

# Comparing the importance of iodine and isoprene on tropospheric photochemistry

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## Abstract

Naturally emitted reactive trace gases are thought to impact tropospheric composition, predominantly through the emission and chemistry of isoprene (C<sub>5</sub>H<sub>8</sub>). Other species are thought to play a less important role. Here the GEOS-Chem model is used to compare the impacts of isoprene and iodine emissions on present-day tropospheric composition. Removing isoprene emissions leads to a 4.4% decrease in tropospheric O<sub>3</sub> burden, a smaller absolute change than the 5.7% increase from removing iodine emissions. Iodine has a negligible impact on global mean OH concentrations and methane lifetime (-0.2% and +0.1%). Isoprene has a substantial impact on both (-7% and +6.5%). Isoprene emissions and chemistry are seen as essential for tropospheric chemistry models, but iodine is often not. We suggest iodine should receive greater attention in model development and experimental research to allow improved predictions of past, present and future tropospheric O<sub>3</sub>.

# Comparing the importance of iodine and isoprene on tropospheric photochemistry

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## Key Points:

- The global impact on  $O_3$  of present-day inorganic iodine and biogenic isoprene emission are similar in magnitude but opposite in sign.
- At the surface, the impact of iodine on  $O_3$  exceeds that for isoprene, and is notably larger in defining the background for Europe and North America.
- Iodine has a negligible impact on OH concentrations and  $CH_4$  lifetimes compared to isoprene.

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## Abstract

Naturally emitted reactive trace gases are thought to impact tropospheric composition, predominantly through the emission and chemistry of isoprene ( $\text{C}_5\text{H}_8$ ). Other species are thought to play a less important role. Here the GEOS-Chem model is used to compare the impacts of isoprene and iodine emissions on present-day tropospheric composition. Removing isoprene emissions leads to a 4.4% decrease in tropospheric  $\text{O}_3$  burden, a smaller absolute change than the 5.7% increase from removing iodine emissions. Iodine has a negligible impact on global mean OH concentrations and methane lifetime (-0.2% and +0.1%). Isoprene has a substantial impact on both (-7% and +6.5%). Isoprene emissions and chemistry are seen as essential for tropospheric chemistry models, but iodine is often not. We suggest iodine should receive greater attention in model development and experimental research to allow improved predictions of past, present and future tropospheric  $\text{O}_3$ .

## Plain Language Summary

Natural emissions from the Earth's surface play a large role in determining the chemistry of the atmosphere, influencing air quality and climate change. Considerable attention is given to land based emissions, notably of isoprene, which are emitted in vast quantities by trees and other vegetation and can impact the concentration of  $\text{O}_3$ , aerosols and the hydroxyl radical. Historically less emphasis has been on the influence of emissions of other compounds. We show that for one aspect of atmospheric composition (the globally averaged  $\text{O}_3$  concentrations) emission of iodine from the ocean are likely at least as important as isoprene emissions and may be more so. As such, there should be an increased focus on better understanding the emissions and chemistry of iodine species (and other halogens) and embed this information into our understanding of the Earth system.

## 1 Introduction

Gaseous emissions arising from terrestrial and marine ecosystems play important roles in regulating tropospheric photochemistry, in turn influencing climate and air quality. For example, oceanic emissions of dimethyl sulfide (DMS) are the dominant source of sulfur into clean background marine air (Yang et al., 2011; Sinha et al., 2007). The resulting sulfur aerosol directly scatters solar radiation as well as acting as cloud con-

denensation nuclei and thus these emissions play an important role in global climate (Shaw, 1983; Andreae & Crutzen, 1997; Ayers & Gillett, 2000).

Natural terrestrial biogenic emissions of volatile organic compounds (VOCs) have significant impacts on atmospheric oxidants (Trainer et al., 1987). Isoprene ( $C_5H_8$ ) forms the largest of these emissions ( $500Tg\ yr^{-1}$  (Guenther et al., 2012, 2006)), and has been the focus of a large body of research in terms of both the rate and controlling factors of its emission and its subsequent atmospheric degradation chemistry. Isoprene emissions come from plants and are dependent on temperature, rainfall, leaf area and other factors, exhibiting large variability both geographically and seasonally (Fuentes & Wang, 1999; Schnitzler et al., 1997; Guenther et al., 2006). Tropical broadleaf trees contribute approximately half of all global isoprene emissions (Guenther et al., 2006), thus the largest isoprene emissions come from areas with the greatest concentration of this plant type, in particular the Amazon rainforest.

The oxidation chemistry of isoprene has been extensively studied in the field (Wiedinmyer et al., 2001; Roberts et al., 1998; Starn et al., 1998; Biesenthal & Shepson, 1997) and in laboratory experiments (Atkinson et al., 1989; Paulson et al., 1992; Paulson & Seinfeld, 1992; Grosjean et al., 1993). Numerical models have been developed to incorporate known degradation chemistry (Trainer et al., 1987; Jenkin et al., 2015; Saunders et al., 2003; Bates & Jacob, 2019). Oxidation of isoprene is typically initiated by the hydroxyl radical (OH) and thus isoprene represents a large, natural, global sink for OH (Lelieveld et al., 2008). It can also react with ozone ( $O_3$ ) and other oxidants (nitrate radical ( $NO_3$ ), chlorine radical (Cl)) (Wennberg et al., 2018). The subsequent photo-chemistry is complex, with a large number of long and short lived species (Wennberg et al., 2018). If isoprene oxidation occurs in the presence of suitable  $NO_x$  concentrations, the production of peroxy radicals ( $RO_2$ ) can lead to net  $O_3$  production. At lower  $NO_x$  concentrations the primary reaction between  $O_3$  and isoprene, together with its ability to produce  $NO_y$  reservoir species and so reduce  $NO_x$  concentrations, can lead to net  $O_3$  loss (Paulot et al., 2012; Horowitz et al., 2007). Transport of  $NO_y$  species produced during isoprene oxidation can increase  $NO_x$  concentrations in remote downwind regions, increasing  $O_3$  production many thousands of kilometers from the isoprene source (Bates & Jacob, 2019). Overall, isoprene is calculated to be a net source of  $O_3$  into the global troposphere (Pierce et al., 1998; Fiore et al., 2011; Squire et al., 2015; Bates & Jacob, 2019).



