

Downward Trend in Methane Detected in a Northern Colorado Oil and Gas Production Region Using AIRS Satellite Data

Patrick Reddy¹ and C Taylor²

¹Independent Research Scientist

²Ramboll US Consulting

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Abstract

The oil and gas (O&G) sector is estimated to be the largest contributor to anthropogenic methane (CH₄) emissions in Colorado. Since 2004, the State of Colorado has implemented multiple regulations to significantly reduce emissions from the O&G sector. The Denver-Julesburg Basin (DJ Basin) is a significant O&G producing region in northern Colorado, and O&G production here has steadily increased over the last decade. To assess CH₄ trends in Northern Colorado, we selected CH₄ retrievals from the NASA Atmospheric Infrared Sounder (AIRS) instrument for 2003-2020. The study grid cell includes Denver, Boulder, and much of the dense O&G production in the DJ Basin. We computed mean June-August ascending node AIRS 700 hPa CH₄ for each year and subtracted mean June-August CH₄ sampled at NOAA's Niwot Ridge (NWR) station, a high-altitude background site. Differences represent estimated enhancement over background. Linear regression shows an annual change of -2.84 ppb +/- 0.8 ppb from 2012-2020 (R-squared 0.90) and an estimated reduction of 56% for 2012-2020, despite substantial increases in O&G production. Local CH₄ enhancement is strongly correlated with surface measurements of ethane at Platteville which is in the center of the O&G fields (correlation coefficient 0.96), and this is evidence that reductions in O&G emissions are driving reductions in CH₄. We conclude that AIRS CH₄ can be used to measure the efficacy of emissions control programs in this region and that regulatory requirements are having an effect.

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2 **Production Region Using AIRS Satellite Data**

3 **P. J. Reddy¹, C. Taylor²**

4 ¹Independent Research Scientist, Saguache County, Colorado

5 ²Ramboll US Consulting, Fort Collins, Colorado

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7 Corresponding author: Patrick Reddy (preddyresearch@gmail.com)

8 **Key Points:**

- 9 • Despite substantial increases in local oil and gas production, methane enhancement in the
10 northern Colorado study area declined significantly from 2012-2020.
11 • Decreases in both methane enhancement and ethane support the conclusion that
12 emissions reduction efforts have been effective.
13 • AIRS satellite methane retrievals can detect changes in boundary-layer concentrations
14 when mixing heights and source emissions are large.

15

16 **Abstract**

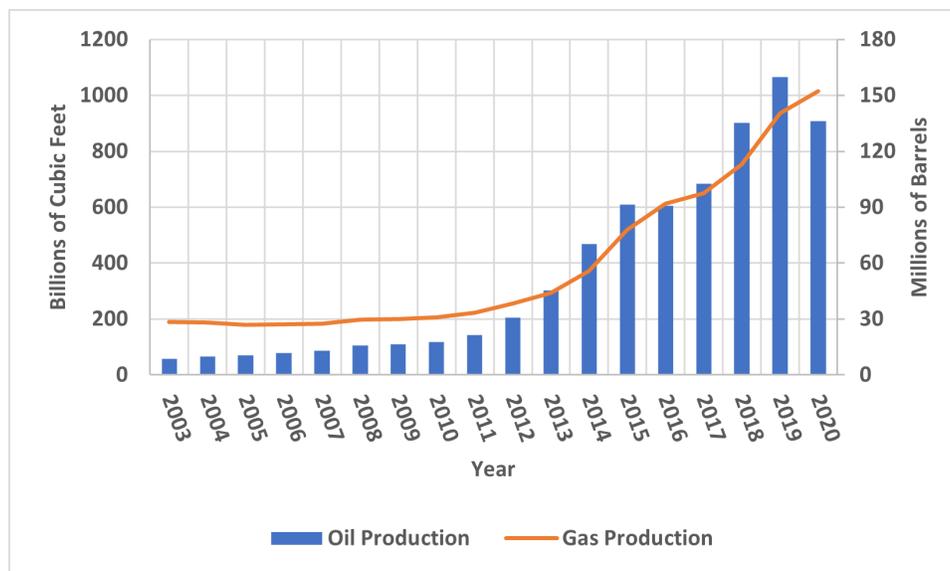
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18 (CH₄) emissions in Colorado. Since 2004, the State of Colorado has implemented multiple
19 regulations to significantly reduce emissions from the O&G sector. The Denver-Julesburg Basin
20 (DJ Basin) is a significant O&G producing region in northern Colorado, and O&G production
21 here has steadily increased over the last decade. To assess CH₄ trends in Northern Colorado, we
22 selected CH₄ retrievals from the NASA Atmospheric Infrared Sounder (AIRS) instrument for
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25 for each year and subtracted mean June-August CH₄ sampled at NOAA's Niwot Ridge (NWR)
26 station, a high-altitude background site. Differences represent estimated enhancement over
27 background. Linear regression shows an annual change of -2.84 ppb +/- 0.8 ppb from 2012-2020
28 (R-squared 0.90) and an estimated reduction of 56% for 2012-2020, despite substantial increases
29 in O&G production. Local CH₄ enhancement is strongly correlated with surface measurements
30 of ethane at Platteville which is in the center of the O&G fields (correlation coefficient 0.96), and
31 this is evidence that reductions in O&G emissions are driving reductions in CH₄. We conclude
32 that AIRS CH₄ can be used to measure the efficacy of emissions control programs in this region
33 and that regulatory requirements are having an effect.

34 **1 Introduction**

35 According to the Colorado Department of Public Health and Environment (CDPHE)
36 (2021), the oil and gas (O&G) sector is estimated to be the largest industrial contributor to
37 anthropogenic methane (CH₄) emissions in Colorado. The Denver-Julesburg Basin (DJ Basin) is
38 a significant O&G producing region in the northern portion of the State. O&G production in the
39 DJ Basin has steadily increased over the last decade. Figure 1 shows O&G production trends for
40 2003-2020 for the Wattenberg Field, the primary resource area in the DJ Basin (data from the
41 Colorado Oil and Gas Commission, October 2021, <https://cogcc.state.co.us/data.html>.) Oil
42 production increased by 343% and gas production by 297% between 2012 and 2020.

43 For a variety of reasons, including the fact that a portion of the state that includes the DJ
44 Basin has been designated an ozone nonattainment area by the Environmental Protection Agency
45 (EPA), Colorado has implemented regulations to significantly reduce emissions of CH₄ and
46 volatile organic compounds (VOCs) from O&G production facilities. Although initial
47 regulations were focused on ozone reduction goals, many control requirements had associated
48 CH₄ emission reduction co-benefits. One of the most significant of these regulatory changes
49 since 2004 was in February 2014 when the Colorado Air Quality Control Commission (AQCC)
50 fully adopted EPA's New Source Performance Standards (NSPS), 40 CFR Part 60, Subpart
51 OOOO, into AQCC Regulation No. 6 and adopted more stringent control requirements for VOCs
52 and hydrocarbons for a variety of O&G sources in AQCC Regulation No. 7. The changes to
53 Regulation No. 7 in early 2014 included expanded control requirements for hydrocarbon liquid
54 storage tanks, pneumatic controllers, glycol dehydrators, and components; implementation of a
55 leak detection and repair program; and limitations to venting associated with maintenance and
56 liquids unloading from storage tanks. Since 2014, the AQCC has continued to adopt additional
57 measures to reduce CH₄ and VOCs from O&G sources, and this is anticipated to continue
58 through 2021 and beyond. The Regional Air Quality Council (RAQC) and CDPHE (2020)

59 projected that O&G VOC emissions would decline from 279.7 to 119 tons per day from 2011 to
 60 2020 within the Denver Metro and North Front Range Ozone Nonattainment Area (a 57%
 61 reduction).



62

63 **Figure 1.** O&G production in the Wattenberg Field within the DJ Basin. The greatest growth
 64 occurs between 2012 and 2020.

65 If these controls have been effective, then we would expect to see reductions in CH₄ in
 66 the region after the implementation of new rules in 2014. In order to understand the efficacy of
 67 past O&G regulations and drivers for future emission control regulations, this analysis uses
 68 satellite and surface measurements to examine CH₄ trends and reductions over the past two
 69 decades in an area of Northern Colorado that includes the Denver metropolitan area and a
 70 portion of the DJ Basin north of Denver that contains a high density of O&G wells. While a
 71 variety of government agencies, researchers, and other groups have assessed CH₄ and VOC
 72 concentrations in Colorado from surface and aerial measurements, satellite CH₄ data has not yet
 73 been widely used to assess CH₄ trends in Colorado.

74 Three recent papers consider trends in surface measurements of CH₄ or non-methane
 75 hydrocarbons (NMHCs) in our study area for shorter time periods within our temporal domain
 76 (2003-2020). Oltmans et al. (2021) recently reported trends in CH₄ using air samples collected at
 77 the National Oceanic and Atmospheric Administration (NOAA) Boulder Atmospheric
 78 Observatory (BAO) tower near Erie in southwestern Weld County. For 2008-2016 they found no
 79 statistically significant trends in CH₄ relative to background values when flows were from the DJ
 80 Basin. Ortega et al. (2021) found a positive trend of 0.9 ppb ± 0.3 % per year for ethane at a site
 81 north of Boulder from 2010-2019 using a solar absorption Fourier Transform InfraRed
 82 instrument. In a source-apportionment study that considered trends from 2013 to 2016, Lyu et al.
 83 (2021) show a reduction of approximately 50% in nonmethane hydrocarbons at the CDPHE
 84 Platteville site which is located within the DJ Basin in Weld County. They report that “new
 85 regulations implemented by the state as well as changes in operating practices made by the
 86 industry for other reasons might explain the observation that NMHC mixing ratios at the

87 Platteville site were lower in 2016 than in 2013.” We will show that our results are generally
88 consistent with those from each of these studies and also point to a statistically significant
89 decline in CH₄ extending from 2012 through 2020.

90 Satellites and aircraft studies are now common tools for calculating emissions fluxes and
91 the contributions of O&G sources to ambient concentrations of CH₄ and VOCs, and because of
92 greater spatial coverage these can yield data more representative of regional emissions than a
93 limited number of surface monitoring sites. In addition, satellite measurements are made on a
94 regular basis and typically provide greater temporal coverage than aircraft studies. Recently, de
95 Gouw et al. (2020) published analyses of CH₄ data from the TROPOspheric Monitoring
96 Instrument (TROPOMI) on board the Copernicus Sentinel-5 Precursor satellite, which was
97 launched in October of 2017. In their paper, the authors have identified and quantified CH₄
98 enhancements over many O&G basins in the United States. TROPOMI has high spatial
99 resolution and greater signal sensitivity than older satellite instruments, but there is not yet a long
100 enough data record to identify long-term trends.

