5281/zenodo.7470815)
 704 doi.org/10.5281/zenodo.7470815. DOI: 10.5281/zenodo.7470815. The inflow time se-
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Figure 1.

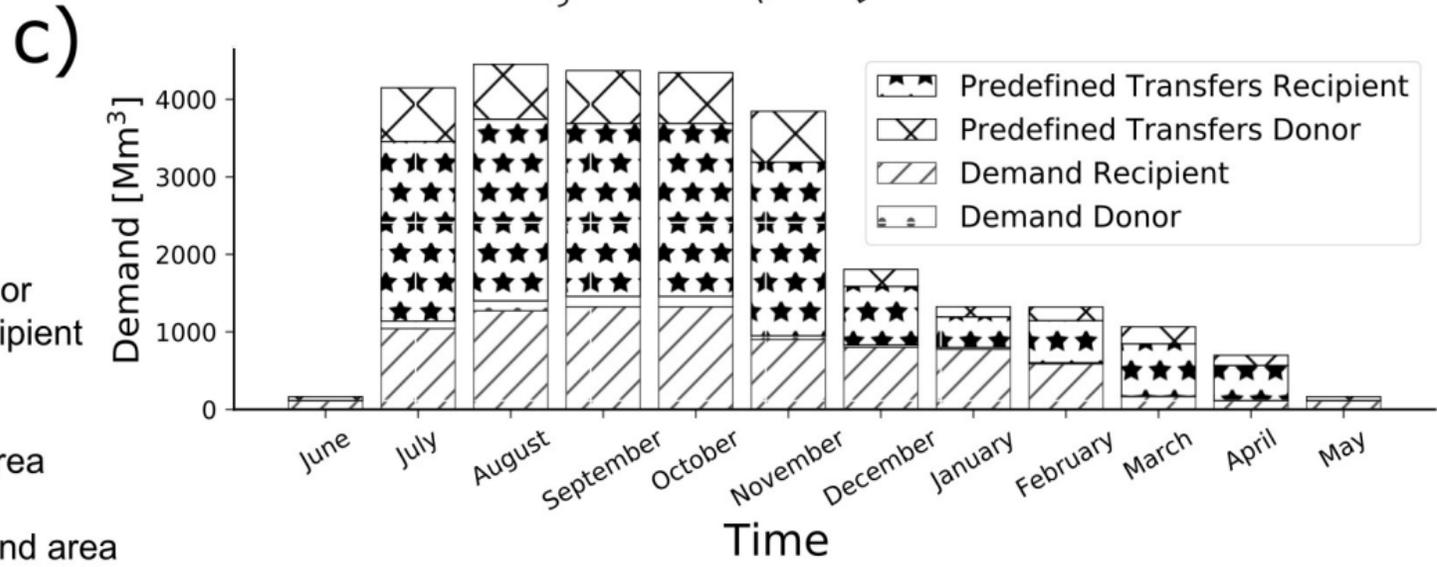
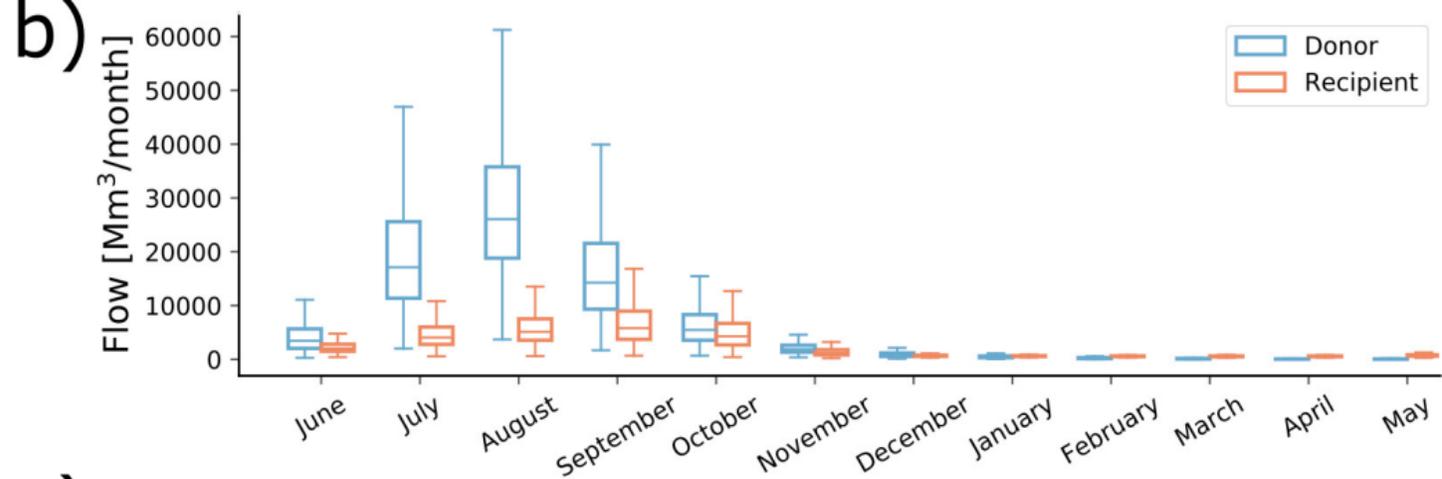
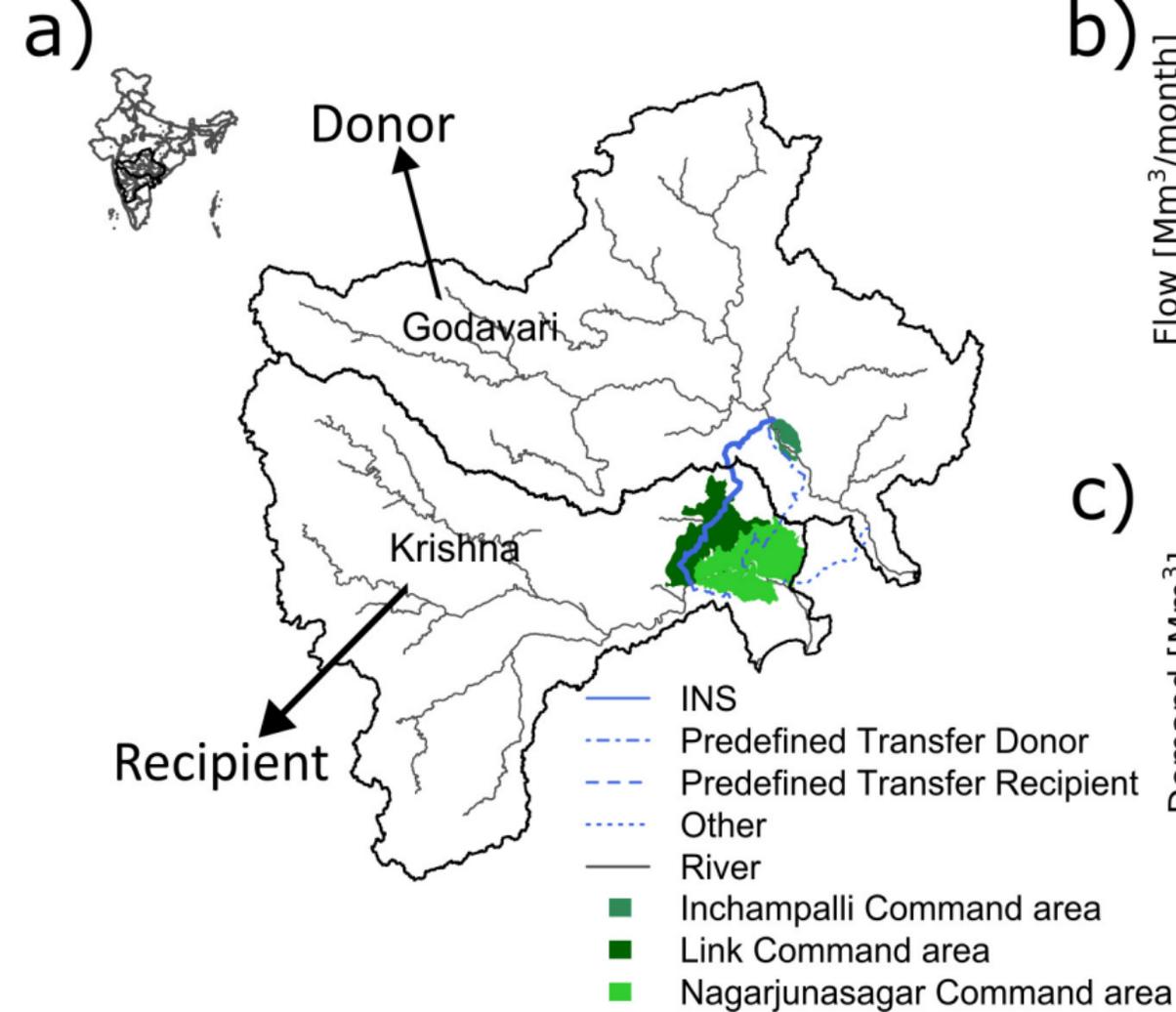


