# Agent-based modelling of alternative futures in the British land use system

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#### Abstract

Socio-economic scenarios such as the Shared Socioeconomic Pathways (SSPs) have been widely used to analyse global change impacts, but representing their diversity is a challenge for the analytical tools applied to them. Taking Great Britain as an example, we represent a set of stakeholder-elaborated UK-SSP scenarios, linked to climate change scenarios (Representative Concentration Pathways), in a globally-embedded agent-based modelling framework. We find that distinct model components are required to account for divergent behavioural, social and societal conditions in the SSPs, and that these have dramatic impacts on land system outcomes. From strong social networks and environmental sustainability in SSP1 to land consolidation and technological intensification in SSP5, scenario-specific model designs vary widely from one another and from present-day conditions. Changes in social and human capitals can generate impacts larger than those of technological and economic change, and comparable to those of modelled climate change. We develop an open-access, transferrable model framework and provide UK-SSP projections to 2080 at 1km2 resolution, revealing large differences in land management intensities, provision of a range of ecosystem services, and the knowledge and motivations underlying land manager decision-making. These differences suggest the existence of large but underappreciated areas of scenario space, within which novel options for land system sustainability could occur.

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# 21 Key Points:

- A national-scale agent-based model is developed to represent paired climatic and socioeconomic scenarios in the land system.
- Key scenario characteristics relate to forms of human behavior, interactions and societal preferences.
- Large differences emerge between scenarios in terms of land management intensities,
   ecosystem service provision and land sparing.
- 28

#### 29 Abstract

Socio-economic scenarios such as the Shared Socioeconomic Pathways (SSPs) have been widely 30 used to analyse global change impacts, but representing their diversity is a challenge for the 31 analytical tools applied to them. Taking Great Britain as an example, we represent a set of 32 stakeholder-elaborated UK-SSP scenarios, linked to climate change scenarios (Representative 33 Concentration Pathways), in a globally-embedded agent-based modelling framework. We find 34 that distinct model components are required to account for divergent behavioural, social and 35 societal conditions in the SSPs, and that these have dramatic impacts on land system outcomes. 36 From strong social networks and environmental sustainability in SSP1 to land consolidation and 37 technological intensification in SSP5, scenario-specific model designs vary widely from one 38

- another and from present-day conditions. Changes in social and human capitals can generate impacts larger than those of technological and economic change, and comparable to those of
- 41 modelled climate change. We develop an open-access, transferrable model framework and
- 42 provide UK-SSP projections to 2080 at 1km2 resolution, revealing large differences in land
- 43 management intensities, provision of a range of ecosystem services, and the knowledge and
- 44 motivations underlying land manager decision-making. These differences suggest the existence
- of large but underappreciated areas of scenario space, within which novel options for land
- 46 system sustainability could occur.
- 47

# 48 **1 Introduction**

- 49 If efforts to mitigate climate change in the coming years are not transformative, then the impacts
- 50 themselves likely will be. The adoption of effective mitigation and adaptation strategies is
- 51 therefore essential, and these depend upon thorough knowledge of possible future conditions
- 52 (Rounsevell et al., 2021). To help generate such knowledge, various sets of scenarios have been
- 53 developed to provide structures within which analyses can be conducted (Schindler & Hilborn,
- 54 2015). Currently, the most widely-used scenario sets for environmental studies are the
- 55Representative Concentration Pathways (RCPs) describing alternative greenhouse gas
- 56 concentration trajectories, and the Shared Socioeconomic Pathways (SSPs) describing alternative
- 57 socio-economic trajectories (O'Neill et al., 2020).
- 58 The RCP-SSP framework has been adopted across disciplines, and a decade's worth of research
- 59 has built upon it (O'Neill et al., 2020). It has proven particularly useful because it allows various
- 60 combinations of climatic and socio-economic conditions to be explored, providing coherent
- storylines of plausible future conditions. RCP-SSP combinations have been defined for
- numerous contexts from global to local scales, often through participatory processes of
- 63 stakeholder engagement (e.g. Kebede et al., 2018; Kok et al., 2019; Wear & Prestemon, 2019).
- 64 Together, these scenarios describe radically different 'worlds' in which societal structures and
- <sup>65</sup> priorities differ, are subject to different modes of governance, and are constrained by different
- 66 socio-economic resources.
- One of the main uses of these scenario storylines has been in computational modelling. This
- 68 modelling supports the identification of pathways towards particular outcomes, such as limiting
- 69 global mean-temperature increases to 1.5°C (Rogelj et al., 2018), or reversing global biodiversity
- declines (Leclère et al., 2020). Model-based implementations of the RCPs and SSPs have

- become the de facto basis for anticipatory policy-making at the international level, effectively
- defining the expected scope of actions and outcomes during the 21st century (O'Neill et al.,
- 73 2020).

Reliance on computational models for quantitative exploration of future conditions is largely 74 inevitable, but is not without drawbacks. Faced with widely divergent SSPs, it would be 75 appropriate to use similarly divergent modelling approaches to fully explore scenario space 76 (Brown et al., 2021; Polasky et al., 2011). However, large-scale land system models have been 77 78 relatively convergent in approaches and assumptions (Brown et al., 2017; Gambhir et al., 2019; Haasnoot et al., 2013; Uusitalo et al., 2015). Most rely on cellular automata, econometric or 79 similar models with statistical transition probabilities between broad land use classes based on 80 observed (past) changes (Brown et al., 2017; Verburg et al., 2019). Only a small subset of 81 scenario components have been explored as a result, usually those related to economic or policy 82 change. Aspects of scenarios most neglected in large-scale land system models relate to human 83 behaviour within the land system, ecosystem services provision, representing land use (as 84 opposed to land cover) alternatives across sectors, and explicit links between global and smaller-85 scale dynamics (Müller et al., 2019; Verburg et al., 2019). As a result, the highly divergent 86 nature of SSP scenarios may be obscured, and important areas of scenario space unexplored 87 (Estoque et al., 2020; Pedde et al., 2019). 88

89 Here we take a set of detailed, stakeholder-developed, qualitative and quantitative SSPs for the United Kingdom, and simulate the development of the British land system throughout these 90 91 scenarios using a flexible agent-based modelling framework driven by national and global scenario storylines. In adapting this framework to each UK-SSP in turn, we highlight the ways in 92 which the scenarios differ from the present day and from one another. We develop a new model 93 application that contains scenario-specific elements and settings, and consider model outputs in 94 the light of the design choices we make and their underlying scenario elements. In doing so we 95 further develop an open-access and transferrable agent-based modelling framework capable of 96 representing paired SSP-RCP scenarios at national to continental scales, and evaluate its 97 application through the comprehensive TRACE protocol (in SI). We also provide new 98 projections to 2080 of the UK-SSPs at 1km2 resolution, accounting for key scenario elements 99 related to human behaviour, ecosystem service valuation and land management intensity. We use 100 our findings to understand potential changes in the British land system in particular, and 101 potential advances in the simulation of SSPs in the land system in general. 102

#### 103 1.1 The UK context

The UK makes a particularly appropriate case study for scenario analysis for a number of 104 reasons. First, its land systems span wide ranges of uses, intensities, environmental and climatic 105 conditions, and economic viabilities - from highly productive arable farming in the south-east to 106 marginal and extensive livestock management in the north west. Second, the UK has well-107 developed data and land system research facilities. Third, land management in the UK faces a 108 particularly uncertain future, with fundamental changes to policy frameworks following the 109 UK's exit from the European Union that are likely to diverge to some extent between the 110 country's four constituent nations. Combined with substantial expected climatic changes and 111 strong remaining links to global markets, these give a notably broad space for scenario 112 exploration. Participatory processes have already been used to explore this space (Holman et al., 113

- 114 2008), most recently with the development of detailed UK-SSP scenarios describing alternative
- social, economic and political trajectories (CEH, 2021; Harmáčková et al., 2022; Pedde et al.,
  2021).
- Nevertheless, modelling of the British land system under alternate scenarios has been limited. 117 Much of the modelling that has been done has focused on the impacts of climate change 118 (Rounsevell & Reay, 2009), and/or has been sub-national in scale and focused on particular 119 scenario elements, issues or ecosystem services (Cantarello et al., 2011; Holman et al., 2005, 120 2016). Bateman et al. (2013) developed an integrated environment-economy model covering 121 different ecosystem services, but their optimisation approach involved constraining economic 122 rules and was only applied to a limited set of scenarios. Policy-oriented reports on UK land use 123 futures therefore have been able to draw on only limited evidence from modelling studies, and 124 none that covers a representative range of British land uses and future scenarios (Foresight Land 125 Use Futures Project, 2010). The UK therefore provides a particularly relevant, well-understood 126 and dynamic analogy for many other national contexts, but one for which limited scenario 127 explorations exist. We aimed to develop a detailed, cross-scale and cross-sectoral model that 128 remains sufficiently efficient and user-friendly to be used in participatory processes for UK 129
- 130 scenario analyses.

131

#### 132 2 Materials and Methods

133 We make use of two main resources in this study: a set of qualitative and quantitative UK-RCP and UK-SSP scenarios described in detail in Harmáčková et al. (2022), Merkle et al. (2022) and 134 Robinson et al. (2022), and a newly-developed UK land use model described below and in the 135 supporting information. By pairing these scenarios and model, we explore potential future land 136 system change in Great Britain prompted by linked climatic and socio-economic conditions 137 (referred to below as the 'UK-RCP-SSPs'). The model is further embedded within a global 138 139 modelling framework to account for global change and the UK's international trade under each scenario. Here we describe the general design and calibration of the model before explaining 140 141 how it was tailored to each of the UK-RCP-SSPs. Full details are contained in a stand-alone 142 methods section and TRACE model evaluation document in the Supporting Information.

143 2.1 Overview

144 We develop CRAFTY-GB, a new agent-based model of the British land system based on a broad

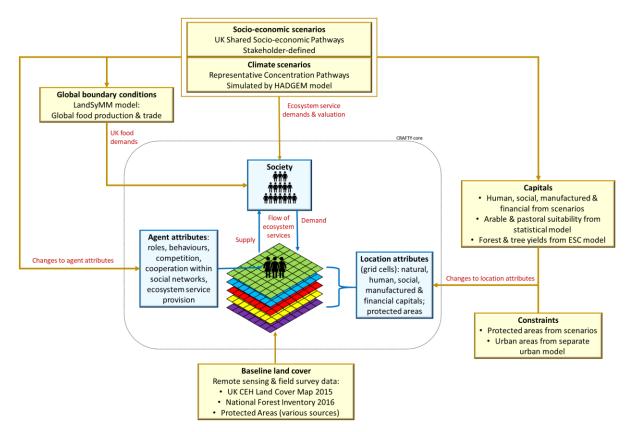
range of available land system data and operating at 1km2 resolution. The range of the model is restricted to Great Britain rather than the UK as a whole because consistent data were not

restricted to Great Britain rather than the UK as a whole because consistent data were not
 available for Northern Ireland. CRAFTY-GB is an application of the CRAFTY agent-based

147 available for Normern netand. CKAPT 1-OD is an application of the CKAPT 1 agent-based 148 modelling framework (Murray-Rust et al., 2014). The core model is therefore the same as in

149 earlier applications of this framework (e.g. to Europe (Brown et al., 2019), Sweden (Blanco et

- al., 2017) and Brazil (Millington et al., 2021)) while the inputs were tailored to the British
- 151 context (Fig. 1).



#### 152

**Figure 1**.: Schematic diagram of CRAFTY-GB structure and information flows. The blue

154 features belong to the generic core of CRAFTY, and the yellow features are specific to the

British model implementation, providing information to the core processes. This external

information is derived from observational, modelled and stakeholder-developed data explained

in the text. Red labels describe particular information exchanges.

158 The basis for modelled land use change in CRAFTY-GB is a set of capitals that describe location

resources or attributes for each 1 km 2 cell (Tables S1 - S3). Each cell is also assigned an agent

- representing a specific form of land management through a modelled process of competition
- 161 with other agents (Table S4). CRAFTY uses the concept of Agent Functional Types (Arneth et
- 162 al., 2014) to create simplified typologies of land managers according to their objectives,
- behaviours, and their forms and intensities of land management. These agents are able to use the
- 164 capitals to produce services that satisfy societal demands, which are exogenously defined.
- Agents are initially distributed using baseline land use data, and then engage in a simulated
- 166 process of competition for cells. This competition is driven by the level of demand for the
- 167 services that different agents provide, and the relative valuation of each of those services.
- 168 Competition outcomes vary with the productive and behavioural characteristics of the agents, as
- 169 well as cooperation between them through modelled social networks.
- 170 This basic model circuit is driven by exogenous scenarios that describe scenario-based climatic
- 171 and socio-economic changes over time. These changes can affect capital values, agent

172 characteristics, societal demand levels, competition processes and policy objectives. The nature

and spatio-temporal properties of modelled land use change therefore depend on the interaction

of these core model components. In this application, scenarios are also used to calibrate the

175 model parameters and to determine which modelled processes are active, which is a novel aspect 176 of the approach. Below we describe model inputs before going on to scenario implementation.

### 177 2.2 Model components

Capitals describing resources or attributes of each individual cell underpin simulated land use change in CRAFTY-GB. Capitals are divided into human, social, manufactured, financial and

natural capitals, with natural capital further divided into yields or suitabilities for arable, pastoral

and forest land uses or species (Tables S1 & S2). Social, human, financial and manufactured

182 capitals were derived from UK-SSP projections of eight socio-economic indicators from (Merkle

et al., 2022) (see Table S2). Forest suitabilities were modelled using the Ecological Site

184 Classification (ESC) yield class model (Forest Research, 2021; Pyatt, 1995), and arable, and

185 improved and semi-natural pastoral suitabilities were modelled statistically (SI section

'Capitals'). Protected areas belonging to 11 different types of national and international
 designation and to five different private land-owning organisations (NGOs) were included in the

designation and to five different private land-owning organisations (NGOs) were includ model and varied according to SSP storylines (Table S3 Fig. S1)

model, and varied according to SSP storylines (Table S3, Fig. S1).

Across the modelled landscape, CRAFTY-GB includes a range of agent types designed to

capture the main forms of land use in Great Britain, including gradations of intensity and

multifunctionality. Agent types were divided between arable land uses (intensive arable for food,

intensive arable for fodder, sustainable arable and extensive arable), pastoral land uses (intensive

pastoral, extensive pastoral, very extensive pastoral), forest land uses (productive native conifer, productive non-native conifer, productive pastoral productive non-native breadlast

productive non-native conifer, productive native broadleaf, productive non-native broadleaf,
 multifunctional mixed woodland and native woodland for conservation), and combined classes

multifunctional mixed woodland and native woodland for conservation), and combined classes
 (bioenergy and agroforestry) (Table S4). Variation in ecosystem service provision within these

197 classes allows them to represent a continuous range of forms of land management rather than

arbitrarily distinct groups. Variations in decision-making behaviour further allow individual

agents and groups of agents to respond differently to modelled changes (SI section 'Behaviour',

Table S5). Urban areas were projected in the scenarios by an independent urban model (more

details in the SI, and full details in Merkle et al., in review). The initial distribution of land uses

202 was based on a range of data sets described in Table S4.

Each modelled land use was represented as providing a range of provisioning, regulating and

cultural ecosystem services and other indicators (e.g. biodiversity, employment) of relevance to

the UK-SSP scenarios. These services are defined in Tables S6 and S7. The potential and required provisioning of these services varied according to the UK-RCP-SSP scenarios. Demand

levels for foods were derived from the LandSyMM (Land System Modular Model;

207 Investion roots were derived from the Land System Modular Model; 208 www.landsymm.earth) global modelling framework running global RCP-SSP scenario

www.landsymm.earur) global modeling framework running global KCP-SSP scenario
 combinations (Rabin et al., 2020), as described in SI section 'Services & demand levels'. Non-

food demands were taken from the UK-SSP scenarios, and are described in (Merkle et al., 2022).

211 Demand levels are shown in the results below, and are available along with all model data (see

- 212 'data availability' section).
- 213 2.3 Scenarios

The SSPs were specified for the UK as described in Pedde et al. (2021), Harmáčková et al. 214 (2022) and Merkle et al. (2022). These substantial extensions of the global SSPs provide detailed 215 narratives and quantifications of social, economic and political developments across the UK until 216 2100. The narratives integrate national stakeholder knowledge on locally-relevant drivers and 217 indicators with higher level information from the European and global SSPs. These narratives 218 219 were simplified and converted into model parameterisations (Fig. 2, Table S8). The UK-SSPs were put in a global context through LandSyMM global land system modelling to provide 220 221 consistency with the broader SSP framework and to account for the UK's international trade. The SSP implementation also utilised the forms of behaviour represented in CRAFTY to capture 222 land management decision-making (Table S6). Of these behaviours, social networks are the only 223 new addition to the CRAFTY framework. These allow agents of the same type to affect one 224 225 another's competitiveness within defined spatial neighbourhoods, to represent the benefits both of improved local knowledge diffusion and of economies of scale. 226

The RCPs were specified for the UK as described in the SI (section 'Scenarios') and (Robinson

et al., 2022). Climatic conditions were taken from the CHESS-SCAPE future climate data set, which extends the regional climate model (RCM) output in the UK Climate Projections 2018

230 (UKCP18) (Lowe et al., 2018; Met Office Hadley Centre (MOHC), 2018) by downscaling them

from 12km to 1km resolution and producing realisations for three RCPs in addition to RCP8.5.

This data set covers several physical climate variables to 2080 at 1 km spatial resolution and time

steps ranging from daily to decadal averages. Spatially and temporally explicit values for several

climate variables were generated for the UK, including temperature and precipitation, potential

evapotranspiration and growing degree days. These variables were then used as inputs to the

crop, grassland and forest modelling to produce annual scenario-specific capital values.

RCP-SSP combinations were chosen to: (i) cover a broad range of uncertainty in both emissions 237 (and hence climate) and socio-economic developments; and (ii) include any combination of SSPs 238 and RCPs that is plausible, meaningful and useful. The six combined scenarios we use (RCP2.6-239 SSP1, RCP4.5-SSP2, RCP4.5-SSP4, RCP6.0-SSP3, RCP8.5-SSP2, RCP8.5-SSP5) cover weak 240 to strong climate change, as well as future societies with high and low challenges to adaptation 241 242 and mitigation. The selection also allows analysis of the effects of different RCPs within the same SSP (RCPs 4.5 and 8.5 with SSP2), and the effects of different SSPs within the same RCP 243 (SSPs 2 and 4 with RCP4.5; SSPs 2 and 5 with RCP8.5). Furthermore, low adaptation challenges 244

(SSP1/5) and high adaptation challenges (SSP3/4) are confronted with different RCPs.