101 To assess long-term CH₄ trends in Northern Colorado, we selected the Atmospheric
102 Infrared Sounder (AIRS) instrument launched in 2002 on the Aqua Satellite, because of its
103 almost 20 years of data, its ongoing use, and its provision of an accurate estimate of the rates of
104 change in CH₄ from year-to-year. Zhang et al. (2020) concluded that AIRS CH₄ data are suitable
105 for analyses of spatial and temporal patterns across the globe. They demonstrated that AIRS CH₄
106 data are closer to surface concentrations than CH₄ data from the SCanning Imaging Absorption
107 spectroMeter for Atmospheric CartographY (SCIAMACHY) satellite instrument or the
108 Greenhouse Gases Observing Satellite (GOSAT). AIRS has been used by researchers to quantify
109 regional and national CH₄ changes and trends over both short and reasonably long-time scales
110 (Rendana et al., 2021; Yang and Wang, 2020; and Ribeiro et al., 2016). Wu et al. (2019) recently
111 completed a comprehensive, long-term, analysis of CH₄ trends across China using AIRS data
112 from 2002-2016. They conclude that AIRS CH₄ concentrations “showed good consistency with
113 the ground measurements of surface CH₄ concentration from the World Data Centre for
114 Greenhouse Gases (WDCGG) ($R^2 = 0.83$, $p < 0.01$), indicating that the remotely-sensed CH₄
115 reflected the spatial and temporal variations of surface CH₄ concentration”. Rendana et al. (2021)
116 also found good agreement between AIRS and surface CH₄ measurements with an R^2 of 0.86.

117 Our goal is to generate an initial estimate of changes in CH₄ and an AIRS-based metric
118 for tracking rates of change in CH₄ that can eventually be linked with potentially more robust
119 trend studies using TROPOMI and future higher resolution satellite instruments.

120 **2 Data Sources and Methods**

121 We used June through August 2020 AIRS Version 7 retrievals for the 700 hPa level for
122 2003-2020 (<https://giovanni.gsfc.nasa.gov/giovanni/>). Details of this product are provided here:
123 https://disc.gsfc.nasa.gov/datasets/AIRS3STM_7.0/summary. Summer season data were chosen
124 in part to optimize the relevance of our results to assessments of trends in local O&G VOC
125 emissions that contribute to ozone exceedances and also because of the higher sensitivities of the
126 AIRS product when surface mixed layers are deep. The AIRS product is particularly well suited
127 for several reasons. Specifically, we selected the AIRS data for this study due to its long period
128 of record, the frequency and timing of overflights, and relatively low uncertainty for CH₄.

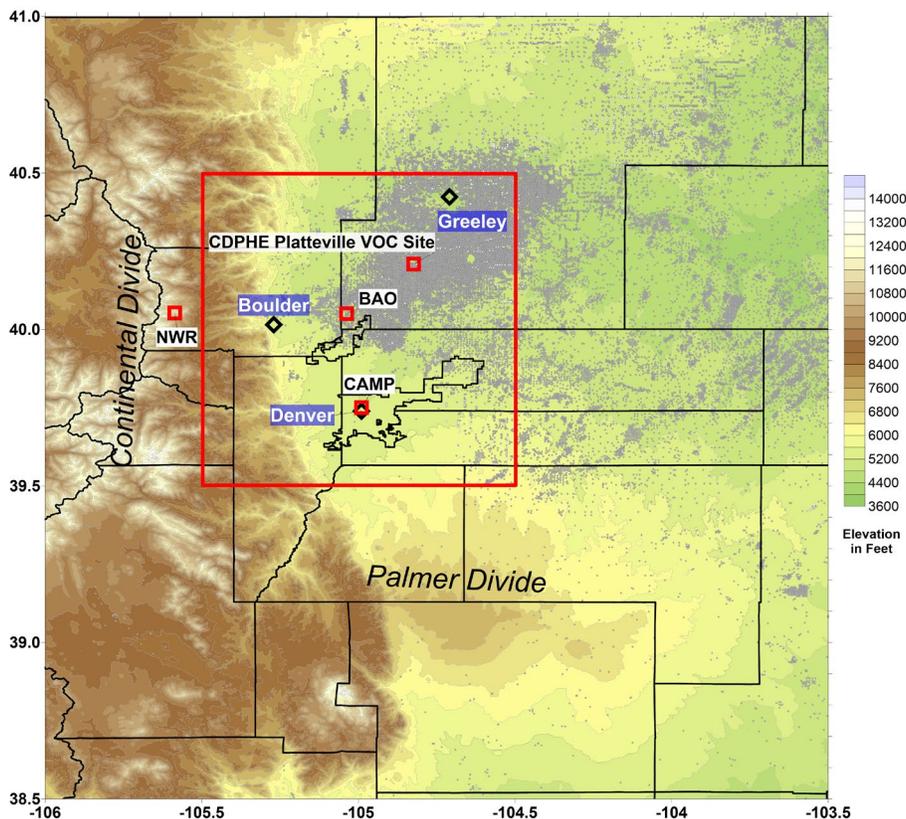
129 We used NOAA Global Monitoring Laboratory (GML) monthly mean CH₄ data collected
130 at Niwot Ridge (<https://gml.noaa.gov/>) (Dlugokencky et al., 2021) to estimate local, continental
131 background concentrations. Niwot Ridge (NWR) is a high-altitude site at 40.053°N, 105.586°W,
132 and 3,523 meters above sea level (MSL) west of the study area. According to Lan et al. (2022),
133 GML calculates monthly means after first extracting weekly values from a smoothing curve
134 applied to all valid data using the method described by Thoning et al. (1989).

135 We also acquired vertical profile data for CH₄ from aircraft flights from the NASA
136 summer 2014 field campaign “Deriving Information on Surface Conditions from Column and
137 Vertically Resolved Observations Relevant to Air Quality” (DISCOVER-AQ) ([https://www-
138 air.larc.nasa.gov/missions/discover-aq/P3B-Profiles.co2014.html](https://www-air.larc.nasa.gov/missions/discover-aq/P3B-Profiles.co2014.html)). With these data we assessed
139 the representativeness of AIRS retrieval concentrations.

140 AIRS CH₄ data were also compared with Colorado Department of Public Health and
141 Environment (CDPHE) median annual ethane data measured at Platteville and the Denver
142 CAMP station (from: https://www.colorado.gov/airquality/tech_doc_repository.aspx). The
143 Platteville site is at 40.209°N, 104.824°W, and 1,469 meters MSL. The CAMP station is at
144 39.75°N, 104.99°W, and 1,593 meters. Monitoring at these sites was initiated in late 2011.
145 Three-hour samples were collected with summa canisters from 6-9 AM every six days, and the
146 chemical constituents were analyzed following EPA Compendium Method TO-15 (USEPA,
147 1999). Ethane and NMHC concentrations at Platteville are largely from O&G sources (Lyu et al.,
148 2021). Ethane trends at this site are representative of O&G emission trends. The
149 representativeness of Platteville monitoring for the O&G fields was tested with the NOAA
150 Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) back trajectories (Stein et
151 al., 2015; Rolph et al., 2017) for 2018.

152 The study area for the trend analysis is presented in Figure 2, and it includes most of the
153 Denver-Boulder metro area and most of the O&G wells in the DJ Basin (the grid cell area will be
154 referred to as the DDJB). The AIRS satellite flies over the DDJB region once every 12 hours.
155 The afternoon overflights occur at approximately 13:30 MST and are referred to as ascending
156 node overflights. We used ascending node data for 2003-2020 for a 1° by 1° grid cell centered at
157 40°N and 105°W. This is a standard grid cell from the NASA AIRS CH₄ product.

158 We obtained mean June-August 500 hPa heights from the National Center for
159 Environmental Prediction (NCEP) Reanalysis ([https://psl.noaa.gov/cgi-
160 bin/data/timeseries/timeseries1.pl](https://psl.noaa.gov/cgi-bin/data/timeseries/timeseries1.pl)) for a 2.5° by 2.5° grid cell centered on 40°N 105°W, and this
161 contains our study area. The NCEP Reanalysis meteorological data product is described by
162 Kalnay et al. (1996). We correlated year-to-year changes in ethane and CH₄ enhancements with
163 year-to-year changes in 500 hPa heights to see if meteorology plays a role in the trends we
164 identified. To provide evidence that the AIRS 700 hPa ascending node CH₄ retrievals for the
165 study area are sensitive to planetary boundary layer (PBL) heights, we calculated the correlation
166 between monthly mean AIRS data and monthly mean PBL heights from the NASA Modern-Era
167 Retrospective analysis for Research and Applications MERRA version 2 (MERRA-2) for 2012-
168 2020 (<https://giovanni.gsfc.nasa.gov/giovanni/>).



169

170 **Figure 2.** Grid cell for the NASA AIRS CH₄ product centered on 40°N and 105°W in northern
 171 Colorado. The grid cell includes most of the Denver-Boulder metro area and most of the O&G
 172 wells (gray dots) in the southwestern portion of the DJ Basin. The grid cell covers mostly lower
 173 elevation terrain (~1,500 meters MSL). The foothills account for a small fraction of the grid cell
 174 area.

175 NASA has averaged AIRS data for nine single-footprint measurements in this grid cell.
 176 The grid cell covers much of the Wattenberg field, an O&G field located in the DJ Basin, as well
 177 as the Denver-Boulder metro area. The high density of O&G wells in the northeast corner of the
 178 AIRS grid cell is associated with the Wattenberg Field. Thermally driven upslope flows prevail
 179 along Colorado's Front Range at 13:30 MST (the satellite overflight time) during the summer
 180 (Toth and Johnson, 1985; Reddy and Pfister, 2016; Pfister et al., 2017; Flocke et al., 2019).
 181 These flows will often transport CH₄ westward and southwestward from the O&G fields into and
 182 across the grid cell by 13:30 MST. For example, using data available from an Iowa State
 183 University website
 184 (https://mesonet.agron.iastate.edu/sites/dyn_windrose.phtml?station=EIK&network=CO_ASOS)
 185 we calculated that winds at Erie at the southern end of the Wattenberg were calm or from the
 186 north through east-southeast 71% of the time from 9 to 13 MST from June-August, 2015-2020.