Figure 2.

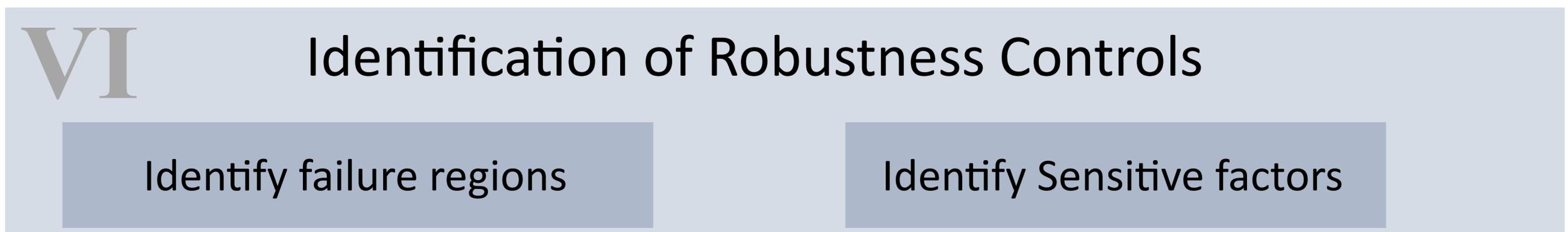
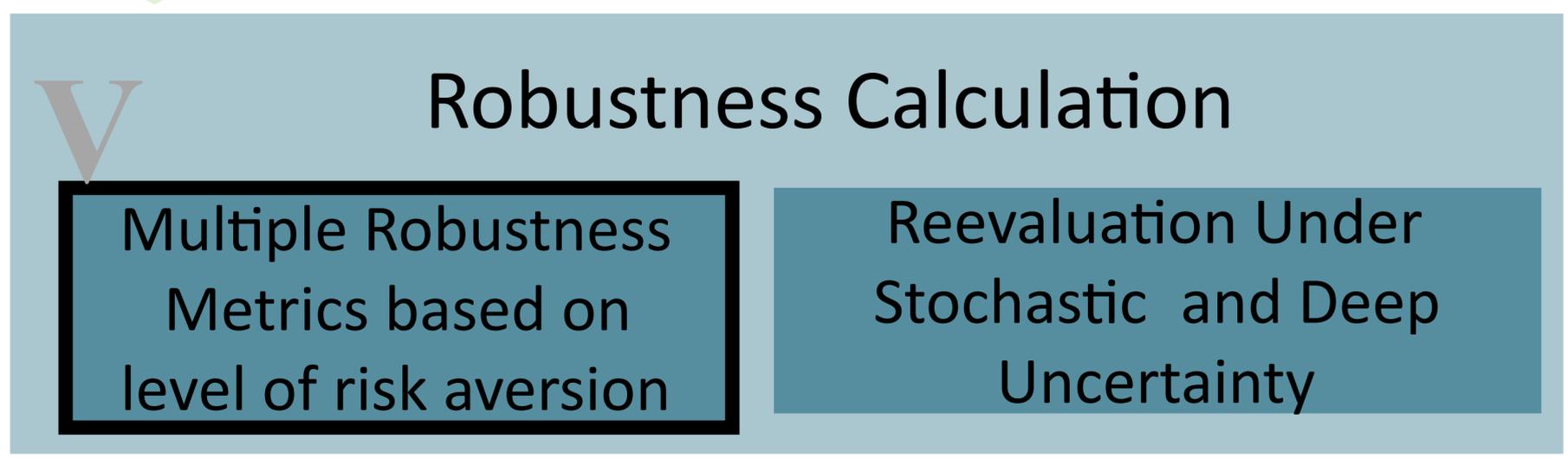
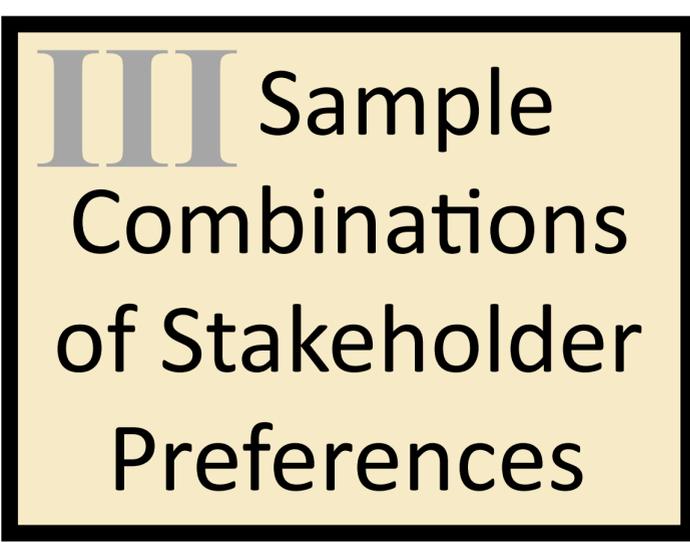