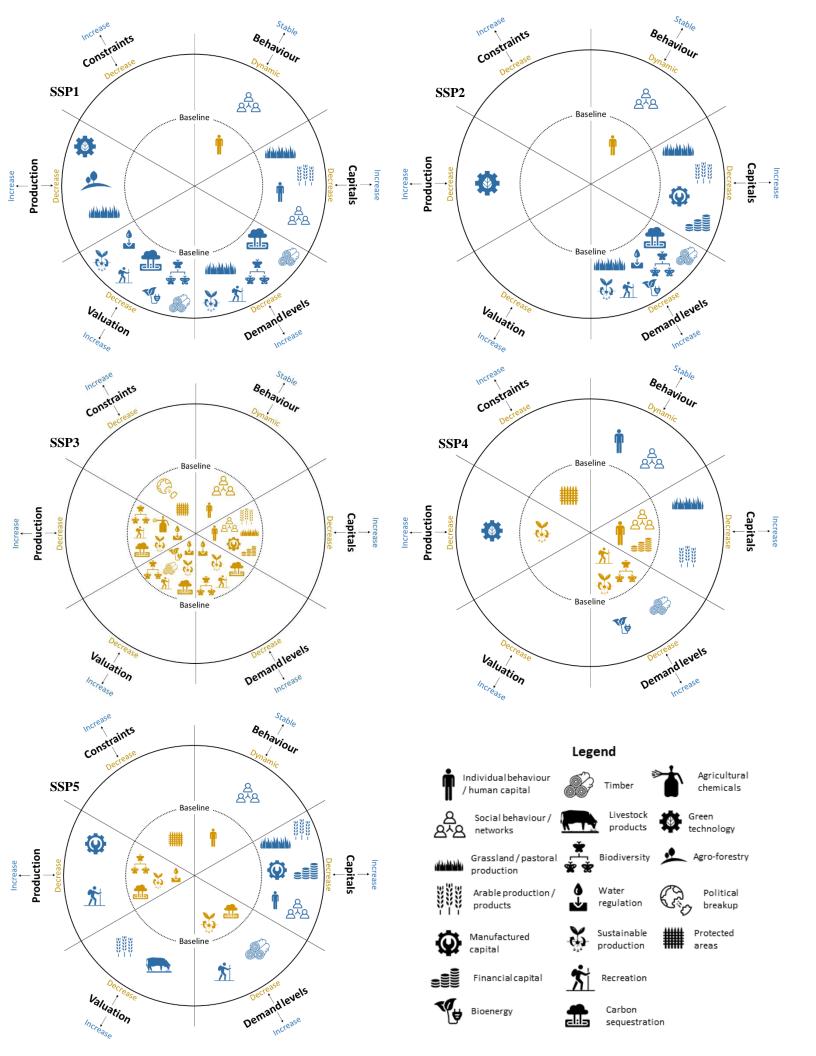


Figure 2: Summary of the implementation of the UK-SSPs in CRAFTY GB. Items included here

- represent main scenario conditions and refer specifically to the CRAFTY-GB implementation,
- relative to the baseline, and are in addition to the broader scenario storylines. Changes in demand
- shown here are per capita and do not represent the overall demand changes summarised in Fig. 4.
  The 'Behaviour' segment in the plots varies between 'stable' and 'dynamic' rather than
- The 'Behaviour' segment in the plots varies between 'stable' and 'dynamic' rather than 'increase' and 'decrease' because behavioural variations are not directional but affect the
- 2/6 Increase and decrease because behavioural variations are not directional but affect beterogeneity and temporal dynamism of agent behaviour (see Table S5)
- heterogeneity and temporal dynamism of agent behaviour (see Table S5).
- 278 2.4 Model evaluation
- 279 Model evaluation is presented in detail in a TRACE ("TRAnsparent and Comprehensive model
- Evaludation") model evaluation document in the SI (Augusiak et al., 2014; Ayllón et al., 2021;
- Grimm et al., 2014; Schmolke et al., 2010), with main components summarised here. The
- 282 CRAFTY framework has been evaluated using combinations of unit tests, sensitivity and
- uncertainty analyses, comparisons to empirical data and to the results of other models, full peer-
- reviewed descriptions of model design and functioning, and full, free access to the model itself
- 285 including interactive online systems for exploring model outputs
- (https://landchange.earth/CRAFTY) (e.g. Alexander et al., 2017; Brown et al., 2014, 2018;
- Holzhauer et al., 2019; Murray-Rust et al., 2014; Synes et al., 2019). The technical
- implementation of this framework through the CRAFTY-GB model and its application to the UK
- 289 RCP-SSPs was evaluated through sensitivity analyses as the model was developed, consultations
- with experts and stakeholders (as described in Merkle et al. (2022)), and finally comparison to
- existing relevant literature on UK land use projections. We did not check CRAFTY-GB's ability
- to reproduce historical land use change within the UK as such change has no definite relevance
   to future changes, and because there is no temporally consistent UK land cover data against
- to future changes, and because there is no temporally consistent UK land cover data against
   which to check modelled change (the UK Land Cover Map data do not allow for comparison of
- all CRAFTY-GB classes across years, and other inputs are unavailable for matching timepoints).
- all CRAFT Y-GB classes across years, and other inputs are unavailable for matching timepoint
- 296 We carried out further evaluation of the representativeness of CRAFTY-GB agent types. The
- baseline allocation of agent types was compared against (semi-)independent datasets to check its
- coverage and interpretation with respect to agricultural and ecological characteristics. These
   datasets were 1) LCM 2015 (Rowland et al., 2017), to provide a summary of the translation of
- datasets were 1) LCM 2015 (Rowland et al., 2017), to provide a summary of the translation of
   LCM classes into CRAFTY-GB classes (Table S4), 2) The standardised European EUNIS
- habitat classification scheme at 100m resolution (European Environment Agency, 2019; Weiss &
- Banko, 2018), 3) The UK CEH Land Cover Plus: Fertilisers and Pesticides data (Jarvis et al.,
- 2020; Osório et al., 2019). Comparison to these data provides an evaluation of the agent typology
- and its initial geographic distribution because it reveals the extent to which the ranges of
- 305 different ecological and agricultural characteristics found in British land systems are captured by
- the typology as a whole, and the extent to which individual agent types can be interpreted as
- 307 representing specific characteristics from those ranges. It is not a targeted validation because the
- agent typology is not designed specifically to achieve these objectives, but it provides a basis

- from which to better interpret model results. On the basis of these and previous evaluations, we believe the model is appropriate for the purpose for which it is used here.
- 311 2.5 Representing levels of management intensity in the model outputs

To improve the interpretability of the results, we developed a land use intensity mapping

approach. This involved the assignment of values on a continuous range for each of the arable,

and each of the pastoral (except very extensive pastoral) classes across the scenarios. Intensity

values were defined as a combination of the use of agricultural inputs (fertilisers, pesticides and

machinery), technology, and modelled production levels. For the purposes of illustration these

are combined multiplicatively here and used to select colour saturation levels in the map figures.
 Alternative representations are possible, and it is important to note that our presentation does not

- distinguish the specific use of technology to reduce the use of chemical inputs, as in UK-SSP1.
- This method does however make scenarios results more comparable and means that differences
- in land management intensities among the scenarios are readily apparent.
- 322

#### 323 **3 Results**

324 3.1 Agent typology evaluation

Results of the comparison between baseline CRAFTY agent types and independent habitat and 325 management maps suggested that the typology has good coverage, with clear but variable 326 associations between agent types and each of the characteristics included (SI section 'Agent 327 typology evaluation', Figs S4-S9). At a basic level, the baseline mapping reproduced the LCM 328 classes that were the primary data used to locate agents geographically (Fig. S2, Table S4). 329 Forest types were the most inconsistent between the CRAFTY-GB baseline and the LCM data, 330 and comparisons at the sub-grid scale reveal that forest types are generally more associated with 331 heterogeneous landscapes compared to intensive arable and pastoral agents (Figs S4 & S5). The 332 ranges of LCM class coverage within each agent type also reflects the mixed nature of land cover 333 in many of the CRAFTY-GB cells. This mixture is reflected in the capitals and service levels 334 335 more fully than in the generic agent type labels, but is also further revealed by the EUNIS habitat comparison. 336

The EUNIS classes were widely distributed between agent types, but with clear associations

(Figs S6 - S8). These were generally as expected, for example with grassland habitats strongly

associated with pastoral areas, farmland habitats with agricultural areas and so on. Woodland

habitats were particularly strongly associated with forested areas in the baseline map, providing

some confirmation of their locations and interpretation. Nevertheless, many different specific
 habitats occurred even within the most intensive agent types at baseline, and these can be

expected to persist or even increase in proportion in most scenarios, with the exception of SSPs 4

and 5 where the scenario storylines include consolidation of farms and fields across larger areas,

implying loss of secondary habitats.

The quality of all of these habitats is also dependent on usage of chemical inputs and machinery.

347 As expected, chemical inputs were most strongly associated with intensive arable areas (within

348 which sustainable arable agents were randomly distributed at baseline, allowing no distinction in

349 levels of chemical application) (Fig. S9). Once again, the association of productive broadleaf

woodlands with agricultural areas was apparent in the elevated levels of chemical inputs within

those cells. Farmland and broadleaf woodland habitats can therefore be expected to be most

affected by the increased application of agricultural chemicals in SSPs 4 and 5.

#### 353 3.2 Scenario results

354 The application of CRAFTY-GB to the UK RCP-SSP scenarios introduced very different driving conditions to the model, which resulted in significant divergence between simulated land use 355 over time (Table 1). Most notably, divergence occurred in terms of intensity of land use. This 356 was partly because intensity was determined by the scenario conditions, and partly because 357 intensity changed as an emergent property of the simulations. For example, the gradual 358 359 restriction of agricultural pesticides in UK-SSP1 led to a direct reduction in management intensity (when defined partly in terms of chemical inputs), but also an indirect reduction as 360 agents that did not require chemical inputs, and were therefore unaffected by the restriction, 361 became more competitive. Such direct and indirect changes in intensity were substantial in all of 362 the scenarios. Overall, these socio-economic effects were far stronger than climatic effects on 363

land use outcomes.

In UK RCP2.6-SSP1 (low emissions coupled with the Sustainability scenario) the emphasis on 365 sustainable agricultural and forestry production and the delivery of multiple ecosystem services 366 led to an overall lower intensity of land management compared to most other scenarios, despite 367 intensification options being available. Reduced meat demand caused a substantial move away 368 from pastoral management in many areas (Fig. 3). However, as the remaining livestock 369 production focused on grass-fed livestock products (as opposed to domestic or imported 370 feedstocks) and other agricultural land uses became more extensive, the area reduction of 371 agricultural management was limited. Intensity gains were simulated in small areas (Fig. 4), but 372 373 overall sustainable and extensive management became more widespread. By 2080, sustainable arable management dominated eastern England, while the British uplands were largely given 374 over to extensive pastoral management (Fig. 3). Nevertheless, substantial areas were also 375 covered by natural vegetation (whether unmanaged or managed for conservation) and, in 376 forestry, native conifer and broadleaf species (Fig. S10). This resulted in some large, contiguous 377 areas under either natural vegetation or native tree cover, especially in south-west England, 378 379 Wales and southern Scotland. Despite the relative increase of extensive, mixed and sustainable land uses, under-supplies of biodiversity, employment, recreation and carbon increased during 380 the simulation, with a slight but persistent over-supply of grass-fed red meat. The UK land 381 system was unable to meet the very high demands for the wide range of ecosystem services in 382 UK-SSP1. 383

UK-SSP2 (the Middle of the Road socio-economic scenario) was run under two climatic 384 scenarios, RCP4.5 and RCP8.5. Overall, the different climatic conditions had limited effects, 385 being most apparent in slightly larger areas of forest under RCP8.5, within which species were 386 more separated between conifer-dominated forests in the south and broadleaf-dominated in the 387 north, following climatic suitability (Fig. S10). In both cases, forests were more widespread than 388 in UK-SSP1 due to increased demands for afforestation to sequester carbon and produce timber. 389 Non-native species dominated these forests, especially in Scotland and in RCP8.5. As a result, 390 the area of natural vegetation was relatively low outside (substantial) areas under conservation 391

392 management. These were possible because of intensification of arable agriculture in particular,

and a decrease in the demand for grass-fed livestock products that allowed food demands to be

met consistently (Fig. 4). This also led to a very large reduction (ca. 60%) in the area of intensive

pastoral management (much of which was converted to forestry; Fig. 5), which also became
 dispersed among other land uses in less productive areas. This was reinforced by a large drop in

meat and milk demand over the first decade of the simulation, and concurrent increase in timber

- demand. The scenario generated very little over-supply, but biodiversity and carbon were slightly
- under-supplied (at around 90% of demand) by the end of the simulation. Intensive arable
- agriculture remained concentrated in the south-east, with extensive pastoral in the north-west
- 401 (Fig. 3).

402 UK RCP6.0-SSP3 (relatively high emissions coupled with the Regional Rivalry scenario) is a

highly dystopian scenario with increasing barriers to trade and widespread social tensions and
 conflict. Overall, simulated land use was highly extensive (more extensive than in any other

405 scenario or even in the baseline) because capitals and inputs supporting agriculture were lacking

- 406 in the storyline. This occurred both within land uses (e.g. decreasing intensity of management
- within 'intensive arable' cells) and between them (e.g. a widespread initial transition from

intensive pastoral to extensive arable management) (Figs. 3-5). Nevertheless, this extensive
 agricultural management occupied large, contiguous areas as growing food for survival becomes

- the primary demand (Fig. 3). Many forest areas were converted to arable agriculture, with
- remaining forests dominated by conifers (Fig. S10). As the scope for intensive management
- decreased during the century, supply levels fell below demands and utilisation of depleted

intensification options increased. Nevertheless, food crops were only able to satisfy around 60%

of demand at some points, with employment levels even lower (Fig. 4). In areas where

415 intensification options were most limited due to low levels of multiple capitals (much of

Scotland and Wales, where independence from England also meant that demands had to be

satisfied domestically), multifunctional alternatives such as agroforestry and sustainable arable

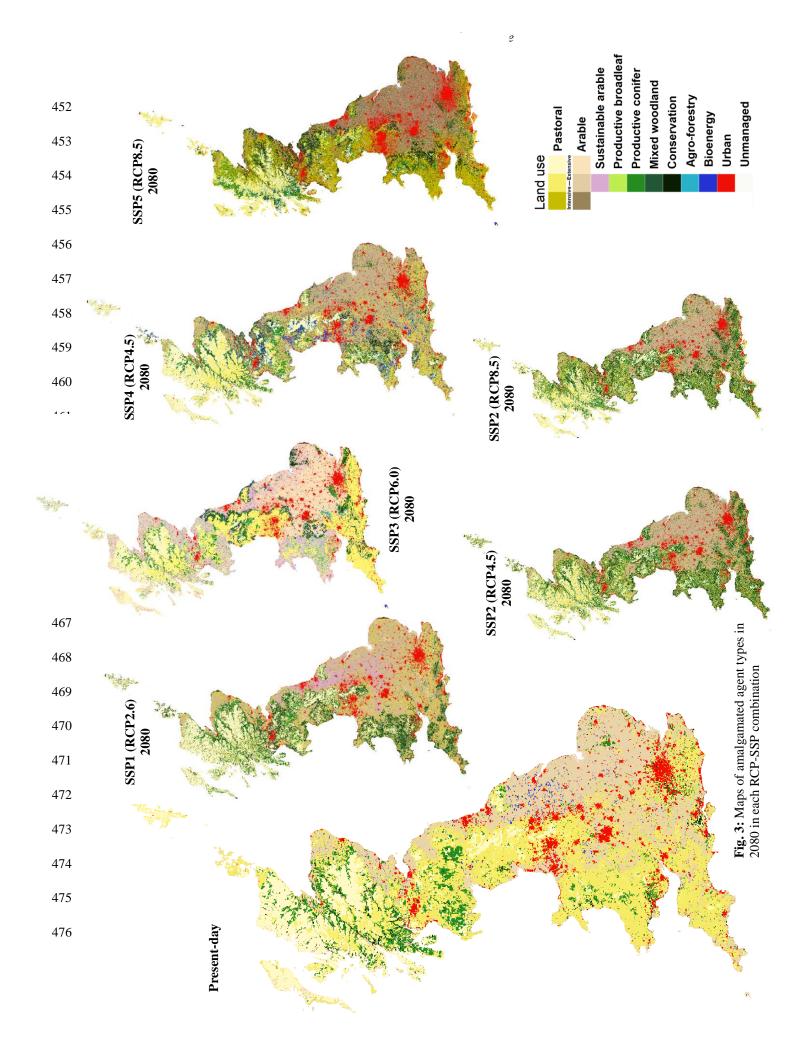
418 production emerged as competitive ways of maintaining some food production.

419 UK RCP4.5-SSP4 (medium emissions coupled with the Inequality scenario) is dominated by a business and political elite who take over much of the British land system and invest in large-420 scale industrial agriculture. This produced a substantially more intensive land system than SSPs 421 1-3, which was especially pronounced in increasing arable extent and intensity (Fig 5). A 422 decrease in the relative demand for grass-fed livestock products led to a reduction in intensive 423 pastoral production from around 2050, but meat and milk were still highly over-supplied at some 424 425 points in time as demand levels fluctuated (with milk supply at more than 150% of demand early and in the middle of the century) (Fig. 4). Conversely, intensive arable production increased as 426 pastoral decreased, as did bioenergy, which was ultimately grown across the country in marginal 427 428 agricultural areas (Fig. 3). This left little room for forest management, but large areas of abandonment and conservation management did emerge in some upland areas, partly due to 429 demand for recreation by the rich elite in the scenario. Within forests, non-native conifers 430 dominated, being used to satisfy timber demand. Large land holdings had a competitive 431 advantage, and land use became particularly homogeneous in productive areas, implying further 432 degradation of habitats. 433

UK RCP8.5-SSP5 (high emissions coupled with the Fossil Fuel Development scenario) was the
 most intensive land use scenario, with massive urban expansion and agricultural intensification

as demand levels increased due to a substantial rise in the UK population and a shift to highly

- individualistic and consumptive lifestyles. Protected areas were removed as concern for the
- 438 environment was low. Declining social capital made marginal production vulnerable to change,
- while strong local networks allowed consolidation of dominant land uses. Nevertheless, there
   was a substantial amount of sustainable arable agriculture and conservation, because these
- 440 was a substantial amount of sustainable alable agriculture and conservation, because these 441 provided multiple low-priority ecosystem services in single cells. Limited forest area was
- 442 concentrated in southern and north-west England, the Welsh borders, and north-west Scotland,
- 443 with native broadleaf and conifer species dominating outside Scotland (Fig. S10). The pastoral
- land area was almost maintained in this scenario due to very high demands for livestock products
- (Fig. 5). Despite some urban expansion into productive land and extensification of unproductive
- land, overall land use intensity increased dramatically (Fig. 4). Food supply increased too, but
- not enough to satisfy demands for grass-fed red meat. There was a general shortfall in supply of
- intangible services, supporting the existence of sustainable and conservation management to
- supply several of these within the intensive landscapes. Land abandonment in the uplands was anemergent response to intensification elsewhere, but this was consistent with the scenario
- emergent response to intensification elsewhere, but this was consistentstoryline of upland rewilding to deliver recreation benefits.
- 451 storyline of upland rewilding to deliver recreation benefits.