187 The AIRS CH₄ uncertainty depends on a variety of factors. Uncertainty for an individual
 188 footprint and a given day will be much greater than for monthly or seasonal averages or gridded
 189 data sets such as the ones we used that include data from multiple footprints. Kulawik et al.
 190 (2021) report decreases in AIRS CH₄ errors with increasing averaging times, and for a single

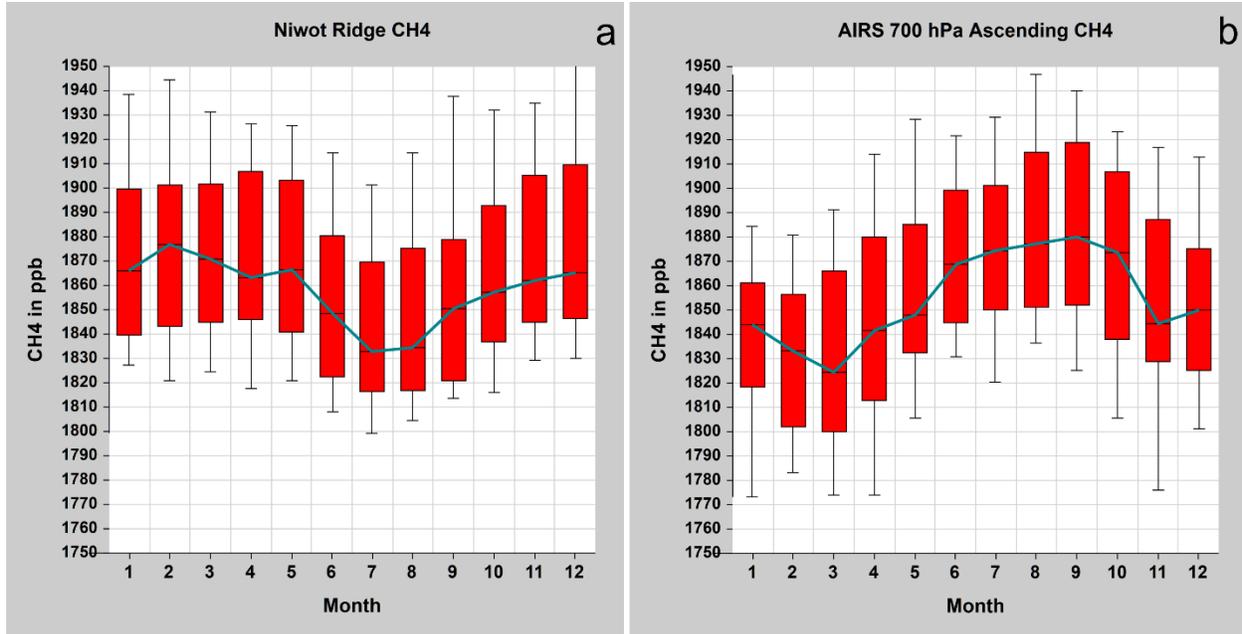
191 footprint uncertainty decreases significantly for seasonal averages. Xiong et al. (2015) report a
192 bias of 0.27 % and a root mean square error (RMS) of 0.87 % for AIRS CH₄ data for 555 hPa to
193 777 hPa based on intercomparisons with aircraft data and using version 6 of the standard NASA
194 AIRS product. Interpreting the findings of Xiong et al. (2015), Kulawik et al. (2021) report a
195 standard deviation of 16 ppb for 555 hPa to 777 hPa when compared with aircraft data and a bias
196 of 5 ppb near the mid-troposphere.

197 We used AIRS monthly average data from 2003 to 2020. We computed June-August
198 averages from these monthly means. We also used the GML monthly averages of weekly flask
199 sample data for NWR to estimate local continental background CH₄ and computed June-August
200 averages from these monthly means. Oltmans et al. (2021) also used NWR data for western
201 continental background estimates in their assessment of CH₄ trends at the BAO tower near Erie,
202 Colorado. They point out that upslope conditions at NWR during July “led to a somewhat
203 elevated median mole fractions that likely overestimate the background concentration”.

204 Butterworth (2011) reports that summer upslope flows at NWR, when they are present,
205 typically begin at 12:00 MST, and upslope persisting for at least two consecutive hours occurs on
206 34.8% of summer days. All but 7 of the NWR sample days for 2012-2020, the key period of our
207 study, were taken prior to 12:00 MST. We subtracted mean June-August Niwot Ridge (NWR)
208 CH₄ concentrations from the mean June-August AIRS DDJB CH₄ values. The difference
209 represents an estimate of the CH₄ enhancement from local and regional sources.

210 Unlike aircraft studies in this area which use upwind and downwind flights to quantify
211 CH₄ enhancements (see Petron et al., 2014), our approach may not account for the influences of
212 upwind surface sources on background and the calculated DDJB CH₄ enhancements. Both the
213 presence of some DDJB CH₄ transport to NWR in July and the transport of CH₄ into the DDJB
214 from the east could introduce biases of unknown magnitude in our estimates of local
215 enhancement. If July NWR overestimates background to some degree, then this could introduce
216 a high bias in the calculated percentages of decline in DDJB CH₄. If the calculated local
217 enhancement includes a relatively unchanging contribution from sources upwind of the DDJB,
218 then this could introduce a low bias in the calculated percentages of decline in DDJB CH₄. In
219 addition, there are likely other unknown biases inherent in the approach and methods we have
220 used.

221 We initially chose June-August AIRS data because it coincides with the summer ozone
222 season when VOC precursors from the DDJB have the greatest impact on area ozone
223 concentrations but also because the signal for local CH₄ enhancement was positive and stronger
224 for this time of year. Box plots of cycles in monthly NWR CH₄ and AIRS 700 hPa ascending
225 CH₄ are plotted in Figure 3. NWR has a minimum in the summer and a peak in the winter
226 through spring. Tropospheric concentrations in the northern hemisphere can increase in the
227 winter because of a seasonal decline in hydroxyl radical concentrations. The hydroxyl radical
228 gradually destroys CH₄ in the summer when photochemistry is more active. This is presumably
229 the cause of winter maxima at this site.



230

231 **Figure 3.** Box plots of (a) monthly NWR CH₄ for 2003-2020 and (b) monthly AIRS 700 hPa
 232 ascending node CH₄ for 2003-2020, with interquartile ranges and lines connecting medians. The
 233 larger signal for AIRS data in the warmer months is attributed to greater PBL heights.

234 In contrast to NWR, the AIRS 700 hPa ascending CH₄ peaks in July through October.
 235 The differences in seasonal cycles between NWR and DDJB AIRS 700 hPa CH₄ are likely the
 236 result of seasonal differences in how the PBL concentrates or disperses local CH₄ as well as the
 237 effects of retrieval constraints on the representativeness of the 700 hPa CH₄ product in the PBL.
 238 Kavitha et al. (2018), for example, described a peak AIRS underestimation of measured surface
 239 concentrations in the winter when shallow boundary layers concentrated CH₄ near the surface.
 240 Rakitin et al. (2015) found lower correlations between AIRS CH₄ and surface measurements at
 241 Beijing when mixing heights were below 700 meters AGL, since the satellite spectrometer is less
 242 sensitive to CH₄ when it is concentrated in a shallow surface layer. They also noted that
 243 correlations between surface and satellite measurements for most of their study sites were higher
 244 in the summer when surface mixed layers were deeper.

245 According to maps of ERA5 reanalysis estimates presented by Zhang et al. (2020b), the
 246 mean daytime peak PBL height over the plains in the DDJB in the winter is 800-1,000 meters
 247 above ground level (AGL), and, in the summer it is greater than 2,400 meters AGL. Lower
 248 elevations of the DDJB are at ~ 1,500 meters MSL, and NWR is at 3,523 meters MSL. This
 249 means that NWR CH₄ is likely representative of the free troposphere in the winter and a well-
 250 mixed troposphere with surface influences in the summer. McGrath-Spangler and Denning
 251 (2012) estimated mean June-August afternoon PBL heights for North America using the North
 252 American Regional Reanalysis (NARR) and the NASA MERRA reanalysis. Maps of their
 253 results show that summer afternoon PBL heights were ~3,000 to ~3,500 meters AGL over the
 254 DDJB. They also calculated mean summer afternoon PBL heights from aerosol backscatter
 255 measurements made by the Cloud-Aerosol Lidar and Infrared Pathfinder Satellite Observations
 256 (CALIPSO) satellite. These PBL heights were ~2,500 meters AGL over the DDJB. AIRS is

257 primarily sensitive to CH₄ above 2,000 meters MSL (Kulawik et al., 2021). Consequently, it
258 should be sensitive to CH₄ from surface sources mixed within the deep summer PBL over the
259 DDJB. The altitude of our study area and resulting deep, summer-afternoon, boundary layers
260 (with mean mixing heights between ~4,000 to ~5,000 meters MSL or ~650 to ~550 hPa) should
261 result in an increased sensitivity of AIRS CH₄ to PBL concentrations during the summer months.

262 In contrast to our data in Figure 3b, Oltmans et al. (2021) show a winter peak and
263 summer minimum in CH₄ at the BAO tower in the Wattenberg Field (Figure S3 in their paper).
264 Their analysis for 2008-2015 shows that the median and 95th percentile values are above 1,900
265 ppb and 2,100 ppb, respectively, during the peak month of February, a month with shallower
266 boundary layers. The February mean at NWR for these years is 1872 ppb, and it is only 1831 ppb
267 for the AIRS 700 hPa product in the DDJB. These high concentrations at BAO are clearly
268 associated with surface sources and emissions concentrated in the boundary layer, and the
269 boundary layer in February is likely too shallow for reasonable representation of surface
270 concentrations of CH₄ in the Platte Valley using the AIRS product.

271 To provide additional evidence that the DDJB AIRS 700 hPa ascending CH₄ retrieval is
272 sensitive to PBL heights, we calculated the correlation between monthly AIRS data and monthly
273 PBL heights from the MERRA-2 reanalysis (Global Modeling and Assimilation Office, 2015)
274 for all months from 2012-2020. Describing the MERRA-2 reanalysis product, Gelaro et al.
275 (2017) state that reanalysis “is the process whereby an unchanging data assimilation system is
276 used to provide a consistent reprocessing of meteorological observations... The process relies on
277 an underlying forecast model to combine disparate observations in a physically consistent
278 manner, enabling production of gridded datasets for a broad range of variables, including ones
279 that are sparsely or not directly observed.” The correlation coefficient between DDJB AIRS CH₄
280 and the MERRA-2 PBL heights in the DDJB was 0.46, suggesting that the retrieval signal is
281 stronger when the boundary layer is not shallow. We would expect this correlation to be higher
282 in a comparison using afternoon-only mixed-layer heights. The monthly mean PBL heights (for
283 all hours) in January and July were 540 meters and 1,507 meters AGL, respectively.