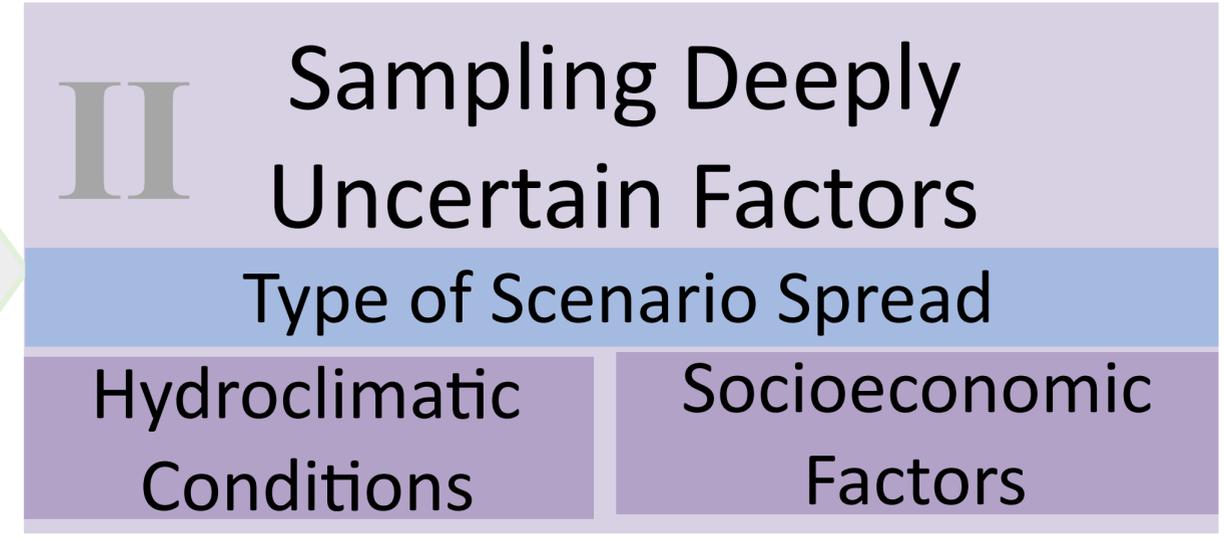
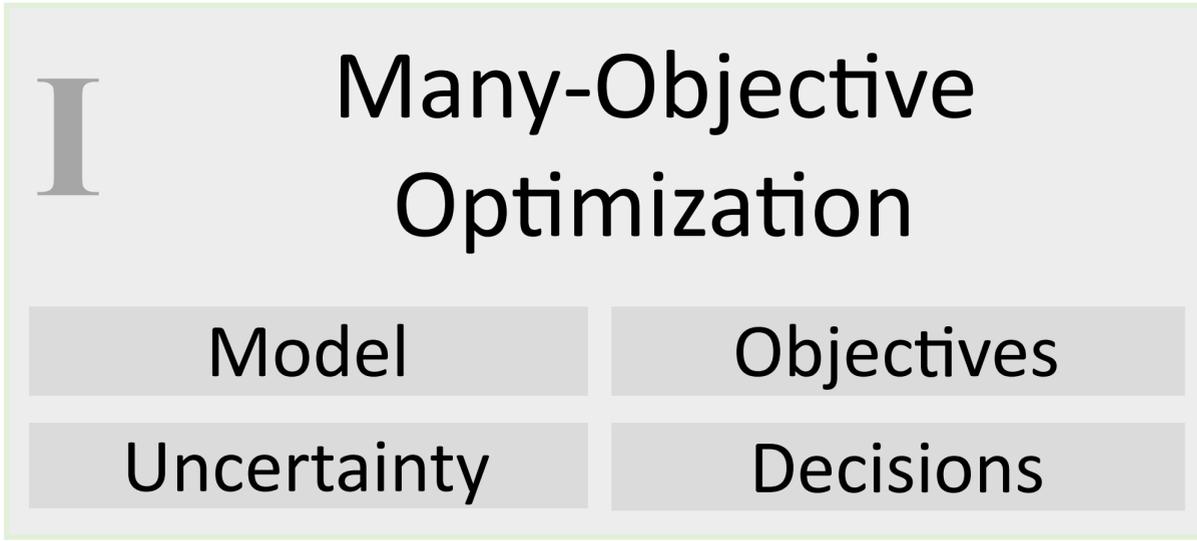
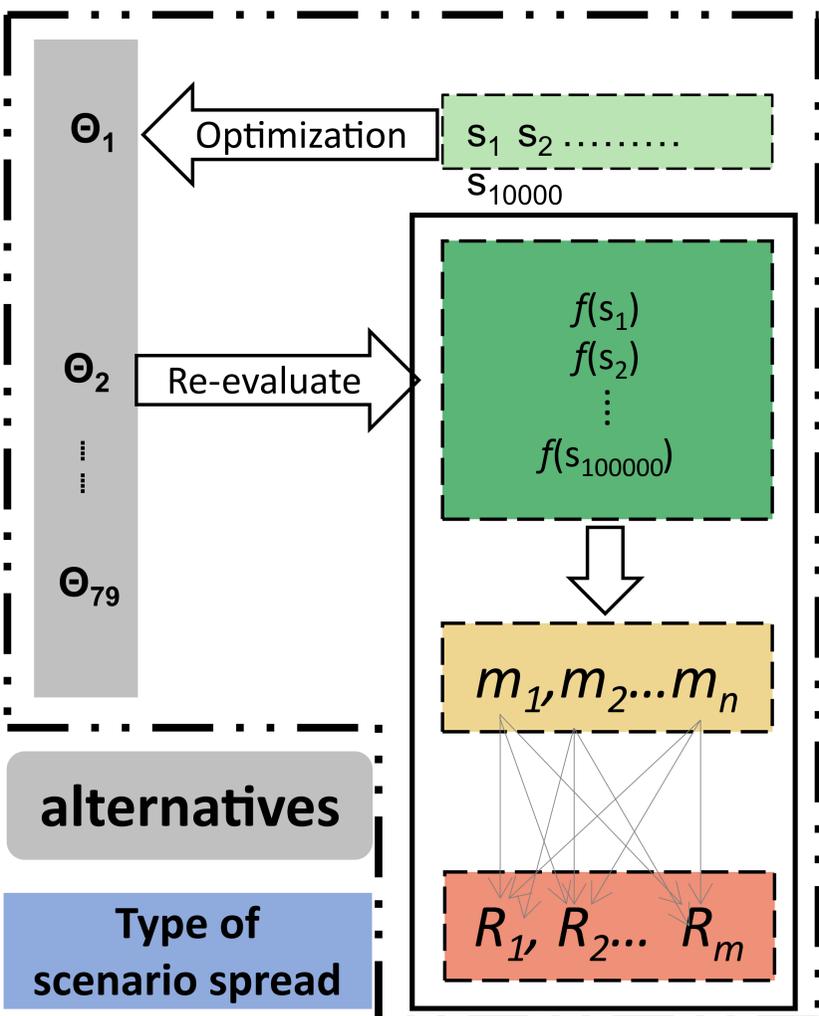