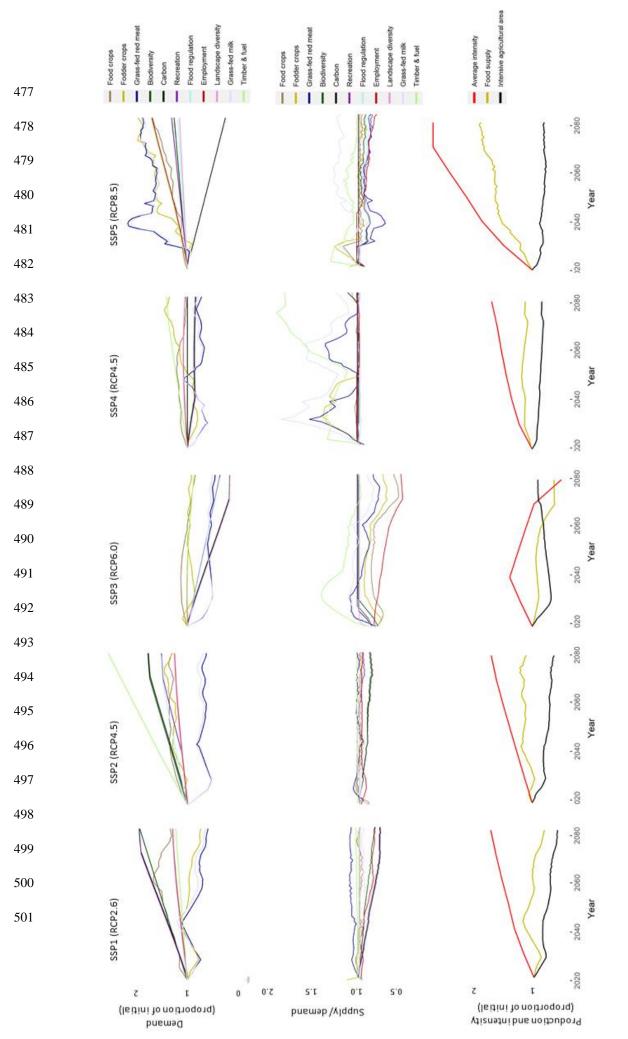


Fig. 4: Demand levels, supply as proportion of demand, and land use intensity, food supply and intensive area throughout each SSP scenario (RCP8.5-SSP2 results were very similar to those shown for RCP4.5-SSP2, and can be found in Fig. S11).

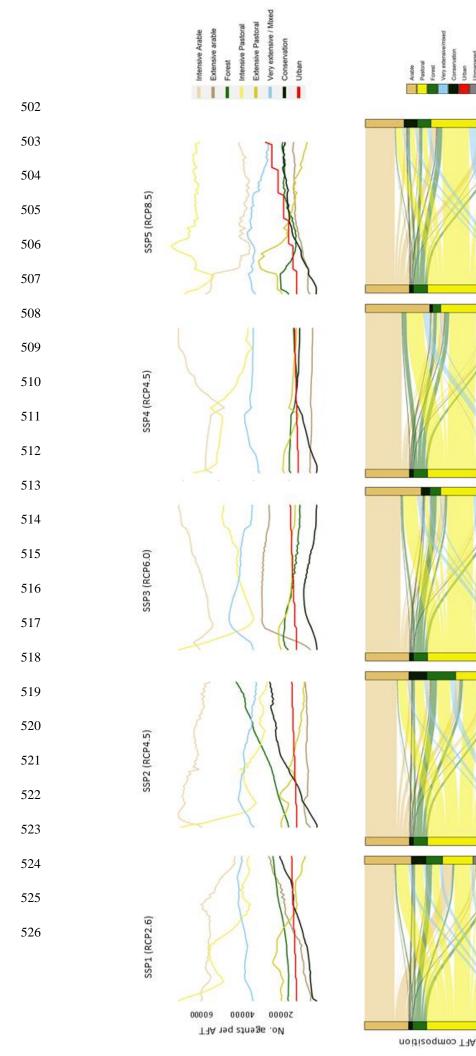


Fig. 5: Agent Functional Type (AFT) dynamics throughout each SSP scenario: numbers of agents within amalgamated AFTs (top) and transitions between broad land use types (bottom). RCP8.5-SSP2 results were very similar to those shown for RCP4.5-SSP2, and can be found in Fig. S11.

Year

Year

Year

Year

Year

| Scenario                                 | Description  | Distinguishing features in<br>CRAFTY-GB  | Main outcomes  |
|--|--|--|--|
| SSP1 -<br>Sustainability                 | UK-SSP1 shows the UK transitioning to a<br>fully functional circular economy as<br>society quickly becomes more egalitarian<br>leading to healthier lifestyles, improved<br>well-being, sustainable use of natural<br>resources, and more stable and fair<br>international relations. It represents a<br>sustainable and co-operative society with a<br>low carbon economy and high capacity to<br>adapt to climate change.                | Novel forms of sustainable<br>agriculture with strong<br>societal support                  | Decreasing area of intensive agriculture, greater<br>multifunctionality of agricultural land   |
|  |  | Low demand levels for<br>livestock products, but<br>preference for grass-fed<br>production | Move away from livestock production and decrease in pastoral<br>area, limited by relatively low-efficiency of pastoral production  |
|  |  | Preference for native tree species in forestry   | Substantial shift towards native species in forests, depending on suitabilities  |
| SSP2 – Middle<br>of the Road             | e  | Established forms of<br>agriculture with potential for<br>intensification                  | Intensification and increasing efficiency of agriculture, leading<br>to intensive area declines  |
|  |  | Increasing demand for timber<br>and forest-based carbon<br>sequestration                   | Large increase in forest area, dominated by non-native tree species  |
|  |  | Low demand for grass-fed livestock products  | Large decrease in intensive pasture area, most livestock<br>production feed-based  |
| SSP3 –<br>Regional<br>rivalry            | The dystopian scenario, UK-SSP3, shows<br>how increasing social and economic<br>barriers may trigger international tensions,<br>nationalisation in key economic sectors,<br>job losses and, eventually a highly<br>fragmented society with the UK breaking<br>apart. It represents a society where rivalry<br>between regions and barriers to trade<br>entrench reliance on fossil fuels and limit<br>capacity to adapt to climate change. | Large decreases in most capitals   | Extensification of production as inputs become unavailable,<br>shortfalls in supply and increasing area with maximum possible<br>intensity                               |
|  |  | Trade barriers reduce food<br>imports. Decreasing demand<br>for most other services        | Food production dominates land uses, with other ecosystem<br>services being by-products of enforced low-intensity<br>management  |
|  |  | Very weak social networks  | Heterogeneous and frequent changes in land use, suboptimal exploitation of available capitals  |
|  |  | Political breakup of the UK  | Divergence in land system trajectories between England, Wales<br>and Scotland, with least intensive production methods being<br>only feasible options in smaller nations |
| SSP4 -<br>Inequality                     | 5  | Economies of scale in agriculture  | Large, homogeneous areas of agriculture emerge, representing<br>large farms with large fields  |
|  |  | High demand for recreation among economic elites   | Conservation/recreation management in upland areas, loss of marginal land uses   |
|  |  | Low demand for grass-fed livestock products  | Decline in pasture, livestock production using crop-based feed   |
|  |  | High demand for bioenergy  | Expansion of bioenergy on arable land in many areas; overall increase in arable area & intensity, at expense of forest areas   |
| SSP5 – Fossil-<br>fuelled<br>development | highly individualistic society where the   | Increasing demands for urban areas and food production                                     | High pressure on land area and strong competition between land uses  |
|  |  | Increasing intensification options   | Very high levels of intensification in agriculture supporting<br>large increases in production   |
|  |  | Removal of Protected Areas<br>and low demands for related<br>ecosystem services            | Expansion of productive land uses into natural areas, with<br>consequent abandonment in upland and marginal areas not under<br>protection.                               |

Table 1: Descriptions of each UK-SSP, the main drivers that distinguish each within CRAFTY-GB, and the results
 of those drivers observed in the model outputs.

#### 529

#### 530 4 Discussion

531 This study targets the gap between detailed stakeholder-developed SSP storylines and their representations in computational models. We attempt to extend scenario modelling using flexible 532 model structures and parameterisations that are not limited to the single pathway established by 533 534 historical land use change (Fig. 2, Table 1). This is not a predictive exercise, but an exploration of possible consequences of alternate futures as envisioned in detail by a group of policy-makers 535 and other stakeholders (Harmáčková et al., 2022; Merkle et al., 2022; Pedde et al., 2021). While 536 537 some aspects of the scenarios remain unrepresented in the model, the substantial scenariospecific modifications we made confirmed some elements of the scenario storylines (e.g. upland 538 539 land abandonment in UK-SSP5), challenged others (e.g. the provision of high-levels of many ecosystem services in UK-SSP1), and revealed further emergent differences not previously 540 anticipated (e.g. extensification of agriculture as a response to altered competition dynamics in 541 UK-SSPs 1 and 5). 542

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The level of land use intensity was the most notable variation between scenario outcomes, in

terms of levels of agricultural inputs and levels of ecosystem service outputs. In UK-SSP1 we

found deliberate extensification (land sharing) leading to some environmental benefits of the
 kind envisioned in the scenario storyline, but still with less success in meeting ecosystem service

kind envisioned in the scenario storyline, but still with less success in meeting ecosystem service demands than some other more intensive (land sparing) scenarios. In the land sparing scenarios

(UK-SSPs 4 and 5), environmental benefits were indirect and, from the point of view of the

agents represented in the model, a by-product of their primary activity. In UK-SSP3 such

550 benefits occurred because strong intensification was not possible given the lack of agricultural

551 inputs (manufactured, chemical, financial and social), but in UK-SSP5 they occurred because

intensification freed up land that could be managed multifunctionally, or abandoned to rewilding.

553 At the same time, substantial increases in farm sizes and agricultural chemical application

implies that environmental quality on farmland declined substantially in UK-SSPs 4 and 5.

555 These changes occurred within a consistent global framework that provided at least some

coherence between the internal and external drivers of British land system change. For instance,

the scenarios took account of global population projections and resultant trade shifts, meaning

that development in Great Britain remains within appropriate global boundary conditions. When

559 implemented in this way, the UK-SSPs had far more substantial effects on land system outcomes

than the climatic UK-RCP scenarios (see also e.g. Brown et al., 2019; Kriegler et al., 2017;

561 Molotoks et al., 2021; Wiebe et al., 2015). Nevertheless, the absence of extreme events from 562 BCP8 5 in particular (because the spatial and temporal resolution of the alimate modelling limits)

562 RCP8.5 in particular (because the spatial and temporal resolution of the climate modelling limits 563 representation of such events) does imply that very large climatic impacts may be missing (Kopp

et al., 2016; Otto et al., 2020). Furthermore, there was no simulated impact of land degradation

on agricultural productivity, potentially arising from climatic extremes, or the high intensity of

use envisaged within the UK-SSP5 storyline. National changes can also be seen in their global

context, for instance in terms of extremely high import levels in UK-SSP5, and for some

commodities in UK-SSPs 1 and 2, suggesting indirect land use change abroad as an externalityof either land sparing or land sharing domestically (Fuchs et al., 2020).

Some of these findings are broadly consistent with the comparable study of (Bateman et al., 570 2013), who found that including ecosystem services in modelling based on economic valuations 571 led to very different balances among service provision. We find a similar importance of the 572 valuation of ecosystem services, and a similar importance of considering spatial and temporal 573 variations in ecosystem service provision levels. In developing a full UK RCP-SSP scenario 574 implementation we also find, however, that policy options and the associated room for 575 manoeuvre are limited by other factors, including the level of international trade, societal 576 tolerance for intensive methods of production, the rate at which land managers become aware of, 577 and adopt, new technologies or practices, and the levels of supporting capitals available to land 578 managers. Two of these, human and social capital, vary enormously across the scenarios, but are 579 usually absent from scenario modelling. Pedde et al. (2019) showed that they are nevertheless 580 essential for major policy targets such as the Paris climate agreement, quite possibly more so 581 than the far-more-studied technological and economic factors. We also concur with earlier 582 studies that concluded that social factors can be more important than climate policy in achieving 583 societal objectives (Liu et al., 2020), because they determine the realised impacts of those 584 policies. 585

Other findings relate to further necessary development. This model, and land use models in 586 general, will have greater utility as they become more closely aligned with biodiversity 587 outcomes, in particular by more fully assessing the role of land management in driving either 588 declines or recovery in terrestrial biodiversity (Leclère et al., 2020; Rounsevell et al., 2018; 589 Urban et al., 2021). More realistic assessment of land-based climate change mitigation is also a 590 priority (Estoque et al., 2020). Both of these will also require improved modelling of forest (and 591 forestry) dynamics, and especially the links between tree species growth, management practices 592 and decisions, and competition within the broader land system (Blanco et al., 2017; Brown et al., 593 2017; Shifley et al., 2017; Vulturius et al., 2017). Together with the development of urban areas, 594 forest management is very sensitive to socio-economic conditions in the SSPs, and in turn has 595 strong implications for the extent of climate change mitigation (Bukovsky et al., 2021). 596

While we propose that these extensions of scenario modelling improve the realism and utility of 597 model outputs, we also acknowledge that they increase uncertainty (revealed uncertainty at least, 598 as the same uncertainty can be said to be hidden in models that do not account for these factors). 599 It has been argued (e.g. by Rosen, 2021) that the SSPs have not been useful for climate 600 mitigation policy analysis because they are implemented differently in different models, leading 601 to a lack of agreement about what different SSPs actually imply. Rosen (2021) suggests a 602 reduction in the number and variance of models used, to develop canonical representations of the 603 SSPs. We disagree with that argument. Instead, we suggest that models should be further 604 605 developed to capture the key elements of SSP scenarios that have been previously neglected – social change, non-economic values of ecosystem services, variations in land use intensity and 606 competition between forms of management. Even then, we suggest that more diversity in models 607 and modelling approaches is needed to properly explore the rich and complex storylines of 608 stakeholder-developed scenarios. The application of multi-model ensembles to explore future 609 scenario space is an especially promising option. Rather than being a recipe for confusion, we 610 611 view this as a way to gradually build up an improved understanding of potential futures and,

- crucially, to support the development of genuinely robust policy pathways towards societal
- 613 objectives.
- 614

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- 632

# 633 **Open Research**

- All output data and model code are freely available through <u>https://landchange.earth/CRAFTY</u>
   and <u>https://doi.org/10.17605/OSF.IO/CY8WE</u>.
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| 645 | Referen | ces |
|-----|---------|-----|
|     |         |     |

| 646 | Alexander, P., Brown, C., Arneth, A., Finnigan, J., & Rounsevell, M. D. A. (2016). Human |
|-----|--|
| 647 | appropriation of land for food: The role of diet. Global Environmental Change: Human     |
| 648 | and Policy Dimensions, 41, 88–98.  |

- Alexander, P., Prestele, R., Verburg, P. H., Arneth, A., Baranzelli, C., Batista e Silva, F., et al.
- (2017). Assessing uncertainties in land cover projections. *Global Change Biology*, 23(2),
  767–781.
- Alexander, P., Rabin, S., Anthoni, P., Henry, R., Pugh, T. A. M., Rounsevell, M. D. A., &
- Arneth, A. (2018). Adaptation of global land use and management intensity to changes in

654 climate and atmospheric carbon dioxide. *Global Change Biology*.

- 655 https://doi.org/10.1111/gcb.14110
- Arneth, A., Brown, C., & Rounsevell, M. D. A. (2014). Global models of human decision-
- making for land-based mitigation and adaptation assessment. *Nature Climate Change*,
  4(7), 550–557.
- Augusiak, J., Van den Brink, P. J., & Grimm, V. (2014). Merging validation and evaluation of
   ecological models to "evaludation": A review of terminology and a practical approach.
   *Ecological Modelling*, 280, 117–128.
- Ayllón, D., Railsback, S. F., Gallagher, C., Augusiak, J., Baveco, H., Berger, U., et al. (2021).
- Keeping modelling notebooks with TRACE: Good for you and good for environmental
   research and management support. *Environmental Modelling & Software*, *136*, 104932.
- 665 Bartkowski, B., & Bartke, S. (2018). Leverage Points for Governing Agricultural Soils: A
- 666 Review of Empirical Studies of European Farmers' Decision-Making. *Sustainability:*
- 667 *Science Practice and Policy*, *10*(9), 3179.

| 668 | Bateman, I. J., Harwood, A. R., Mace, G. M., Watson, R. T., Abson, D. J., Andrews, B., et al.  |
|-----|--|
| 669 | (2013). Bringing ecosystem services into economic decision-making: land use in the             |
| 670 | United Kingdom. Science, 341(6141), 45–50.   |
| 671 | Berger, T. (2001). Agent-based spatial models applied to agriculture: a simulation tool for    |
| 672 | technology diffusion, resource use changes and policy analysis. Agricultural Economics,        |
| 673 | 25(2-3), 245–260.  |
| 674 | Blanco, V., Holzhauer, S., Brown, C., Lagergren, F., Vulturius, G., Lindeskog, M., &           |
| 675 | Rounsevell, M. D. A. A. (2017). The effect of forest owner decision-making, climatic           |
| 676 | change and societal demands on land-use change and ecosystem service provision in              |
| 677 | Sweden. Ecosystem Services, 23(December 2016), 174–208.  |
| 678 | Blanco, V., Brown, C., Holzhauer, S., Vulturius, G., & Rounsevell, M. D. A. (2017). The        |
| 679 | importance of socio-ecological system dynamics in understanding adaptation to global           |
| 680 | change in the forestry sector. Journal of Environmental Management, 196, 36-47.                |
| 681 | Boumans, R., Costanza, R., Farley, J., Wilson, M. A., Portela, R., Rotmans, J., et al. (2002). |
| 682 | Modeling the dynamics of the integrated earth system and the value of global ecosystem         |
| 683 | services using the GUMBO model. Ecological Economics: The Journal of the                       |
| 684 | International Society for Ecological Economics, 41(3), 529–560.                                |
| 685 | Brown, C., Holzhauer, S., & Metzger, M. J. (2018). Land managers' behaviours modulate          |
| 686 | pathways to visions of future land systems. Regional Environmental Change. Retrieved           |
| 687 | from https://link.springer.com/article/10.1007/s10113-016-0999-y                               |
| 688 | Brown, C, Murray-Rust, D., Van Vliet, J., Alam, S. J., Verburg, P. H., & Rounsevell, M. D.     |
| 689 | (2014). Experiments in globalisation, food security and land use decision making. PloS         |
| 690 | One, 9(12). https://doi.org/10.1371/journal.pone.0114213                                       |

| 691 | Brown, C, Brown, K., & Rounsevell, M. (2016). A philosophical case for process-based    |
|-----|---|
| 692 | modelling of land use change. <i>Modeling Earth Systems and Environment</i> , 2(2), 50. |

- Brown, C, Alexander, P., Holzhauer, S., & Rounsevell, M. D. A. (2017). Behavioral models of
- 694 climate change adaptation and mitigation in land-based sectors. *Wiley Interdisciplinary*