In contrast to isoprene, iodine emissions lead to the destruction of  $\text{O}_3$ .  $\text{O}_3$  from the atmosphere can be transported into the ocean’s surface microlayer (SML) where it can react with iodide ( $\text{I}^-$ ) to produce HOI and  $\text{I}_2$  (Carpenter et al., 2021). These emissions are estimated to supply  $2 \text{ Gg yr}^{-1}$  of iodine to the global atmosphere. An additional  $0.6 \text{ Gg yr}^{-1}$  of iodine occurs through the emission of iodinated hydrocarbons ( $\text{CH}_3\text{I}$ ,  $\text{CH}_2\text{I}_2$ ,  $\text{CH}_2\text{IBr}$  and  $\text{CH}_2\text{ICl}$ ) (Jones et al., 2010; MacDonald et al., 2014; Prados-Roman et al., 2015). This reaction between  $\text{I}^-$  and  $\text{O}_3$  in the SML is also responsible for a significant fraction of the dry deposition of  $\text{O}_3$  to the ocean (Fairall et al., 2007; Carpenter et al., 2013; Luhar et al., 2017; Pound et al., 2020).

$\text{I}^-$  in the SML is formed from the thermodynamically more stable iodate ( $\text{IO}_3^-$ ) via biological reduction processes (Amachi, 2008; Chance et al., 2007) and as such could display sensitivity to both seasonal and climate timescales (Carpenter et al., 2021). Ice core samples show that the atmospheric iodine abundance has increased since pre-industrial times and significantly accelerated through the end of the 20<sup>th</sup> century, which is mainly attributed to increased atmospheric  $\text{O}_3$  driving higher HOI and  $\text{I}_2$  emissions (Cuevas et al., 2018; Legrand et al., 2018).

Tropospheric lifetimes of the emitted gaseous iodine compounds are relatively short (on the order of minutes to days), photolyzing to produce atomic iodine ( $\text{I}$ ). The subsequent catalytic iodine cycles are an efficient chemical loss route of  $\text{O}_3$ . Iodine atoms are rapidly oxidised by  $\text{O}_3$  to form iodine oxide ( $\text{IO}$ ), which can then further self-react to form higher oxides or cycle back to atomic  $\text{I}$  (Sommariva et al., 2012). Further reactions of  $\text{IO}$  can impact both  $\text{HO}_x$  ( $\text{OH} + \text{HO}_2$ ) and  $\text{NO}_x$  ( $\text{NO} + \text{NO}_2$ ) concentrations (Sommariva et al., 2012; Sherwen, Evans, et al., 2016). The inclusion of  $\text{I}$  chemistry in model simulations has been shown to reduce surface  $\text{O}_3$  concentrations and lower background  $\text{O}_3$  (Sarwar et al., 2019; Sherwen, Evans, et al., 2016). Recent work also shows that iodine containing trace compounds can be exported from the troposphere into the stratosphere where they may play a role in modulating the concentration of stratospheric  $\text{O}_3$  (Koenig et al., 2020; Cuevas et al., 2022).

Observations of reactive inorganic iodine compounds in the atmosphere are limited mainly to  $\text{IO}$ , which has been measured using a number of optical techniques (Prados-Roman et al., 2015; Volkamer et al., 2015; S. Wang et al., 2015; Koenig et al., 2020; Gómez Martín et al., 2013; Großmann et al., 2013; Mahajan et al., 2012, 2010). Although these emissions are sparse, model simulations of  $\text{IO}$  generally compare well to these observations

(X. Wang et al., 2021). Observations have also been made of  $I_2$  (Lawler et al., 2014) however these are even more limited in their spatial distribution than those of IO.

Although the mass of isoprene ( 500 Tg yr<sup>-1</sup> (Guenther et al., 2012)) and iodine ( 4 TgI yr<sup>-1</sup> (Sherwen, Schmidt, et al., 2016)) emitted into the atmosphere differ significantly, both can have a profound impact on the composition of the troposphere. Assessing the relative impacts on troposphere composition based on previous literature is difficult as these assessments have been made in different models over different timescales and have focused on the impact of only one of these sources at a time. Thus, assessing the relative importance of isoprene and iodine emissions on tropospheric photo-chemistry is difficult. Here we use the GEOS-Chem model to compare the relative impacts of iodine and isoprene on the tropospheric abundance of  $O_3$  and OH, and the impact of both iodine and isoprene on surface  $O_3$  mixing ratios. Thus, we compare the overall impacts of iodine and isoprene on atmospheric composition and present the argument that iodine should be considered, analogously to isoprene, as an important natural control on atmospheric composition.

## 2 Model description

This work uses the GEOS-Chem model (Bey et al., 2001) version 13.1.1 (GCC13.1.1, 2021) run globally at a spatial resolution of 2°x2.5° on the reduced vertical grid (47 vertical levels), running with full chemistry in both the troposphere and stratosphere. Meteorological data for these runs used MERRA-2 (Gelaro et al., 2017).

Isoprene emissions in GEOS-Chem are from MEGAN v2.1 (Guenther et al., 2012) which varies isoprene emissions depending on plant functional type, leaf area index, temperature, and photosynthetically active radiation. The subsequent isoprene oxidation chemistry in GEOS-Chem is from Bates and Jacob (2019) which has been used since model version 12.8.

The halogen (Cl, Br, I) chemistry scheme in GEOS-Chem was recently updated in version 12.9 by X. Wang et al. (2021). Organic iodine emissions are from Ordóñez et al. (2012). Inorganic iodine emissions follow Carpenter et al. (2013) as implemented by Sherwen, Evans, et al. (2016) and are given by equations 1 and 2 where  $ws$  is the wind speed [m/s],  $[O_{3(g)}]$  is the  $O_3$  concentration in the atmosphere at the interface with the surface [ppbv], and  $[I_{aq}^-]$  is the oceanic iodide concentration [Mol].

$$F_{HOI} = [O_{3(g)}] \sqrt{[I_{(aq)}^-]} \left( \frac{3.56 \times 10^5}{ws} - 2.16 \times 10^4 \right) \quad (1)$$

$$F_{I_2} = [O_{3(g)}][I_{(aq)}^-]^{1.3} (1.74 \times 10^9 - 6.54 \times 10^8 \ln(ws)) \quad (2)$$

The ocean surface iodide concentration is given by the parameterization of MacDonald et al. (2014), given in equation 3, where  $T$  is the sea surface temperature [K].

$$[I_{(aq)}^-] = 1.46 \times 10^6 \times \exp\left(\frac{-9134}{T}\right) \quad (3)$$

Three model runs were conducted from the 1<sup>st</sup> of January 2015 to the 1<sup>st</sup> of July 2017. The first year and a half of each simulation was considered the spin up to allow the composition to reach equilibrium. Analysis was thus performed on the period 1<sup>st</sup> July 2016 to 1<sup>st</sup> July 2017. For the first simulations no changes were made to the model. For the second simulation iodine emissions from the ocean were set to zero, and the concentration of iodine containing compounds in the model initial condition (for 2015-01-01) were set to zero. In the third simulation, isoprene emissions were set to zero, and the concentration of isoprene derived in the initial condition (for 2015-01-01) were also set to zero.