284 We did not use descending node AIRS data because local flow regimes are not typically
285 moving DJ Basin emissions into and across the grid cell at night (see Toth and Johnson, 1985),
286 and shallow boundary layers can concentrate CH₄ near the surface. Similarly, we did not use
287 colder season data since the AIRS retrievals might underestimate daytime CH₄ when persistent
288 shallow inversions concentrate CH₄ near the surface.

289 AIRS CH₄ retrieval concentrations for specific levels of the atmosphere are not
290 independent of concentrations at other levels in the atmosphere and are not a true measure of an
291 ambient concentration at a specific level. In the mid-latitudes of the northern hemisphere, the
292 AIRS CH₄ retrievals at 400 hPa have the highest sensitivity to CH₄ concentrations. The AIRS
293 400 hPa CH₄ retrieval concentrations also have the greatest independence from CH₄ at other
294 levels in the atmosphere (Xiong et al., 2015). AIRS 400 hPa CH₄, however, is not as sensitive to
295 surface mixed layer concentrations as retrievals at lower levels of the atmosphere.

296 The amount that each vertical region of the atmosphere contributes to an AIRS retrieval
297 for a standard level is characterized by a curve called an averaging kernel. In their Figure 2,
298 Kulawik et al. (2021) show the averaging kernels for nominal AIRS altitude levels for a tropical

299 location. The shapes of these curves vary by region, but, in general the 700 hPa product level
300 (which is actually designated as 681 hPa) has greater sensitivity to concentrations from 800 hPa
301 to 500 hPa than products for higher up in the atmosphere. Averaging kernel sensitivities to near-
302 surface CH₄ approach zero with increasing altitudes.

303 For all these reasons we selected AIRS CH₄ 700 hPa AIRS product. The AIRS 700 hPa
304 retrievals are representative of a deep layer of the atmosphere and are characterized by the broad
305 averaging kernel. These retrievals are sensitive to near-surface concentrations as well as
306 concentrations in layers extending up to the stratosphere. The summer 700 hPa level data are
307 anticipated to be strongly correlated with concentrations at the surface in the DDJB.

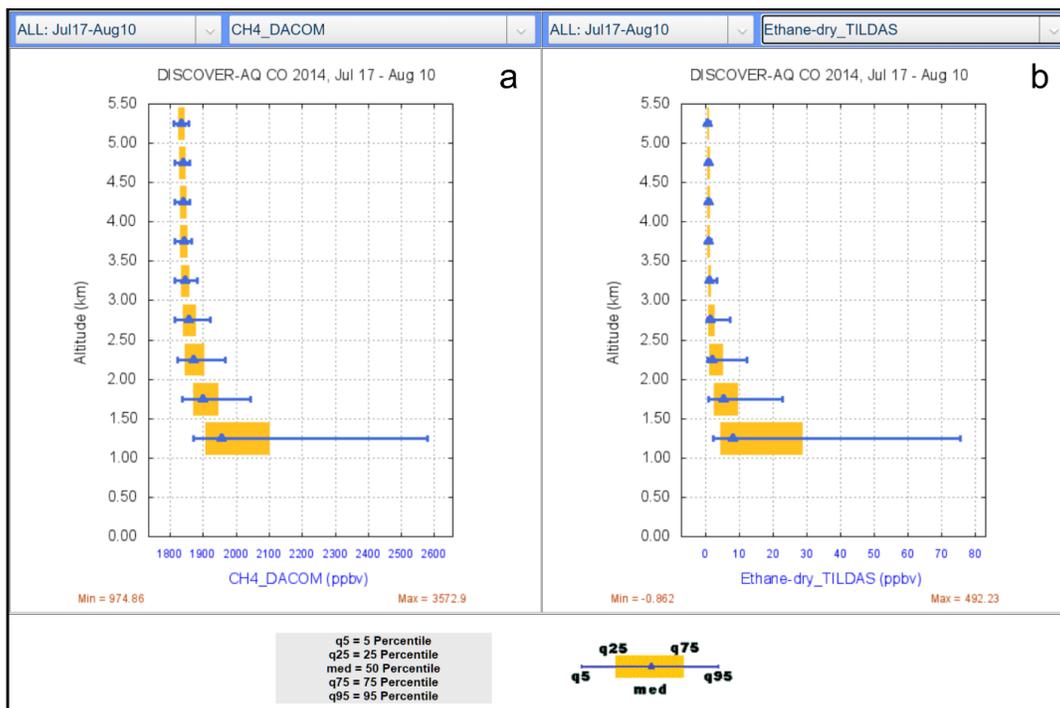
308 **3 Results and Discussion**

309 The sensitivity of AIRS 700 hPa CH₄ to local near-surface conditions was evaluated
310 using aircraft data collected by the DISCOVER-AQ campaign in July and August of 2014.
311 NASA completed 16 flights with repeated vertical spirals over the DDJB. All 6 DISCOVER-AQ
312 vertical spiral or airport missed-approach sites were in the AIRS DDJB grid cell, except for the
313 Ft Collins spiral site. Roughly 50% of the flight samples from the surface to 3,000 meters were
314 collected between 10:30 and 15:15 MST. Figure 4 presents statistics for the DISCOVER-AQ
315 flight measurements of CH₄ and ethane for 500-meter altitude bins. These represent all flight
316 hours and not just the hour of peak PBL height which is generally 13 MST (Zhang et al., 2020b).
317 The median CH₄ concentration for the bin centered at the 1,750-meter MSL level was ~1,900
318 ppb which is the closest of all the altitude bin medians to the July-August 2014 average AIRS
319 700 hPa CH₄ of 1,886 ppb. This suggests that the AIRS 700 hPa retrieval concentrations were
320 sensitive to CH₄ in the mixed layer and can be used to track changes associated with emissions
321 into the boundary layer.

322 The representativeness of NWR CH₄ concentrations for background conditions is further
323 supported by comparison of NWR CH₄ to aircraft CH₄ measurements made during DISCOVER-
324 AQ. In Figure 4, CH₄ above 3,000 meters drops to background levels and ethane approaches zero
325 indicating that the 3,000 to 5000-meter level was generally free from strong influences from
326 boundary layer emissions from O&G sources. The median CH₄ concentrations measured by
327 DISCOVER-AQ from 3,000 to 5,000 meters MSL are ~1830-1840 ppb, and these are
328 comparable to the NWR July-August 2014 average CH₄ concentration of 1,852 ppb at 3,523
329 meters MSL. This is additional evidence that NWR is a reasonable choice for background CH₄ in
330 our study.

331 Three-hour HYSPLIT back trajectories were obtained for the highest 20 percentile, 40-60
332 percentile, and lowest 20 percentile Platteville ethane concentration samples for all of 2018 using
333 40 km Eta Data Assimilation System (EDAS) meteorological fields
334 (<https://www.ready.noaa.gov/edas40.php>). These coincided with the 6-9 AM MST sample
335 window for each daily measurement. These show that the Platteville site is representative of
336 emissions and ambient concentrations in the southern Wattenberg field. Figure 5 presents a map
337 of the 3-hour back trajectory points for 6-9 AM MST which indicates that Platteville ethane and
338 VOC samples are representative of an area that extends far beyond the local footprint of the
339 monitoring site. This is clear evidence that even under the reduced surface wind speed and
340 vertical mixing regimes expected in the early morning when samples are collected, the Platteville

341 site is representative of emissions across a wide area of the DJ Basin between Denver and
 342 Greeley.

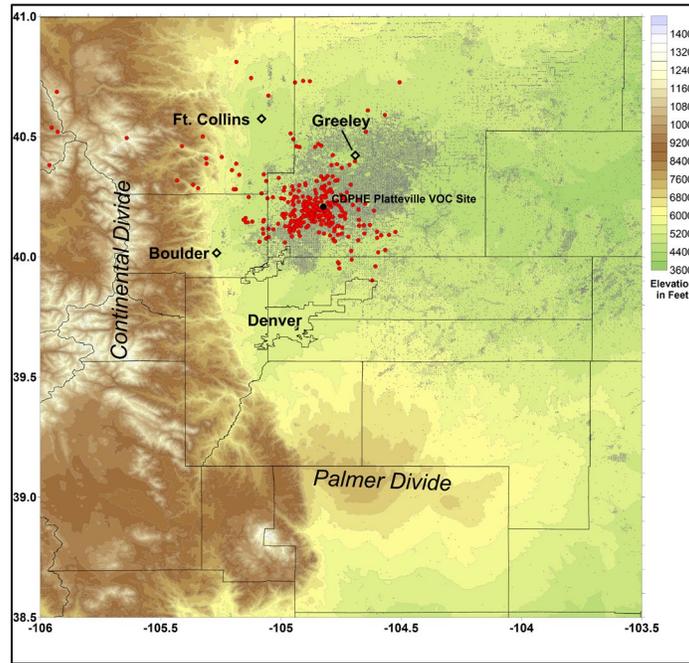


343
 344 **Figure 4.** Percentile plots of CH₄ (left) and ethane (right) for all flights and spirals of the
 345 DISCOVER-AQ P3-B in July and August 2014. Blue triangles are medians, orange bars
 346 represent the 25th through 75th percentile range, and horizontal blue lines represent the 5th
 347 through 95th percentile range. The 1.25 km level represents missed runway approaches just
 348 above ground level near Greeley.

349 We acquired plots of ethane concentrations and winds from the experimental forecasting
 350 system that is run by the National Center for Atmospheric Research (NCAR) and based on the
 351 Weather Research and Forecasting (WRF) model with chemistry (WRF-Chem) (Kumar et al.,
 352 2020) (<https://www.acom.ucar.edu/firex-aq/forecast.shtml>). Plots are for 5:00 MST on 30
 353 January 2022 and 23 July 2021. These are shown in Figure 6. Elevated ethane concentrations are
 354 largely from O&G sources. Toth and Johnson (1985) and Neff (1997) describe the features of
 355 early morning slope and drainage flows as they move eastward from the foothills west of Denver
 356 and northward from the Palmer Divide which is south of Denver and converge as southerly
 357 drainage flows along the Platte River Valley from Denver to Platteville. Drainage winds
 358 eventually curve eastward following the river valley east of Greeley. These conditions are typical
 359 when stronger synoptic-scale winds are not present.