695 *Reviews: Climate Change*. https://doi.org/10.1002/wcc.448

- Brown, C, Alexander, P., & Rounsevell, M. (2018). Empirical evidence for the diffusion of
  knowledge in land use change. *Journal of Land Use Science*, *13*(3), 269–283.
- Brown, C, Holzhauer, S., Metzger, M. J., Paterson, J. S., & Rounsevell, M. (2018). Land
- managers' behaviours modulate pathways to visions of future land systems. *Regional Environmental Change*, 18(3), 831–845.
- Brown, C, Seo, B., & Rounsevell, M. (2019). Societal breakdown as an emergent property of
   large-scale behavioural models of land use change. *Earth System Dynamics Discussions*,
   (May), 1–49.
- Brown, C, Kovács, E., Herzon, I., Villamayor-Tomas, S., Albizua, A., Galanaki, A., et al.
- (2020). Simplistic understandings of farmer motivations could undermine the
   environmental potential of the common agricultural policy. *Land Use Policy*, 105136.
- Brown, C, Holman, I., & Rounsevell, M. (2021). How modelling paradigms affect simulated
  future land use change. *Earth System Dynamics*, *12*, 211–231.
- 709 Bukovsky, M. S., Gao, J., Mearns, L. O., & O'Neill, B. C. (2021). SSP-based land-use change
- scenarios: A critical uncertainty in future regional climate change projections. *Earth's*
- 711 *Future*, 9(3). https://doi.org/10.1029/2020ef001782

| 712 | Burton, V., Moseley, D., Brown, C., Metzger, M. J., & Bellamy, P. (2018). Reviewing the          |
|-----|--|
| 713 | evidence base for the effects of woodland expansion on biodiversity and ecosystem                |
| 714 | services in the United Kingdom. Forest Ecology and Management, 430, 366-379.                     |
| 715 | Cantarello, E., Newton, A. C., & Hill, R. A. (2011). Potential effects of future land-use change |
| 716 | on regional carbon stocks in the UK. Environmental Science & Policy, 14(1), 40-52.               |
| 717 | CEH. (2021). UK Shared Socioeconomic Pathways (UK-SSPs). Retrieved November 18, 2021,            |
| 718 | from https://uk-scape.ceh.ac.uk/our-science/projects/SPEED/shared-socioeconomic-                 |
| 719 | pathways   |
| 720 | DEFRA. (2016a). Crops Grown For Bioenergy in England and the UK: 2015. Retrieved from            |
| 721 | https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_           |
| 722 | data/file/578845/nonfood-statsnotice2015i-19dec16.pdf  |
| 723 | DEFRA. (2016b). Organic farming statistics 2015. Retrieved from                                  |
| 724 | https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_           |
| 725 | data/file/524093/organics-statsnotice-19may16.pdf  |
| 726 | Douglas, P. H. (1976). The Cobb-Douglas Production Function Once Again: Its History, Its         |
| 727 | Testing, and Some New Empirical Values. Economy, Journal of Political, 84(5), 903-               |
| 728 | 916.   |
| 729 | Estoque, R. C., Ooba, M., Togawa, T., & Hijioka, Y. (2020). Projected land-use changes in the    |
| 730 | Shared Socioeconomic Pathways: Insights and implications. Ambio, 1–10.                           |
| 731 | European Environment Agency. (2019, February 7). Ecosystem types of Europe. Retrieved            |
| 732 | November 16, 2021, from https://www.eea.europa.eu/data-and-maps/data/ecosystem-                  |
|     |  |

733 types-of-europe-1

- 734 EUROSTAT. (2013). Meeting of Providers of OECD Income Distribution Data 2.2
- 735 *Comparability of OECD with other international and national estimates on income*
- *inequality and poverty*. EU. Retrieved from
- 737 https://www.oecd.org/els/soc/2.2b%20Eurostat-EUSILC-Comparability.pdf
- EUROSTAT. (2018). Methodology for data validation 2.0 Revised Edition 2018. Retrieved from
- https://ec.europa.eu/eurostat/ramon/statmanuals/files/methodology\_for\_data\_validation\_v
   2\_0\_rev2018.pdf
- EUROSTAT. (2022). Data validation Eurostat. Retrieved March 3, 2022, from
- 742 https://ec.europa.eu/eurostat/data/data-validation
- Foresight Land Use Futures Project. (2010). Land Use Futures: Making the most of land in the
- 744 *21st century*. The Government Office for Science, London. Retrieved from
- 745 https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment\_
- 746 data/file/288845/10-634-land-use-futures-summary.pdf
- Forest Research. (2021). Ecological Site Classification Decision Support System (ESC-DSS).
- 748 Retrieved June 28, 2021, from https://www.forestresearch.gov.uk/tools-and-
- 749 resources/fthr/ecological-site-classification-decision-support-system-esc-dss/
- Forestry Commission. (2021). Forestry Commission Open Data. Retrieved June 28, 2021, from
   https://data-
- 752 forestry.opendata.arcgis.com/search?q=national%20forest%20inventory%202016
- Fuchs, R., Brown, C., & Rounsevell, M. (2020). Europe's Green Deal offshores environmental
  damage to other nations. *Nature*, 586(7831), 671–673.
- <sup>755</sup> Fulginiti, L. E., & Perrin, R. K. (1998). Agricultural productivity in developing countries.

756 *Agricultural Economics*, *19*(1), 45–51.

| 757 | Gorton, M., Douarin, E., Davidova, S., & Latruffe, L. (2008). Attitudes to agricultural policy and |
|-----|--|
| 758 | farming futures in the context of the 2003 CAP reform: A comparison of farmers in                  |
| 759 | selected established and new Member States. Journal of Rural Studies, 24(3), 322–336.              |
| 760 | Grimm, V., Augusiak, J., Focks, A., Frank, B. M., Gabsi, F., Johnston, A. S. A., et al. (2014).    |
| 761 | Towards better modelling and decision support: Documenting model development,                      |
| 762 | testing, and analysis using TRACE. Ecological Modelling, 280, 129–139.                             |
| 763 | Harmáčková, Z., Pedde, S., Bullock, J. M., Dellaccio, O., Dicks, J., Linney, G., et al. (2022,     |
| 764 | February 16). Improving Regional Applicability of the UK Shared Socioeconomic                      |
| 765 | Pathways Through Iterative Participatory Co-Design.  |
| 766 | https://doi.org/10.2139/ssrn.4010364   |
| 767 | Harrison, P. A., Holman, I. P., Cojocaru, G., Kok, K., Kontogianni, A., Metzger, M. J., &          |
| 768 | Gramberger, M. (2013). Combining qualitative and quantitative understanding for                    |
| 769 | exploring cross-sectoral climate change impacts, adaptation and vulnerability in Europe.           |
| 770 | Regional Environmental Change, 13(4), 761–780.   |
| 771 | Hastie, T. J., & Tibshirani, R. J. (1990). Generalized additive models (Vol. 1, pp. 297–318).      |
| 772 | CRC Press.   |
| 773 | Holman, I. P., Rounsevell, M. D. A., Shackley, S., Harrison, P. A., Nicholls, R. J., Berry, P. M., |
| 774 | & Audsley, E. (2005). A Regional, Multi-Sectoral And Integrated Assessment Of The                  |
| 775 | Impacts Of Climate And Socio-Economic Change In The Uk. Climatic Change, 71(1–2),                  |
| 776 | 9–41.  |
| 777 | Holman, I. P., Rounsevell, M., Berry, P. M., & Nicholls, R. J. (2008). Development and             |
| 778 | application of participatory integrated assessment software to support local/regional              |
| 779 | impact and adaptation assessment. Climatic Change, 90(1), 1-4.                                     |

| 780 | Holman, Ian P., Harrison, P. A., & Metzger, M. J. (2016). Cross-sectoral impacts of climate and     |
|-----|---|
| 781 | socio-economic change in Scotland: implications for adaptation policy. Regional                     |
| 782 | Environmental Change, 16(1), 97–109.  |
| 783 | Holzhauer, S., Brown, C., & Rounsevell, M. (2019). Modelling dynamic effects of multi-scale         |
| 784 | institutions on land use change. Regional Environmental Change, 19(3), 733-746.                     |
| 785 | IUCN National Committee United Kingdom. (2012). Putting Nature on the Map: identifying              |
| 786 | protected areas in the UK. Retrieved from   |
| 787 | https://portals.iucn.org/library/sites/library/files/documents/2012-102.pdf                         |
| 788 | Jarvis, S. G., Redhead, J. W., Henrys, P. A., Risser, H. A., Da Silva Osório, B. M., & Pywell, R.   |
| 789 | F. (2020). CEH Land Cover plus: Pesticides 2012-2017 (England, Scotland and Wales)                  |
| 790 | [Data set]. NERC Environmental Information Data Centre.   |
| 791 | https://doi.org/10.5285/99a2d3a8-1c7d-421e-ac9f-87a2c37bda62  |
| 792 | JNCC. (2020). UK Protected Area Datasets for Download. Retrieved June 28, 2021, from                |
| 793 | https://jncc.gov.uk/our-work/uk-protected-area-datasets-for-download/                               |
| 794 | Kebede, A. S., Nicholls, R. J., Allan, A., Arto, I., Cazcarro, I., Fernandes, J. A., et al. (2018). |
| 795 | Applying the global RCP–SSP–SPA scenario framework at sub-national scale: A multi-                  |
| 796 | scale and participatory scenario approach. The Science of the Total Environment, 635,               |
| 797 | 659–672.  |
| 798 | Kok, K., Pedde, S., Gramberger, M., Harrison, P. A., & Holman, I. P. (2019). New European           |
| 799 | socio-economic scenarios for climate change research: operationalising concepts to                  |
| 800 | extend the shared socio-economic pathways. Regional Environmental Change, 19(3),                    |
|     |   |

801 643–654.

| 802 | Kopp, R. E., Shwom, R. L., Wagner, G., & Yuan, J. (2016). Tipping elements and climate-            |
|-----|--|
| 803 | economic shocks: Pathways toward integrated assessment. Earth's Future, 4(8), 346-                 |
| 804 | 372.   |
| 805 | Kriegler, E., Bauer, N., Popp, A., Humpenöder, F., Leimbach, M., Strefler, J., et al. (2017).      |
| 806 | Fossil-fueled development (SSP5): An energy and resource intensive scenario for the                |
| 807 | 21st century. Global Environmental Change: Human and Policy Dimensions, 42, 297-                   |
| 808 | 315.   |
| 809 | Leclère, D., Obersteiner, M., Barrett, M., Butchart, S. H. M., Chaudhary, A., De Palma, A., et al. |
| 810 | (2020). Bending the curve of terrestrial biodiversity needs an integrated strategy. Nature,        |
| 811 | 585(7826), 551–556.  |
| 812 | Liu, JY., Fujimori, S., Takahashi, K., Hasegawa, T., Wu, W., Geng, Y., et al. (2020). The          |
| 813 | importance of socioeconomic conditions in mitigating climate change impacts and                    |
| 814 | achieving Sustainable Development Goals. Environmental Research Letters: ERL [Web                  |
| 815 | <i>Site]</i> , <i>16</i> (1), 014010.  |
| 816 | Lowe, J. A., Bernie, D., Bett, P., Bricheno, L., Brown, S., Calvert, D., et al. (2018). UKCP18     |
| 817 | Science Overview Report. Retrieved December 9, 2021, from  |
| 818 | https://www.metoffice.gov.uk/pub/data/weather/uk/ukcp18/science-reports/UKCP18-                    |
| 819 | Overview-report.pdf  |
| 820 | Lynn, P., & Knies, G. (2016). UNDERSTANDING SOCIETY The UK Household Longitudinal                  |
| 821 | Study Waves 1-5 Quality Profile. Institute for Social and Economic Research University             |
| 822 | of Essex. Retrieved from   |
| 823 | https://www.understandingsociety.ac.uk/sites/default/files/downloads/documentation/mai             |
| 824 | nstage/quality-profile.pdf   |

| 825 | Martin, W., & Mitra, D. (2001). Productivity Growth and Convergence in Agriculture versus |
|-----|---|
| 826 | Manufacturing. Economic Development and Cultural Change, 49(2), 403–422.                  |

- Meijer, J. R., Huijbregts, M. A. J., Schotten, K. C. G., & Schipper, A. M. (2018). Global patterns
- of current and future road infrastructure. *Environmental Research Letters: ERL [Web*
- *Site*], *13*(6), 064006.
- 830 Merkle, M., Dellaccio, O., Dunford, R., Harmáčková, Z., Harrison, P. A., Mercure, J.-F., et al.
- (2022, February 16). Creating Quantitative Scenario Projections for the UK Shared
   Socioeconomic Pathways. https://doi.org/10.2139/ssrn.4006905
- Met Office Hadley Centre (MOHC). (2018). UKCP18 Regional Projections on a 12km grid over
- the UK for 1980-2080 [Data set]. Retrieved from
- https://catalogue.ceda.ac.uk/uuid/589211abeb844070a95d061c8cc7f604
- Millington, J. D. A., Katerinchuk, V., Bicudo da Silva, R. F., de Castro Victoria, D., & Batistella,
- 837 M. (2021). Modelling drivers of Brazilian agricultural change in a telecoupled world.
- *Environmental Modelling & Software*, 105024.
- Molotoks, A., Smith, P., & Dawson, T. P. (2021). Impacts of land use, population, and climate
- change on global food security. *Food and Energy Security*, *10*(1).
- 841 https://doi.org/10.1002/fes3.261
- Murphy, J. M., Harris, G. R., Sexton, D. M. H., Kendon, E., Bett, P., Clark, R., & Yamazaki, K.
  (2018). UKCP18 land projections: science report. Met Office. Met Office.
- 844 Murray-Rust, D., Brown, C., van Vliet, J., Alam, S. J., Robinson, D. T., Verburg, P. H., &
- 845 Rounsevell, M. (2014). Combining agent functional types, capitals and services to model
- land use dynamics. *Environmental Modelling & Software*, 59, 187–201.

| 847 | Murray-Rust, Dave, Dendoncker, N., Dawson, T. P., Acosta-Michlik, L., Karali, E., Guillem, E.,  |
|-----|---|
| 848 | & Rounsevell, M. (2011). Conceptualising the analysis of socio-ecological systems               |
| 849 | through ecosystem services and agent-based modelling. Journal of Land Use Science,              |
| 850 | 6(2–3), 83–99.  |
| 851 | National Trust. (2021). National Trust Open Data. Retrieved June 28, 2021, from https://uk-     |
| 852 | nationaltrust.opendata.arcgis.com/  |
| 853 | National Trust for Scotland. (2015). National Trust for Scotland Property Boundaries. Retrieved |
| 854 | June 28, 2021, from https://marine.gov.scot/information/national-trust-scotland-property-       |
| 855 | boundaries  |
| 856 | Natural England. (2017). Heritage Coasts (England). Retrieved June 28, 2021, from               |
| 857 | https://naturalengland-defra.opendata.arcgis.com/datasets/heritage-coasts-                      |
| 858 | england/explore?location=52.802383%2C-2.195731%2C6.95&showTable=true                            |
| 859 | Natural England. (2020a). Areas of Outstanding Natural Beauty (England) [Data set]. Retrieved   |
| 860 | from https://data.gov.uk/dataset/8e3ae3b9-a827-47f1-b025-f08527a4e84e/areas-of-                 |
| 861 | outstanding-natural-beauty-england  |
| 862 | Natural England. (2020b). Energy Crops Scheme Agreements Tranches 1 2 [Data set]. Retrieved     |
| 863 | from https://data.gov.uk/dataset/363474ab-0d45-4dff-8857-5fcd35cdf3db/energy-crops-             |
| 864 | scheme-agreements-tranches-1-2  |
| 865 | Natural England. (2020c). National Parks (England) [Data set]. Retrieved from                   |
| 866 | https://data.gov.uk/dataset/334e1b27-e193-4ef5-b14e-696b58bb7e95/national-parks-                |
| 867 | england   |

| 868 | Natural England. (2021a). Local Nature Reserves (England) [Data set]. Retrieved from               |
|-----|--|
| 869 | https://data.gov.uk/dataset/acdf4a9e-a115-41fb-bbe9-603c819aa7f7/local-nature-                     |
| 870 | reserves-england   |
| 871 | Natural England. (2021b). National Nature Reserves (England) [Data set]. Retrieved from            |
| 872 | https://data.gov.uk/dataset/726484b0-d14e-44a3-9621-29e79fc47bfc/national-nature-                  |
| 873 | reserves-england   |
| 874 | Natural England. (2021c). Sites of Special Scientific Interest (England). Retrieved June 28, 2021, |
| 875 | from https://naturalengland-   |
| 876 | defra.opendata.arcgis.com/datasets/f10cbb4425154bfda349ccf493487a80_0/explore?loca                 |
| 877 | tion=52.837148%2C-2.496337%2C6.94  |
| 878 | Natural Resources Wales. (2017a). Heritage Coasts. Retrieved June 28, 2021, from                   |
| 879 | https://datamap.gov.wales/layers/inspire-nrw:NRW_HERITAGE_COAST                                    |
| 880 | Natural Resources Wales. (2017b). National Parks [Data set]. Retrieved from                        |
| 881 | https://data.gov.uk/dataset/949976cb-f952-4405-9fa1-bf531fdca0f5/national-parks                    |
| 882 | Natural Resources Wales. (2018). Local Nature Reserves (LNRs) [Data set]. Retrieved from           |
| 883 | https://data.gov.uk/dataset/c0c66de2-ef27-471f-a501-ebf2713f8649/local-nature-                     |
| 884 | reserves-Inrs  |
| 885 | Natural Resources Wales. (2020). SSSIs. Retrieved June 28, 2021, from                              |
| 886 | https://naturalresourceswales.sharefile.eu/share/view/s7097d5022294fc5b/foe8deca-f112-             |
| 887 | 4e5e-af93-02b2fc71ade3   |
| 888 | Natural Resources Wales. (2021a). Areas of Outstanding Natural Beauty (AONBs) [Data set].          |
| 889 | Retrieved from https://data.gov.uk/dataset/b40871c7-ab45-44f1-8989-                                |
| 890 | 47f872e4a9da/areas-of-outstanding-natural-beauty-aonbs   |

- Natural Resources Wales. (2021b). National Nature Reserves (NNRs) [Data set]. Retrieved from
   https://data.gov.uk/dataset/ce3bdae3-cc24-4fa9-8db0-a1fc2217e995/national-nature-
- 893 reserves-nnrs
- OECD. (2013). Income distribution. https://doi.org/10.1787/data-00654-en
- O'Neill, B. C., Carter, T. R., Ebi, K., Harrison, P. A., Kemp-Benedict, E., Kok, K., et al. (2020).
- Achievements and needs for the climate change scenario framework. *Nature Climate Change*, 1–11.
- 898 ONS. (2017). Health expectancies QMI. Retrieved March 3, 2022, from
- 899 https://www.ons.gov.uk/peoplepopulationandcommunity/healthandsocialcare/healthandli
- 900 feexpectancies/methodologies/healthexpectanciesqmi
- 901 ONS. (2022). Wealth and Assets Survey QMI. Retrieved March 3, 2022, from
- https://www.ons.gov.uk/peoplepopulationandcommunity/personalandhouseholdfinances/
   debt/methodologies/wealthandassetssurveyqmi
- 904 Osório, B., Redhead, J. W., Jarvis, S. G., May, L., & Pywell, R. F. (2019). CEH Land Cover
- 905 plus: Fertilisers 2010-2015 (England) [Data set]. NERC Environmental Information Data
  906 Centre. https://doi.org/10.5285/15f415db-e87b-4ab5-a2fb-37a78e7bf051
- Otto, C., Piontek, F., Kalkuhl, M., & Frieler, K. (2020). Event-based models to understand the
  scale of the impact of extremes. *Nature Energy*, 5(2), 111–114.
- Pearson, R. G., Dawson, T. P., & Liu, C. (2004). Modelling species distributions in Britain: a
  hierarchical integration of climate and land-cover data. *Ecography*, 27(3), 285–298.
- 911 Pedde, S., Kok, K., Hölscher, K., Frantzeskaki, N., Holman, I., Dunford, R., et al. (2019).
- Advancing the use of scenarios to understand society's capacity to achieve the 1.5 degree
- 913 target. *Global Environmental Change: Human and Policy Dimensions*, 56, 75–85.