Figure 3.

a) Well Characterized Uncertainty



alternatives

Type of scenario spread

Ψ vector of deep uncertain factors

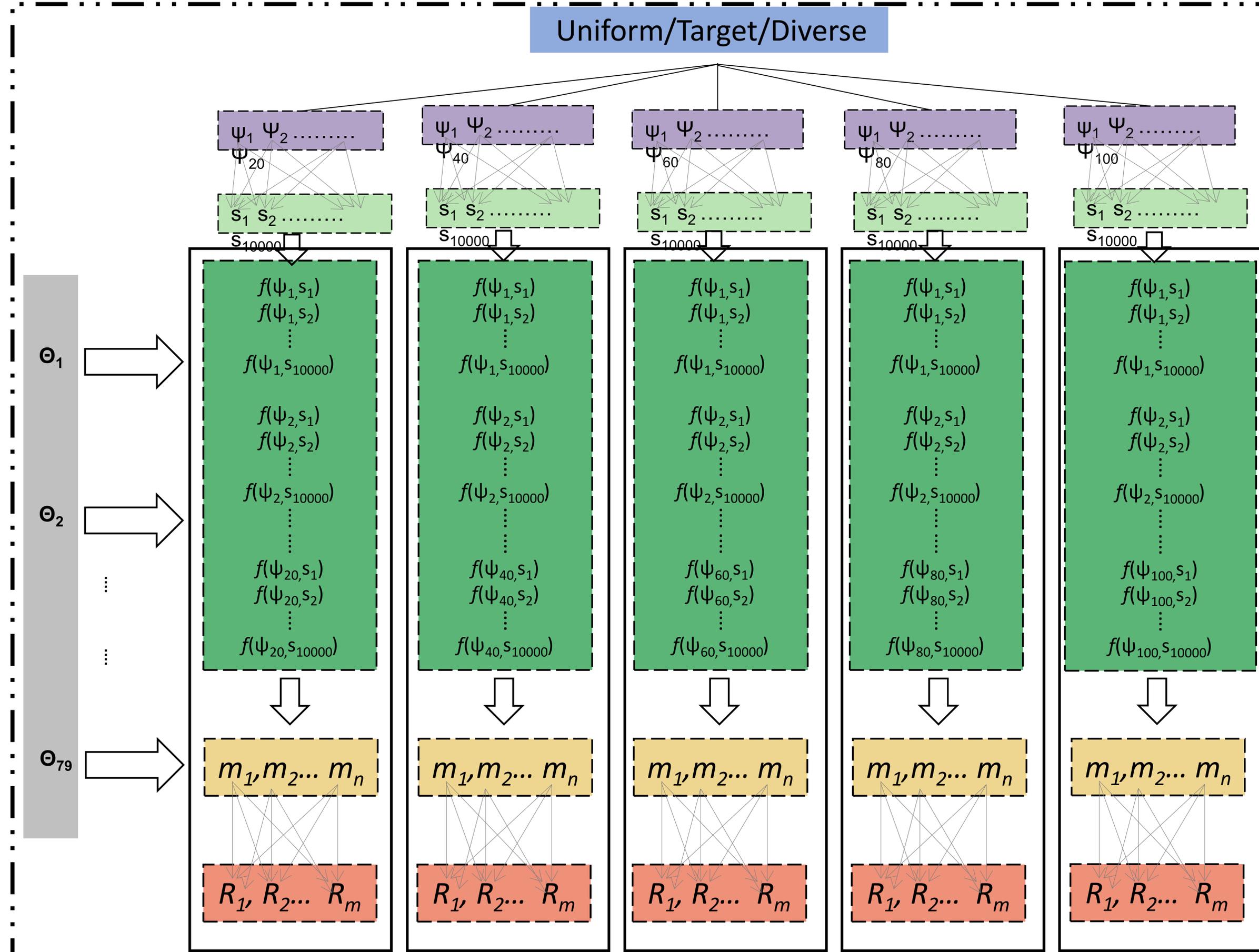
s streamflow realisation used for optimization