The model output daily average diagnostics with data processing performed in python using xarray (Hoyer & Hamman, 2017), numpy (Harris et al., 2020), cartopy (Met Office, 2010 - 2015) and matplotlib (Hunter, 2007).

### 3 Results

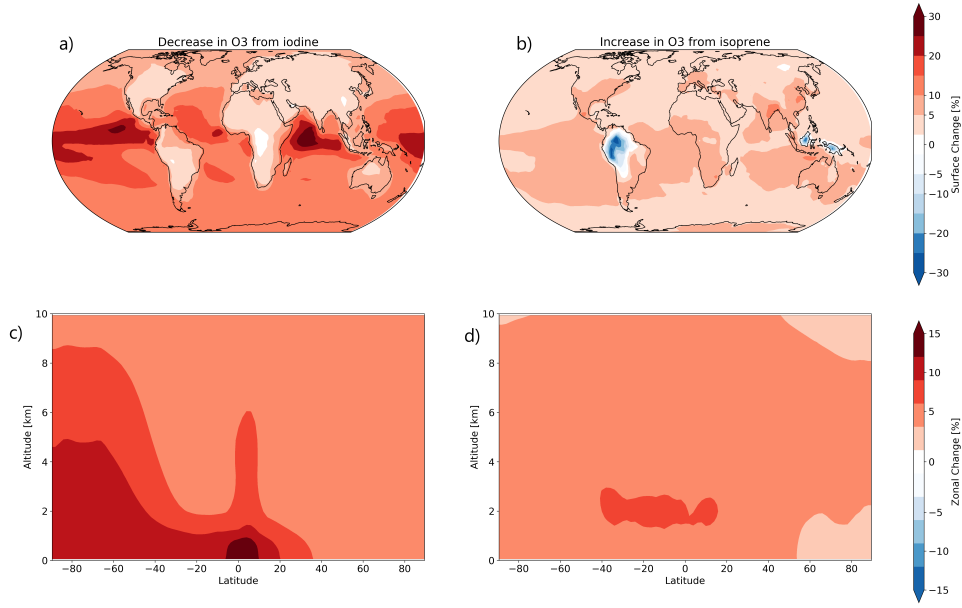
#### 3.1 Impacts on O<sub>3</sub>

Figure 1 shows the percentage change in annual mean concentrations of surface and zonal O<sub>3</sub> concentrations from switching off iodine (left) and isoprene emissions (right). Equivalent analyses for CO, NO<sub>x</sub> and NO<sub>y</sub> are shown in the supplementary material (figures S1 to S3). Iodine emissions reduce the global tropospheric O<sub>3</sub> burden from 332 Tg/yr to 315 Tg/yr (5% reduction). The largest decreases occur within the tropical marine boundary layer ( $\geq 20\%$ ) above tropical waters where iodide concentrations are the greatest (Chance et al., 2014), resulting in the highest iodine emissions (Sherwen, Evans, et al., 2016). Due to rapid atmospheric convection over the tropics, this region of depleted O<sub>3</sub> extends up to around 6km altitude. Another region of fractionally significant iodine-initiated O<sub>3</sub> loss

is the Southern Ocean where the large ocean surface area provides widespread iodine emission. However due to the comparatively low  $O_3$  concentrations over the Southern Ocean, these large percentage changes do not correspond to large changes in absolute concentration. A percentage decrease of 10% at 2km here translates to 2 ppbv reduction in  $O_3$ .

Isoprene emissions create a more complex distribution of changes. Over Amazonia and Oceania, the locations of the largest isoprene emission,  $O_3$  concentrations decrease at the surface. This is for two reasons. Firstly the direct reaction between  $O_3$  and isoprene increases the chemical loss of  $O_3$  (10% of global isoprene emissions are oxidised by  $O_3$  (Bates & Jacob, 2019)). Secondly the concentrations of  $NO_x$  in the region decrease as  $NO_x$  is shifted to reservoir species ( $NO_y$ , poly-aromatic nitrates (PAN) and organic nitrates) which reduces the chemical production of  $O_3$  over the region ( $NO_x$  and  $NO_y$  are shown in the supplementary material (figures S2 and S3)). Outside of these regions, the isoprene driven shift of  $NO_x$  to  $NO_y$  species over emission regions contributes to the global picture of increased  $O_3$ . Away from these emission regions, transported  $NO_y$  increase  $NO_x$  concentrations over the remote oceans downwind of high isoprene emissions, increasing  $O_3$  production (figure 1). This is most noticeable in the southern hemisphere. Globally the increase in  $O_3$  from isoprene emissions is dominated by this increase in  $NO_x$ . Any increase in  $O_3$  due to an increase in VOC concentration is limited because only a small amount of the world is VOC limited (mostly polluted cities in America, Europe and Asia) with the majority of the worlds  $O_3$  production being  $NO_x$  limited (Ivatt et al., 2022).

Globally the impact of isoprene and iodine on  $O_3$  are similar but opposite. Isoprene increases the tropospheric  $O_3$  burden by 13.6 Tg (4.1%) whereas iodine decreases it by 16.5 Tg (5.4%). Close to the surface (0-1km), iodine's impact on  $O_3$  (2.4 Tg, 8.9% decrease) is significantly larger than that of isoprene (1.2 Tg, 4.4% increase). Both are more important in the southern hemisphere than the northern, although iodine reduces the  $O_3$  burden by almost 50% more than isoprene increases it. The relatively larger role of both iodine and isoprene in southern hemispheric  $O_3$  reflects the increased importance of natural processes compared to anthropogenic emissions there.

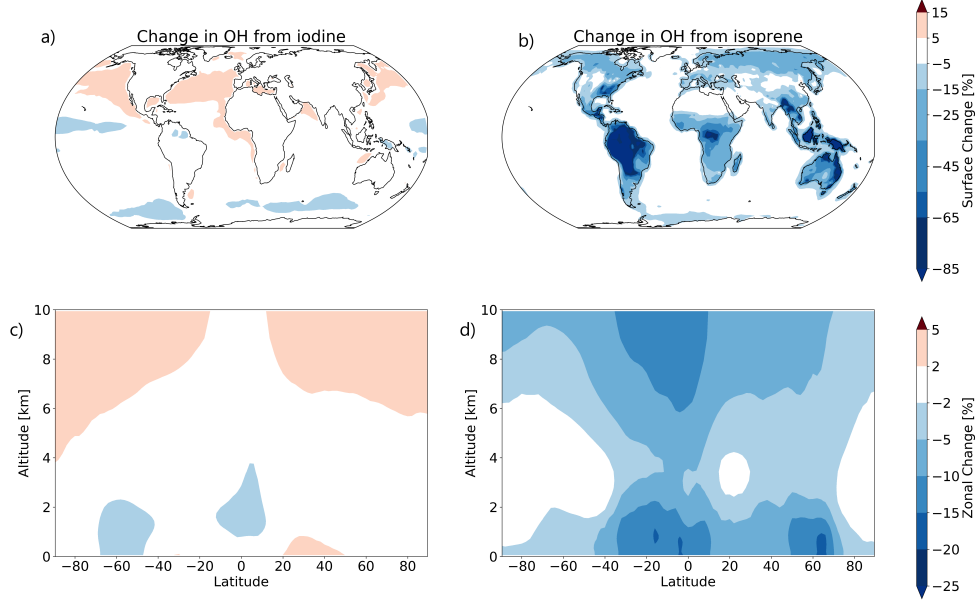


**Figure 1.** Annual average percentage decrease in surface [a) and b)] and zonal [c) and d)] O<sub>3</sub> from iodine emissions [a) and c)] and annual average percentage increase in surface and zonal O<sub>3</sub> from isoprene emissions [b) and d)].