| 914 | Pedde, S., | . Harrison | . P. A. | . Holman | . I. P. | . Powney. | . G. D. | . Lofts. | S. | , Schmucki | . R. | . et al. | (2021) | ). |
|-----|------------|------------|---------|----------|---------|-----------|---------|----------|----|------------|------|----------|--------|----|
|     |            |            |         |          |         |           |         |          |    |            |      |          |        |    |

- Enriching the Shared Socioeconomic Pathways to co-create consistent multi-sector
  scenarios for the UK. *The Science of the Total Environment*, 756, 143172.
- Polhill, J. G., Gotts, N. M., & Law, A. N. R. (2001). Imitative versus nonimitative strategies in a
- 918 land-use simulation. *Cybernetics and Systems*, *32*(1–2). Retrieved from
- 919 http://www.citeulike.org/user/jamesdamillington/article/2850188
- 920 Pyatt, G. (1995). An ecological site classification for forestry in Great Britain (No. 260).
- 921 Forestry Commission Research Division. Retrieved from
- 922 https://www.forestresearch.gov.uk/documents/4950/RIN260.pdf
- Rabin, S. S., Alexander, P., Henry, R., Anthoni, P., Pugh, T. A. M., Rounsevell, M., & Arneth,
- A. (2020). Impacts of future agricultural change on ecosystem service indicators. *Earth System Dynamics*, 11(2), 357–376.
- Robinson, E. L., Huntingford, C., Semeena, V. S., & Bullock, J. M. (2022). CHESS-SCAPE:
- 927 Future projections of meteorological variables at 1 km resolution for the United Kingdom
- 928 1980-2080 derived from UK Climate Projections 2018 [Data set]. NERC EDS Centre for
- 929 Environmental Data Analysis.
- 930 https://doi.org/10.5285/8194b416cbee482b89e0dfbe17c5786c
- Robinson, Emma L., Blyth, E., Clark, D., Comyn-Platt, E., Finch, J., & Rudd, A. (2017). Climate
- hydrology and ecology research support system meteorology dataset for Great Britain
- 933 (1961-2015) [CHESS-met] v1.2. https://doi.org/10.5285/b745e7b1-626c-4ccc-ac27-
- 934 56582e77b900>

| 935 | Robinson, Emma L., Blyth, E. M., Clark, D. B., Finch, J., & Rudd, A. C. (2017). Trends in     |
|-----|---|
| 936 | atmospheric evaporative demand in Great Britain using high-resolution meteorological          |
| 937 | data. Hydrology and Earth System Sciences, 21(2), 1189–1224.                                  |
| 938 | Rogelj, J., Popp, A., Calvin, K. V., Luderer, G., Emmerling, J., Gernaat, D., et al. (2018).  |
| 939 | Scenarios towards limiting global mean temperature increase below 1.5 °C. Nature              |
| 940 | <i>Climate Change</i> , 8(4), 325–332.  |
| 941 | Rolo, V., Roces-Diaz, J. V., Torralba, M., Kay, S., Fagerholm, N., Aviron, S., et al. (2021). |
| 942 | Mixtures of forest and agroforestry alleviate trade-offs between ecosystem services in        |
| 943 | European rural landscapes. Ecosystem Services, 50, 101318.                                    |
| 944 | Rosen, R. A. (2021). Why the shared socioeconomic pathway framework has not been useful for   |
| 945 | improving climate change mitigation policy analysis. Technological Forecasting and            |
| 946 | Social Change, 166, 120611.   |
| 947 | Rounsevell, M., & Reay, D. (2009). Land use and climate change in the UK. Land Use Policy,    |
| 948 | 26, S160–S169.  |
| 949 | Rounsevell, M., Robinson, D. T., & Murray-Rust, D. (2012). From actors to agents in socio-    |
| 950 | ecological systems models. Philosophical Transactions of the Royal Society of London.         |
| 951 | Series B, Biological Sciences, 367(1586), 259–269.  |
| 952 | Rounsevell, Mark, Fischer, M., Torre-Marin Rando, A., & Mader, A. (2018). The Regional        |
| 953 | Assessment Report on Biodiversity and Ecosystem Services for Europe and Central Asia.         |
| 954 | Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services              |
| 955 | (IPBES).  |
|     |   |

- 956 Rounsevell, Mark, Arneth, A., Brown, C., Cheung, W. W. L., Gimenez, O., Holman, I., et al.
- 957 (2021). Identifying uncertainties in scenarios and models of socio-ecological systems in
  958 support of decision-making. *One Earth*, 4(7), 967–985.
- 859 Rowland, C. S., Morton, R. D., Carrasco, L., McShane, G., O'Neil, A. W., & Wood, C. M.
- 960 (2017). Land Cover Map 2015 (1 km percentage target class, GB). NERC Environmental
  961 Information Data Centre.
- 962 RSPB. (2021). RSPB Reserves. Retrieved June 28, 2021, from https://opendata-
- 963 rspb.opendata.arcgis.com/datasets/6076715cb76d4c388fa38b87db7d9d24\_0/explore?loca
   964 tion=55.360270%2C-3.252783%2C5.99
- Schindler, D. E., & Hilborn, R. (2015). Sustainability. Prediction, precaution, and policy under
  global change. *Science*, *347*(6225), 953–954.
- 967 Schmolke, A., Thorbek, P., DeAngelis, D. L., & Grimm, V. (2010). Ecological models
- 968 supporting environmental decision making: a strategy for the future. *Trends in Ecology & Evolution*, 25(8), 479–486.
- 970 Scoones, I. (1998). Sustainable Rural Livelihoods: A Framework for Analysis. Institute of
- 971 Development Studies.
- 972 Scottish Government. (2020a). Local Nature Reserves (Scotland) [Data set]. Retrieved from

973 https://data.gov.uk/dataset/ff131012-8777-42c9-a263-97cead27ddee/local-nature974 reserves-scotland

- 975 Scottish Government. (2020b). National Nature Reserves (Scotland) [Data set]. Retrieved from
- 976 https://data.gov.uk/dataset/5dae8e31-3ef3-4a2e-8c6c-31068e354c83/national-nature977 reserves-scotland

| 978 | Scottish Government. (2021a). Cairngorms National Park Designated Boundary [Data set].             |
|-----|--|
| 979 | Retrieved from https://data.gov.uk/dataset/8a00dbd7-e8f2-40e0-bcba-                                |
| 980 | da2067d1e386/cairngorms-national-park-designated-boundary  |
| 981 | Scottish Government. (2021b). Loch Lomond and The Trossachs National Park Designated               |
| 982 | Boundary [Data set]. Retrieved from https://data.gov.uk/dataset/6f63d73d-c45d-4947-                |
| 983 | 8ad0-2d6f52b200ff/loch-lomond-and-the-trossachs-national-park-designated-boundary                  |
| 984 | Scottish Government. (2021c). National Scenic Areas [Data set]. Retrieved from                     |
| 985 | https://data.gov.uk/dataset/8d9d285a-985d-4524-90a0-3238bca9f8f8/national-scenic-                  |
| 986 | areas  |
| 987 | Scottish Wildlife Trust. (2016, September 19). Our data. Retrieved June 28, 2021, from             |
| 988 | https://scottishwildlifetrust.org.uk/our-work/our-evidence-base/our-data/                          |
| 989 | Shifley, S. R., He, H. S., Lischke, H., Wang, W. J., Jin, W., Gustafson, E. J., et al. (2017). The |
| 990 | past and future of modeling forest dynamics: from growth and yield curves to forest                |
| 991 | landscape models. Landscape Ecology, 32(7), 1307–1325.   |
| 992 | Siebert, R., Toogood, M., & Knierim, A. (2006). Factors Affecting European Farmers'                |
| 993 | Participation in Biodiversity Policies. Sociologia Ruralis, 46(4), 318–340.                        |
| 994 | Smith, B., Wårlind, D., Arneth, A., Hickler, T., Leadley, P., Siltberg, J., & Zaehle, S. (2014).   |
| 995 | Implications of incorporating N cycling and N limitations on primary production in an              |
| 996 | individual-based dynamic vegetation model. Biogeosciences, 11(7), 2027-2054.                       |
| 997 | SNH. (2020). SNH Natural Spaces - Sites of Special Scientific Interest. Retrieved June 28, 2021,   |
| 998 | from https://gateway.snh.gov.uk/natural-spaces/dataset.jsp?dsid=SSSI                               |

- 999 Synes, N. W., Brown, C., Watts, K., White, S. M., Gilbert, M. A., & Travis, J. M. J. (2016).
- Emerging Opportunities for Landscape Ecological Modelling. *Current Landscape Ecology Reports*, 1(4), 146–167.
- 1002 Synes, N. W., Brown, C., Palmer, S. C. F., Bocedi, G., Osborne, P. E., Watts, K., et al. (2019).
- 1003 Coupled land use and ecological models reveal emergence and feedbacks in socio-1004 ecological systems. *Ecography*, 42(4), 814–825.
- Taylor, K. E., Stouffer, R. J., & Meehl, G. A. (2012). An Overview of CMIP5 and the
   Experiment Design. *Bulletin of the American Meteorological Society*, *93*(4), 485–498.
- 1007 UK Centre for Ecology & Hydrology. (2016). Land Cover Map 2015. Retrieved June 28, 2021,
- 1008 from https://www.ceh.ac.uk/services/land-cover-map-2015
- 1009 UNESCO. (2017). Biosphere Reserves around the World. Retrieved June 28, 2021, from
   1010 http://ihp-
- 1011 wins.unesco.org/layers/mab\_biosphere\_reserves:geonode:mab\_biosphere\_reserves
- 1012 Urban, M. C., Travis, J. M. J., Zurell, D., Thompson, P. L., Synes, N. W., Scarpa, A., et al.
- 1013 (2021). Coding for Life: Designing a Platform for Projecting and Protecting Global

1014 Biodiversity. *Bioscience*. https://doi.org/10.1093/biosci/biab099

- 1015 Vulturius, G., André, K., Swartling, Å. G., Brown, C., Rounsevell, M., & Blanco, V. (2017). The
   1016 relative importance of subjective and structural factors for individual adaptation to
- 1017 climate change by forest owners in Sweden. *Regional Environmental Change*, 1–10.
- 1018 van Vuuren, D. P., Edmonds, J., Kainuma, M., Riahi, K., Thomson, A., Hibbard, K., et al.
- 1019 (2011). The representative concentration pathways: an overview. *Climatic Change*,
- 1020 *109*(1), 5.

| 1021 | Wear, D. N., & Prestemon, J. P. (2019). Spatiotemporal downscaling of global population and  |
|------|--|
| 1022 | income scenarios for the United States. PloS One, 14(7), e0219242.                           |
| 1023 | Weiss, M., & Banko, G. (2018). Ecosystem Type Map v3. 1Terrestrial and marine ecosystems.    |
| 1024 | Technical Paper, 11, 2018.   |
| 1025 | Wiebe, K., Lotze-Campen, H., Sands, R., Tabeau, A., van der Mensbrugghe, D., Biewald, A., et |

- al. (2015). Climate change impacts on agriculture in 2050 under a range of plausible
- 1027 socioeconomic and emissions scenarios. *Environmental Research Letters: ERL [Web*
- 1028 *Site]*, *10*(8), 085010.

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#### Earth's Future

#### Supporting Information for

#### Agent-based modelling of alternative futures in the British land use system

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### Additional Supporting Information (Files uploaded separately)

TRACE ("TRAnsparent and Comprehensive model Evaludation") document uploaded separately to preserve its independent formatting.

#### Introduction

This Supporting Information document provides more details on the methods and results of the CRAFTY-GB implementation. The text builds on and extends the main paper.

#### Methods

#### Model components

#### CAPITALS

Capitals are divided into human, social, manufactured, financial and natural capitals, with natural capital further divided into yields or suitabilities for arable, pastoral and forest land uses or species (Tables S1 & S2). Social, human, financial and manufactured capitals were derived from UK-SSP projections of eight socio-economic indicators from Merkle et al. (2022) (Table S1). Natural capitals were created in two distinct steps. Forest suitabilities were modelled using the Ecological Site Classification (ESC) model originally developed by (Pyatt, 1995) and since used frequently in forestry modelling for the UK (Forest Research, 2021). This model uses data on accumulated temperature, continentality, wind risk, moisture deficit, soil moisture regime, and soil nutrient regime to predict biophysical suitability and associated potential yield class (timber growth) for a range of tree species. In the scenarios, these data were derived from UK-specific RCPs (Robinson et al. 2022).

To project land suitability for arable and pastoral land a General Additive Model (GAM; (Hastie & Tibshirani, 1990) was produced to link land cover classes from Land Cover Map 2015 (Rowland et al., 2017) to UK-RCP covariates. Land Cover target class 3 (arable) and 4 (improved Grassland) were used as the training maps for arable and improved grassland, respectively, whilst semi-natural grassland was trained on LCM target classes 5-7, 9 and 10 (neutral, calcareous and acid grassland; heather; and heather grassland). UK-RCP derived bioclimatic variables for growing degree days (GDD), minimum and maximum temperature, and soil moisture deficit (SMD) and surplus (SMS) were used as covariates, following Pearson et al., ((Pearson et al., 2004). Urban areas were masked out in advance of model training. The baseline map of arable suitability was further processed to take into consideration changes in agricultural yields through time as modelled by the IMPRESSIONS European integrated assessment model (Harrison et al., 2019) and these augmented arable layers were used as a capital layer within the CRAFTY-UK modelling. The two grassland suitability maps were used directly as capital layers within the CRAFTY-UK modelling.

| Capital      | Indicator Variables  | Linear Reso           | aling Thresho      | lds                   |                     |                        |  |
|--------------|--|-----------------------|--------------------|-----------------------|---------------------|------------------------|--|
|              |  | Very Low<br>[0 ; 0.2] | Low<br>[0.2 ; 0.4] | Medium<br>[0.4 ; 0.6] | High<br>[0.6 ; 0.8] | Very High<br>[0.8 ; 1] |  |
| Social       | Income quintile<br>ratio (S80/S20)   | 60 ; 25               | 25 ; 10            | 10;5                  | 5;2                 | 2;1                    |  |
|              | Proportion of<br>people who agree<br>to "people around<br>here are willing to<br>help their<br>neighbours" | 0;30                  | 30 ; 50            | 50 ; 70               | 70 ; 90             | 90 ; 100               |  |
| Human        | Life expectancy at<br>birth  | 30 ; 50               | 50 ; 60            | 60 ; 70               | 70 ; 80             | 80;110                 |  |
|              | Proportion of<br>people aged 25 – 64<br>with tertiary<br>education   | 0;10                  | 10 ; 20            | 20 ; 30               | 30 ; 45             | 45 ; 80                |  |
| Financial    | Household Income<br>per capita [EUR<br>PPS]  | 0;5                   | 5;10               | 10 ; 25               | 25 ; 50             | 50 ; 80                |  |
|              | Proportion of<br>people who agree<br>to "I can save any<br>amount of my<br>income"                         | 0;20                  | 20 ; 30            | 30 ; 40               | 40 ; 50             | 50 ; 100               |  |
| Manufactured | Gross Fixed Capital<br>Formation per Area<br>[mEUR/km <sup>2</sup> ]                                       | 0;0.75                | 0.75 ; 1.25        | 1.25 ; 3              | 3;10                | 10 ; 500               |  |
|              | Average of total<br>speed-weighted<br>road length [Speed-<br>weighted km/km <sup>2</sup> ]                 | 0;0.1                 | 0.1 ; 0.2          | 0.2 ; 0.3             | 0.3 ; 1             | 1;4                    |  |

**Table S1**: Description of socio-economic capitals. For each of the capitals, individual values per area and time slice were formed as means between two indicator variables interpolated between decadal values, and subsequently normalised to [0,1]. Full details of the indicator variables underlying the socio-economic capitals are given in Merkle et al. (2022). Natural capital, split into 11 suitabilities, is described in Table S2.

| Suitability         | Explanation  | Source/reference               |
|---------------------|--|--------------------------------|
| Arable suitability  | GAM-projected arable suitability index (0 to 1)      | GAMs (Hastie and Tibshirani,   |
|                     | based on relationship between bioclimatic            | 1990)                          |
|                     | covariates and LCM target class 3, modified by       | LCM 2015 (Rowland et al.,      |
|                     | changes in arable yields from IMPRESSIONS            | 2017)                          |
|                     | integrated model.                                    | IMPRESSIONS IAP (Harrison et   |
| Improved            | GAM-projected semi-natural grassland suitability (0- | al., 2019)                     |
| grassland           | 1 index) based on relationship between bioclimatic   | Biophysical covariates Pearson |
| suitability         | covariates and LCM target class 4.                   | et al., (2002). See capitals   |
| Semi-natural        | GAM-projected semi-natural grassland suitability (0- | section for full description.  |
| grassland           | 1 index) based on relationship between bioclimatic   |                                |
| suitability         | covariates and LCM target classes 5-7,9 and 10.      |                                |
| Natural: Short      | ESC modelling: Willow yield                          | ESC (Forest Research, 2021)    |
| Rotation Coppice    |  |                                |
| (SRC) suitability   |  |                                |
| Natural: Agro-      | ESC modelling: Sycamore yield                        |                                |
| forestry tree       |  |                                |
| suitability         |  |                                |
| Natural: Non-       | ESC modelling: Sitka spruce yield                    | 1                              |
| native conifer      |  |                                |
| suitability         |  |                                |
| Natural: Non-       | ESC modelling: Beech yield                           |                                |
| native broadleaf    |  |                                |
| suitability         |  |                                |
| Natural: Native     | ESC modelling: Scots pine yield                      | 1                              |
| conifer suitability |  |                                |
| Natural: Native     | ESC modelling: Sessile Oak yield                     | 1                              |
| broadleaf           |  |                                |
| suitability         |  |                                |
| Natural: Native     | ESC modelling: Silver Birch yield                    | 1                              |
| broadleaf           |  |                                |
| suitability         |  |                                |
| Natural: General    | ESC modelling: Combination of all other yields       | 1                              |
| tree species        | ,  |                                |
| suitability         |  |                                |

**Table S2:** Description of Suitabilities comprising natural capital. All are normalised to a [0,1] scale at baseline and are linked to empirical production values through supply normalisation (described below). Abbreviations are as follows: GAM - General additive model; LCM – Land Cover Map: IAP – Integrated Assessment Platform; ESC – Ecological Site Classification.

