The Global Distribution and Drivers of Grazing Dynamics Estimated from Inverse Modelling

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Abstract

We examine how zooplankton influence phytoplankton bloom phenology from the top-down, then use inverse modelling to infer the distribution and drivers of mean community zooplankton grazing dynamics based on the skill with which different simulated grazing formulations are able to recreate the observed seasonal cycle in phytoplankton biomass. We find that oligotrophic (eutrophic) biomes require more (less) efficient grazing dynamics, characteristic of micro- (meso-) zooplankton, leading to a strong relationship between the observed mean annual phytoplankton concentration in a region and the optimal grazing parameterization required to simulate it's observed phenology. Across the globe, we found that a type III functional response consistently exhibits more skill than a type II response, suggesting the mean dynamics of a coarse model grid-cell should offer stability and prey refuge at low biomass concentrations. These new observationally-based global distributions will be invaluable to help constrain, validate and develop next generation of biogeochemical models.

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Key Points:

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13	•	Oligotrophic (eutrophic) biomes exhibit more (less) efficient mean community graz-
14		ing dynamics, characteristic of micro- (meso-) zooplankton.
15	•	There is a strong relationship between mean phytoplankton biomass and the graz-
16		ing dynamics required to simulate its observed seasonal cycle.
17	•	A type III response does a consistently better job of recreating observed phyto-

plankton phenology compared to a type II response.

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19 Abstract

We examine how zooplankton influence phytoplankton bloom phenology from the top-20 down, then use inverse modelling to infer the distribution and drivers of mean commu-21 nity zooplankton grazing dynamics based on the skill with which different simulated graz-22 ing formulations are able to recreate the observed seasonal cycle in phytoplankton biomass. 23 We find that oligotrophic (eutrophic) biomes require more (less) efficient grazing dynam-24 ics, characteristic of micro- (meso-) zooplankton, leading to a strong relationship between 25 the observed mean annual phytoplankton concentration in a region and the optimal graz-26 ing parameterization required to simulate it's observed phenology. Across the globe, we 27 found that a type III functional response consistently exhibits more skill than a type II 28 response, suggesting the mean dynamics of a coarse model grid-cell should offer stabil-29 ity and prey refuge at low biomass concentrations. These new observationally-based global 30 distributions will be invaluable to help constrain, validate and develop next generation 31 of biogeochemical models. 32

³³ Plain Language Summary

To improve our predictions of the ocean's ability to feed a growing human popu-34 lation and buffer a changing climate we must improve our understanding of what hap-35 pens to carbon once it is absorbed into the ocean. One of the largest gaps in marine car-36 bon cycling is the role of zooplankton grazing. The rate at which zooplankton graze (or 37 consume) phytoplankton modifies the size and seasonal evolution of both populations 38 and in turn their associated rates of net primary production (the base of the food chain), 39 secondary production (an indicator of fisheries catch) and export production (the bio-40 logical sequestration of carbon). However, regional differences in in-situ grazing dynam-41 ics, which cannot be directly measured outside of a laboratory, remain poorly constrained 42 by observations and thus difficult to model. Here, we run many model simulations, each 43 of which simulate grazing dynamics in a different way, then compare the results to in-44 fer which type of grazing dynamics are required to match observations. We find that there 45 is dramatic spatial variability in how zooplankton appear to be grazing and that this vari-46 ability maps well onto the observed phytoplankton concentration, suggesting that the 47 type of zooplankton present may be determined by the amount of prey available. 48

49 **1** Introduction

Marine net primary production (NPP) accounts for roughly half of global carbon 50 fixation (Falkowski et al., 2000) and supports the biological export of carbon (de la Rocha, 51 2006) and base of the marine food chain (Armengol et al., 2019). Although oceanogra-52 phers have historically focused on light (Sverdrup, 1953) and nutrients (Howarth, 1988), 53 increasing experimental (Lima-Mendez et al., 2015; Guidi et al., 2016), observational (Behrenfeld 54 et al., 2013) and modelling (Hashioka et al., 2013; Prowe et al., 2012; Laufkötter et al., 55 2015; Rohr, Richardson, Lenton, Chamberlain, & Shadwick, 2022) work has highlighted 56 zooplankton grazing as a critical control on NPP. However, zooplankton grazing dynam-57 ics are poorly constrained (Everett et al., 2017), difficult to model (Petrik et al., 2022), 58 and sensitive to environmental change (Richardson, 2008). This uncertainty can lead to 59 large biases in export and secondary production in global marine biogeochemical (BGC) 60 models (Rohr, Richardson, Lenton, Chamberlain, & Shadwick, 2022). 61

Empirical laboratory experiments have shown that grazing dynamics (i.e. the manner in which specific grazing rates increase with prey concentration) vary substantially across zooplankton species, age, and size (Hansen et al., 1997; Hirst & Bunker, 2003). However, these studies, which consider the idealized behavior of a single species in a wellmixed environment, are unlikely to be representative of the open ocean. To parameterize relatively coarse global models, oceanographers must understand the mean dynamics of many species, averaged across a patchy ocean, which may diverge dramatically from

the dynamics of individual zooplankton (Rohr, Richardson, Lenton, & Shadwick, 2022). 69 Unfortunately, unlike chlorophyll and phytoplankton biomass, which can be inferred from 70 ocean optical properties measured remotely (Sathyendranath et al., 2019; Westberry et 71 al., 2008), zooplankton biomass cannot be measured from satellites. Instead, zooplank-72 ton must be measured at sea using a variety of different methods (Pakhomov et al., 2020; 73 Pinkerton et al., 2020; Benfield et al., 1998), each with large uncertainties and disparate 74 units (Moriarty et al., 2013), making it difficult to describe the global distribution with-75 out large levels of statistical inference (Everett, n.d.; Heneghan et al., 2020). 76

Without robust global data sets of zooplankton biomass and growth rates, it is difficult, if not impossible, for modellers to prescribe the correct grazing dynamics. This
likely contributes to persistent uncertainty in futrue projections NPP (Tagliabue et al., 2021), export production (Fu et al., 2016) and zooplankton biomass (Petrik et al., 2022).
Given the increasingly under-constrained nature of heavily parameterized BGC models
(Schartau et al., 2017), it is imperative to build out an observationally-informed under-standing of the distribution and drivers of grazing dynamics to help constrain them.