### 3.2 Impact on OH

The similarity in the influence of isoprene and iodine on O<sub>3</sub> is not seen for OH. Figure 2 shows that iodine emissions in the model have negligible changes to tropospheric OH concentration (-0.24%), whereas isoprene emissions decrease tropospheric OH by around 7%. This difference in the response reflects different chemistry. The impact of iodine on OH has previously been found to be small due to compensating effects (Sherwen, Evans, et al., 2016). The reduction in O<sub>3</sub> concentrations from iodine leads to lower primary OH production, this is however offset by increased conversion of HO<sub>2</sub> to OH cycling via HOI. The global increase in O<sub>3</sub> from isoprene increases the primary chemical production of OH, however globally the increase in the chemical sink from reactions of OH with isoprene and its degradation products, is dominant and OH is decreased. The largest decreases in OH coincide with the regions of greatest isoprene emissions (Amazonia and Oceania). Due to efficient convection over these locations, the reduction in OH is observed throughout the troposphere. The change in model OH concentrations driven by isoprene

210 and iodine emissions results in changes in methane lifetimes of similar importance. The  
 211 reduction in OH concentrations from isoprene emissions increases the methane lifetime  
 212 by 6.5% (from 8.7 to 9.3 yrs). The negligible changes in OH caused by iodine result in  
 213 a minimal impact on methane lifetime (0.1% increase).



**Figure 2.** Annual average percentage change in surface [a) and b)] and zonal [c) and d)] OH from iodine emissions [a) and c)] and from isoprene emissions [b) and d)].

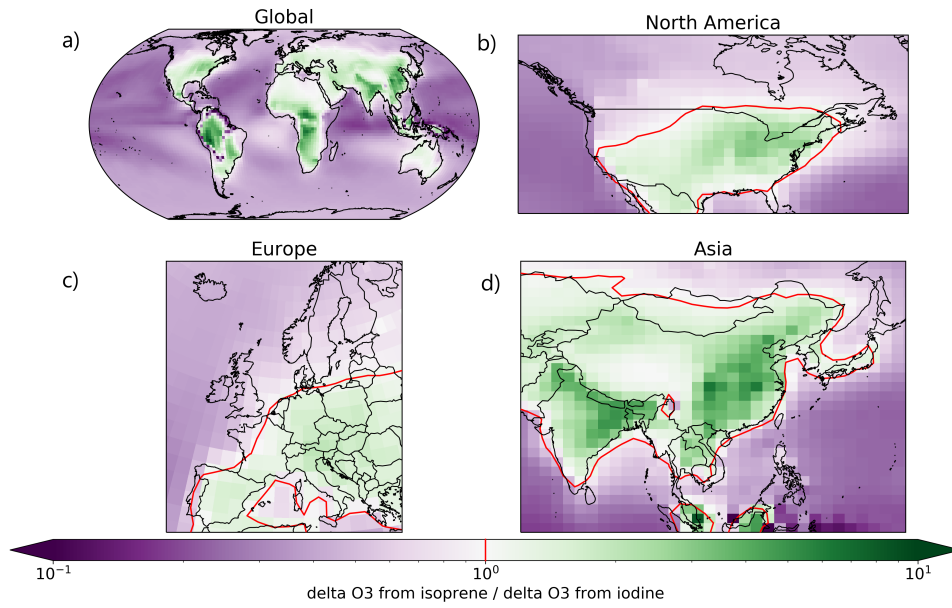
### 3.3 Importance of iodine and isoprene emissions on background O<sub>3</sub>

Figure 3 shows the ratio of change in surface O<sub>3</sub> from isoprene and iodine emissions. Additional seasonal plots, both global and regional, are shown in figures S4-7 of the supplementary material. This ratio allows for a comparison of the relative importance of iodine or isoprene emissions to be determined on surface O<sub>3</sub> concentration.

Although isoprene emissions lead to O<sub>3</sub> production in remote regions via increasing NO<sub>y</sub> concentration, O<sub>3</sub> loss due to iodine is more important in the marine environment. Iodine is thus more important than isoprene in determining the background concentration of O<sub>3</sub> at inflow regions (west coast of America and northern Europe). Iodine emissions are less important for Asia as transport of airmasses into this region spend no or negligible amounts of time over the ocean, with inflow coming from Europe.

Northern hemisphere winter O<sub>3</sub> in both terrestrial and oceanic environments has a significantly greater dependence on iodine emissions than isoprene emissions. This is largely due to minimal isoprene emissions and little O<sub>3</sub> production. This is not the case in the southern hemisphere where high isoprene emissions from South America and much of Africa maintain the dependence on isoprene. Changes to wind direction in summer and autumn result in iodine becoming important for O<sub>3</sub> into central Asia as the airmass entering this region switches from continental to oceanic in origin.

An important driver of the seasonal variation in the relative importance of iodine and isoprene is seasonality in their respective emissions. Iodine emissions only have a weak seasonal dependence when compared to isoprene emissions. Monthly iodine emissions in the northern hemisphere increase by 23% from minimum to maximum whereas isoprene increases by 270%.



**Figure 3.** Average surface ratio of fractional change in  $O_3$  from isoprene to fractional change in  $O_3$  from iodine, globally [a)], North America [b)], Europe [c)] and Asia [d)]. Red contour lines on regional plots are drawn based on the value of the ratio being 1 (iodine and isoprene emissions have equal impact on surface  $O_3$ )



## 4 Conclusions

Globally, iodine and isoprene emissions have a similar impact on the tropospheric  $O_3$  burden but in an opposing direction. The relative importance of each depends on location and season. Although iodine has the larger impacts on  $O_3$ , its impact on OH and methane lifetime are negligible compared to isoprene. Iodine has significantly more impact on surface  $O_3$  concentrations than isoprene. This has specific importance when considering background  $O_3$  air quality at inflow regions where the air mass has been transported over the marine environment, such as the western coast of north America and Europe. The emissions and subsequent chemistry of iodine should be considered in the same way as isoprene.