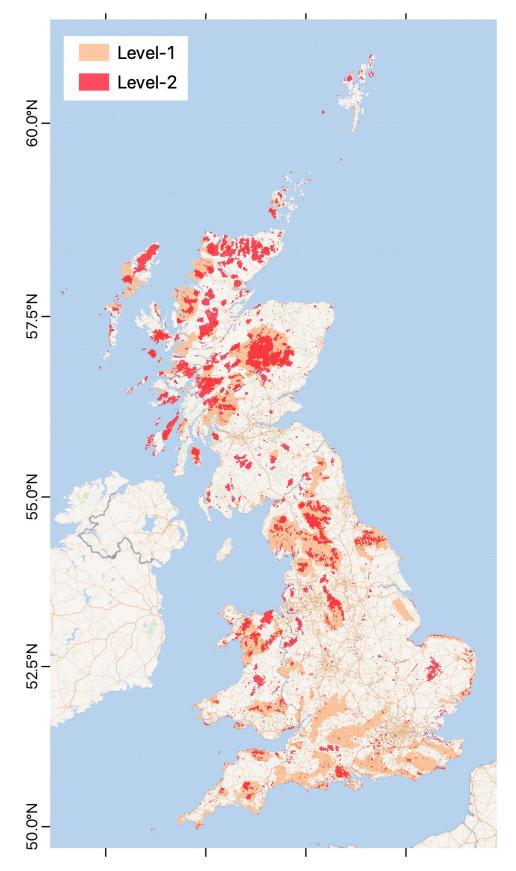
#### PROTECTED AREAS

Protected areas belonging to 11 different types of national and international designation and to 5 different private land-owning organisations (NGOs) were included in the model (Table S3, Fig. S1). Each protected area was first categorised into IUCN Protected Area Management Categories according to the existing categorisation of the (IUCN National Committee United Kingdom, 2012) or, where no existing categorisation was found, according to landowners' stated objectives. Two broad levels of protected area emerged from this classification: IUCN category IV and V areas where many forms of land use are found (all of the officially designated protected areas in the UK), and IUCN category II areas where land use is more tightly controlled (most of the NGO-owned protected areas). We therefore adopted two forms of constraint within the protected areas, with all land use except the most intensive

being permitted in the first group, and no land use change except to the most extensive or conservation management permitted in the second. We also prevented land use change on areas classified as water, bare rock, coastal sediment and marsh in the baseline land use map. Institutions were used to enforce land use protections, and were represented as having complete power and knowledge with which to do so.

| Type of protected area                                   | IUCN<br>category | Data source   | Effect in<br>CRAFTY-GB |
|--|------------------|---|------------------------|
| International  |                  |   |                        |
| Biosphere Reserves                                       | IV               | (UNESCO, 2017)  | Not intensive          |
| Ramsar site  | IV               | (JNCC, 2020)  |                        |
| Special Area of Conservation (SAC)                       | IV               |   |                        |
| Special Protection Area<br>(SPA)                         | IV               |   |                        |
| National   |                  |   |                        |
| Area of Outstanding Natural<br>Beauty (AONB)             | V                | (Natural England, 2020a; Natural Resources<br>Wales, 2021a)                                       | Not intensive          |
| Site of Special Scientific<br>Interest (SSSI)            | IV               | (Natural England, 2021c; Natural Resources<br>Wales, 2020; SNH, 2020)                             |                        |
| Heritage Coast (HC)                                      | V                | (Natural England, 2017; Natural Resources<br>Wales, 2017a)  | _                      |
| Local Nature Reserve (LNR)                               | IV               | (Natural England, 2021a; Natural Resources<br>Wales, 2018; Scottish Government, 2020a)            | _                      |
| National Nature Reserve<br>(NNR)                         | IV               | (Natural England, 2021b; Natural Resources<br>Wales, 2021b; Scottish Government, 2020b)           |                        |
| National Park (NP)                                       | V                | (Natural England, 2020c; Natural Resources<br>Wales, 2017b; Scottish Government, 2021a,<br>2021b) |                        |
| National Scenic Area (NSA)                               | V                | (Scottish Government, 2021c)  |                        |
| NGOs   |                  |   |                        |
| John Muir Trust (JMT)                                    | П                | JMT, personal communication   | No Change              |
| National Trust / National<br>Trust for Scotland (NT/NTS) | V                | (National Trust, 2021; National Trust for Scotland, 2015)   |                        |
| RSPB   | П                | (RSPB, 2021)  |                        |
| Scottish Wildlife Trust                                  | II               | (Scottish Wildlife Trust, 2016)   |                        |
| Other NGO  | 11               | Trees for Life, personal communication  |                        |

**Table S3:** Types of protected area included in the model, their equivalent IUCN ranking (taken from (IUCN National Committee United Kingdom, 2012) or determined based on management objectives), data sources and the modelled constraint each type of protected area places on land use change.



**Fig. S1:** Protected areas applied in CRAFTY-GB. The map is projected using OSGB1936 / British National Grid coordinate reference system (EPSG: 27700). The background map is provided by Wikimedia (https://maps.wikimedia.org/)

#### LAND USES (AGENT TYPES)

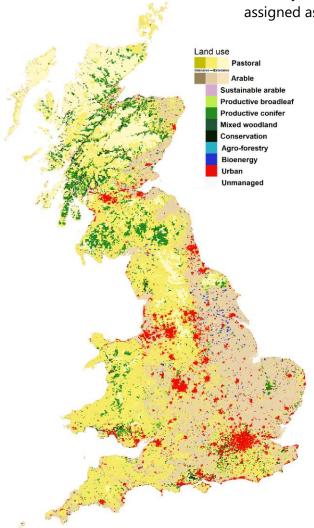
CRAFTY-GB includes a range of agent types designed to capture the main forms of land use in Great Britain, including gradations of intensity and multi-functionality. Agent types were divided between arable land uses (intensive arable for food, intensive arable for fodder, sustainable arable and extensive arable), pastoral land uses (intensive pastoral, extensive pastoral, very extensive pastoral), forest land uses (productive native conifer, productive nonnative conifer, productive native broadleaf, productive non-native broadleaf, multifunctional mixed woodland and native woodland for conservation), and combined classes (bioenergy and agroforestry) (Table S4). Variation in ecosystem service provision within these classes allows them to represent a continuous range of forms of land management rather than arbitrarily distinct groups.

Allocation of the initial distribution of land uses was based on the 2015 Land Cover Map (LCM2015) produced by the (UK Centre for Ecology & Hydrology, 2016) (Rowland et al., 2017) and the National Forest Inventory (NFI) 2010-2015 (Forestry Commission, 2021). Further datasets were used to define the extent and location of specific land uses, and full details are given in Table S4. Urban areas were derived from land cover data at the baseline (LCM 2015) and then projected in the scenarios by an independent urban model (described in detail in Merkle et al., in review). This model created 1km gridded urban surface projections through a newly developed urban allocation algorithm based on a neighbourhood density function, SSP-specific sprawl parameter settings, and SSP-specific land exclusions of protected areas and flood risk areas. Land not otherwise used was modelled as unmanaged. In some cases, input data were incomplete and had to be further processed before being used. This was true of some coastal areas and islands (particularly estuaries and the Shetland isles). Data gaps in Shetland were filled using a simple regression model using topographic variables (i.e. elevation, slope, and aspect) trained upon the data of the nearest Orkney island. Gaps in coastal areas were filled using nearest-neighbour values. We also used 5x5 moving average interpolation to smooth hard boundaries between administrative units in the capitals. Finally, where scenario input data for 2020 were not consistent with baseline data, those data series were normalised by the equivalent baseline values.

| Land use (agent)           | Notes  | Initial allocation   |
|----------------------------|--|--|
|                            |  |  |
| Intensive arable           | Farmers managing intensively for crop production for           | Allocated to LCM2015 aggregate class 'Arable'. Food  |
| (food)                     | food.  | and fodder types distributed randomly within that  |
| Intensive                  | Farmers managing intensively for crop production for           | according to (modelled) baseline demand levels to  |
| <b>arable</b> (fodder)     | livestock fodder, ultimately producing meat and milk.          | provide the required amount of each  |
| Sustainable                | Farmers managing organically or otherwise less                 | Allocated to LCM2015 aggregate class 'Arable' to   |
| arable                     | intensively for crop production for food                       | give an area coverage equal to the 2015 area of  |
|                            |  | organic arable in the UK (as reported by (DEFRA,   |
|                            |  | 2016a), with specific cells chosen spatially randomly  |
| Extensive arable           | Farmers managing with few inputs for limited crop              | Allocated to the LCM2015 aggregate class 'Arable'  |
|                            | production for food; equivalent to subsistence                 | cells within the lowest 10% of modelled suitability  |
|                            | production where capitals are very low                         | for arable crops   |
| Intensive                  | Farmers managing intensively for livestock                     | LCM2015 Improved grassland   |
| pastoral                   |  |  |
| Extensive                  | Farmers managing extensively for livestock on semi-            | LCM2015 Semi-natural grassland   |
| pastoral                   | natural grassland  |  |
| Very extensive<br>pastoral | Minimal management involving some grazing                      | LCM2015 Mountain, heath, bog and LCM2015 Semi-   |
| •                          | Dedicated modulation of Chart Datation Councils (              | natural grassland (Fen Marsh Swamp)  |
| Bioenergy                  | Dedicated production of Short Rotation Coppice /<br>Miscanthus | Assigned to LCM2015 aggregate class 'Arable' to  |
|                            | Wiscantinus  | cover the 2015 extent of arable bioenergy land   |
|                            |  | (DEFRA, 2016b), assigned to locations of Energy<br>Crops Scheme (Tranche 2) agreements 2013-2015 |
|                            |  | (Natural England, 2020b)   |
| Agroforestry               | Farmers practicing silvo-pastoral or silvo-arable forms of     | NFI 'low-density' class when otherwise unassigned.   |
| Agroiorestry               | agroforestry, combining trees with either grazing or           | NET low-density class when otherwise dhassigned.   |
|                            | crops, for timber, crop and livestock production.              |  |
| Productive non-            | Production-focused forest managers with non-native             | LCM2015 Coniferous woodland class, sub-divided by  |
| native conifer             | conifer plantations. Primary objective is softwood             | NFI Conifer class, located where modelled suitability  |
| native conner              | timber production.   | is higher for non-native than for native species   |
| Productive non-            | Production focused forest managers with non-native             | LCM2015 Broadleaf woodland, sub-divided by NFI   |
| native broadleaf           | broadleaf plantations. (Not currently common, but felt         | broadleaf class, located where modelled suitability is   |
|                            | to have potential importance in the future). Primary           | higher for non-native than for native species  |
|                            | objective is hardwood timber production.                       | higher for hor harve than for harve species  |
| Productive native          | Production focused forest managers with native conifer         | ICM2015 Coniferous woodland, sub-divided by NFI  |
| conifer                    | plantations. Primary objective is softwood timber              | Conifer class, located where modelled suitability is   |
|                            | production.  | higher for native than for non-native species  |
| Productive native          | Production focused forest managers with native                 | LCM2015 Broadleaf woodland, sub-divided by NFI   |
| broadleaf                  | broadleaf plantations. Primary objective is hardwood           | broadleaf class, located where modelled suitability is   |
|                            | timber production.   | higher for native than for non-native species.   |
| Multifunctional            | Forest managers with mixed woodlands and multiple              | LCM2015 Broadleaf or Coniferous woodland,  |
| mixed woodland             | objectives practising low-intensity management                 | subdivided by NFI mixed classes  |
| Native woodland            | Conservation focused forest managers. Primary                  | LCM2015 Broadleaf or coniferous woodland,  |
| (conservation)             | objective is to conserve biodiversity.                         | excluding NFI classes indicating active management   |
| , ,                        |  | or no forest cover, and located where modelled   |
|                            |  | broadleaf suitability is within the lowest 50% or  |
|                            |  | modelled conifer suitability is within the lowest 10%  |
| Urban                      | Urban and industrial areas                                     | Modelled separately  |
| Unmanaged                  | Represents areas with minimal to no management,                | Unassigned cells   |
|                            | often where biophysical conditions preclude significant        |  |
|                            | productivity e.g. high montane or deep peat areas              |  |
|                            |  |  |

**Table S4:** Allocation of initial distribution of land uses. Levels of intensity are assigned discretely in terms of agent types, but modelled continuously across these types according to availability and usage of agricultural inputs and production levels. The resulting allocation is shown in Fig. S2

**Figure S2:** Simplified baseline allocation of land uses. Pastoral and arable uses are presented on coloured intensity gradients, and are concentrated towards the extensive end of the gradient because intensity becomes greater in some scenarios (intensity values are assigned as described in the text).



#### **BEHAVIOURS**

CRAFTY-GB is designed to represent many forms of behaviour relating to land management decision-making through a small number of generic parameters, described in Table S5.

| Parameterised<br>behaviour | Description   | Interpretation  |
|----------------------------|---|---|
| Capital<br>sensitivities   | Quantification of agent dependence on each<br>capital for the production of a service.<br>Variation at individual and typological levels.   | Represents agent abilities to utilise<br>capitals (e.g. through particular<br>production methods), reliance on<br>supporting capitals (e.g. social support<br>systems) and access to personal<br>resources (e.g. additional labour).                              |
| Productive<br>abilities    | The maximum potential service production an<br>agent can achieve under perfect capital<br>conditions. Variation at individual and<br>typological levels.  | Represents the ability and willingness of<br>agents to provide ecosystem services,<br>including potential decisions about<br>trade-offs between services made on<br>the basis of agent preferences.   |
| Search ability             | Comprising three parameters: the number of<br>search iterations an agent type can undertake<br>per timestep, the number of cells considered<br>for competition during each search iteration,<br>and the order (random or ranked) in which<br>those cells are competed for. Variation at<br>typological level. | Represents the ability and willingness of<br>agents to seek new land to manage, and<br>their knowledge about the potential<br>productivity of that land.  |
| Abandonment<br>threshold   | Minimum benefit level an agent will accept<br>before abandoning land. Variation at<br>individual and typological levels.  | Represents agents' dedication to their<br>land use in the absence of more<br>beneficial alternatives. Can incorporate<br>risk aversion, 'traditionalist' attitudes,<br>cultural norms etc.  |
| Competition<br>threshold   | Maximum relative competitive disadvantage<br>in benefit values that an agent will tolerate<br>before relinquishing land to another land use<br>agent. Variation at individual and typological<br>levels.  | Represents agents' dedication to their<br>land use under competition from more<br>beneficial alternatives. Can incorporate<br>similar factors as the abandonment<br>threshold, as well as opportunity costs<br>and more specific aversions to other<br>land uses. |
| Social networks            | An additional component of the model,<br>representing social links between agents of<br>each type located within a defined circular<br>neighbourhood of one another. Settings<br>control neighbourhood radius, other<br>parameters that effects act upon, and<br>magnitudes of those effects.                 | Represents social support or norms,<br>knowledge diffusion, economies of scale<br>or any other spatially-mediated<br>interaction between agents   |

#### Table S5: Behavioural effects included in CRAFTY-GB

Of these behaviours, social networks are the only new addition to the CRAFTY framework, and function as follows. Agent types each have a defined neighbourhood within which influences can occur. Neighbourhoods have a default 20 km radius, based on evidence that a neighbourhood of this size best captures diffusion effects in the uptake of land management options and policies in the UK (Brown et al., 2018). Within each neighbourhood, the density of agents of the same type is calculated at each timestep, and this density is used to rescale

other parameter values. Here, density affects the competitiveness of agents, with increasing competitiveness when density is high to represent the benefits both of improved local knowledge diffusion and of economies of scale. The magnitude of this effect and the size of the social neighbourhood are varied in the scenarios as described in the main text (and below).

#### SERVICES & DEMAND LEVELS

A range of provisioning, regulating and cultural ecosystem services and other indicators (e.g. biodiversity, employment) of relevance to the UK-SSP scenarios were modelled. These services are defined in Table S6, and their provision by different agent types based on capital levels is presented in Table S7. In this implementation, the relative calibration of service provision is approximate and largely assumption-based, though informed by empirical or modelled evidence where possible.

| Services               | Details   |
|------------------------|---|
| Food crops             | Crops for human consumption                                 |
| Fodder crops           | Crops for consumption by ruminant and monogastric livestock |
| Grass-fed meat         | Red meat produced in pastoral systems                       |
| Grass-fed milk         | Milk produced in pastoral systems                           |
| Bioenergy fuel         | Bioenergy crops; short rotation coppice & miscanthus        |
| Softwood               | Softwood (conifer) timber                                   |
| Hardwood               | Hardwood (broadleaf) timber                                 |
| Biodiversity           | Biological diversity  |
| Landscape diversity    | Diversity of landscape elements                             |
| Carbon sequestration   | Quantity of carbon sequestered (above & below ground)       |
| Recreation             | Recreation potential  |
| Flood Regulation       | Land ability to store water                                 |
| Employment             | Potential for employment associated with land management    |
| Sustainable production | Abstract service providing sustainability in agriculture    |

**Table S6:** Goods and services modelled in CRAFTY-GB. The ability of agents to produce these services given certain capital values is presented in Table S7.

In modelling production of crops and livestock products, we assume divisions between crop production for direct human consumption, crop production for livestock consumption, and grass-fed livestock production. We assume that pastoralist agents produce grass-fed milk (intensive pastoral only) and red meat, while 'arable for fodder' agents effectively produce crop-fed red and white meat, and milk. Monogastrics are gramivores, so are fed only from cropland. Evidence for production levels includes an existing application of the CRAFTY framework to Scotland (Burton et al., in prep), and literature evidence on ecosystem services provision in different land use types (Burton et al., 2018; Rolo et al., 2021).