Here, were use inverse modelling to reach an initial estimate of the global distri-84 bution and drivers of grazing dynamics. In the absence of robust observations of zoo-85 plankton biomass, we rely on the well-established influence of grazing dynamics on phy-86 toplankton population dynamics (Gentleman & Neuheimer, 2008; Truscott et al., 1994; 87 Steele, 1974) to asses model skill from the top-down. We run a suite of simulations in 88 a global, coupled ocean-BGC model, parameterized with a wide range of grazing param-89 eters. We determine the optimal parameters required to best match the phenology of the 90 observed phytoplankton seasonal cycle to infer the global distribution of grazing dynam-91 ics (Sec. 3.1). We then show how this distribution appears to be driven by regional vari-92 ability in phytoplankton biomass (Sec. 3.2) and explain the underlying mechanisms (Sec. 93 **3.3**). Finally, we address the limitations of these estimates and means to improve them 94 (Sec. 4.1), before discussing how they could improve model validation and design (Sec. 95 **4.2**). 96

⁹⁷ 2 Materials and Methods

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2.1 The Grazing Formulation in BGC Models

In the simplest BGC models zooplankton grazing occurs between one group of zooqq plankton on one group of phytoplankton and can be described with a single-prey func-100 tional response curve (Gentleman & Neuheimer, 2008). Qualitatively, BGC models are 101 largely split in the type of functional response curve (type II vs III) they use (Rohr, Richard-102 son, Lenton, & Shadwick, 2022). The primary difference between them is that the type 103 II response increases linearly at low prey concentrations, while the type III increases ex-104 ponentially (**Figure 1**). Either way, this curve, q([P]), can be parameterized with a sat-105 uration grazing rate $(g_{max}; 1/d)$, which describes the rate when prey is not limiting, and 106 half saturation concentration $(K_{1/2}; mmolC/m^3)$, which describes how much prey is re-107 quired to get there (i.e. $g([K_{1/2}]) = 0.5 * g_{max}$). Here we focus on $K_{1/2}$ because it has 108 been shown to have a stronger influence on population dynamics than g_{max} (Rohr, Richard-109 son, Lenton, & Shadwick, 2022). 110

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2.2 Influence of Grazing on Phytoplankton Phenology

Grazing dynamics can influence the shape of seasonal phytoplankton biomass accumulation via the curvature of the functional response, which has either a stabilizing or destabilising influence on phytoplankton population dynamics depending on its concavity (Steele, 1974; Truscott et al., 1994; Gentleman & Neuheimer, 2008). If the functional response is concave upward, then phytoplankton specific loss rates to grazing increase with the size of the population. This creates a negative feedback loop which dampens changes in the size of the phytoplankton population, thereby exerting a stabilizing
 influence on phenology. Alternatively, downward concavity means phytoplankton spe cific loss rates to grazing decline with population growth, creating a destabilizing, pos itive feedback which amplifies changes in the size of the phytoplankton population.

The shape of the functional response, and thus its stabilizing influence on phenol-122 ogy, is determined by it's response type (II or III) and parameters (particularly $K_{1/2}$). 123 While the parameter values determine the magnitude of curvature and thus the strength 124 of the stabilizing influence, the response type determines the direction. A type II response 125 is always concave downward and thus always destabilizing. However, a type III response 126 has upward concavity below $K_{1/2}$ and thus stabilizing properties a low phytoplankton 127 concentrations. In turn, the grazing formulation has been shown to exert a strong in-128 fluence on population stability, and thus the size, likelihood, and possibility of phyto-129 plankton blooms, sub-seasonal oscillations, and extinction events (Dunn & Hovel, 2020; 130 Steele, 1974; Adjou et al., 2012; Hernández-García & López, 2004; Truscott et al., 1994; 131 Malchow et al., 2005). 132

2.3 Model Set up

Ecologically, the true shape of the mean functional response curve in a given swath 134 of the ocean is determined by both the physiological characteristics of individual zoo-135 plankton as well as their relative distribution (Rohr, Richardson, Lenton, & Shadwick, 136 2022). This is difficult to measure in-situ, but can be inferred through inverse modelling 137 via its the top-down influence of grazing dynamics on the simulated phytoplankton sea-138 sonal cycle. Here we use a global, coupled ocean-BGC to determine which $K_{1/2}$ values 139 and response types are required to best match the observed phytoplankton seasonal cy-140 cle. 141

The BGC model used, the Whole Ocean Model of Biogeochemistry and Trophic-142 dynamics (WOMBAT) (Law et al., 2017), is part of the Australian Earth Systems Model 143 (ACCESS-ESM1.5) (Ziehn et al., 2020) and has been used extensively in previous stud-144 ies (Mortenson et al., 2021; Kwiatkowski et al., 2020; Ziehn et al., 2017; Oke et al., 2013). 145 The ocean model is the global configuration of Modular Ocean Model version 5 (Griffies, 146 2012). WOMBAT has a relatively simple BGC structure ((Rohr, Richardson, Lenton, 147 Chamberlain, & Shadwick, 2022)) with 1 phytoplankton and 1 zooplankton group. More 148 complex models include multiple zooplankton grazing on multiple prey types, but we are 149 interested in basin-scale variability in the mean dynamics of the entire plankton com-150 munity. These can be inferred by tuning single-prey grazing response toward the observed 151 community-averaged phytoplankton phenology as observed from space. 152

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2.4 Model Experiments

We ran a total of 36 global simulations, each with a different grazing formulation. 154 To isolate the influence of grazing, each run was initialized from the same state, embed-155 ded in an identical repeat-climatological physical ocean and forced with identical sur-156 face flux and freshwater runoff from the Japanese 55-year atmospheric reanalysis sur-157 face dataset, JRA55-do (Tsujino et al., 2020). After initialisation, each run was spun up 158 for 5 years to a quasi-steady state, long enough to equilibrate with changes to its graz-159 ing formulation. Model output is reported from the fifth year of the simulation and can 160 be considered climatological. 161

We ran two suites of experiments, using a type III and II functional response. Within each suite we tested 18 different parameters combination: $K_{1/2} = 0.5$, 1, 2, 4, 8, 16 $(mmolC/m^3)$ and $g_{max} = 0.5$, 1, 2 (1/d). These values are consistent with the range that has been derived empirically and used in models historically (Rohr, Richardson, Lenton, ¹⁶⁶ & Shadwick, 2022). All other parameters were kept constant and are identical to those ¹⁶⁷ in Law et al. (2017).