The processes leading to inorganic iodine flux from the ocean surface are complex, much like isoprene, however, the representation of these emissions is currently simplistic. Iodine emissions are dependent on the downward flux of  $O_3$  from the atmosphere into the SML, aqueous iodine chemistry, turbulence and the physical processes in the ocean surface as well as biological factors. Previous experimental constraints of the  $O_3 + I^-$  reaction and the role of organic chemistry are poorly constrained due to lack of experimental data and experimental data not reflecting real world SML concentrations. A more advanced representation of oceanic iodine emissions for use in global models should couple the chemical, physical and biological processes in the SML which drive the flux of iodine into the atmosphere. This will more accurately represent the production and subsequent emission of iodine and further improve our understanding of the role ocean atmosphere exchange plays in modulating tropospheric photochemistry.

## 5 Open Research

GEOS-Chem source code is openly available on GitHub (<https://github.com/geoschem/geos-chem>). This work used model version 13.1.1 (GCC13.1.1, 2021).

Analysis code used to produce statistics and figures in this paper are available at <https://doi.org/10.5281/zenodo.7016985> (Pound, 2022).

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# Supporting Information for ”Comparing the importance of iodine and isoprene in tropospheric photochemistry”

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## Contents of this file

1. Text S1 and S2
2. Figures S1 to S7

**Introduction** This supplemental material covers additional figures and description of results for the paper Comparing the importance of iodine and isoprene in tropospheric photochemistry.

## Text S1. - Additional plots for tropospheric changes

Iodine has a negligible impact on tropospheric CO concentration (figure S1). Iodine’s impacts CO via CO lifetime, as iodine has a negligible impact on tropospheric OH, a negligible change in CO is also the result. Isoprene significantly increases global CO concentrations, and as with OH and O<sub>3</sub>, the increase is more pronounced in the southern

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hemisphere. The increase in CO concentration from isoprene emissions is due to the oxidation and subsequent reactions of isoprene.

Isoprene increases  $\text{NO}_x$  in remote regions and decreases  $\text{NO}_x$  over tropical regions (figure S2). This is through conversion of  $\text{NO}_x$  into  $\text{NO}_y$  (figure S3) in the tropical locations which is then transported and subsequently reacts in remote regions, increasing  $\text{NO}_x$  there. Increase in  $\text{NO}_x$  and  $\text{NO}_y$  from iodine emissions are likely a result of nitrogenated iodine compounds. For figure S3  $\text{NO}_y$  was defined as the sum of species (with GEOS-Chem species name where needed) NO,  $\text{NO}_2$ ,  $\text{NO}_3$ ,  $\text{HNO}_2$ ,  $\text{HNO}_3$ ,  $\text{HNO}_4$ ,  $\text{BrNO}_2$ ,  $\text{BrNO}_3$ ,  $\text{ClNO}_2$ ,  $\text{ClNO}_3$ , ethanal nitrate (ETHLN), ethyl nitrate (ETNO3), monoterpene organic nitrate (HONIT), isoprene nitrates and dinitrates (ICN, IDN, IHN1, IHN2, IHN3, IHN4, INPB, INPD, ITCN, ITHN), hydroxynitrate from methacrolein (MCRHN, MCRHNB), methyl nitrate (MENO3), organic nitrate from monoterpene (MONITA, MONITS, MONITU), peroxyethacroyl nitrate (MPAN), methyl peroxy nitrate (MPN), hydroxynitrate from methyl vinyl ketone (MVKN),  $\text{N}_2\text{O}_5$ , n-propyl nitrate (NPRNO3), peroxyacetylnitrate (PAN), peroxypropionynitrate (PPN), propanone nitrate (PROPNN) and alkylnitrates greater than or equal to C4 (R4N2).

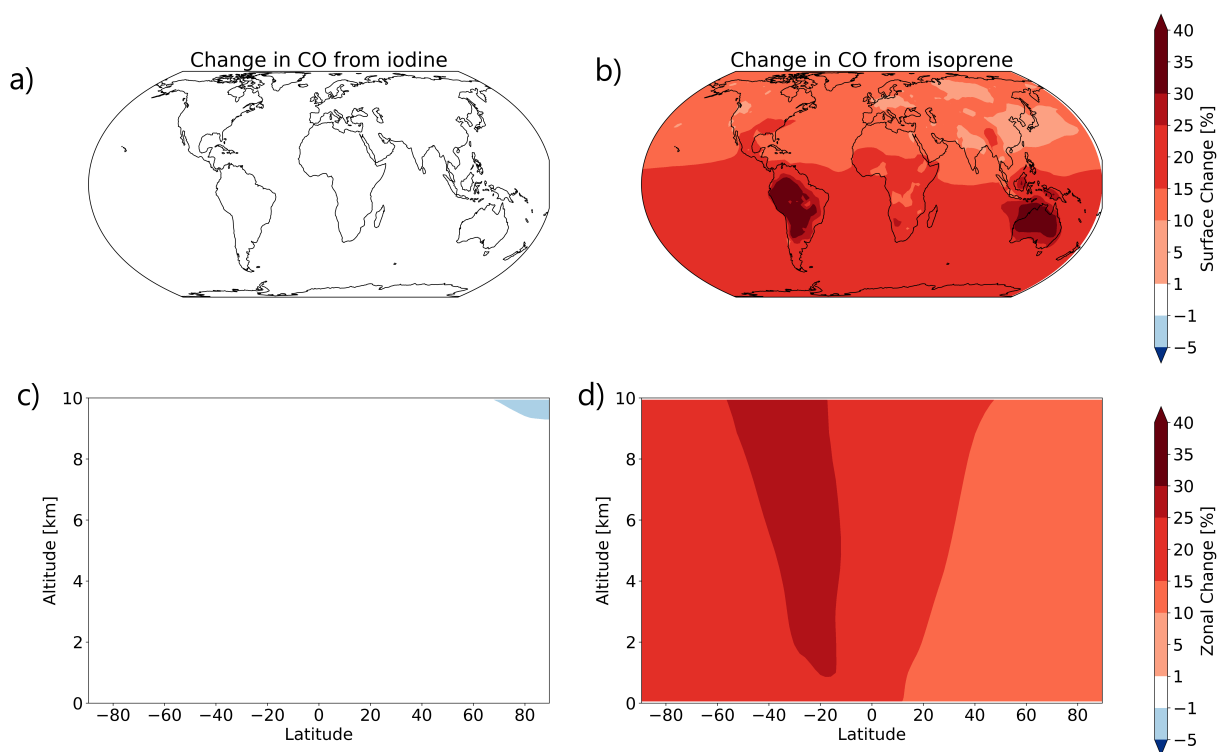
**Text S2. - Seasonal plots for ratio of iodine and isoprene changes to surface ozone**

Figures S4 - S7 show the seasonal variation in each region and globally from figure ??.