**Table S7:** Capital sensitivities and service production levels of each modelled land use (agent type). Capital sensitivities determine how reliant each form of land use is on certain characteristics of the land system (as defined in the Capitals section), and service production levels determine the relative quantity of each service produced when capitals are not limiting. Scales here are approximate and relative within each capital, and are subject to small amounts of variation across scenarios; these are described further in the Scenarios section and complete absolute values are given in the relevant production files (see data availability section). Urban and unmanaged land uses are not actively modelled and do not use capitals or produce services. Abbreviations are as follows: Capitals H=human capital; S=social capital; M=manufactured capital; F=financial capital; Ar=arable suitability; IG=intensive grassland suitability; SNG=semi-natural grassland suitability; Bi=bioenergy suitability; AF=agro-forestry suitability; NNC= non-native conifer suitability; ND=non-native broadleaf suitability; Tr=tree suitability. Services Food=food crops; Fodder=fodder crops; GF meat=grass-fed meat; GF milk=grass-fed milk; Fuel=bioenergy fuel; SW= softwood; HW=hardwood; BD=biodiversity; LD=landscape diversity; C=carbon sequestration; Rec=recreation; FI. reg.=flood regulation; Emp= employment; SusP=sustainable production.

| Land use (agent)                    |   |   | vity |      |    |    |     |    |    |     |     |    |    |    | Produc | ction of se | rvices |      |      |    |    |    |    |   |     |      |     |          |
|-------------------------------------|---|---|------|------|----|----|-----|----|----|-----|-----|----|----|----|--------|-------------|--------|------|------|----|----|----|----|---|-----|------|-----|----------|
|                                     | Н | S | -    | <br> | Ar | IG | SNG | Bi | AF | NNC | NNB | NC | NB | Tr | Food   | Fodder      | GF     | GF   | Fuel | SW | HW | BD | LD | С | Rec | FI.  | Emp | SusP     |
|                                     |   | - |      | _    |    |    |     |    |    |     |     |    |    |    |        |             | meat   | milk |      |    |    |    |    |   | -   | reg. |     | <u> </u> |
| Intensive arable<br>(food)          |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Intensive<br>arable (fodder)        |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Sustainable arable                  |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Extensive arable                    |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Intensive pastoral                  |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    | 1 |     |      |     |          |
| Extensive pastoral                  |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Very extensive                      |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| pastoral                            |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Bioenergy                           |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Agroforestry                        |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Productive non-<br>native conifer   |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Productive non-<br>native broadleaf |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Productive native<br>conifer        |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Productive native<br>broadleaf      |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Multifunctional<br>mixed woodland   |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Native woodland<br>(conservation)   |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Urban                               |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |
| Unmanaged                           |   |   |      |      |    |    |     |    |    |     |     |    |    |    |        |             |        |      |      |    |    |    |    |   |     |      |     |          |

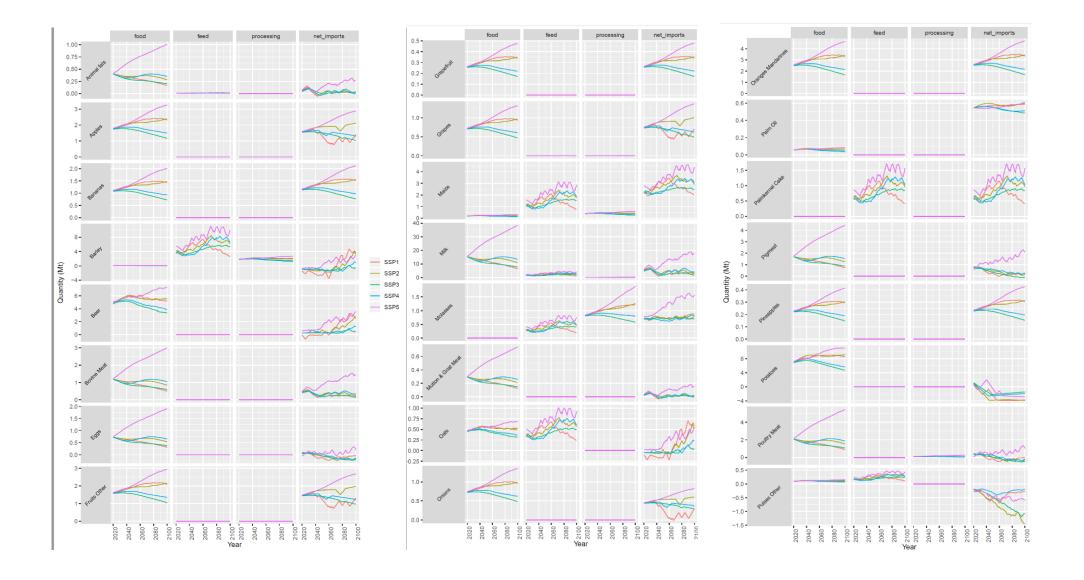
#### Approximate relative value



Non-food demands were taken from the stakeholder-defined scenarios, and are described in (Merkle et al. 2022). Demand levels for foods were derived from the LandSyMM (Land System Modular Model; www.landsymm.earth) global modelling framework (Rabin et al., 2020). Within LandSyMM, the dynamic global vegetation model LPJ-GUESS simulates physiological, demographic, and disturbance processes for a variety of plant functional types (Smith et al., 2014), while the land use model PLUM simulates land use and management based on global trade and cell-level (0.5°) productivity (Alexander et al., 2018). Food demand was calculated from scenario projections of country-level population and gross domestic product (GDP), using the historical relationship of per capita GDP to consumption of each of six crop types -C3 cereals, C4 cereals, rice, oil crops, pulses, and starchy roots - plus ruminant and monogastric livestock. Separate demand levels were calculated for food crops for human consumption and for feed for monogastric livestock and ruminant livestock not raised on pasture. Both types of demand account for crops used for processing, seed stocks, and for losses sustained during the production process. Demands were also adjusted to take account of imports and exports, as calculated by PLUM. Demand levels broken down by food type are shown in Fig. S3.

In the case of CRAFTY-GB, the total food production of the UK simulated by LandSyMM was taken as the national demand (i.e. aggregated from the 0.5° grid that LandSyMM uses). Because the simulated LandSyMM baseline (representing the year 2020) is not based on land cover data, while the baseline land allocation of CRAFTY-GB is, all LandSyMM demands were normalised relative to their 2020 values, giving a continuous series of annual changes in demand levels as proportions of 2020 demand.

First the domestic production of feed and food crops was calculated. Food crops scale with the production of agents in CRAFTY, from a baseline quantity of 35.65Mt of crops (an average of 771 tonnes for each of the 46,252 purely arable agents in CRAFTY-GB at the baseline, including subsequent losses, processing and seeds etc.). Feed crops were converted to livestock products through product-specific Feed Conversion Ratios taken from (Alexander et al., 2016). Monogastrics are fed exclusively on these feed crops (including those imported), meaning that the demands for Mt of pork, poultry and eggs could be immediately converted into demands for Mt of feed crops. Ruminant livestock (according to demands for Beef, Mutton, Goat and milk) were similarly converted, and the remaining available feed crops were assigned proportionally to them. Leftover demand for these livestock products was converted to a pasture demand by scaling from the baseline, and for comparison by using an additional pasture food conversion ratio.



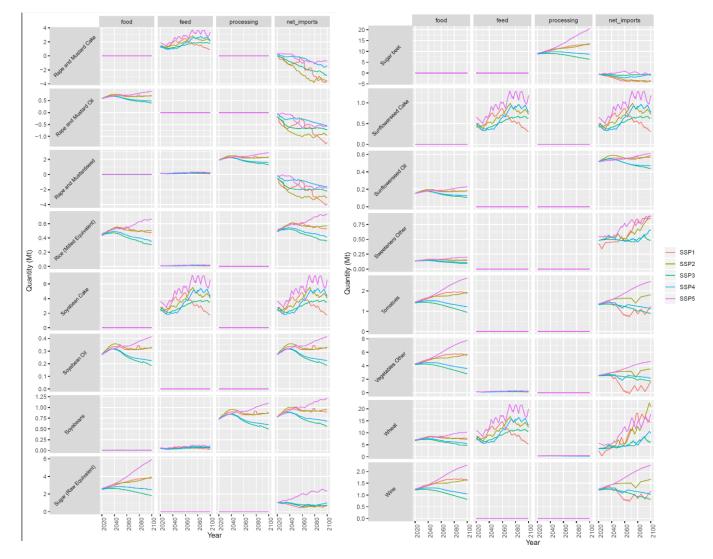


Figure S3: Food commodity demand levels supplied by the PLUM model as part of LandSyMM, prior to conversion for use in CRAFTY-GB

#### Scenarios

We use combinations of the Representative Concentration Pathways (RCP) climate scenarios (van Vuuren et al., 2011) and Shared Socioeconomic Pathways (SSP) socio-economic scenarios (O'Neill et al. 2017). A combined set of these scenarios was specified for the British context through a combination of stakeholder engagement and computational or statistical modelling.

The SSPs were specified for the UK as described in Pedde et al. (2021), Harmáčková et al. (2022) and Merkle et al. (2022). These substantial extensions of the global SSPs provide detailed narratives of social, economic and political developments across the UK until 2100. The narratives integrate national stakeholder knowledge on locally-relevant drivers and indicators with higher level information from the European and global SSPs. These narratives were simplified and converted into model parameterisations (Fig. 2, Table S8). SSPs were put in a global context through LandSyMM global land system modelling to provide consistency with the broader SSP framework and to account for the UK's international trade.

The SSP implementation also utilised the forms of behaviour represented in CRAFTY to capture land management decision-making (Table S5). Of these behaviours, social networks are the only new addition to the CRAFTY framework. These allow agents of the same type to affect one another's competitiveness within defined spatial neighbourhoods, to represent the benefits both of improved local knowledge diffusion and of economies of scale, and are described above (SI section 'BEHAVIOURS').

Climatic conditions are taken from the CHESS-SCAPE data set, which provides several climate variables at 1 km2 spatial resolution and several temporal resolutions, from daily to decadal. CHESS-SCAPE is derived from the 12 km2 resolution UKCP18 regional predictions for the UK. UKCP18 regional predictions were obtained by running a perturbed parameter ensemble of a regional climate model (RCM), nested within a global climate model (GCM) for RCP8.5 (Murphy et al., 2018). CHESS-SCAPE was derived from this regional data set by: (i) downscaling from 12 km2 to 1 km2 using a modified version of the CHESS methodology (Emma L. Robinson, Blyth, Clark, Finch, et al. 2017); (ii) bias-correcting to observed historical climate using the CHESS-met dataset (Emma L. Robinson, Blyth, Clark, Comyn-Platt, et al. 2017); and (iii) time-shifting and pattern scaling to provide RCPs 2.6, 4.5 and 6.0, using members of the CMIP5 ensemble to define target trajectories of global temperature change (Taylor et al., 2012). Full details can be found in (Robinson et al. 2022). The highest temporal resolution of CHESS-SCAPE is daily. From these were calculated 20-year mean-monthly climatologies, at a 10-year time-step, giving spatially and temporally explicit values for several climate variables for the UK, including temperature and precipitation. The climate variables were used to calculate Penman-Monteith potential evapotranspiration with interception correction (PETI), following the method of Robinson et al. (2017). This is potential evapotranspiration calculated for a short grass, with a correction applied on rain days to account for the greater efficiency of evaporation of water from the canopy surface before it can reach the soil. The air temperature was used to calculate growing degree days (GDD), which is a count of the number of days for which mean air temperature was greater than 5°C. The air temperature, precipitation, PETI and GDD were then used as inputs to the crop, grassland and forest modelling to produce annual scenario-specific capital values.

RCP-SSP combinations were chosen to: (i) cover a broad range of uncertainty in both emissions (and hence climate) and socio-economic developments; and (ii) include any combination of SSPs and RCPs that is plausible, meaningful and useful. The six combined scenarios we use (RCP2.6-SSP1, RCP4.5-SSP2, RCP4.5-SSP4, RCP6.0-SSP3, RCP8.5-SSP2, RCP8.5-SSP5) cover weak to strong climate change, as well as future societies with high and low challenges to adaptation and mitigation. The selection also allows analysis of the effects of different RCPs within the same SSP (RCPs 4.5 and 8.5 with SSP2), and the effects of different SSPs within the same RCP (SSPs 2 and 4 with RCP4.5; SSPs 2 and 5 with RCP8.5). Furthermore, low adaptation challenges (SSP1/5) and high adaptation challenges (SSP3/4) are confronted with different RCPs.

The model components and inputs described above were used to produce coherent representations of the UK-RCP-SSPs. These representations are summarised in Fig. 2 and in Table S8.

| Scenario                        | Description   | Implementation   |  |   |   |   |                |  |  |  |  |  |  |
|---------------------------------|---|--|--|---|---|---|----------------|--|--|--|--|--|--|
|                                 |   | Behaviour  | Capitals   | Demand levels   | Valuation   | Production  | Other          |  |  |  |  |  |  |
| SSP1 -<br>Sustainability        | UK-SSP1 shows the UK<br>transitioning to a fully functional<br>circular economy as society<br>quickly becomes more egalitarian<br>leading to healthier lifestyles,<br>improved well-being, sustainable<br>use of natural resources, and<br>more stable and fair international<br>relations. It represents a<br>sustainable and co-operative<br>society with a low carbon<br>economy and high capacity to<br>adapt to climate change.                | Social networks<br>add up to 10%<br>to agent<br>competitiveness<br>Agents more<br>likely to change<br>or abandon<br>land use, except<br>for very<br>extensive<br>management<br>consistent with<br>conservation | Arable and<br>intensive<br>grassland<br>productiviti<br>es +20% by<br>2070<br>Social<br>capital<br>increases | At least 60% of<br>ruminant<br>products from<br>grass-fed systems<br>Higher demand<br>for sustainable<br>food<br>Higher per capita<br>demands for<br>timber,<br>biodiversity,<br>carbon, recreation   | Benefit<br>values for<br>non-food<br>services<br>x1.5 | Grass-fed meat & milk<br>productivity +10%<br>Agro-forestry agents +10%<br>productivity of main services<br>Sustainable/extensive<br>production levels benefit<br>more from increases in<br>manufactured capital<br>Extensive & multifunctional<br>agents have less dependence<br>on financial capital (-20%)                       |                |  |  |  |  |  |  |
|                                 |   |  |  | Higher demand<br>for sustainable<br>agriculture   |   |   |                |  |  |  |  |  |  |
| SSP2 –<br>Middle of the<br>Road | UK-SSP2 is a world in which strong<br>public-private partnerships<br>enable moderate economic<br>growth but inequalities persist. It<br>represents a highly regulated<br>society that continues to rely on<br>fossil fuels, but with gradual<br>increases in renewable energy<br>resulting in intermediate<br>adaptation and mitigation<br>challenges.  | Agents more<br>likely to change<br>or abandon<br>land use<br>Social networks<br>add up to 4% to<br>competitiveness   | Arable &<br>intensive<br>grass<br>productiviti<br>es +20% by<br>2070   | Increased per<br>capita demands<br>for timber (+40%),<br>carbon (+40%),<br>bioenergy (+20%),<br>water regulation<br>(+20%) &<br>recreation (+20%)<br>and sustainable<br>ag. Products<br>(+50%)<br>Min. 50% of<br>ruminant<br>products from<br>grass-fed systems |   | Sustainable/extensive<br>production levels benefit<br>more from increases in<br>manufactured capital  |                |  |  |  |  |  |  |
| SSP3 –<br>Regional<br>rivalry   | The dystopian scenario, UK-SSP3,<br>shows how increasing social and<br>economic barriers may trigger<br>international tensions,<br>nationalisation in key economic<br>sectors, job losses and, eventually<br>a highly fragmented society with<br>the UK breaking apart. It<br>represents a society where rivalry<br>between regions and barriers to<br>trade entrench reliance on fossil<br>fuels and limit capacity to adapt<br>to climate change. | Individual-level<br>randomness in<br>agent<br>characteristics<br>Social networks<br>operate over<br>smaller (5km)<br>radius, with<br>smaller (max<br>+2%) effect on<br>competitiveness                         | Arable &<br>intensive<br>grass<br>suitabilities<br>-20% by<br>2100   | Demand for<br>sustainable ag,<br>biodiversity,<br>carbon, I diversity<br>-80%   | Food<br>productio<br>n benefit<br>5x non-<br>food     | Heavy reliance on<br>manufactured capital to<br>follow input availability<br>(agricultural products)<br>Services can only be supplied<br>within-nation; no trade<br>between parts of UK<br>(demands scaled by<br>population)<br>Intensive agents produce -<br>50% secondary services<br>Biodiversity lower production<br>all (-50%) | PAs<br>removed |  |  |  |  |  |  |

| SSP4 -<br>Inequality                     | UK-SSP4 shows how a society<br>dominated by business and<br>political elites may lead to<br>increasing inequalities by<br>curtailing welfare policies and<br>excluding the majority of a<br>disengaged population. The<br>business and political elite<br>facilitate low carbon economies<br>but large differences in income<br>across segments of UK society<br>limits the adaptive capacity of the<br>masses.      | Intensive agents<br>less likely to<br>give up or give<br>in<br>Social networks<br>add up to 10%<br>to<br>competitiveness | Arable &<br>intensive<br>grass<br>suitability<br>values<br>+20% by<br>2070 | Fuel (bioenergy),<br>timber demands<br>200% by 2070<br>Recreation &<br>biodiversity -20%<br>by 2040, static<br>thereafter<br>Sustainable ag<br>demand -50% | All<br>services<br>have<br>lower<br>benefit<br>due to<br>lack of<br>ability to<br>pay (-<br>50%) | Extensive agents produce<br>less due to lack of support (-<br>10%)<br>Greater reliance on (benefit<br>from) manufactured capital<br>in forestry (+20%)  | PAs<br>removed<br>in 2050 |
|--|--|--|--|--|--|---|---------------------------|
| SSP5 – Fossil-<br>fuelled<br>development | UK-SSP5 shows the UK<br>transitioning to a highly<br>individualistic society where the<br>majority become wealthier<br>through the exploitation of<br>natural resources combined with<br>high economic growth. It<br>represents a technologically<br>advanced world with a strong<br>economy that is heavily<br>dependent on fossil fuels, but<br>with a high capacity to adapt to<br>the impacts of climate change. | Social networks<br>add up to 10%<br>to<br>competitiveness<br>thresholds<br>allow more<br>change                          | Arable &<br>intensive<br>grass<br>suitability<br>values<br>+40% by<br>2070 | Recreation<br>demand +20%<br>Sustainable ag<br>demand -80%   | Food<br>productio<br>n benefit<br>3x non-<br>food  | Intensive production more<br>reliant on manufactured<br>capital (+20%)<br>Recreation not reliant on<br>infrastructure (manufactured<br>capital)<br>Lower levels of secondary<br>services in intensive<br>agriculture (-10%) | PAs<br>removed            |

Table S8: Descriptions and summary of the implementation of the UK-SSPs

#### INTENSITY REPRESENTATION

To improve the interpretability of the results, we developed a land use intensity mapping approach. This involved the assignment of values on a continuous range for each of the arable, and each of the pastoral (except very extensive pastoral) classes across the scenarios. Intensity values were defined as a combination of the use of agricultural inputs (fertilisers, pesticides and machinery), technology, and modelled production levels. For the purposes of illustration these are combined multiplicatively here and used to select colour saturation levels in the map figures.

Alternative representations are possible, and it is important to note that our presentation does not distinguish the specific use of technology to reduce the use of chemical inputs, as in UK-SSP1. This method does however make scenarios results more comparable and means that differences in land management intensities among the scenarios are readily apparent.

#### Results

Agent typology evaluation

This section describes the comparison of the CRAFTY-GB agent typology with different (semi-) independent datasets on land cover (LCM 2015), habitat characteristics (EUNIS habitats), and agricultural intensity (UK CEH Land Cover Plus: Fertilisers and Pesticides data). All comparisons were made between the baseline AFT map (1km<sup>2</sup> spatial resolution) and maps of the respective datasets at their native resolution.

#### LCM 2015

The first comparison was between the baseline AFT map (1km<sup>2</sup>) and the LCM 2015 fractional land cover (1km<sup>2</sup>) to check the consistency of the baseline AFT allocation as described in Table S4. Full results of the comparison are shown in Figure S4, with a summary provided in Figure S5.

As the LCM 2015 dominant land cover map was the main source for the baseline AFT allocation there is generally a good agreement between AFTs and LCM land-cover classes, although with large variations across individual grid cells (Fig S5). Intensive agricultural AFTs show the highest fractions of arable land or improved grassland with only small contributions from woodlands and other classes, indicating a good representation of rather homogeneous agricultural landscapes within these AFTs. In contrast extensive agricultural AFTs are often associated with a mixture of agricultural and different semi-natural LCM classes (Fig S4). The AFT 'Sustainable arable' has been allocated randomly within the agricultural cells in the baseline, therefore not showing substantial differences to the intensive types. Broadleaved forest types are usually associated with a substantial amount of arable land and improved grasslands, but less mixed with conifer classes, indicating a clear distinction between the forest types in the allocation. However, broadleaved woodlands seem to represent more heterogeneous landscapes compared to coniferous woodlands. As expected, the most heterogeneous landscapes (with regard to land-cover composition) were found in multifunctional and native woodland agents. The agroforestry AFT does not have an equivalent in the LCM 2015 data and is mostly associated with LCM water classes, indication some room for improvement.