2.5 Model Skill Assessment

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We used two metrics to evaluate the model's ability to recreate the observed phy-169 toplankton seasonal cycle, the correlation coefficient (CC) and coefficient of variation 170 (CV). The CC measures the co-variability between the simulated and observed clima-171 tologies, while the CV measures the magnitude of variability relative to the mean (i.e. 172 standard deviation divided by the mean). Together they capture the shape (CC) and 173 strength (CV) of the seasonal cycle. Importantly, both metrics are normalized by (or 174 in the case of CC agnostic to) the mean annual phytoplankton population size to con-175 trol for the influence of grazing rates on the mean state and isolate its influence on phe-176 nology. 177

For each metric, the seasonal cycle of simulated surface phytoplankton biomass was 178 compared to an 18-year remote sensing climatology (July 2002 - April 2021) from the 179 Carbon-based Productive Model (CbPM) (Westberry et al., 2008). The remote sensing 180 record was interpolated onto the model grid in space (2160x4320 to 300x360) and time 181 (8-day to 5-day resolution) and all time series were centered on the summer solutioe. We 182 used observed carbon instead chlorophyll because WOMBAT does not explicitly resolve 183 chlorophyll. However we repeated the entire analysis using the VIIRS chlorophyll record 184 (comparing model carbon to observed chlorophyll) and found qualitatively similar re-185 sults (Supplemental Section 1). 186

The cost function for model skill was quantified for each run in each grid-cell by subtracting the absolute difference between the modelled (CV_{mod}) and observed (CV_{obs}) coefficient of variation from the correlation coefficient $(CC_{mod,obs})$,

Model Skill =
$$norm(CC_{mod,obs}) - norm(|CV_{mod} - CV_{obs}|)$$
 (1)

Note, both metrics are normalized across all model output such that their contribution
 to the cost function is equally weighted and cost function scores can provide a direct com parisons between the skill of runs between the type II and III experiment suites.

For each response type, we consider three sets of 6 runs, with each run using a different $K_{1/2}$ values (0.5,1,2,4,8,16), but a constant g_{max} value across the set. Within each set the cost function score is interpolated between $K_{1/2}$ values using a piece-wise cubic polynomial and the $K_{1/2}$ value with the maximum score is identified (see **Figure 3**). This value is then averaged across all three sets, each of which use a different g_{max} value (0.5,1,2). Finally, very high latitude regions below -55S or above 55N with limited remote sensing coverage were excluded.

200 3 Results

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3.1 Global distribution and drivers of grazing dynamics

The distribution of observed mean annual surface phytoplankton biomass estimated 202 observationally from CbPM (Figure 1A) has a striking co-variability with the distri-203 bution of grazing dynamics inferred by the optimal $K_{1/2}$ value required to match the ob-204 served seasonal cycle (Figure 1B, \mathbf{C}). We find that more oligotrophic regions with low 205 mean annual phytoplankton biomass require smaller $K_{1/2}$ values to best match the ob-206 served phenology (Figure 1A-C; bluer), while more eutrophic regions with high mean 207 annual phytoplankton biomass require larger $K_{1/2}$ values (Figure 1A-C; greener). Qual-208 itatively, this pattern generally holds regardless of whether a type II (Figure 1B) or III 209 (Figure 1C) functional response is used to described grazing dynamics or whether re-210

mote sensing biomass (Figure 1A) or chlorophyll (Figure S1) is used to described the observed phenology.

This is remarkably consistent with the community composition which we would ex-213 pect to inhabit those respective biomes (Everett, n.d.; Heneghan et al., 2020). Ecolog-214 ically, the value of $K_{1/2}$ at a fixed g_{max} is related to the rate at which zooplankton can 215 capture (rather than consume) prey. Physiologically, the zooplankton which have high 216 prey capture rates are typically rapidly-grazing microzooplankton and filter feeders. How-217 ever, these zooplankton species are generally unable to consume anything larger than 218 219 small flagellates, ciliates and cyanobacteria, exactly the sort of phytoplankton that tend to dominate less productive, nutrient-poor, regions such as the gyres. On the other hand 220 slowly-grazing euphausiids, copepods and macrozooplankton tend to have much slower 221 capture rates but are capable of consuming much larger prey, such as dinoflagellates and 222 diatom assemblages, which tend to dominate more productive, nutrient-rich, regions like 223 the coasts and higher latitudes. Our results not only agree that there is substantial di-224 versity in the zooplankton community across the globe but suggest that these distinc-225 tions are essential in mechanistically shaping phytoplankton phenology (see Sec. 3.2). 226

Plotting the distribution of mean annual observed surface biomass against the op-227 timal $K_{1/2}$ values required to match its seasonal cycle further clarifies the relationship 228 between bulk phytoplankton biomass and the mean dynamics with which it appears to 229 be grazed (Figure 1D). Regardless of functional response type (II - circles; III - diamonds)) 230 or g_{max} value (thin lines), larger $K_{1/2}$ values are required to recreate the seasonal phe-231 nology of biomes with more mean annual phytoplankton biomass. Switching from a type 232 III to type II response or increasing g_{max} both increase mean grazing rates and the y-233 intercept of the regression but have relatively little influence on its slope (Table S1; Fig-234 **ure 1D**). A similar response is seen when using chlorophyll as the predicting variable 235 (Figure S1D). Plotting the optimal $K_{1/2}$ value averaged across all three sets of g_{max} 236 values yields our best guess at the relationship between $K_{1/2}$ and mean phytoplankton 237 biomass ([Phyto]). 238

Type II :
$$K_{1/2} = 15.8[Phyto] - 7.3$$

Type III : $K_{1/2} = 15.2[Phyto] - 10.2$ (2)

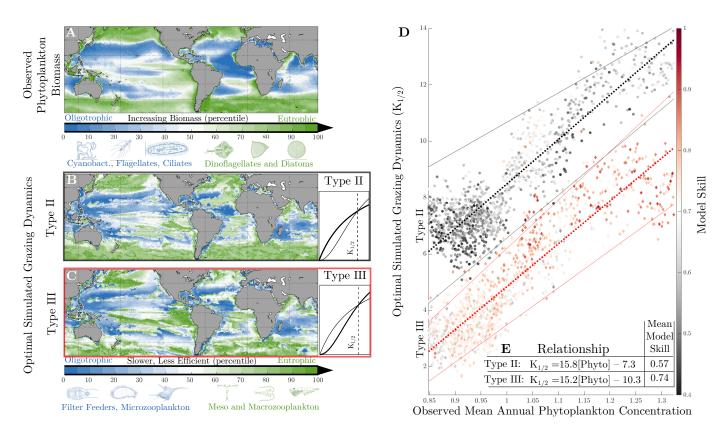
The slope of this relationship was not particularly sensitive to the selection of g_{max} 239 when a type III response was used (Supplemental Section 2. Note that the top and 240 bottom 20% of biomass grid cells are excluded because the relatively coarse ocean model 241 is less suited to resolve end-member biogeochemical scenarios. However, extending the 242 range to biomes considered suggests that the the optimal $K_{1/2}$ value begins to asymp-243 totically approach a minima and maxima in low and high biomass biomes, respectively. 244 This could imply two ecologically distinct biomes rather than a continuum, suggesting 245 a piece wise fit may be more appropriate (Supplemental Section 3). 246