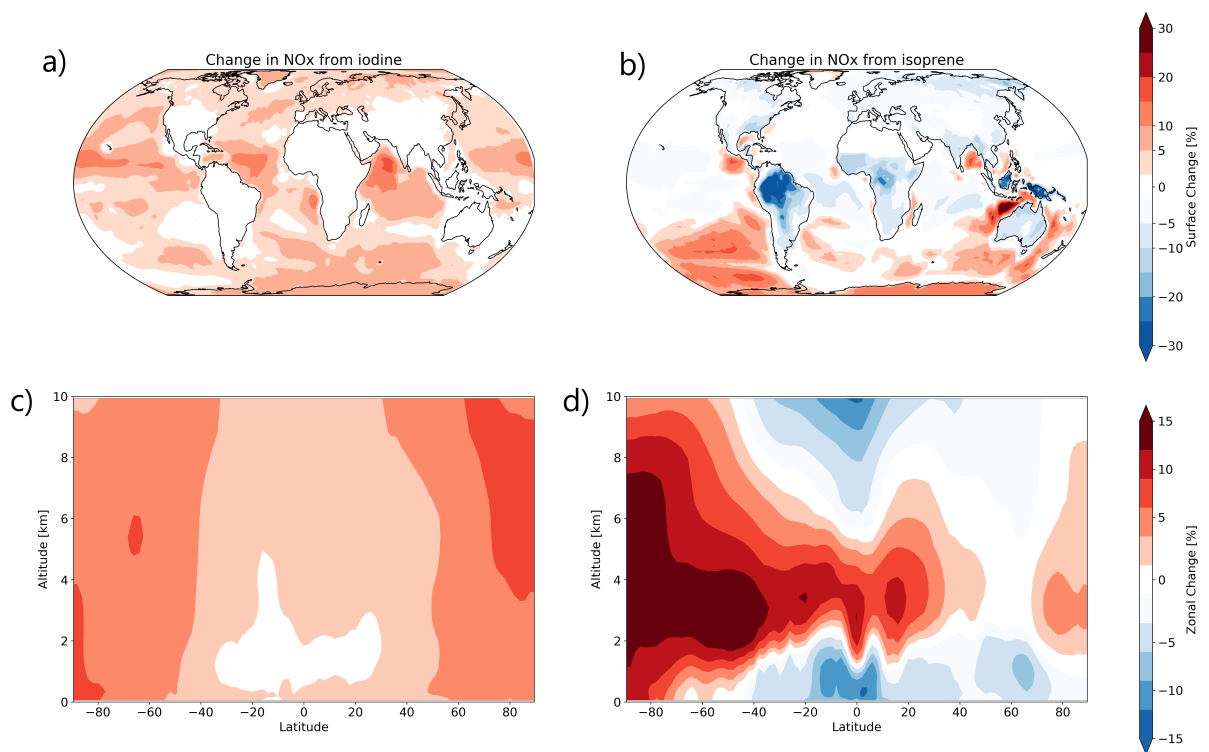
These are northern hemisphere seasons, with winter defined as December, January and

February. Spring is defined as March, April and May. Summer is June, July and August.

Autumn is September, October, November.

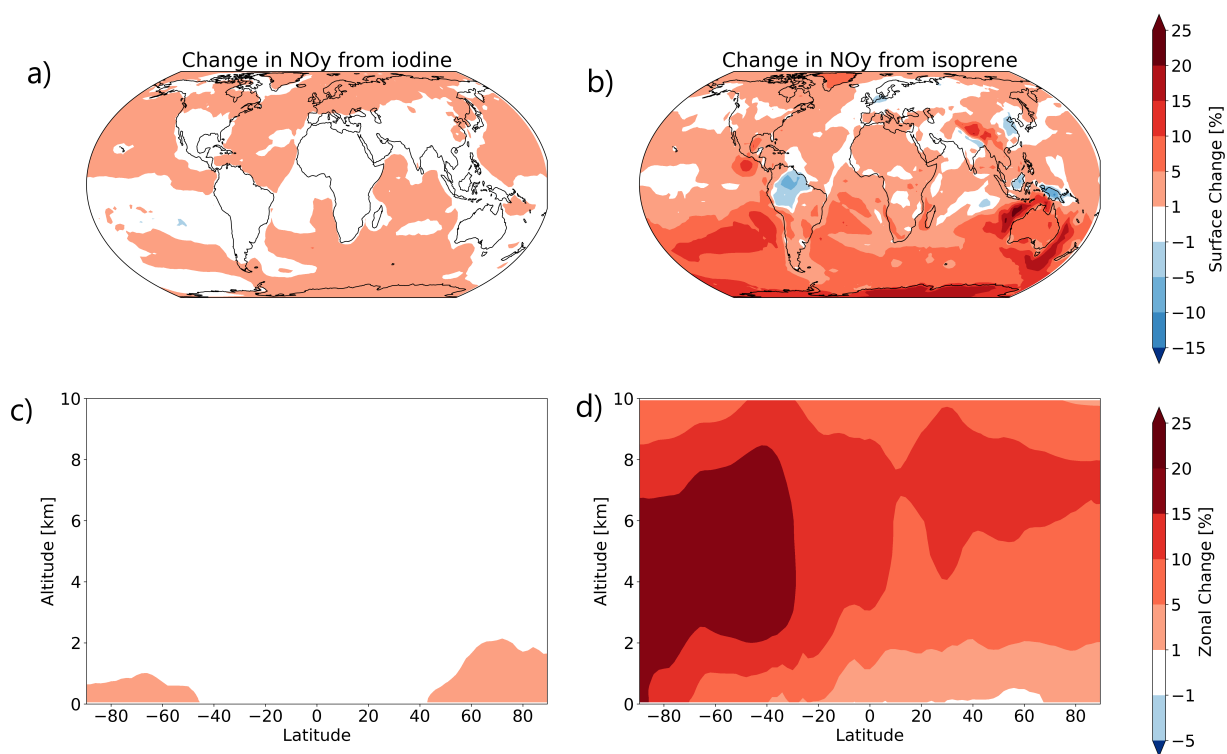


**Figure S1.** Annual average percentage change in surface [a) and b)] and zonal [c) and d)] CO from iodine emissions [a) and c)] and isoprene emissions [b) and d)].