f vector of performance measures on evaluating IBT model

m_i combinations of performance measures

R Robustness metric value

b) Deep Uncertainty

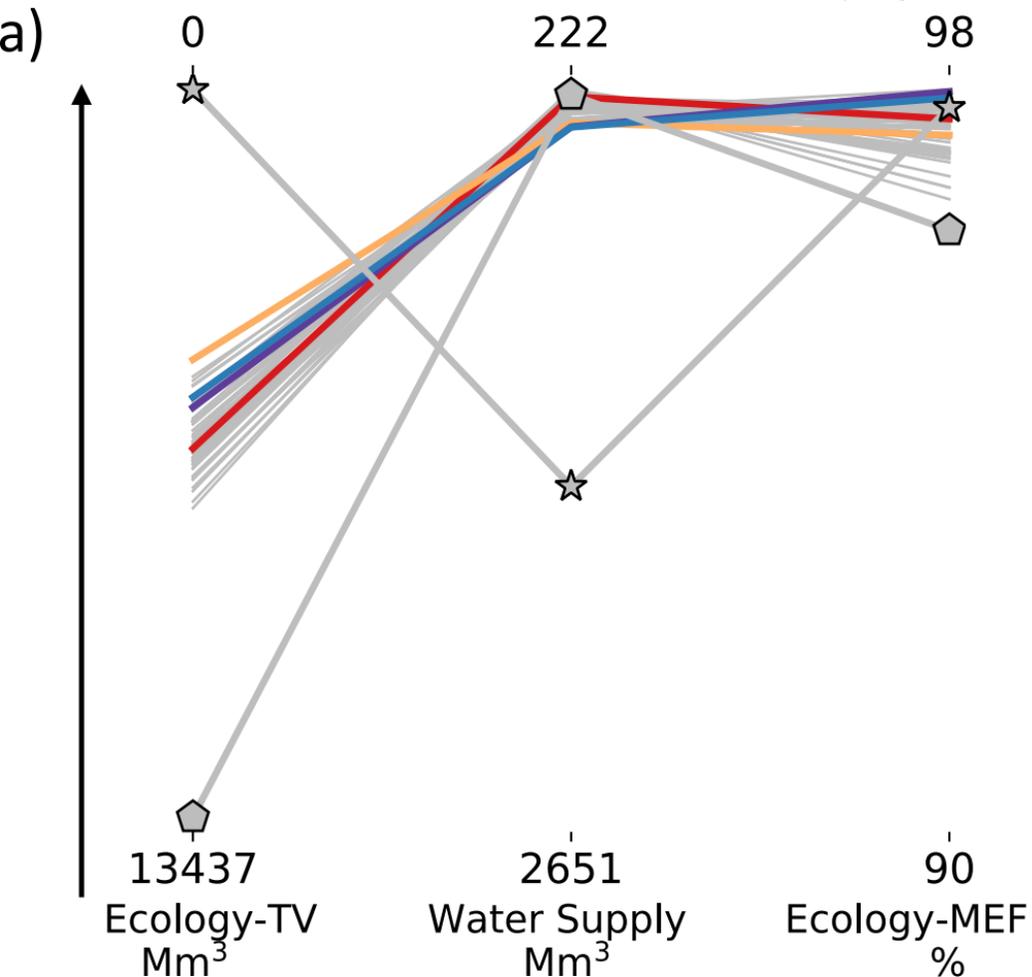


Uniform/Target/Diverse

Θ_1
 Θ_2
 \dots
 Θ_{79}

Figure 4.

Well-characterised uncertainty (WCU)



Deep uncertainty (DU)

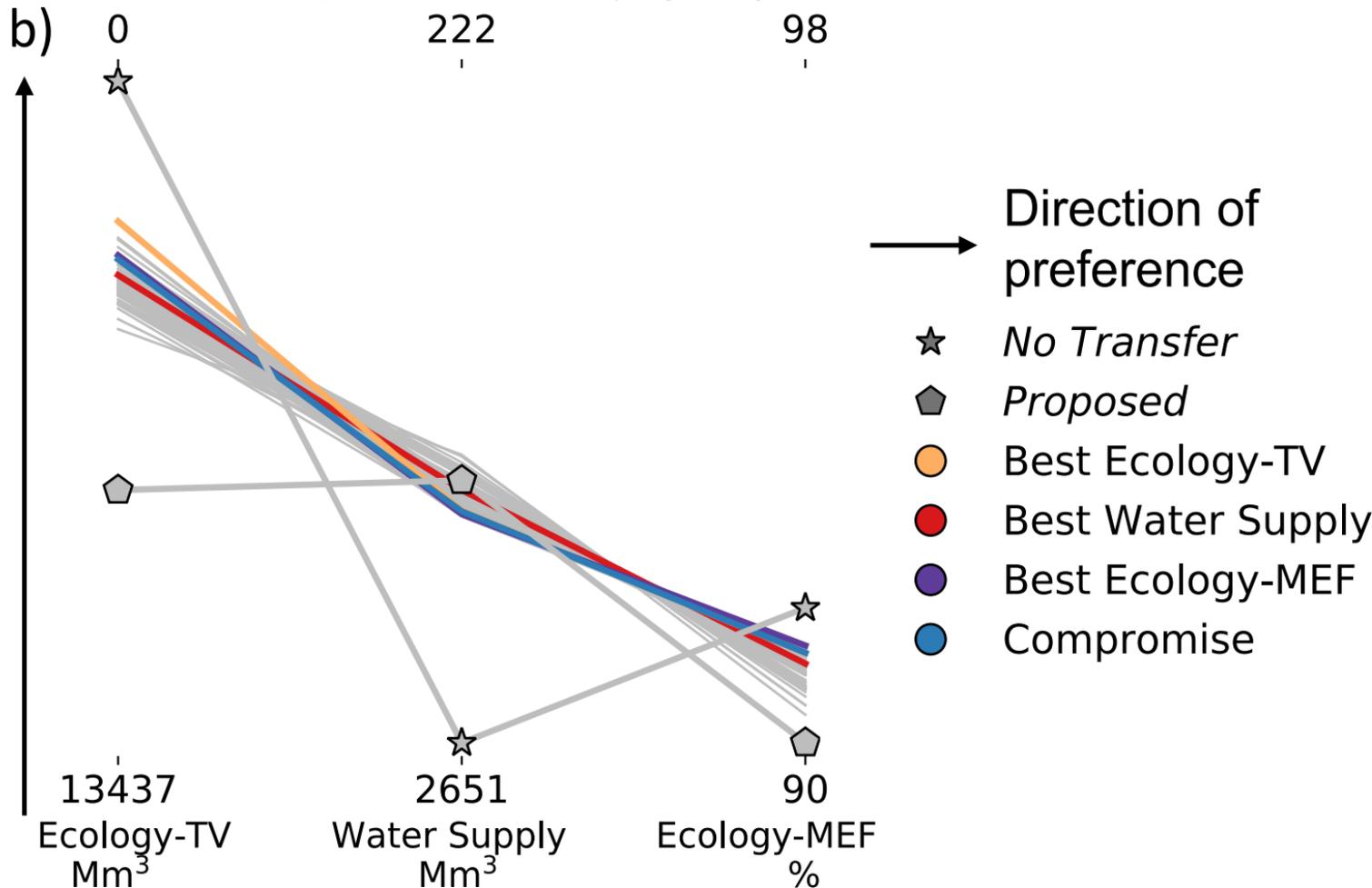


Figure 5.