Finally, regardless of g_{max} value or biome, using a type III function response does 247 a consistently better job of recreating seasonal phenology compared to a type II response (Figure 248 **1D**), scoring 30% better on average (**Table S1**). This is consistent with in-situ (Morozov 249 et al., 2008) and mathematical (Rohr, Richardson, Lenton, & Shadwick, 2022; Morozov, 250 2010) observations that the downward concavity, prey refuge, and stabilizing properties 251 associated with a type III response may be a better empirical representation of the mean 252 state of a patchy ocean, even if individual zooplankton graze with a type II response (Hansen 253 et al., 1997; Hirst & Bunker, 2003). 254

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3.2 Influence of the grazing formulation on regional bloom phenology

²⁵⁶ When $K_{1/2}$ is large, phytoplankton phenology tends to exhibit a stronger, well-defined ²⁵⁷ seasonal cycle with less high frequency variability (**Figure 2**; green lines). This is be-



The distribution and drivers of grazing dynamic. A) The observed climatological Figure 1. mean annual phytoplankton biomass concentration is plotted as a percentile of the full spatial distribution. Below, the corresponding $K_{1/2}$ parameter required to best recreate the observed phytoplankton seasonal cycle using a B) Type II or B) III response function is also plotted as a percentile for direct qualitative comparison. Beside each is an example functional response curve for their respective (bolded) response types, both parameterized with the same $K_{1/2}$ and g_{max} values. Below A) and C) is a schematic of the characteristic phytoplankton associated with low and high biomass waters and the characteristic zooplankton associated with low and high $K_{1/2}$ values. C) The optimal $K_{1/2}$ required with a type II (circles; B)) and type III (diamonds; C)) response are plotted against the corresponding phytoplankton biomass. Points are colored by their mean cost function score, with redder colors indicating increasing model skill. Each point represents then mean of roughly 30 grid cells, binned based on their percentile biomass, with the lowest and highest 20th percentiles excluded. All points and the thick dashed regression were computed by averaging the optimal $K_{1/2}$ value across three experiment suites, each using a different g_{max} values. Uncertainty bounds (thin solid lines) are provided by calculating the relationship using just lowest (0.5) and highest (2) g_{max} values. E) The linear regression (dashed lines) and mean model skill is quantified in the inset table and included for all g_{max} values in Table S1.

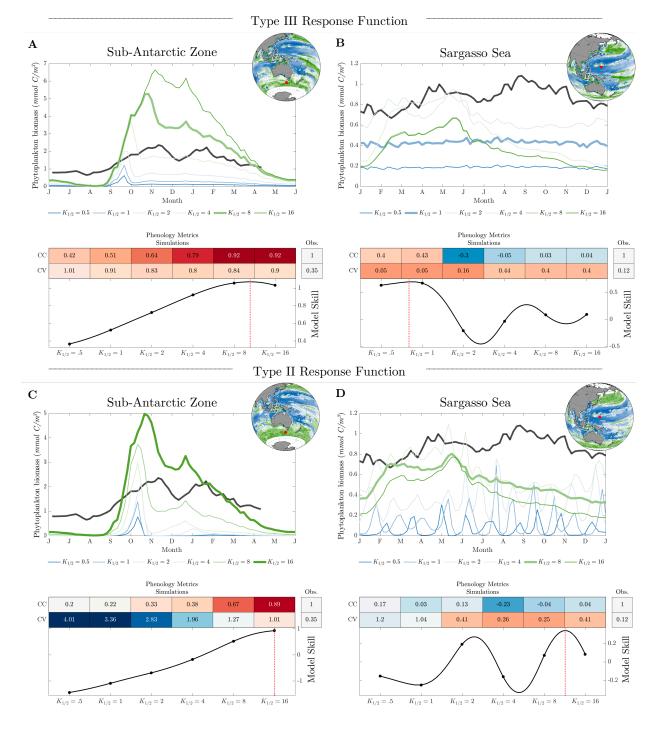


Figure 2. Influence of $K_{1/2}$ and response type on regional phenology. Example seasonal cycles and their model skill are provided from the **A**, **B** Subantarctic zone south of Australia and **B**, **D** Sargasso sea. The emergent phenology for all six $K_{1/2}$ values is include for both a **A**, **B** Type III and **C**, **D** type II functional response. All simulations use the same identical g_{max} value of 1/d. **A**) In the upper panel of each subplot the observed (black) and simulated (blue-green) seasonal phytoplankton biomass cycles are plotted with the simulation which best matches the observed phenology boldened. In the lower panel the cost function scoring is demonstrated for each simulation, with the corresponding CC and CV shown above the combine cost function score. Red (blue) color shading indicates better (worse) model skill. The optimal $K_{1/2}$ value is determined by the interpolated maximum of the cost function scores (red line). In the inset map (top right corner) the distributions are qualitatively identical to Figure 1, with the same percentile-based colorbar, and differs only in that they show results from runs with $g_{max} = 1$ to be consistent with the traces, rather than the mean of all g_{max} values. **B-D** Identical to **A**) but a different location or response function.

cause the grazing formulation does not heavily influence the stability of the system (Gentleman 258 & Neuheimer, 2008; Rohr, Richardson, Lenton, & Shadwick, 2022), allowing bloom phe-259 nology to be driven primarily by bottom-up controls, such as light and nutrient avail-260 ability, which generally follow a stronger seasonal cycle linked to vertical mixing cycle 261 and the length of the day. In turn, regional phenology is not as sensitive to the whether 262 a type II (Figure 2A, C) or III (Figure 2B, D) response is used. However, as $K_{1/2}$ 263 decreases, the grazing formulation has a stronger influence on the stability on the sys-264 tem. This influence is stabilizing if a type III response is used but destabilizing if a type 265 II response is used (see **Supplemental Section 4**), resulting in substantively different 266 phenologies depending on the response type used (**Figure 2**; blue tracers). To demon-267 strate this mechanistic influence of the grazing formulation we consider two cases stud-268 ies in the sub-Antarctic zone and Sargasso sea, which are generally representative of the 269 seasonal variability in high and low biomass biomes, respectively (see **Figure S4**). 270