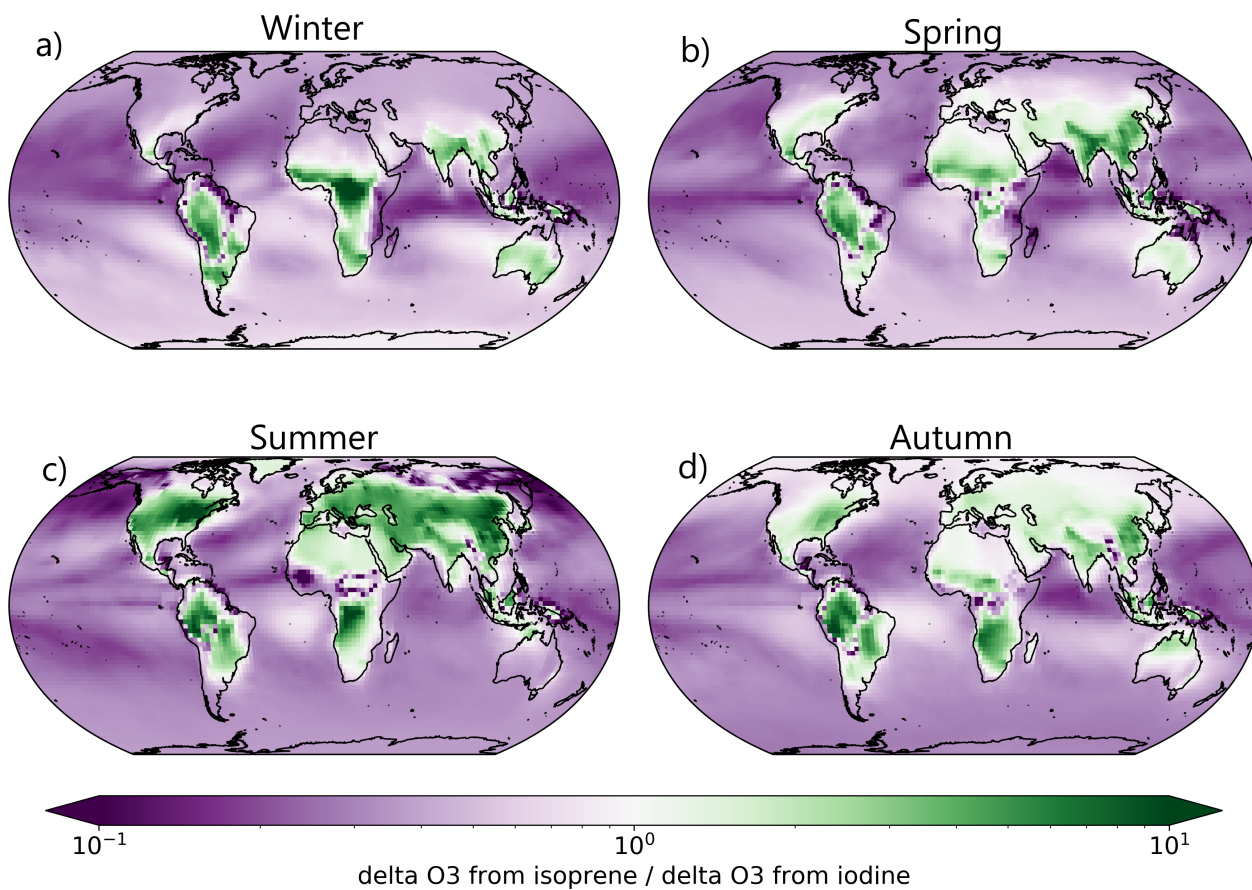


**Figure S2.** Annual average percentage change in surface [a) and b)] and zonal [c) and d)]  $\text{NO}_x$  (NO +  $\text{NO}_2$ ) from iodine emissions [a) and c)] and isoprene emissions [b) and d)].