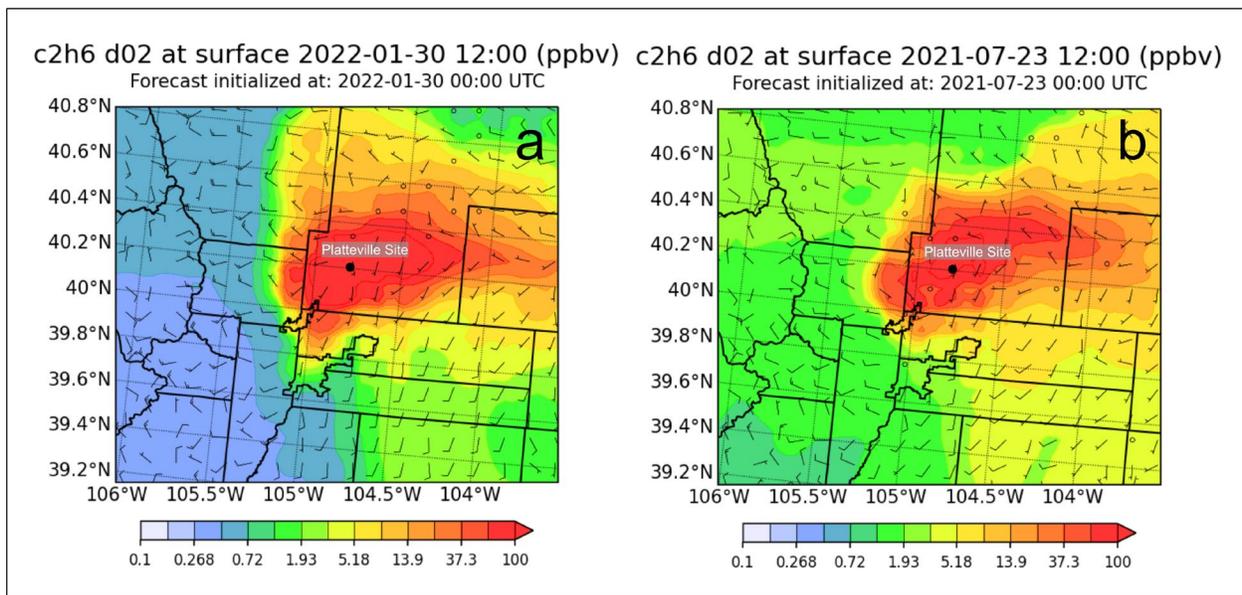
360 Modeled ethane concentrations and winds capture these flows and their effects on ethane
 361 and show that ethane pools in the area surrounding Platteville and the O&G field. In addition,
 362 Pfister et al. (2017) show average modeled O&G tracer concentrations for 6:00-12:00 MST
 363 during the 2014 Front Range Air Pollution and Photochemistry Experiment in Figure 8 of their
 364 paper. The O&G tracer plot places Platteville in the center of a large area with fairly uniform
 365 high concentrations, and the spatial distribution is comparable to the fairly uniform high-

366 concentration footprints in Figure 6. Platteville is suitably located to sample ethane and other
 367 O&G trace gases within the O&G field during the morning hours.



368

369 **Figure 5.** HYSPLIT three-hour back trajectory points (shown as red dots) for 6-9 AM MST for
 370 the CDPHE Platteville monitor in 2018. Oil and gas wells are plotted as grey dots.



371

372 **Figure 6.** NCAR WRF-Chem surface wind and ethane concentration forecasts for (a) 5:00 MST
 373 30 January 2022 and (b) 5:00 MST 23 July 2021. These show the effects of typical morning
 374 slope and drainage flows that pool ethane near Platteville and transport ethane eastward near
 375 Greeley following the curve in the Platte River valley. The Platteville location and label have
 376 been added.

377 Annual ethane means, medians, and statistics for 2012-2020 are presented for Platteville
 378 and Denver CAMP in Tables 1 and 2, respectively. We used annual data for increased sample
 379 sizes and because summer data for 2016 were limited. Table 3 lists the June-August mean DDJB
 380 AIRS 700 hPa CH₄, the number of AIRS retrievals, the June-August NWR CH₄, the calculated
 381 DDJB enhancement in CH₄, and Platteville annual median ethane. We used the median ethane
 382 data in this case to reduce the influence of high values associated with shallow surface mixed
 383 layers. Table 3 shows that the estimated DDJB CH₄ enhancement decreased by 55% between
 384 2013 and 2020. Platteville median ethane decreased by 73% during the same period. The choice
 385 of 2013 as the starting year was based on the peak enhancement in 2013. Since there is the
 386 possibility of influences from unexplained cofactors each year, the true enhancement peak might
 387 have occurred in either 2012 or 2013. Later we show a slightly more conservative estimate for
 388 CH₄ based on a linear regression for 2012-2020, and this reduction is 56%. Similarly, the
 389 reduction in median Platteville ethane based on the linear regression for 2012-2020 was 77%.

390 **Table 1.** Platteville Ethane Annual Means, Medians, and Statistics for 2012-2020.

Year	Mean ethane (ppb)	Median ethane (ppb)	Standard deviation (ppb)	95% confidence level (ppb)	Number of samples
2012	213.9	196	155.4	32.4	91
2013	277.4	212	250.5	70.5	51
2014	187.5	132	222.0	59.4	56
2015	155.0	143	108.9	31.6	48
2016	155.3	119	117.5	30.9	58
2017	177.3	79.6	358.4	91.8	61
2018	100.7	87.2	77.8	19.9	61
2019	97.5	65.3	90.8	23.5	60
2020	81.2	57.3	62.4	16.9	55

391 **Table 2.** Denver CAMP Ethane Annual Means, Medians, and Statistics for 2012-2020.

Year	Mean ethane (ppb)	Median ethane (ppb)	Standard deviation (ppb)	95% Confidence level (ppb)	Number of samples
2012	23.5	18.6	15.3	3.2	88
2013	29.5	20.9	23.4	6.2	57
2014	24.3	18.5	19.0	5.0	58
2015	27.8	21.3	18.3	5.2	50
2016	34.9	23.8	27.0	7.0	60
2017	28.0	23.5	19.1	4.9	61
2018	32.2	24.0	24.5	6.3	61
2019	32.5	23.0	27.8	7.1	61
2020	33.6	23.6	24.4	7.8	40

392

393

394 **Table 3.** AIRS DDJB CH₄, NWR CH₄, DDJB Enhancement, and Platteville Ethane for 2003-
395 2020.

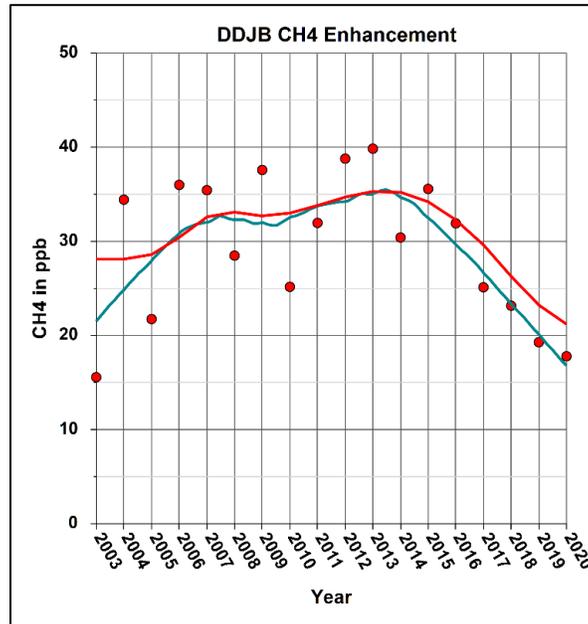
Year	Mean June-August AIRS CH ₄ DDJB (ppb)	Number of AIRS retrievals June-August	Mean June-August Niwot Ridge CH ₄ (ppb)	DDJB CH ₄ minus Niwot Ridge CH ₄ (ppb)	Platteville annual median ethane (ppb)
2003	1829.7	76	1814.1	15.6	
2004	1840.3	73	1805.9	34.4	
2005	1841.6	72	1819.8	21.7	
2006	1841.6	76	1805.6	36.0	
2007	1848.8	71	1813.3	35.4	
2008	1854.9	73	1826.4	28.5	
2009	1864.2	78	1826.6	37.6	
2010	1860.1	68	1834.9	25.2	
2011	1871.5	68	1839.5	32.0	
2012	1877.8	71	1839.0	38.8	196
2013	1881.4	74	1841.6	39.8	212
2014	1884.0	76	1853.6	30.4	132
2015	1896.0	75	1860.4	35.6	143
2016	1904.3	71	1872.4	31.9	119
2017	1907.8	72	1882.7	25.1	79.6
2018	1920.0	67	1896.8	23.2	87.2
2019	1916.9	75	1897.6	19.3	65.3
2020	1927.8	60	1910.1	17.8	57.3
Percent Reduction 2013 to 2020				55%	73%

396 Peischl et al. (2018) and Petron et al. (2014) conclude that O&G sources account for 75%
397 of local CH₄ sources. Consequently this 55-56% reduction may underestimate the percent
398 reduction in CH₄ from O&G sources. It is reasonable that the percent reduction in ethane was
399 greater than the percent reduction in CH₄, since O&G sources account for almost all the ethane at
400 Platteville.

401 In addition, CDPHE estimates that O&G systems accounted for 59% of CH₄ statewide in
402 2019 (CDPHE, 2021, Exhibit 1-5). While emissions estimates are not specific to the DDJB grid
403 cell, only the statewide O&G sector is estimated to have meaningful emissions reductions from
404 2010 to 2019 (CDPHE, 2021). All other sectors have CH₄ emissions changes that are either
405 negligible (e.g., residential, commercial, industrial, transportation, and electric power), slightly
406 increasing (e.g., agriculture and waste management), or are not emitting CH₄ in the DDJB grid
407 cell (e.g., coal mining). Consequently, it is reasonable to assume that the trends in the DDJB CH₄
408 enhancement are predominantly affected by changes within the O&G sector.

409 Figure 7 shows the estimated long-term trends in DDJB CH₄ enhancement based on
410 AIRS measurements in the DDJB grid cell. NWR flask samples are collected weekly for a total
411 of ~13 samples for June-August, while an average of 72 AIRS retrievals are available for each
412 June-August period. In many cases the sample and retrieval days do not match, and the means of

413 each dataset can represent different conditions of continental background and transport for
 414 Colorado. This could introduce noise in the DDJB CH₄ enhancement time series. Consequently,
 415 a four-point median smoother and a LOESS curve (Cleveland, 1979) have been fitted to the data
 416 to show the general trends throughout the entire period.