In the sub-Antarctic zone south of Australia (Figure 2A, C) the observed evo-271 lution of biomass (black line) exhibits a strong seasonal cycle with an amplitude $\sim 20\%$ 272 greater than its mean and relatively little sub-seasonal variability. It is best recreated 273 using larger $K_{1/2}$ values and exhibits slightly more model skill when a type III response 274 is used. When using a type III response (Figure 2A), lower $K_{1/2}$ values reduce the mean 275 biomass but do not systematically modify the CV, leaving the ratio of summer to win-276 ter biomass roughly proportional. This occurs because the lower $K_{1/2}$ values systemat-277 ically increase mean grazing rates, driving biomass down, but increase the first order sta-278 bility (Figure S4A), largely preserving the shape of the seasonal cycle. Alternatively, 279 when using a type II response (**Figure 2C**) response, decreasing $K_{1/2}$ delays bloom ini-280 tiation but amplifies its acceleration once initiated, leading to smaller, shorter, sharper 281 features and systematically higher CV_s . The delayed initiation relative to the type III 282 response occurs because type II response disproportionately increases grazing rates a low 283 biomass concentrations compared to a type III response. The shorter, sharper bloom du-284 ration occurs because lowering $K_{1/2}$ in a type II response destabilizes the system (Figure 285 $\mathbf{S4B}$), allowing phytoplankton biomass to accumulate exponentially until other stabi-286 lizing factors kick in form the bottom-up (i.e. nutrients, self-shading) and rapidly ter-287 minate the bloom. 288

In the Sargasso Sea (Figure 2B, D) the observed evolution of biomass (black line) 289 exhibits a much weaker seasonal cycle, with more high frequency variability and an am-290 plitude less than half the size of its mean. It is best recreated using smaller $K_{1/2}$ val-291 ues and exhibits much more skill when a type III response is used. When using a type 292 III response (Figure 2A), highly stable, lower $K_{1/2}$ damp out any seasonality in bottom-293 up controls and prevent biomass from accumulating. Increasing $K_{1/2}$ systematically in-294 creasing the CV by allowing a small seasonal cycle to emerge which is not well correlated 295 with the observed phenology. On the other hand, when using a type II response, while 296 the phenology is nearly identical at high $K_{1/2}$ values, it diverges substantially at low ones. 297 Here, decreasing stability induces unstable predator-prey dynamics which drives higher 298 frequency, sub-seasonal oscillations, leading to higher CVs and poor correlations with 299 the observed phenology. 300

While the high latitudes and gyres clearly require different $K_{1/2}$ both perform better when a type III response is used. In the sub-Antarctic, a type II grazing leads to blooms that are offset (worse CC) and much sharper (higher CV) than what is observed while in the Sargasso type II grazing can lead to unnatural, sub-seasonal oscillation that are not observed in the remote sensing record. These results further support the use of a type III to represent the complex changes in plankton community composition and its relative distribution, which must be averaged over a coarse model grid cell.

308 4 Discussion

Of the dominant controls on NPP and subsequent carbon cycling, we have a fairly advanced understanding of the global distribution and drivers of light, temperature, nutrients and deep vertical mixing. However, the distribution and drivers of loss rates, particularly to grazing, remain largely a blind spot. We have made initial estimates using inverse modelling and discuss their limitations and future directions (4.1) in addition to their potential utility to improve BGC and climate models (4.2).

4.1 Limitations and future directions

The largest limitation likely stems from the fidelity of the model we have sought 316 to optimize. We have worked to control model bias by repeating experiments in an iden-317 tical physical ocean. However, if their is a systematic bias somewhere in the simulated 318 seasonal light and nutrient supply, then it is possible that the 'wrong' grazing dynam-319 ics could combine with the 'wrong' bottom-up controls to produce the correct phenol-320 ogy, leading us to incorrectly infer that those grazing dynamics were optimal. This could 321 be the case along the equator where there is a disproportionately large model bias in model 322 phytoplankton biomass relative to that of NPP (Figure S5), suggesting phytoplank-323 ton specific growth rates are too low due to some systematic bias in the bottom-up con-324 trols. This may explain why we found slower grazing dynamics (Figure 1B; greener) 325 in the equatorial Indian, Atlantic, and Pacific basins than we would have expected from 326 the low mean annual phytoplankton biomass (Figure 1A; bluer) observed there. If the 327 model is misrepresenting bottom-up controls as too weak, it make sense that slower than 328 expected grazing is needed to permit biomass accumulation and match the observed phe-329 nology. Additional biases may stem from the remote sensing products, which interpo-330 late over cloud cover and cannot capture the complete seasonal cycle at high latitudes, 331 as well as the exact nature of the link between trophic controls and bloom phenology (Behrenfeld 332 et al., 2013; Rohr et al., 2017). Considering these potential biases, it is essential to fur-333 ther refine estimates of the distribution of grazing dynamics with more observationally-334 focused methods. 335

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4.2 Model Utility

Despite their limitations, our results demonstrate that grazing dynamics vary largely 337 in space and agree qualitatively with our best observational understanding of how zoo-338 plankton species are distributed across the ocean (Everett, n.d.; Heneghan et al., 2020). 339 Considering the sensitivity of carbon cycling to grazing (Rohr, Richardson, Lenton, Cham-340 berlain, & Shadwick, 2022), it is critical for models to build in the mechanisms to both 341 recreate this distribution and allow it to respond to changing physical and environmen-342 tal drivers. As warming, stratification, and faster wind transform the surface ocean, the 343 ensuing balance of light and nutrients will reshape marine ecosystems, likely favouring 344 different zooplankton species, in different places, with vastly different grazing dynam-345 ics. For instance, if increasing stratification dominates increasing wind stress leading to 346 shallower summer mixed layers in the Southern Ocean, then we may expect a shift to-347 ward smaller phytoplankton, which have higher light, but lower nutrient, requirements 348 (Pörtner et al., 2019). This shift would be followed by a shift in zooplankton which are 349 better adapted to graze on smaller phytoplankton, such microzooplankton and salpes. 350 Already a southward shift of salpes into regions previous dominated by Euphausiids has 351 been observed (Henschke & Pakhomov, 2019; Steinberg & Landry, 2017). As faster graz-352 ers push south the global distribution of net primary, secondary and export production 353 354 will be substantially altered (Rohr, Richardson, Lenton, Chamberlain, & Shadwick, 2022). This shift must be captured in BGC models if Earth system and ecosystem models hope 355 to reduce the uncertainty in projections of NPP (Tagliabue et al., 2021), export produc-356 tion (Fu et al., 2016), and fisheries catch (Tittensor et al., 2021) which will be required 357

to accurately predict changes in the oceans capacity to buffer a changing climate and feed a growing population.