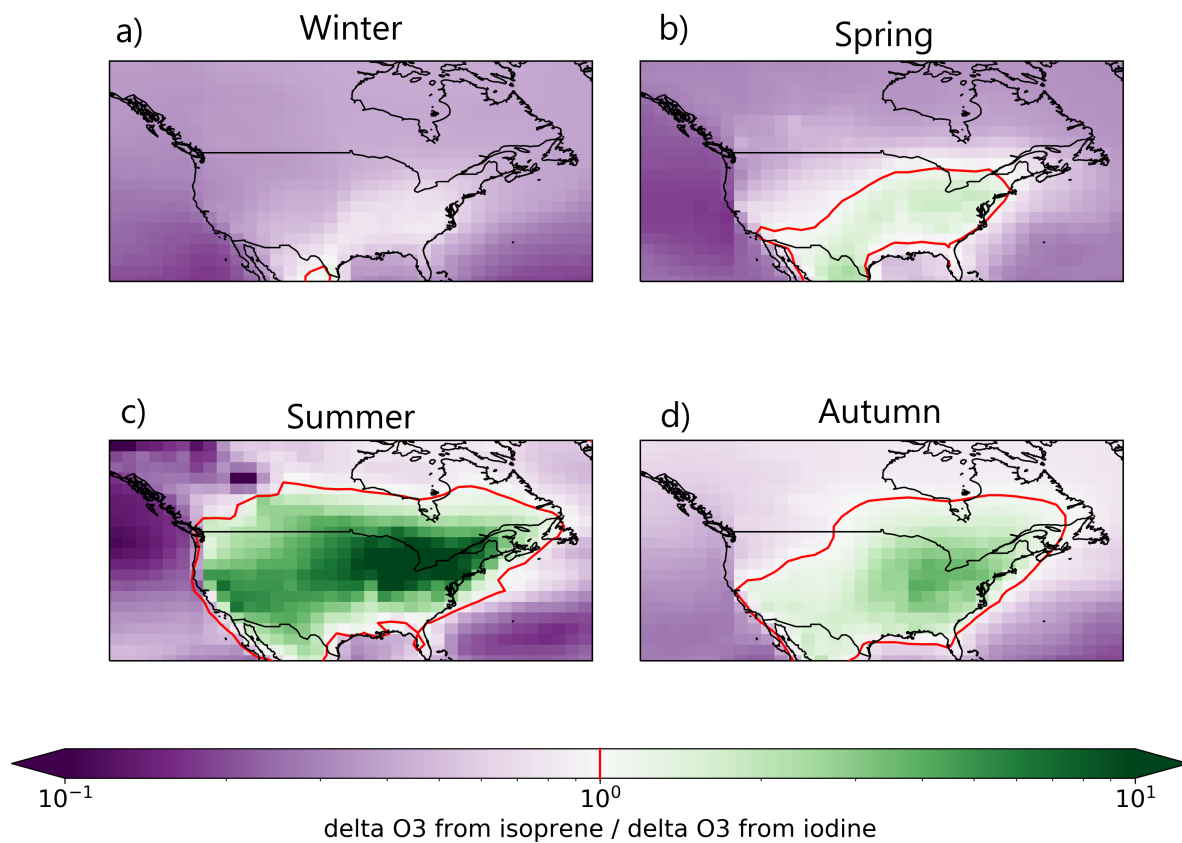




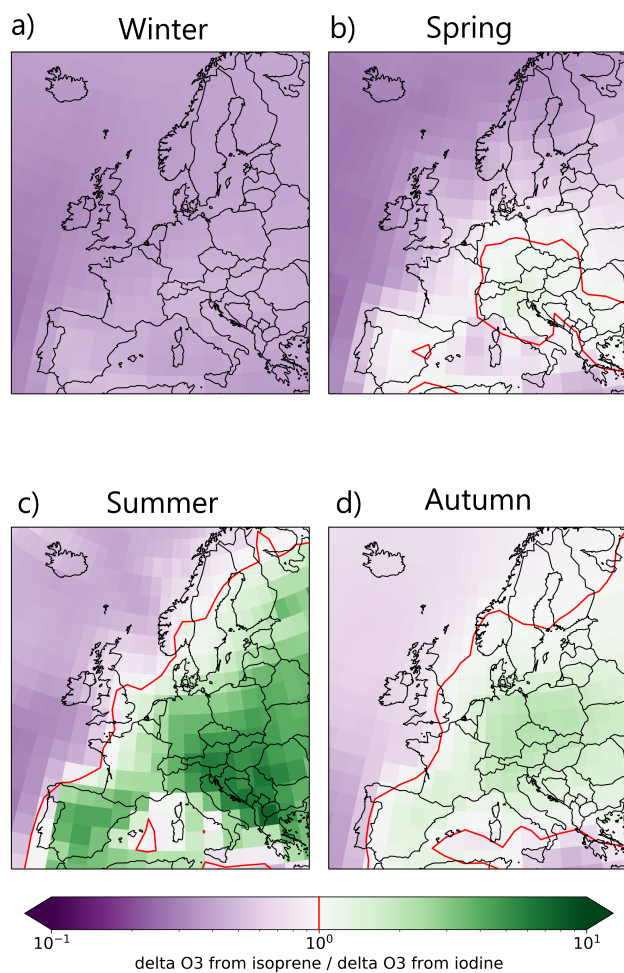
**Figure S3.** Annual average percentage change in surface [a) and b)] and zonal [c) and d)]  $\text{NO}_y$  from iodine emissions [a) and c)] and isoprene emissions [b) and d)].



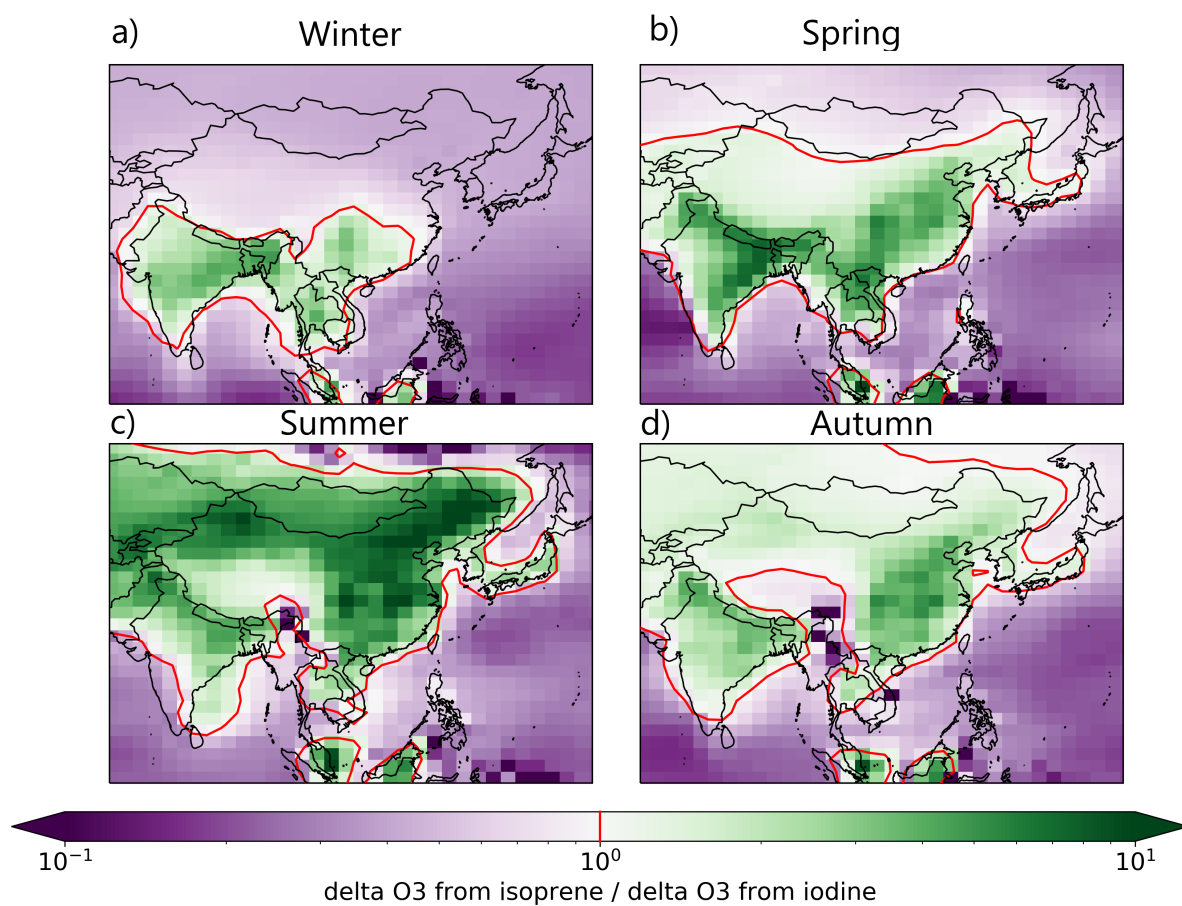
**Figure S4.** Global seasonal surface ratio of magnitude of change in  $O_3$  from isoprene to magnitude of change in  $O_3$  from iodine. Red contour lines represent value of the ratio as 1 (iodine and isoprene have the same magnitude impact on surface  $O_3$ )



**Figure S5.** Seasonal surface ratio of magnitude of change in  $O_3$  from isoprene to magnitude of change in  $O_3$  from iodine over North America. The red contour line represents value of 1 (iodine and isoprene have the same magnitude impact on surface  $O_3$ )



**Figure S6.** Seasonal surface ratio of magnitude of change in  $O_3$  from isoprene to magnitude of change in  $O_3$  from iodine over Europe. The red contour line represents value of 1 (iodine and isoprene have the same magnitude impact on surface  $O_3$ )



**Figure S7.** Seasonal surface ratio of magnitude of change in  $O_3$  from isoprene to magnitude of change in  $O_3$  from iodine over Asia. The red contour line represents value of 1 (iodine and isoprene have the same magnitude impact on surface  $O_3$ )