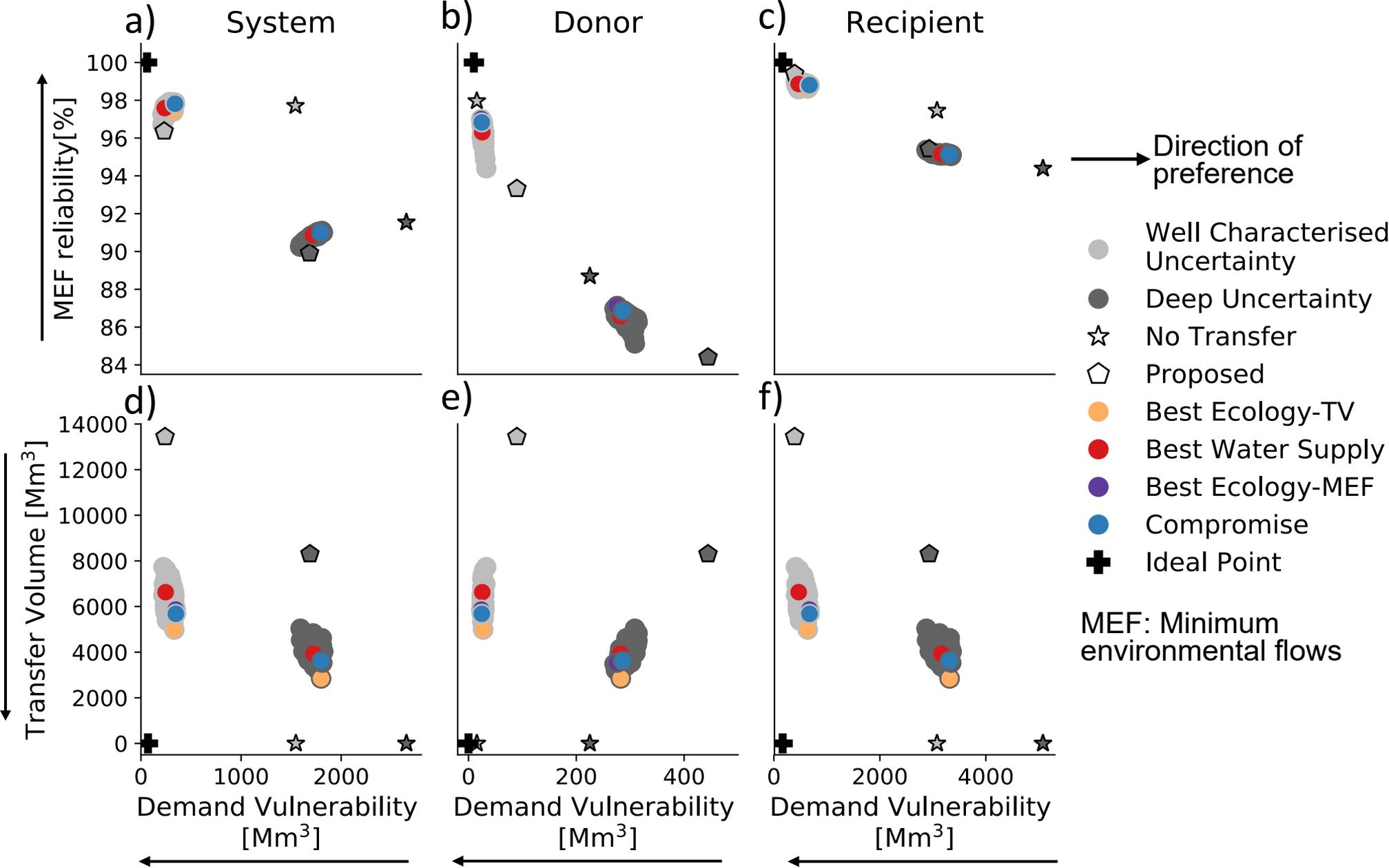


Figure 6.