4.2.1 Model Validation

Given the large uncertainty in future projections of zooplankton biomass (Petrik 361 et al., 2022) it is clear that the environmental controls on mean grazing dynamics are 362 not well constrained. There is little convergence in how state-of-the-art BGC represent 363 zooplankton. Although some recent BGC models include 10+ plankton groups (Negrete-364 García et al., 2022; Sommer et al., 2022; Cael et al., 2021), most CMIP6-class models 365 used in 100+ year ensemble climate projections only include 1-3 zooplankton and phy-366 toplankton groups (Kearney et al., 2021; Rohr, Richardson, Lenton, Chamberlain, & Shad-367 wick, 2022). It is thus critical to know if competition between these limited functional 368 groups is sufficient to drive an emergent distribution in mean community grazing dynam-369 ics that matches reality. Even with the limitations discussed above, we believe that di-370 agnostically computing the apparent mean grazing dynamcis (i.e. fitting a functional re-371 sponse curve to mean grazing rate vs. total prey biomass) and comparing these qual-372 itative distributions to Figure 1 will help assess model skill and constrain marine car-373 bon cycling in increasingly under-constrained BGC models (Schartau et al., 2017). If the 374 apparent grazing dynamics differ dramatically, or worse, are spatially homogeneous (Law 375 et al., 2017; Zahariev et al., 2008), then it should be clear that other spatially hetero-376 geneous bottom-up controls must be over-tuned to recreate observed heterogeneity in 377 NPP. 378

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4.2.2 Model Development

If explicit competition between limited functional groups is insufficient to resolve 380 the emergent distribution of community averaged grazing dynamics and a sufficiently com-381 plex food web is not compatible with the computational cost of high resolution, fully cou-382 pled projections (Neelin et al., 2010), then it may be possible to parameterize zooplank-383 ton community composition using the relationship inferred in **Figure 1C**. That is, mod-384 ellers could implicitly represent changes in zooplankton community composition by mod-385 ifying the $K_{1/2}$ value of a single group as a function of phytoplankton abundance using 386 equation 2 (or those in Supplemental Section 2). This would allow modelers to ex-387 plicitly modify the mean grazing dynamics, rather than explicitly resolving each of its constitute species, allowing the mean attributes of the zooplankton community to respond 389 dynamically to changing environmental conditions without increasing the number of state 390 variables the model must carry. 391

Underlying this parameterization are the assumptions that a) bulk phytoplankton 392 biomass follows community composition, with less (more) productive waters inhabited 393 by smaller (larger) phytoplankton (Roy et al., 2013), b) zooplankton community com-394 position is determined by the composition of the prey field (Kiørboe & Hirst, 2014) c) 395 different species of zooplankton graze with systematically different dynamics (Hansen 396 et al., 1997). All three assumptions are generally well supported individually by obser-397 vations and are together consistent with the emergent relationship between observed phy-398 toplankton biomass and the inferred grazing dynamics required to recreated its phenol-300 ogy (**Figure 1C**). Although there are several key challenges to implementing this rela-400 tionship (see **Supplemental Section 5**), leveraging it to parameterize functional dif-401 ferences driven by competition could extend well beyond those associated with grazing 402 and allow modellers to vary a collection of biogeochemically important attributes asso-403 404 ciated with distinctive zooplankton communities at a low computation cost (see **Sup**plemental Section 5). 405

406 5 Conclusions

These results present a novel, observationally-informed, map of global mean com-407 munity grazing dynamics, which likely differ dramatically from the behaviour of indi-408 vidual zooplankton species long measured in laboratories. Further refining the observed 409 distribution and drivers of grazing, and how to replicate them in models, will require close 410 collaboration with zooplankton ecologists, but presents an exciting new frontier in chem-411 ical oceanography focused on a rigorous understanding of how NPP is controlled from 412 the top-down, a perspective often ignored. Moreover, improving the representation of 413 414 zooplankton, which have for too long been treated simply as closure term, could realize dramatic improvements in marine BGC models and our predictions of future ocean 415 states. 416

417 6 Open Research

All model output needed to reproduce the results and associated documentation can be found at the CSIRO Data Access Portal (https://doi.org/10.25919/wn09-6j31). The remote sensing products used in Figure 4 can be found at http://orca.science .oregonstate.edu/2160.by.4320.8day.hdf.carbon2.m.php. We'd like to thank readers of this document for their attention, and invite them to address any questions to Tyler Rohr, at tyler.rohr@csiro.au.

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Supporting Information for "The Global Distribution and Drivers of Grazing Dynamics Estimated from Inverse Modelling"

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Text S1. Distribution and drivers of grazing dynamics using the VIIRS chlorophyll record

In Figure 1 we used phytoplankton carbon biomass estimate remotely from the Carbonbased Productivity Model (Westberry et al., 2008) to compare directly to prognostic phytoplankton biomass resolved in the simulation. However, estimating carbon biomass from space using particle back-scattering involves a different set of assumptions than traditional estimate of phytoplankton abundance which infer chlorophyll concentrations from ocean color. To confirm these differences did not influence our results we repeated the analysis comparing the seasonal cycle of modelled phytoplankton carbon to that of remotely sensed chlorophyll from VIIRS (Sathyendranath et al., 2019). The results were largely consistent (Figure S1).

Note, model skill scores appear higher for VIIRS than CbPM (**Table S1**; last column); however, model skill was normalized across all runs using chlorophyll (VIIRS) and all run using carbon (CbPM) independently. Thus, the higher scores for VIIRS do not necessarily mean the modelled seasonal phytoplankton cycle better reflects observed chlorophyll compared to carbon, but rather that the difference between the model skill achieved with the optimal $K_{1/2}$ values compared to sub-optimal $K_{1/2}$ values is larger when comparing to observed chlorophyll.

Text S2. Sensitivity of the relationship between phytoplankton abundance and optimal $K_{1/2}$ value to the functional response type, g_{max} value, and indicator of phytoplankton abundance

The relationship between mean annual phytoplankton abundance and the $K_{1/2}$ parameter required to best recreate its phenology was computed using different functional response types (II, III), g_{max} values (0.5,1,2) and observed indicators for phytoplankton abundance variables (carbon, chlorophyll) and reported in **Table S1**. Across all type III configurations, the relationship was qualitatively consistent. Changing g_{max} in a type III response slightly modified the y-intercept (or mean optimal $K_{1/2}$) but did not substantially influence the slope of the relationship, regardless of if biomass of chlorophyll was being used as the metric for observed phytoplankton abundance. Although, when using a type II response was used, the slope of the relationship was much more sensitive to the value of g_{max} , indicative of its much stronger influence on grazing rates at low prey concentrations and thus population dynamics (Rohr et al., 2022). However, regardless of the g_{max} value used, the type II response consistently exhibited less skill than the type III response when averaged globally across all grid-cells, suggesting it is less suited to represent the mean dynamics in coarse models.

Text S3. Piece-wise linear model Instead of fitting a continuous linear regression to the relationship between phytoplankton abundance and $K_{1/2}$ (Figure 1D, S1D) we now consider a piece-wise linear fit. Here we expand the range of biomes included to the 5-90th percentiles, but still exclude very high and low biomass regimes where the model does not perform as well as in the open ocean. Looking at the expanded relationship (Figure S2), there are two clear asymptotes when using CbPM biomass as an indicator of phytoplankton abundance but only one when using VIIRS chlorophyll. It is not entirely clear why there is no low asymptote for chlorophyll but it may have to do with

the detection threshold for ocean colour versus backscatter or the fact that at low phytoplankton concentrations the particle back scatter signal may no longer be dominate by phytoplankton.