417

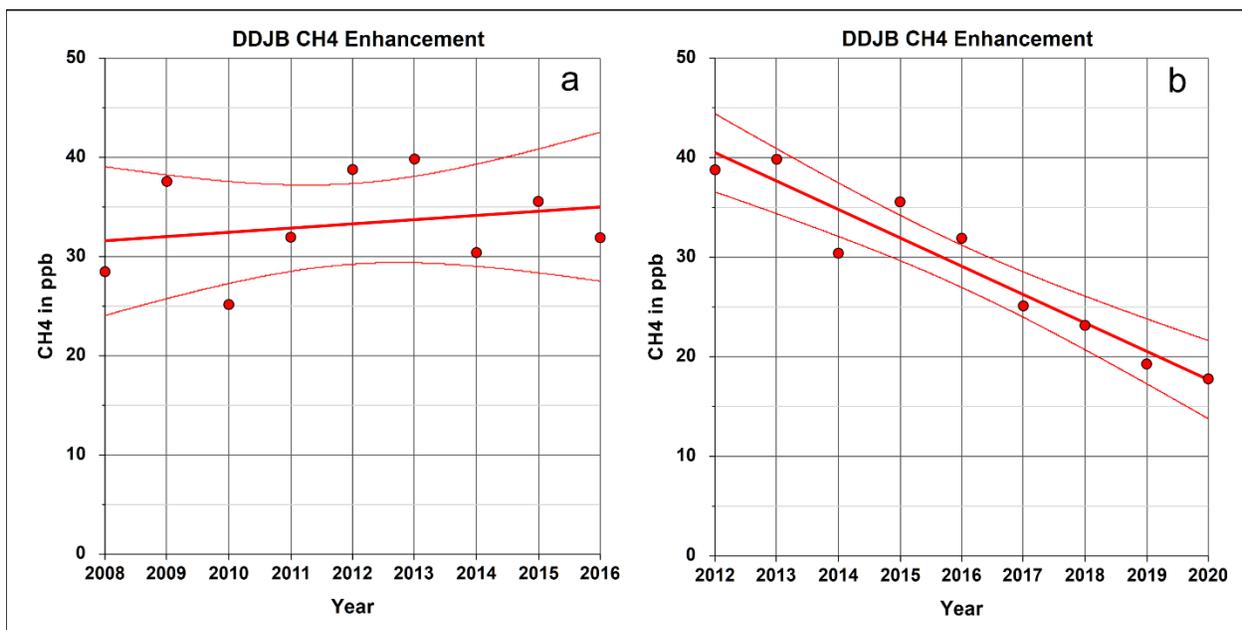
418 **Figure 7.** DDJB June-August CH₄ enhancement from 2003 through 2020 fitted with a 4-point
 419 median smoother (red) and a LOESS curve with a 0.4 smoothing parameter (green).

420 A gradual increase which may be associated with growth in O&G extraction activities is
 421 followed by a steep downward trend after 2013, and we attribute the decline to DJ Basin O&G
 422 emissions reductions. A linear regression through this data for 2003-2012 yields a statistically
 423 insignificant trend of 1.25 +/- 1.8 ppb. Oltmans et al. (2021) found no significant trend in BAO
 424 CH₄ near Erie in the southwest portion of the Wattenberg from 2008 through mid-2016. We also
 425 found no significant trend for the DDJB AIRS enhancement from 2008 through 2016. A plot of
 426 our results for this period is presented in Figure 8(a). When we consider 2012-2020 (see Figure
 427 8(b)), however, we find that there is a nearly linear and statistically significant downward trend
 428 with an R² of 0.90. CH₄ declined in the DDJB grid cell at a rate of -2.84 +/- 0.8 ppb (based on
 429 the 95% confidence interval) per year. In addition, a linear regression of the ratio of DDJB CH₄
 430 to NWR CH₄ against year (which is independent of local enhancement estimates) shows a
 431 decline from 1.022 to 1.008 during this period (R² of 0.91).

432 The trends in 2012-2020 annual median Platteville and Denver CAMP ethane (see Figure
 433 9) show a statistically significant linear decrease in Platteville ethane of -19.1 ppb +/- 5.9 ppb per
 434 year (based on the 95% confidence interval, R² of 0.90) and a statistically significant linear
 435 increase in Denver CAMP ethane of 0.7 ppb +/- 0.4 ppb per year (based on the 95% confidence
 436 interval, R² of 0.68). Lyu et al. (2021) report a 33%-56% contribution from O&G sources to 6-9
 437 AM MST Denver CAMP NMHC. Ortega et al. (2021) found a positive trend of 0.9 ppb ± 0.3 %
 438 per year for ethane north of Boulder from 2010-2019. Denver and Boulder ethane have been

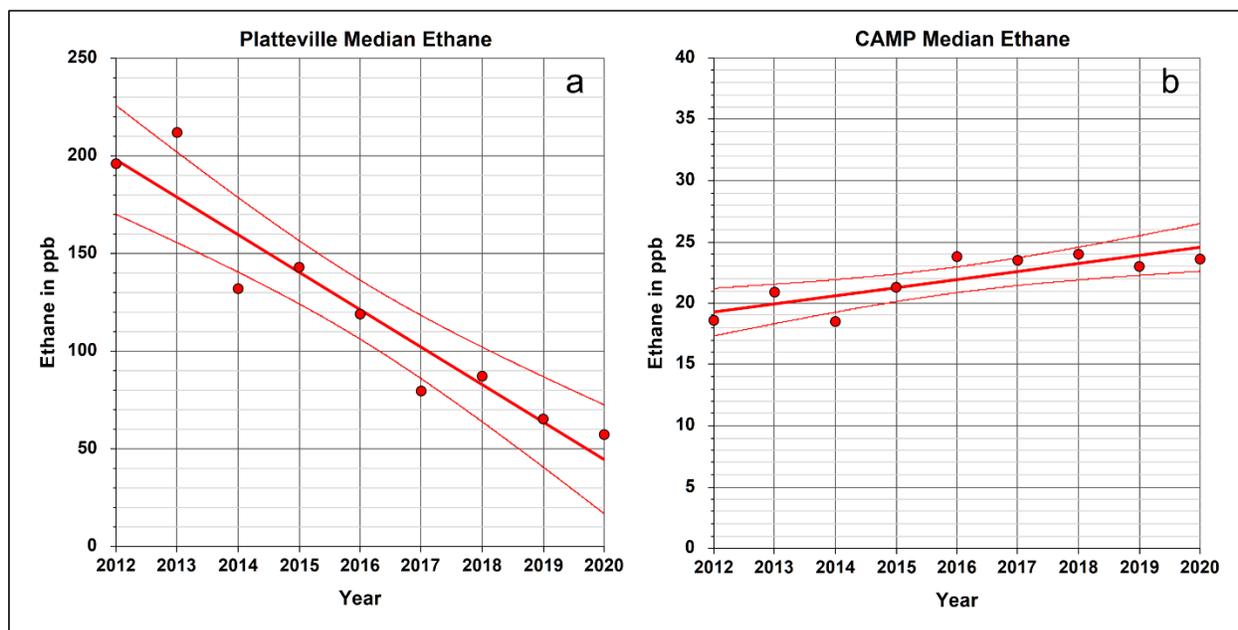
439 increasing at a similar rate, although the rate for the former is based on morning-only
 440 measurements.

441 Rossabi et al. (2021) analyzed west-east gradients in CH₄ at several sites for 9 to 11
 442 three-day periods during July and August of 2014. They found a high Pearson correlation
 443 coefficient of 0.77 between CH₄ and ethane in eastern Boulder County and concluded that it is
 444 highly likely that natural gas extraction activities are the dominant source of ethane in the region.
 445 Their sites extended from the higher terrain west of Boulder, across Boulder, and into the
 446 southwestern edge of the Wattenberg Field. Figure 3 of their study shows that CH₄ at the
 447 Boulder County Public Health site in Boulder was ~ 6 ppb higher than estimated background.
 448 Their Figure 3 also shows that CH₄ at two sites in or near the southwestern edge of the
 449 Wattenberg Field was ~12 ppb higher than background. Their results, other work we have cited,
 450 and our results for ethane at Platteville and Denver show steep spatial gradients in O&G impacts
 451 across the region. Nominal increases in ethane at Boulder and Denver are not necessarily
 452 inconsistent with steep declines at Platteville, which is in the center of the Wattenberg Field; and
 453 we have concluded that the Platteville site is representative of the Wattenberg, based on
 454 HYSPLIT back trajectories and cited studies.



455

456 **Figure 8.** DDJB June-August CH₄ enhancement versus year with linear regression and 95%
 457 confidence limits for (a) 2008-2016 and (b) 2012-2020. No statistically significant trend was
 458 found for CH₄ from 2008-2016 at BAO (Oltmans et al., 2021) or in our analyses for the DDJB,
 459 but a robust downward trend is evident for the DDJB from 2012-2020.

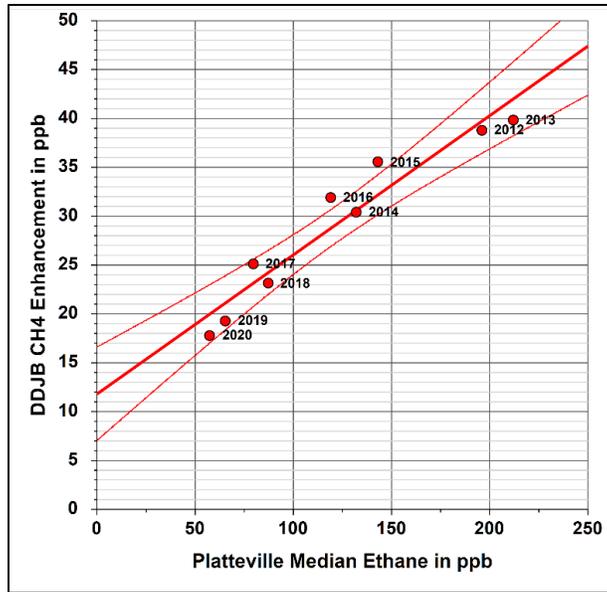


460

461 **Figure 9.** Linear regression and 95% confidence limits for (a) annual median Platteville ethane
 462 and (b) annual median Denver CAMP ethane, versus year for 2012-2020. Trends for both sites
 463 are statistically significant.