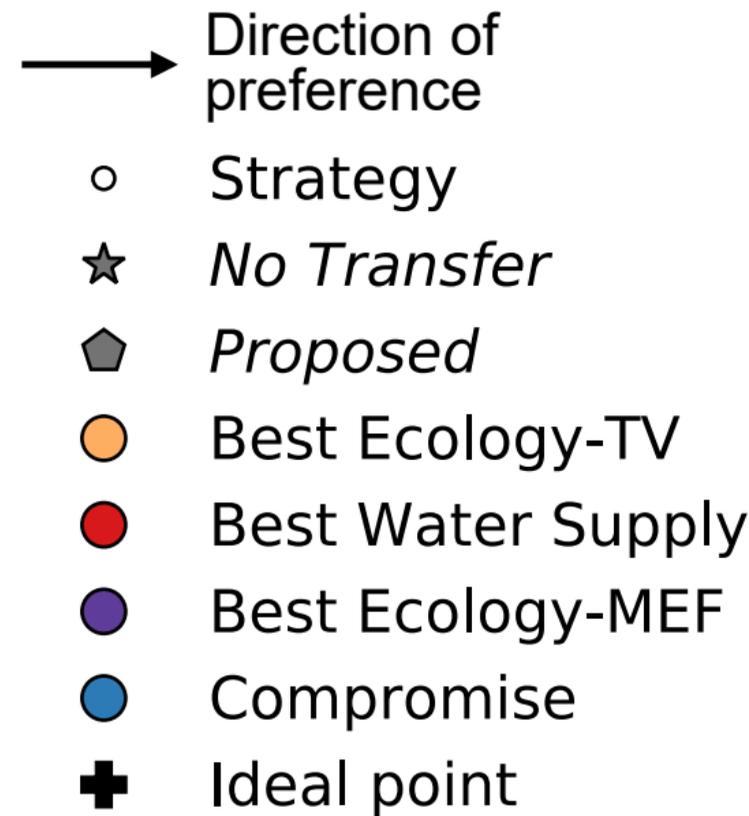
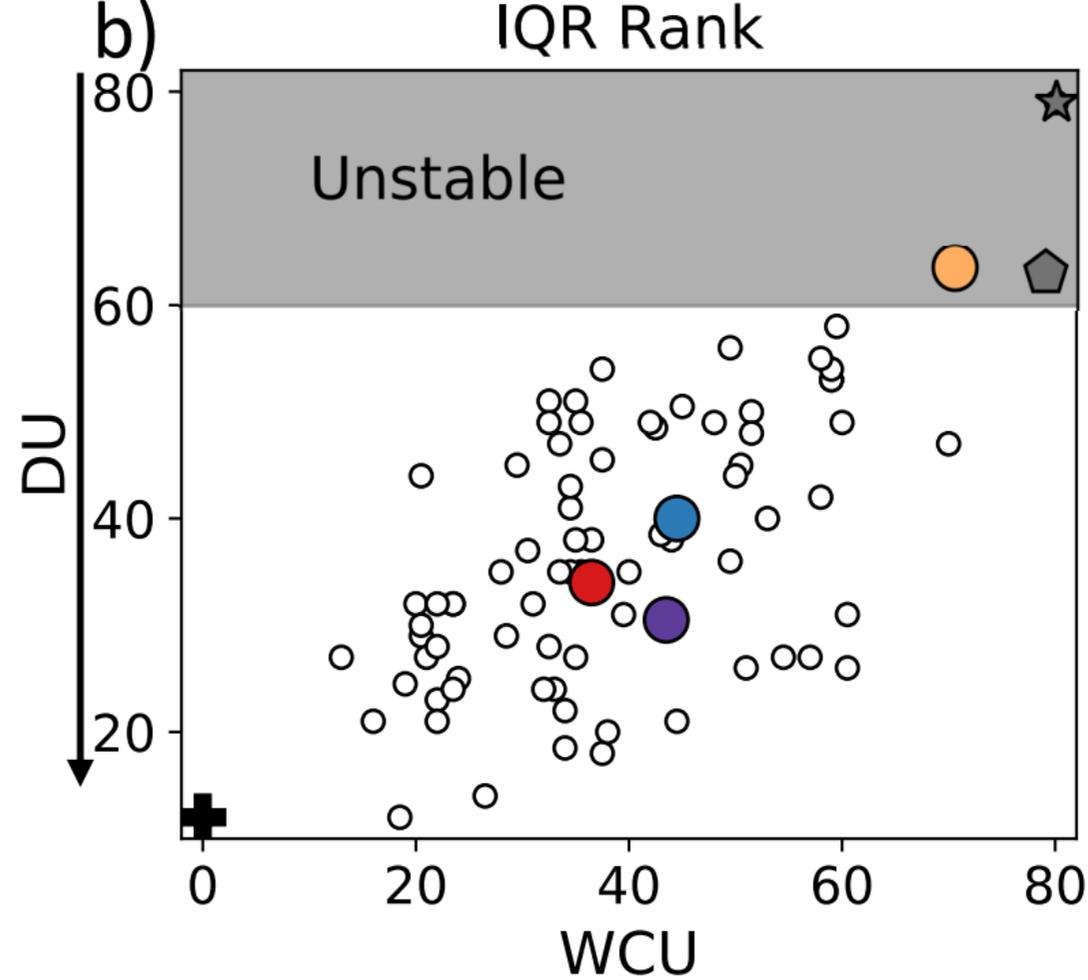
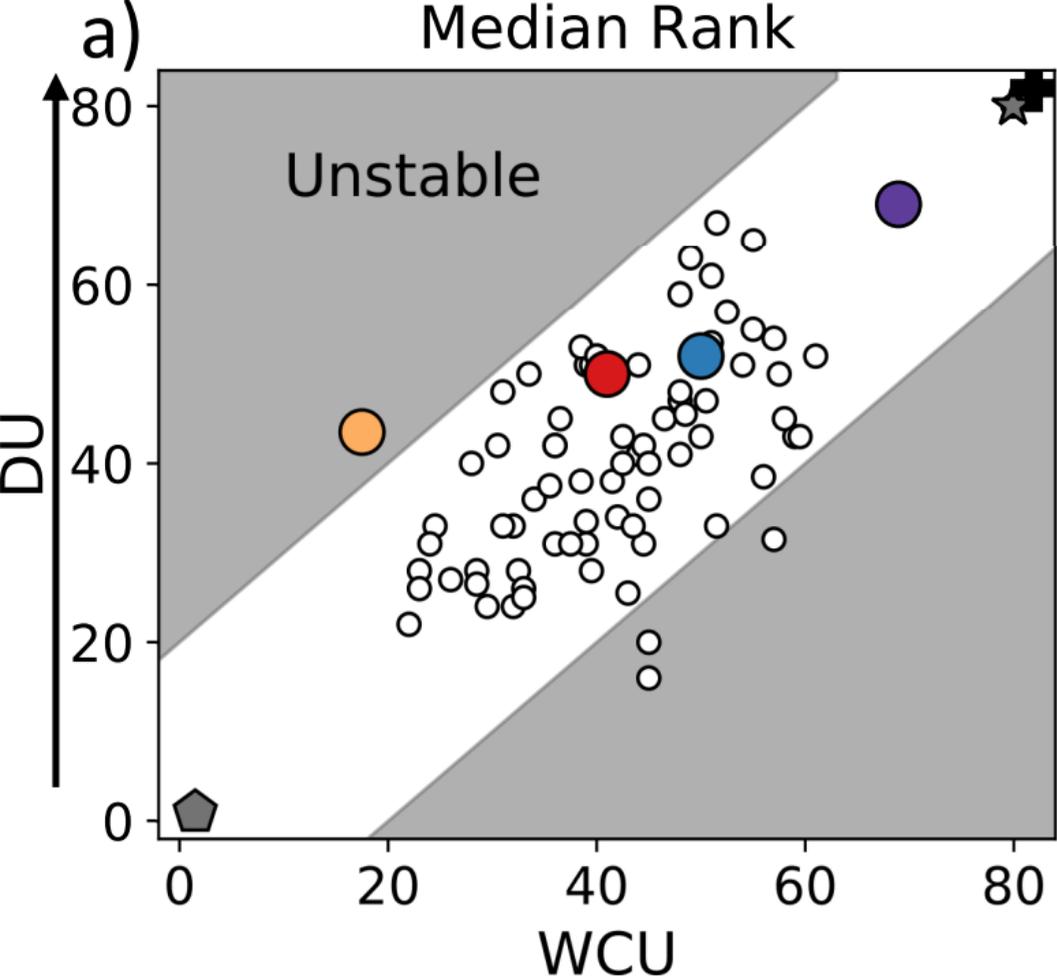


Figure 7.