We then used Bayesian ensemble algorithm for change-point detection and time series decomposition (Zhao et al., 2019) to identify change-points where the relationship began to approach an asymptote. We then fit a piece-wise linear model to the data by assuming a slope of zero across each asymptote beyond the statistically identified change-point and forcing continuity between the asymptotes. These relationships are plotted in **Figure S2**. While they remain qualitatively consistent with the finding that higher $K_{1/2}$ are required drive the phenology of higher biomass biomes, they suggest the transition may occur rather rapidly between two states dominated by slow or rapidly grazing zooplankton.

Collectively, it appears clear that there is an upper and lower bound on realistic $K_{1/2}$ values: roughly 2-10 $mmol/m^3$ when using a type III response (as recommended) or roughly 5-15 when using a type II response (not recommended).

Text S4. First order stability of the functional response

The shape of the functional response curve for zooplankton grazing (g([P])) influences the shape of phytoplankton phenology primarily via its stabilizing or destabilizing influence on phytoplankton population dynamics (Gentleman & Neuheimer, 2008). The stabilizing influence of grazing is determined by how phytoplankton specific loss rates to grazing (i.e. clearance rates; Cl = g([P])/[P]) change in response to changing phytoplankton biomass. If phytoplankton accumulation decreases clearance rates, thereby promoting further population growth, that is a positive feed back with a destabilizing influence.

Alternatively, if phytoplankton accumulation increases clearance rates, thereby damping further population growth, that is a negative feed back with stabilizing influence. The stabilizing influence of the functional response at a given phytoplankton concentration can thereby be quantified is thereby determined by the sign of the first derivation of the clearance rate with respect to the phytoplankton concentration (i.e. $\frac{dCl}{d[P]}$). The value of $\frac{dCl}{d[P]}$ is determined both by the parameters (i.e. $K_{1/2}$) the prescribe the shape of the curve (g([P])) as well as the prognostic phytoplankton concentration which is variable in space in time (Rohr et al., 2022).

To capture a mean sense of the stabilizing influence of the functional response across a complete model run and many different phytoplankton concentrations, we define the first order stability as the value of $\frac{dCl}{d[P]}$ at the mean [P] of a given run. The annuallyaveraged first order stability of our experiments was consistently, necessarily, negative (destabilizing) when a type II response was employed (Figure S3B) and positive (stabilizing) when a type III response was employed (Figure S3A). In both cases, large $K_{1/2}$ values stretch out the response curve, leading to the depression and linearization of the functional response at low (but common) prey concentrations, slow and steady clearance rates, and very little influence on the stability of the system. Decreasing $K_{1/2}$ in the type II formulations monotonically decreases the first order stability by both directly altering the shape of the functional response curve and indirectly decreasing the prognostic phytoplankton population via increased grazing pressure. Decreasing $K_{1/2}$ in the type III formulations monotonically increases the first order stability of the system. This occurs because increasing Grazing Pressure keeps the annually-averaged phytoplankton

concentration below $K_{1/2}$, where the first order stability increases as $K_{1/2}$ decreases. Note, however, that model configurations with a very low g_{max} or very strong bottom-up growth conditions could buoy phytoplankton populations above $K_{1/2}$, such that decreasing $K_{1/2}$ decreases the first stability of the system, even with a type III response.

Text S5. Challenges and advantages of implementation in a BGC model

First and foremost, the best implementation of this parameterization (4.3.2) will require better constraining the relationship between phytoplankton biomass and $K_{1/2}$ in addition to the strength and co-variability of other drivers of zooplankton bio-geogrpahy such as temperature (Brandão et al., 2021) or the relative distribution of prey in models with multiple phytoplankton groups. Using the former may expedite the increase in grazing rates associated with traditional metabolic temperature limitation (Laufkötter et al., 2015) as warmer water also favors more efficient grazing zooplankton species as well (Richardson, 2008). Using the later would largely obviate the assumption that the relative distribution phytoplankton size co-varies with the bulk concentration.

The second major challenge will be determining the space and time scales over which to assume that specific grazing rates should change due to the influence of food scarcity on individual zooplankton versus the influence of zooplankton community composition on mean grazing dynamics. That is, while the value of $K_{1/2}$ determines the instantaneous response of zooplankton grazing rates to food scarcity, it should take longer for $K_{1/2}$ itself to evolve. This is because $K_{1/2}$ reflects the mean physiological characteristics of the entire zooplankton community and can only change at the rate with which community composition can evolve. This timescale likely varies globally and as a function of other

environmental drivers such as temperature (Richardson, 2008). For example, much shorter time periods are needed in communities dominated by asexually-reproducing zooplankton such as salps compared to those dominated by zooplankton with complex, multi-year, life histories, such as euphausiids (Steinberg et al., 2015).

However, getting this right could realize dramatic improvements in BGC models and our predictions of changes to marine carbon cycling. Extending from the assumption that a given optimal $K_{1/2}$ reflects the mean behavior of a particular zooplankton community, other attributes of that community could be additionally parameterized. For instance, crustaceans associated with slower grazing (and larger $K_{1/2}$ values) are typically stronger swimmers. They tend to vertically migrate on daily and seasonal timescales, allowing them to actively transport carbon much faster than microzooplantkon (Steinberg & Landry, 2017). This could be represented by increasing the flux of carbon from zooplankton into the sinking detritus pool (i.e. POC) at low $K_{1/2}$ values, without explicitly including the important role of Diel-vertical migration in carbon transport (see (Archibald et al., 2019)). Other important BGC attributes that vary with zooplankton community composition include the recalcitrance of their detritus and thus the remineralization rates of what they contribute to export production, their sensitivity to temperature, their stoichiometry and carbon content, and their response to seasonal change in the depth of the surface mixed layer.

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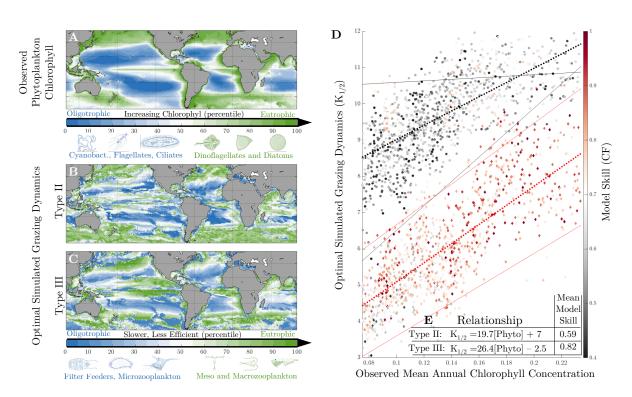


Figure S1. Identical to **Figure 1**, except using VIIRS chlorophyll instead of CbPM carbon biomass to track the observed phytoplankton phenology.