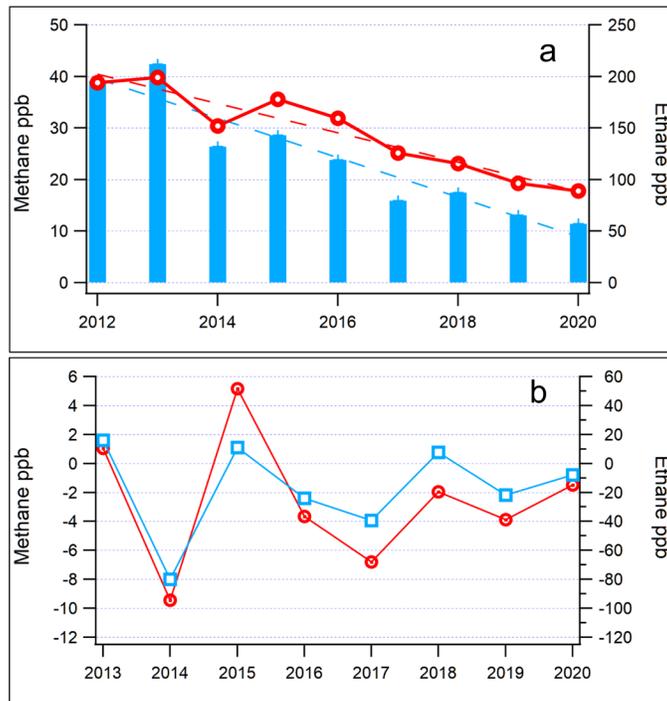
464 A regression between the June-August DDJB CH₄ enhancement and annual median
 465 Platteville ethane are plotted in Figure 10. The regression is statistically significant and has an R²
 466 of 0.93, and the Pearson correlation coefficient is 0.96. If Platteville ethane which is largely from
 467 O&G sources were reduced to zero, the regression in Figure 10 would yield a remaining CH₄
 468 enhancement of 12 ppb or 30% of the 2012 regression value in Figure 8(b). This is comparable
 469 to the 25% local contribution from non-O&G sources in 2012 estimated by Petron et al. (2014).

470 Figure 11 shows the time series of DDJB CH₄ enhancement and Platteville ethane and the
 471 year-to-year changes in each. Results in Figures 10 and 11 provide compelling evidence that the
 472 calculated DDJB CH₄ enhancement is sensitive to and tracks with surface boundary layer
 473 concentrations. These results also show that DDJB CH₄ has been responsive to reductions in
 474 O&G emissions within the grid cell. Recall that Rossabi et al. (2021) concluded that it is highly
 475 likely that natural gas extraction activities are the dominant source of ethane in the region. In
 476 addition, using positive matrix factorization and chemical mass balance methods, Lyu et al.
 477 (2021) found that O&G activities accounted for 92%-96% of NMHC at Platteville for 2013-
 478 2016. The strong correlation and the fact that there have not been substantial changes in ethane
 479 in Denver or Boulder suggest that reductions in O&G emissions in the Wattenberg are driving
 480 the observed declines in AIRS CH₄ within the DDJB. The correlation between DDJB CH₄
 481 enhancement and median Denver ethane is -0.66.



482

483 **Figure 10.** Linear regression between June-August mean 2012-2020 DDJB CH₄ enhancement
 484 and annual median Platteville ethane (with plotted 95% confidence limits). The fit is robust with
 485 an R² of 0.93.



486

487 **Figure 11.** Trends in June-August DDJB CH₄ enhancement and annual median Platteville ethane
 488 in ppb. Panel (a) illustrates June-August mean 2012-2020 DDJB CH₄ enhancement in red with
 489 open circles and annual median Platteville ethane shown with blue bars (with plotted regression
 490 lines) and panel (b) shows year-to-year changes in these, with CH₄ in red with open circles and
 491 ethane in blue with open squares.

492 Feng et al. (2022) find that interannual variability in meteorology (especially wind
493 speeds) plays a significant role in trends in CH₄ in the US. Reddy and Pfister (2016)
494 demonstrated that maximum surface ozone concentrations in and near the DDJB are highly
495 correlated with 500 hPa heights. They state that “upper level ridges decrease winds and allow
496 cyclic terrain-driven circulations to reduce transport away from sources”. We obtained mean
497 June-August 500 hPa heights from the NCEP Reanalysis for a grid cell centered on 40°N
498 105°W, and this contains the DDJB. The correlation coefficients for interannual differences in
499 Platteville ethane, DDJB CH₄ enhancement, and NWR CH₄ versus the interannual differences in
500 500 hPa heights were calculated for 2013-2020; and these are 0.69, 0.74, and 0.26 for Platteville,
501 the DDJB, and NWR, respectively. Even though we used annual data for ethane, it appears that
502 the impacts of summer meteorology are strong enough to have a significant effect on variations
503 in yearly values. The correlation is very high for DDJB CH₄ enhancement and low for NWR
504 CH₄. While reductions of emissions in the DDJB are likely responsible for the overall trends, the
505 more subtle variations in Figure 11 are probably caused, at least in part, by year-to-year changes
506 in meteorology. The strong correlations with weather for both ethane and CH₄ enhancement
507 support the hypothesis that the AIRS product is detecting real changes in year-to-year CH₄
508 within the PBL.

509 Oltmans et al. (2021) estimated background propane, a tracer for O&G emissions, at the
510 BAO tower near Erie in the DJ Basin and subtracted this from ambient data. Using only data
511 from 2008-2015 and scaling the annual changes in local contribution to 2011 data, they projected
512 a -1.5% change per year in propane through 2021 and compared this with the RAQC’s projected
513 emissions inventory reductions of -6.5% per year. We have not estimated background ethane,
514 also an O&G tracer, at the Platteville monitor, but we used the regression statistics for median
515 Platteville ethane versus year and scaled the -19.1ppb slope by the 2012 regression value of
516 197.8 ppb to obtain an average annual decline of -9.7% from 2012 through 2020 (with a range of
517 -12.6% to -6.7%, based on the lower and upper 95% confidence intervals for the slope).

518 This rate of decline is not based on data from an earlier period but on an analysis of data
519 through 2020, and the rate of decline of the emissions inventory is just above the upper limit of
520 the calculated range for 2012-2020 at Platteville. Trends in both Platteville ethane and the DDJB
521 CH₄ enhancement provide evidence that the regulatory requirements reflected in the emissions
522 inventories are having a significant effect.

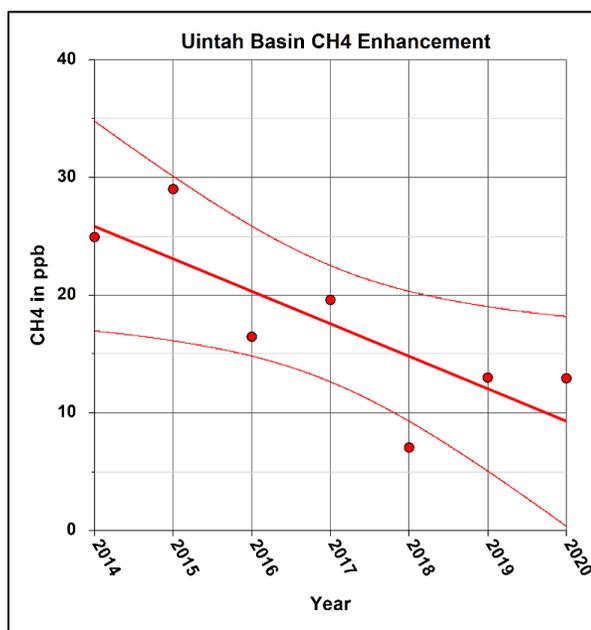
523 The focus of this paper is CH₄ trends within the DDJB, but to further verify our method
524 and the ability of the AIRS 700 hPa CH₄ product to detect warm-season trends in the high-
525 altitude western US, we completed a preliminary analysis of trends in CH₄ enhancements for the
526 Uintah Basin O&G fields in Utah, a major O&G production region in the western US. The floor
527 of the basin is at ~1,500 meters MSL, and summer mean peak PBL heights are comparable to
528 those in the DDJB (McGrath-Spangler and Denning, 2012; Zhang et al., 2020b). We used the
529 standard AIRS version 7 product for the two AIRS grid cells that contain the basin (see
530 supplementary Figure S1). We calculated June-August mean AIRS CH₄ and subtracted the mean
531 June-August NWR CH₄ data calculated for our study.

532 The resulting trend for the Uintah Basin CH₄ enhancement for 2014-2020 is plotted in
533 Figure 12. Lin et al. (2021) show that basin natural gas production increased rapidly, peaking in
534 2012 and 2013, and then declined rapidly. The number of producing gas wells peaked in 2015

535 and then declined rapidly. They also show that oil production and the number of oil wells peaked
 536 in 2014-2015, and then declined unevenly after this. They found that decreases in CH₄ after 2015
 537 were related to declining O&G production and not to a reduction in CH₄ leaks. We show a
 538 statistically significant linear decrease of -2.77 ppb +/- 2.47 ppb (R² of 0.62) from 2014-2020.

539 Lin et al. (2021) calculated basin CH₄ enhancement by taking the difference between
 540 surface concentrations at a site in the center of the O&G field and one on the far western edge of
 541 the basin. For a trend analysis with linear regression, they used afternoon-only concentrations for
 542 April-September to avoid cold pooling events that concentrate CH₄ in shallow surface layers.
 543 Using monthly means, they found a decreasing trend of -4.63 ppb with an R² of 0.241 for 2015-
 544 2020. Their enhancements (from the regression line) range between ~20 ppb to ~45 ppb, whereas
 545 the AIRS-based summer season enhancements are from 7 ppb to 29 ppb.

546 Using a combination of measurements and modeling, Lin et al. (2021) report that
 547 emissions decreased by approximately 50% during this period, and we calculate a 60% reduction
 548 in CH₄. This percentage was derived from the regression values for 2015 and 2020. There is
 549 reasonable agreement between the AIRS-based results and theirs. This is further evidence that
 550 the AIRS 700 hPa CH₄ product can be used to detect changes consistent with surface
 551 concentrations in and near a major O&G production area for periods and locations with deep
 552 vertical mixing.



553

554 **Figure 12.** AIRS 700 hPa CH₄ enhancement for the Uintah Basin in Utah with a linear
 555 regression and 95% confidence limits for 2014-2020. The fit is statistically significant with an R²
 556 of 0.62.