Low Level of risk aversion High

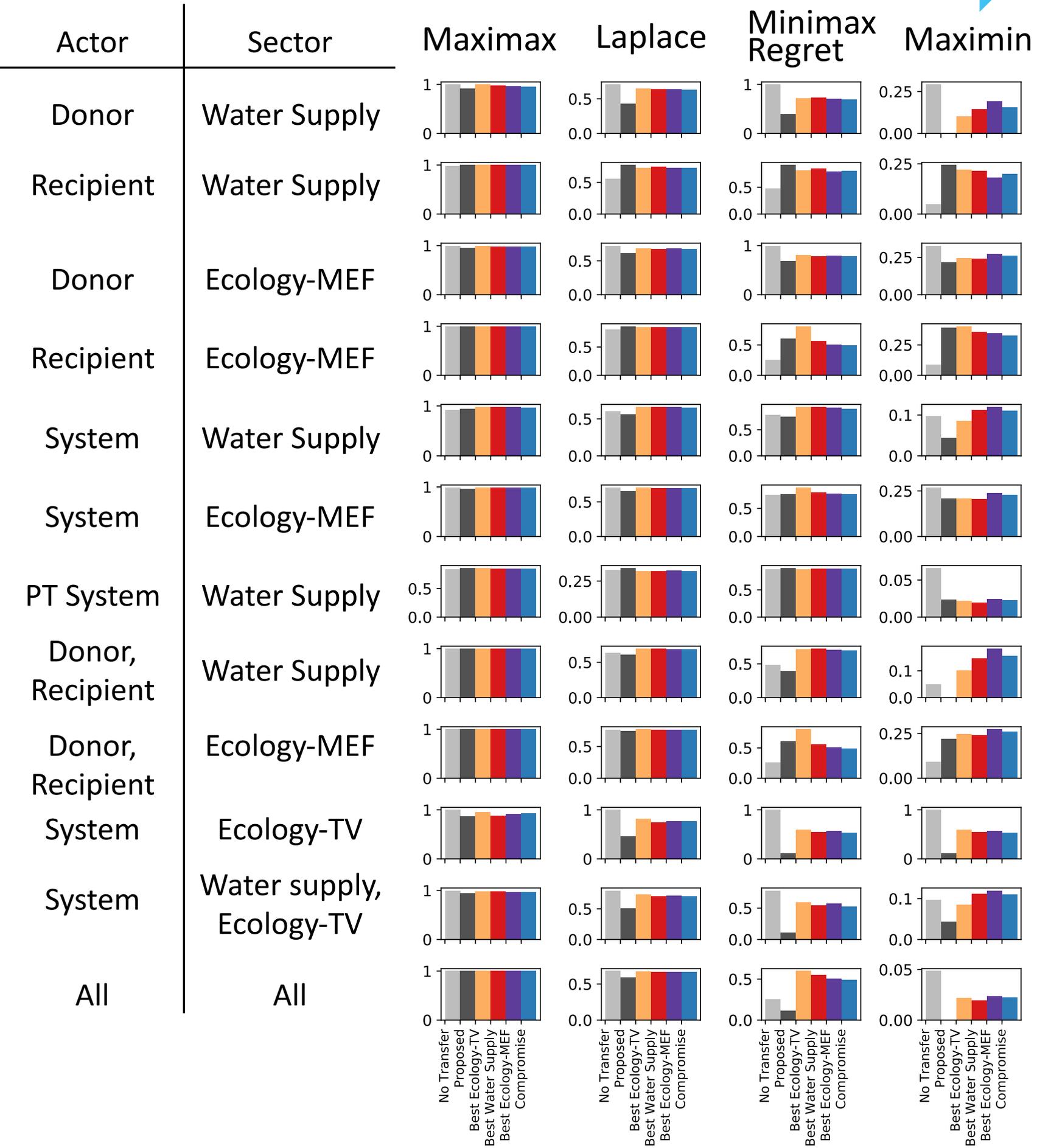
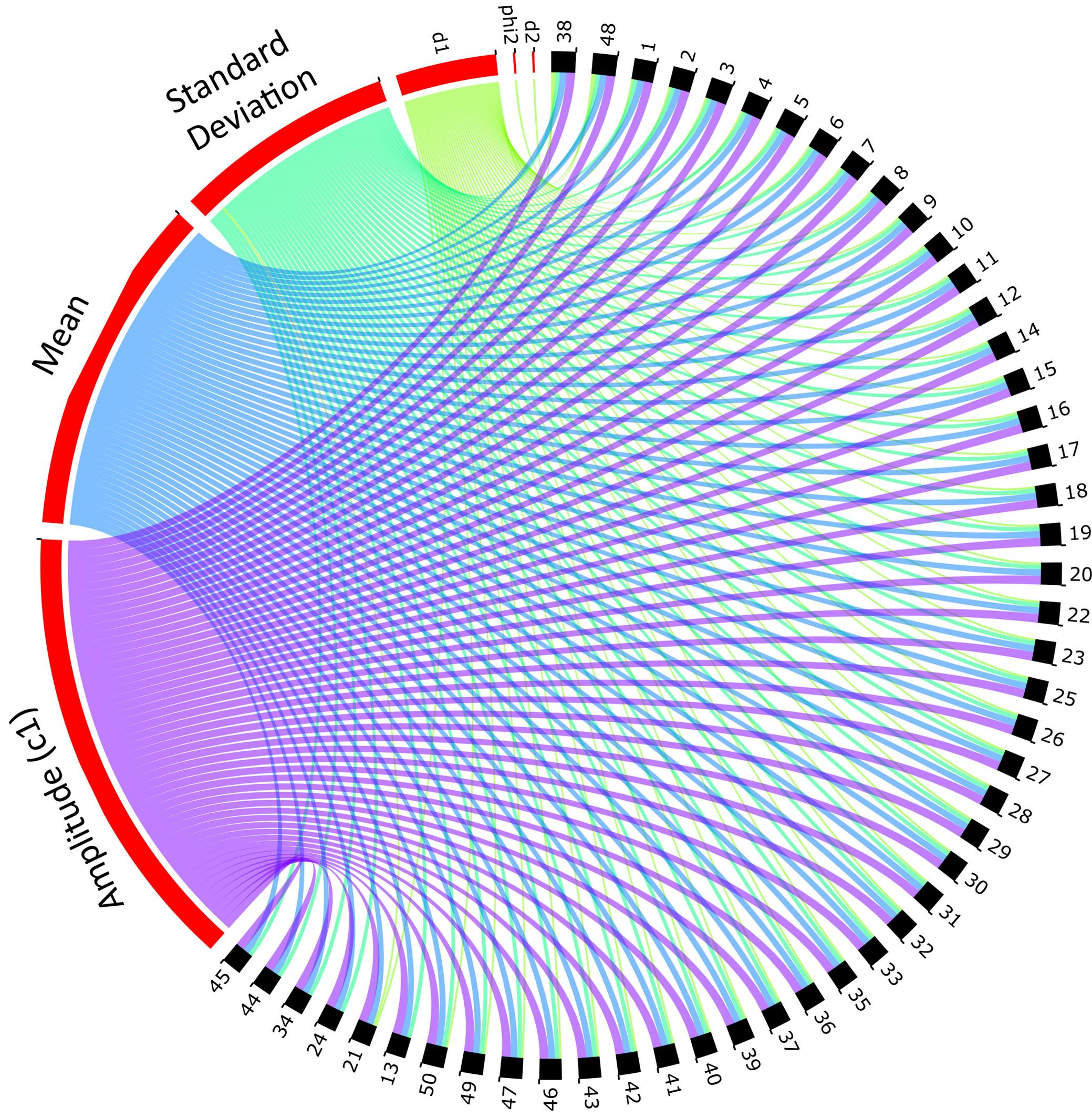


Figure 8.



- d1 Donor demand
- d2 Recipient demand
- phi2 Semi-annual monsoonal shift
- █ Strategy number
- █ Control factor