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Response Type	$\mathop{\mathrm{g}_{\mathrm{max}}}_{\mathrm{(d^{-1})}}$	Linear Regression (middle 60 percentile)	${f Correlation} {f Coefficient}$	Mean Model Skill
$K_{1/2} \text{ (mmol C/m^3)}$ as a function of Observed Phytoplankton Biomass (CbPM; Figure 1)				
Type II	0.5	$K_{1/2} = \ 15.3 [Phyto] - 8.7$	0.74	0.57
Type II	1	$K_{1/2} = 21.7 \ [Phyto] - 13.5$	0.81	0.61
Type II	2	$K_{1/2} = 10.4 [Phyto] + 0.2$	0.56	0.56
Type II	Mean	$K_{1/2} = 15.8 [Phyto] - 7.3$	0.83	0.57
Type III	0.5	$K_{1/2} = 13.2 [Phyto] - 9.7$	0.80	0.75
Type III	1	$K_{1/2} = \ 15.2 [Phyto] - 10.3$	0.74	0.74
Type III	2	$K_{1/2} = 17.2 [Phyto] - 10.9$	0.74	0.73
Type III	Mean	$K_{1/2} = \ 15.2 [Phyto] - 10.34$	0.85	0.74
$ m K_{1/2}(mmol/r)$	n ³) as a fun	ction of Observed Phytoplankton	Chlorophyl (VIIR	S; Figure S2)
Type II	0.5	${\rm K}_{1/2}=33.7[{\rm Chl}]+3.1$	0.65	0.65
Type II	1	${\rm K}_{1/2} = 23.1 [{\rm Chl}] + 7.6$	0.41	0.64
Type II	2	${\rm K}_{\rm 1/2}{\rm = \ 2.1[Chl] + 10.4}$	0.00	0.47
Type II	Mean	${\rm K}_{1/2} = 19.7 [{\rm Chl}] + 7.0$	0.50	0.59
Type III	0.5	${\rm K}_{\rm 1/2} = 22.7 [{\rm Chl}] + 1.3$	0.51	0.79
Type III	1	${\rm K}_{\rm 1/2} = 28.5 [{\rm Chl}] + 2.1$	0.56	0.81
Type III	2	${\rm K}_{\rm 1/2} = 28.5 [{\rm Chl}] + 3.9$	0.48	0.85
Type III	Mean	${\rm K}_{1/2}{=}26.4[{\rm Chl}]{+}2.5$	0.57	0.82

Table S1. The relationship between mean annual phytoplankton abundance and the $K_{1/2}$ parameter required to best recreate its phenology. Different relationships refer to different response functions (II,III), g_{max} values (0.5,1,2) and observed phytoplankton variables (Carbon, Chlorophyll). Mean model skill refers to the average cost function score of the optimal $K_{1/2}$ across all grid cells in a given configuration

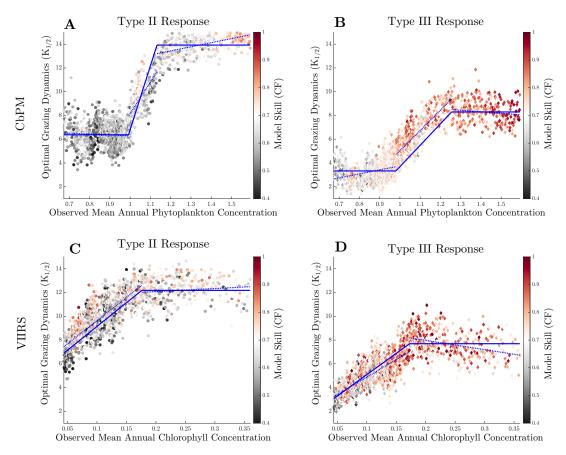


Figure S2. A piece-wise linear model is fit to the relationship between phytoplankton abundance and the optimal $K_{1/2}$ required to recreate its seasonal cycle when using **A**, **C**) a type II and **B**, **D**) a type III response function and **A**, **B**) CbPM carbon and **C**, **D**) VIIRS Chlorophyll to represent phytoplankton abundance

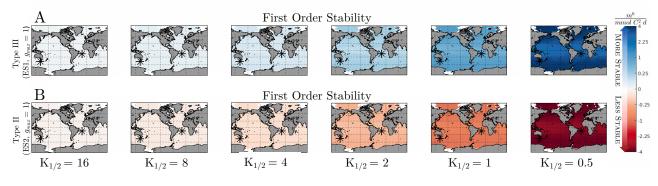


Figure S3. Sensitivity of ecosystem stability to $K_{1/2}$. Global distributions of the meanannual first order stability is plotted for all $K_{1/2}$ values, each with a consistent $g_{max} = 1$, and a **A**) type III and **b**) type II functional response.

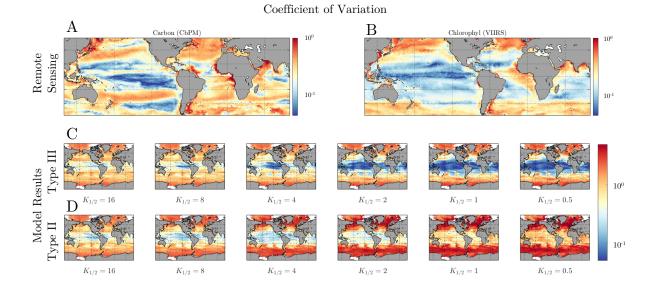


Figure S4. Sensitivity of the strength of the phytoplankton seasonal cycle to $K_{1/2}$. A, B) The global distribution of observed coefficients of variation, computed from the seasonal phytoplankton A) biomass and B) chlorophyll cycles is plotted above the C, D) the distribution of simulated coefficients of variation for all all $K_{1/2}$ values, each with a consistent $g_{max} = 1$, and a C) type III and D) type II functional response.

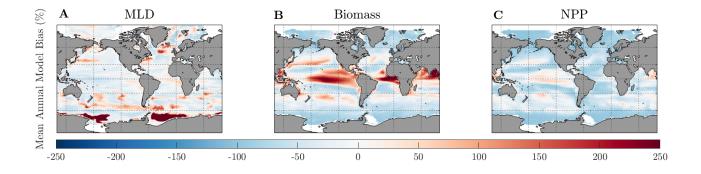


Figure S5. The mean annual model bias is plotted for the A) mixed layer depth (MLD) relative to HYCOM reanalysis, A) Phytoplankton biomass relative to CbPM and C) NPP relative to NPP.