557

558 4 Conclusions

559 We selected a one-degree NASA AIRS grid cell centered at 40°N and 105°W in northern
560 Colorado to assess changes in CH₄ from 2003 through 2020. This grid cell includes Denver,
561 Boulder, and Greeley, Colorado, and much of the Wattenberg O&G field in the Denver
562 Julesburg O&G basin. O&G emissions from this area are a significant source of CH₄, ethane, and
563 VOC precursors to ozone. Colorado implemented regulations to significantly reduce emissions
564 from O&G sources in early 2014 and has implemented additional emissions control requirements
565 since (Colorado Regulation No. 3, No. 7, and No. 22, [https://cdphe.colorado.gov/aqcc-](https://cdphe.colorado.gov/aqcc-regulations)
566 [regulations](https://cdphe.colorado.gov/aqcc-regulations)). Our study was designed to use AIRS CH₄ to assess the effects of these controls (and
567 operational changes implemented by the industry voluntarily) on ambient CH₄ and to determine
568 whether this approach could be used to track the effectiveness of emissions controls. Surface
569 measurements alone have limited spatial coverage compared with satellite instruments. Aircraft
570 studies are typically limited in temporal coverage. Newer and more advanced satellite
571 instruments such as TROPOMI do not provide data prior to about 2016-2017. Our study shows
572 that AIRS data can serve as a bridge to more robust analyses based on TROPOMI and other
573 satellite instruments that will soon become operational.

574 We computed mean June-August ascending node AIRS 700 hPa CH₄ for each year and
575 subtracted June-August mean NWR CH₄. NWR is a high-altitude surface monitoring site
576 immediately west of our grid cell. The differences represent the enhancement over background
577 within the grid cell. Vertical profiles of CH₄ from aircraft flights within the grid cell during the
578 July-August 2014 DISCOVER-AQ field campaign were used to verify the sensitivity of the 700
579 hPa AIRS product to near-surface concentrations and to confirm that the Niwot Ridge site is a
580 reasonable choice for background concentrations.

581 We find 55% and 73% reductions in DDJB CH₄ enhancement and Platteville ethane,
582 respectively, between 2013 and 2020 and a 56% reduction in DDJB CH₄ enhancement using a
583 linear regression for 2012-2020. These significant reductions occurred even though oil
584 production increased by 343% and gas production by 297% from 2012-2020. The reduction in
585 CH₄ is comparable to the projected 57% reduction in O&G VOC emissions for 2011-2020 from
586 the 2020 draft ozone State Implementation Plan for the ozone nonattainment area that includes
587 the DDJB (RAQC and CDPHE, 2020). We are unable to calculate biases or uncertainties for the
588 55-56% CH₄ reduction, but our estimates of CH₄ enhancements for the Uintah basin suggest that
589 biases inherent in this approach may be minimal.

590 Smoothing curves applied to the grid cell enhancement show a gradual increase in CH₄
591 through 2013 followed by a noticeable decline. A linear regression of grid cell enhancement
592 versus year shows a 2.84 ppb +/- 0.8 ppb decrease per year from 2012 through 2020 with an R²
593 value of 0.90. Grid cell enhancement is strongly correlated with Platteville ethane. Platteville is
594 in the center of the O&G fields. Using HYSPLIT back trajectories for 2018, we demonstrated
595 that this site is representative of a large area of the Wattenberg field, a major source region for
596 CH₄. The correlation coefficient between grid cell CH₄ enhancement and Platteville ethane was
597 0.96, and the R² for a linear regression between these was 0.93. Platteville ethane dropped from
598 196.0 ppb in 2012 to 57.3 ppb in 2020. Based on a linear regression, we calculate an annual rate
599 of change of -9.7% relative to 2012 (with a range of -12.6% to -6.7%). Denver median ethane

600 concentrations ranged from 18.5 to 24.0 ppb during the period with a slight increasing trend of
601 0.7 ppb +/- 0.4 ppb per year.

602 A preliminary analysis of trends in CH₄ enhancement for the Uintah Basin O&G
603 production area based on AIRS 700 hPa CH₄ and NWR data compares well with the decline of
604 surface CH₄ described by Lin et al. (2021). This is additional evidence that the AIRS product can
605 detect changes directly related to surface concentrations for periods and locations with deep
606 vertical mixing.

607 The grid cell CH₄ enhancement should be considered an estimate that is representative of
608 and proportional to true enhancement above background within the boundary layer. The strong
609 correlation between grid cell CH₄ enhancement and Platteville ethane and the relative stability of
610 Denver and Boulder ethane concentrations over time suggest that reductions in O&G emissions
611 are driving the reductions in CH₄ that are evident in the AIRS CH₄ product. We conclude that
612 AIRS CH₄ can be used to measure the efficacy of O&G emissions control programs in this
613 region and that these controls, as well as other voluntary design changes such as tankless systems
614 and process centralization, are having an effect. Further intercomparisons between AIRS and
615 TROPOMI CH₄ trends for the years of overlapping records would help scientists quantify
616 percent reductions in CH₄ in recent years and understand the biases and uncertainties in these
617 estimates. A comprehensive picture of the effects of emissions controls will require continued
618 analyses of surface, aircraft, and satellite measurements for this region. Evidence and results
619 presented here also suggest that the AIRS CH₄ data could be useful in tracking progress in
620 achieving DDJB greenhouse gas emissions reductions.

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628 website, NCEP Reanalysis data from the NOAA Physical Sciences Laboratory website, monthly
629 CH₄ data from the NOAA Global Monitoring Laboratory website, surface ethane data from the
630 CDPHE repository, vertical chemical profiles from the NASA DISCOVER-AQ program, and
631 online HYSPLIT back trajectories from the NOAA Air Resources Laboratory website. We are
632 grateful for access to WRF-Chem forecast plots provided online by NCAR which is funded by
633 the National Science Foundation (NSF).

634 **Open Research**

635 The primary datasets used in this study include AIRS and MERRA-2 reanalysis data from the
636 NASA Giovanni website, NCEP Reanalysis data from the NOAA Physical Sciences Laboratory,
637 ethane monitoring data from the CDPHE Air Pollution Control Division, and monthly CH₄ data
638 for Niwot Ridge from the NOAA Global Monitoring Laboratory. Key datasets are available in a
639 data repository (<https://doi.org/10.5281/zenodo.7038756>). Statistical calculations were
640
641

642 completed with NCSS 8.0 and Microsoft Excel. Graphs, plots, and maps were prepared with
 643 NCSS 8.0, Surfer 12, Excel, and Igor Pro.

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Supporting Information for

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**Downward Trend in Methane Detected in a Northern Colorado Oil and Gas
Production Region Using AIRS Satellite Data**

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P. J. Reddy¹, C. Taylor²

855

¹Independent Research Scientist, Saguache County, Colorado

856

²Ramboll US Consulting, Fort Collins, Colorado

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858 **Contents of this file**

859

Text S1

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Figure S1

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862 **Introduction**

863

This document contains supplementary figure S1 which shows the location of the standard AIRS version 7 product grid cells that contain the Uintah Basin oil and gas production region in Utah.

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866 **S1. The AIRS Version 7 Grid Cells Over the Uintah Basin**

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AIRS Version 7 grid cells were obtained from the NASA Giovanni website (<https://giovanni.gsfc.nasa.gov/giovanni/>). Details of this product can be found here: https://disc.gsfc.nasa.gov/datasets/AIRS3STM_7.0/summary. The two cells are 1° by 1° in size and have been joined in Figure S1. One grid cell is centered on 40°N, 110°W, and the other is centered on 40°N, 109°W.

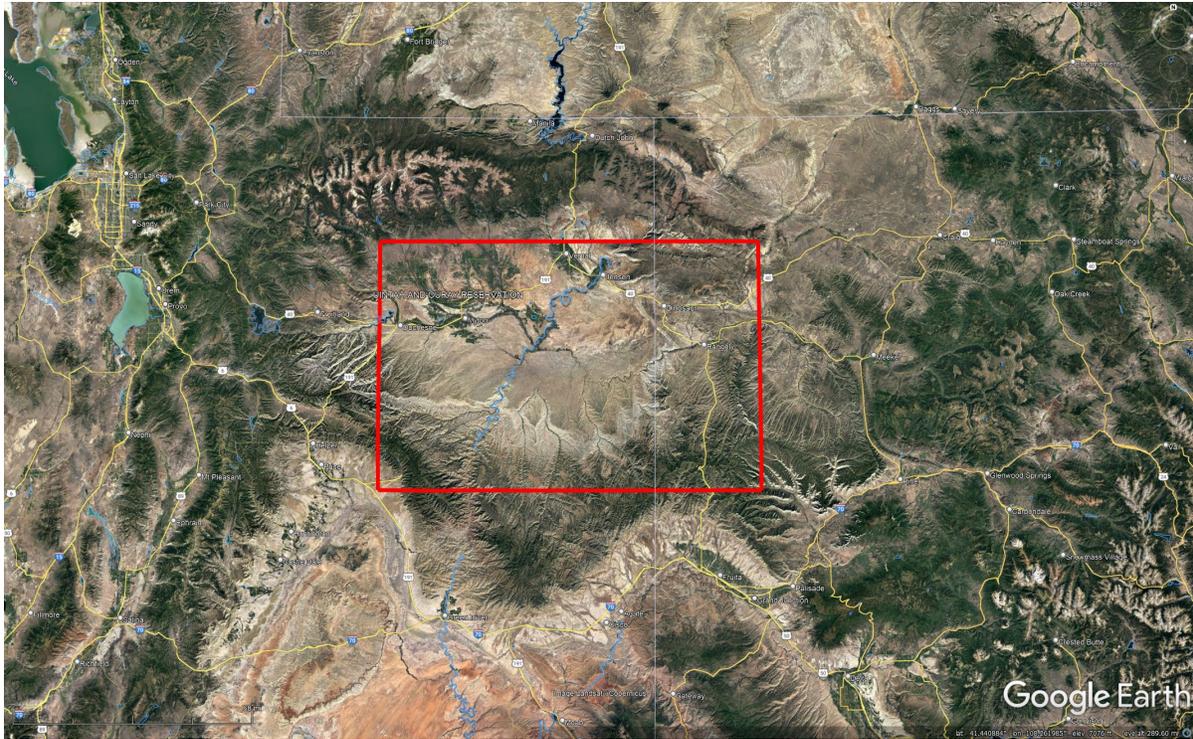
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874 **Figure S1.** Merged AIRS grid cells covering the Uintah Basin in Utah. The map source is Google
875 Earth.

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