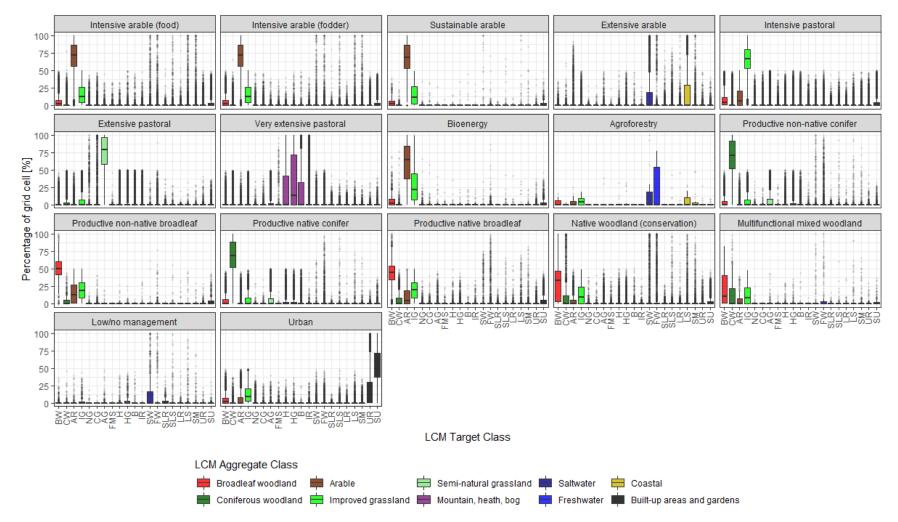
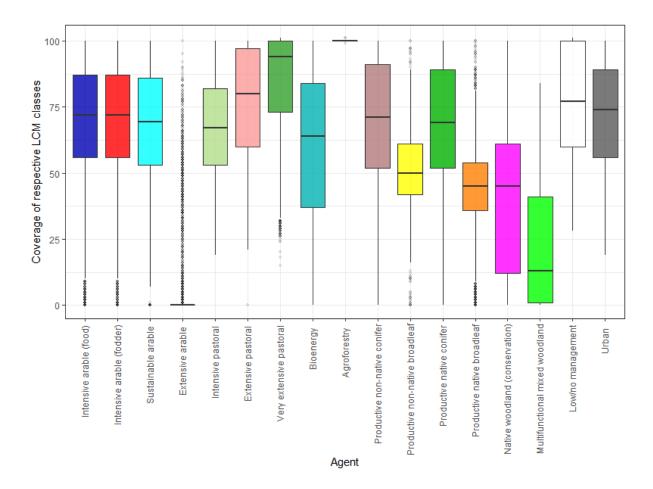


Figure S4: Sub-grid scale distribution of LCM classes within CRAFTY-GB agents. LCM classes as follows: BW Broadleafed woodland. CW Coniferous woodland. AR Arable and horticulture. IG Improved grassland. NG Neutral grassland. CG Calcareous grassland. AG Acid grassland. FMS Fen, marsh, and swamp. H Heather. HG Heather grassland. B Bog. IR Inland rock. SW Saltwater. FW Freshwater. SLR Supra-littoral rock. SLS Supra-littoral sediment. LR Littoral rock. LS Littoral sediment. SM Saltmarsh. UR Urban. SU Suburban



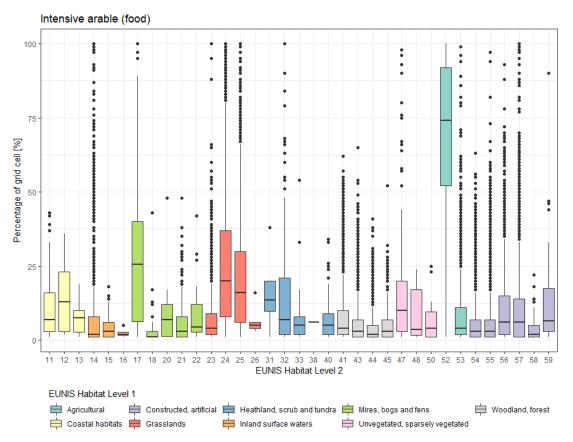
**Figure S5:** Sub-grid scale distribution of the LCM classes that correspond to CRAFTY-GB agents according to Table S4 (not accounting for additional data that has been used for allocation). For example, for agent 'Intensive arable (food)' fractions of LCM 'Arable' have been considered. As 'Agroforestry' has no equivalent in the LCM data, all classes were considered corresponding classes, explaining the 100% match. Lower (median) percentages indicate more heterogeneous landscapes, as higher percentages of 'non-matching' LCM classes can be found in the respective grid cells.

#### EUNIS habitats

The second comparison was between the baseline AFT map (1km resolution) and the EUNIS Habitat Map (100m resolution) representing ecosystem types across Europe (European Environment Agency, 2019). In this comparison, the distribution of EUNIS habitat types within each CRAFTY-GB cell was recorded. Because they are derived from different sources, these two maps were not expected to align closely. Nevertheless, the comparison was intended to characterise CRAFT-GB agents with regard to provision of habitat diversity and illustrate the scope for translation between the two.

Results showed good agreement between classes in each dataset, but with large variation within types. This can reflect heterogeneity within 1km cells, mismatches between the datasets used, as well as variation within AFTs that would be apparent in service levels but not in their labels. Figure S6 shows an example comparison between the CRAFTY-GB agent type 'Intensive arable (food)' and the EUNIS classes, revealing a clear association with arable

habitats, but also the less frequent presence of several other habitat types within those cells. Full results are shown in Figure S7 and summarised in Figure S8.



**Figure S6:** The distribution of EUNIS habitat types within the CRAFTY-GB 'intensive arable (food)' class. Habitat identities are explained in Table S9.

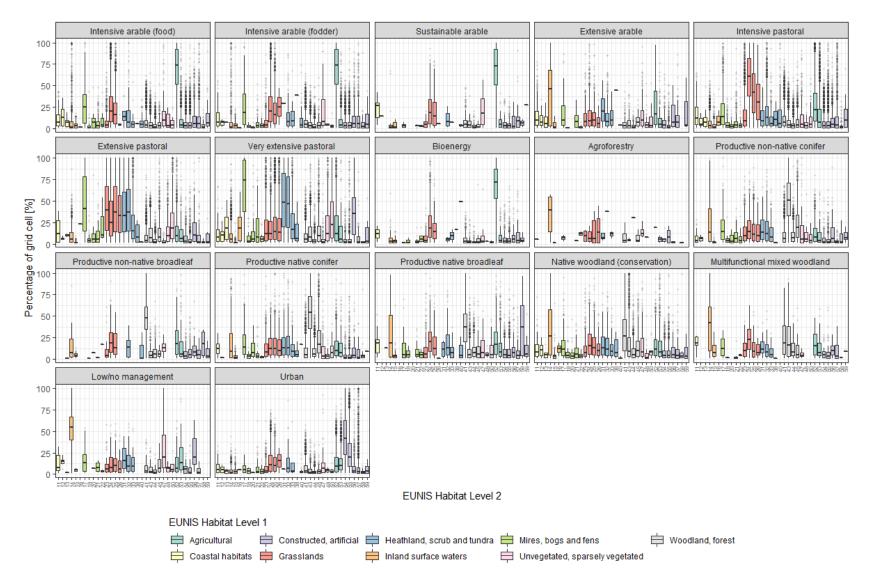


Figure S7: The distribution of EUNIS habitat types within the CRAFTY-GB agents. Habitat identities are explained in Table S9.

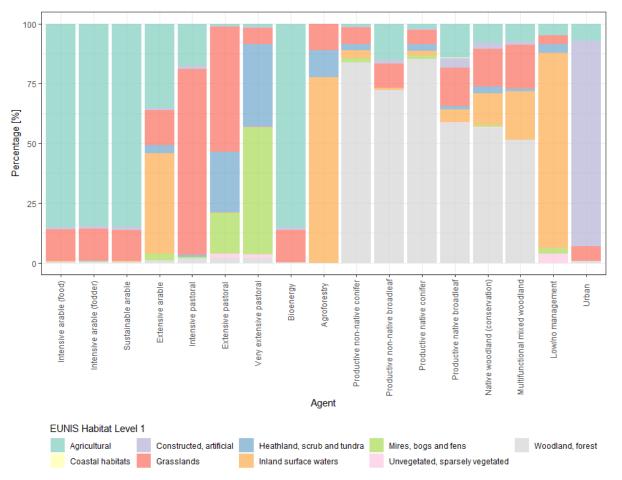


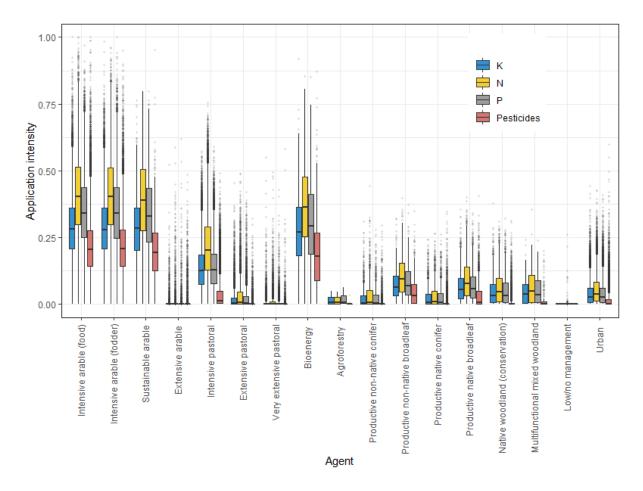
Figure S8: The distribution of broad EUNIS habitat types within the CRAFTY-GB agents.

| ID | EUNIS Level 2  | EUNS Level 1                    |
|----|--|---------------------------------|
| 11 | Coastal dunes and sandy shores                                 |                                 |
| 12 | Coastal shingle  | Coastal habitats                |
| 13 | Rock cliffs, ledges and shores, including the supralittoral    |                                 |
| 14 | Surface standing waters  |                                 |
| 15 | Surface running waters   | Inland surface waters           |
| 16 | Littoral zone of inland surface waterbodies                    |                                 |
| 17 | Raised and blanket bogs  |                                 |
| 18 | Valley mires, poor fens and transition mires                   |                                 |
| 20 | Base-rich fens and calcareous spring mires                     | Mires, bogs and fens            |
| 21 | Sedge and reedbeds, normally without free-standing water       |                                 |
| 22 | Inland saline and brackish marshes and reedbeds                |                                 |
| 23 | Dry grasslands   |                                 |
| 24 | Mesic grasslands   |                                 |
| 25 | Seasonally wet and wet grasslands                              | Grasslands                      |
| 26 | Alpine and subalpine grasslands                                |                                 |
| 31 | Arctic, alpine and subalpine scrub                             |                                 |
| 32 | Temperate and mediterranean-montane scrub                      |                                 |
| 33 | Temperate shrub heathland                                      |                                 |
| 38 | Riverine and fen scrubs  | Heathland, scrub and tundra     |
| 40 | Shrub plantations  |                                 |
| 41 | Broadleaved deciduous woodland                                 |                                 |
| 43 | Coniferous woodland  |                                 |
| 44 | Mixed deciduous and coniferous woodland                        |                                 |
|    | Lines of trees, small anthropogenic woodlands, recently felled | Woodland, forest                |
| 45 | woodland, early-stage woodland and coppice                     |                                 |
| 47 | Screes   |                                 |
| 48 | Inland cliffs, rock pavements and outcrops                     |                                 |
|    | Miscelanneous inland habitats with very sparse or no           | Unvegetated, sparsely vegetated |
| 50 | vegetation   |                                 |
| 52 | Arable land and market gardens                                 |                                 |
| 53 | Cultivated areas of gardens and parks                          | Agricultural                    |
| 54 | Buildings of cities, towns and villages                        |                                 |
| 55 | Low density buildings  |                                 |
| 56 | Extractive industrial sites                                    |                                 |
|    | Transport networks and other constructed hard-surfaced         | Constructed, artificial         |
| 57 | areas  |                                 |
| 58 | Highly artificial man-made waters and associated structures    |                                 |
| 59 | Waste deposits   |                                 |
|    | <b>S9:</b> EUNIS habitat identities.                           |                                 |

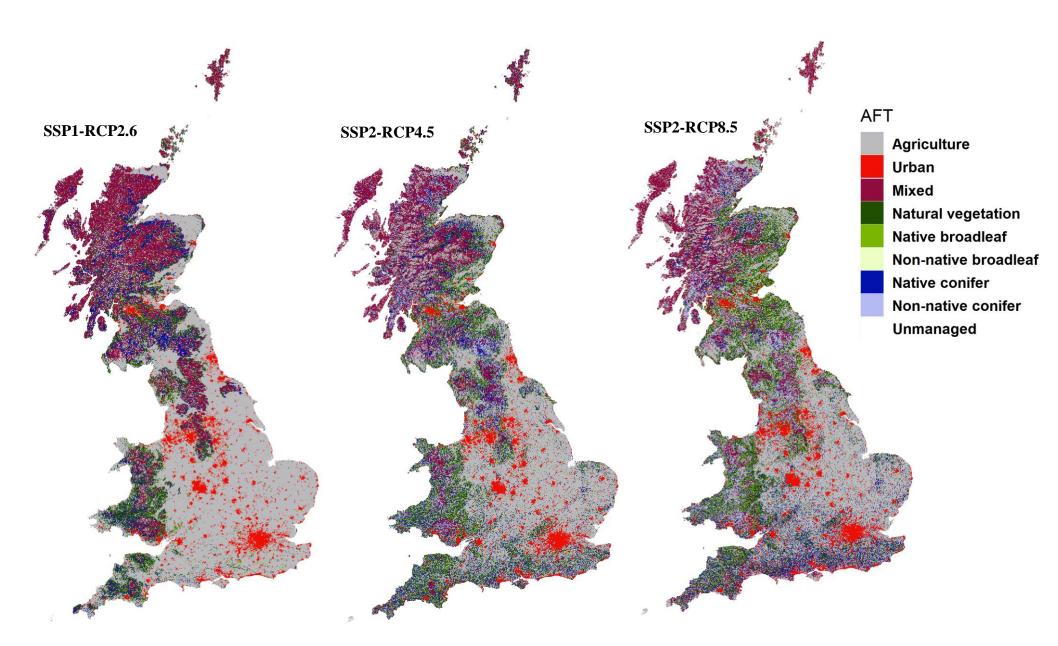
#### CEH Pesticides and Fertilisers

The third comparison of the baseline AFT map was to the 'CEH Land Cover® Plus: Pesticides v2.0' and 'CEH Land Cover® Plus: Fertilisers 2010-2015 (England)' datasets. These datasets report annual application intensity per km<sup>2</sup> grid cell of 162 ingredients for pesticides and nitrogen (N), phosphorus (P), and potassium (K) for fertilisers. Both datasets are gridded products based on the interpolation of survey data to crop type data (Jarvis et al. 2020; Osório et al. 2019). The sum of all 162 ingredients per grid cell was used as an indicator for pesticide application intensity. Both this indicator and the fertiliser data were min-max normalized to 0-1 in order to display comparable measures in Fig. S9. Although there is again a large variation of intensity levels within individual cells assigned to an AFT, which represents to some extent real-world variability (and is depicted in CRAFTY-GB by variable levels of capitals and services), average intensity levels of the individual AFTs show up as expected.

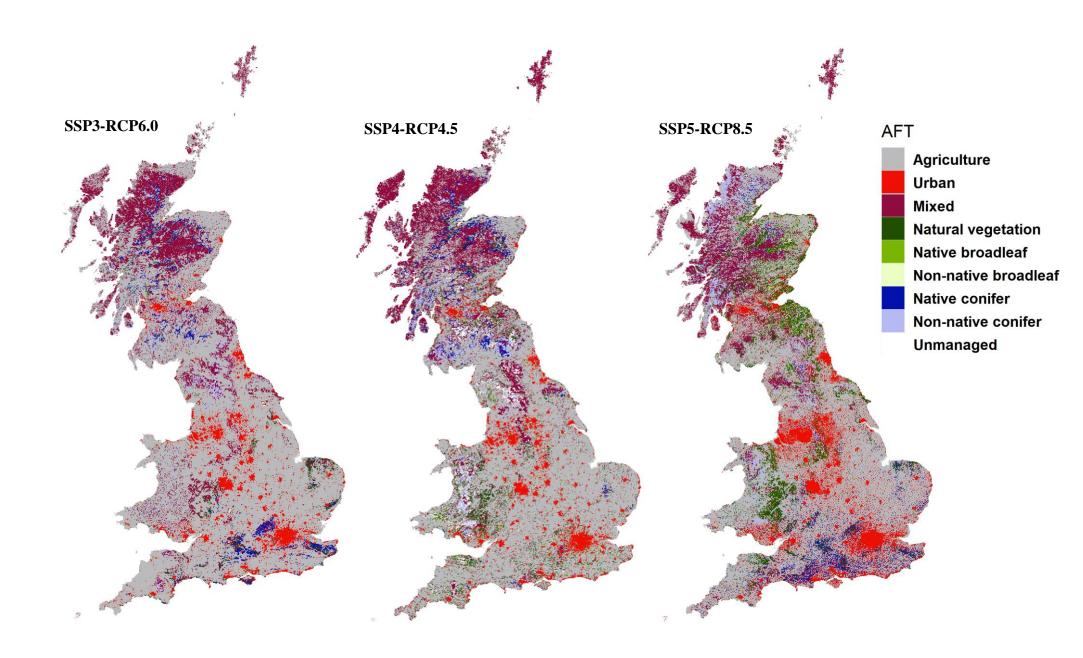
Intensive agricultural AFTs show the highest application intensities of both pesticides and fertilisers, while the application is substantially lower in extensive AFTs (both arable and pastoral). Due to the random allocation of 'Sustainable Arable' within cropland in the baseline map, there is no distinction to the intensive agricultural AFTs at this initial timepoint. Broadleaf woodland AFTs show higher rates of pesticides and fertilisers compared to coniferous woodland AFTS, most probably due to the higher association of broadleaf systems with intensive arable land (as discussed above).

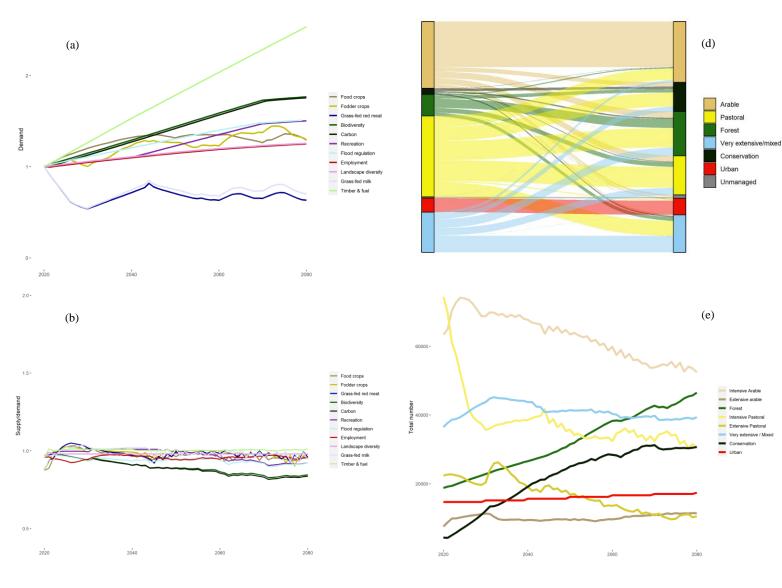


**Figure S9:** Associations between CRAFTY-GB agent types and application of fertilisers and pesticides as described in independent baseline data.

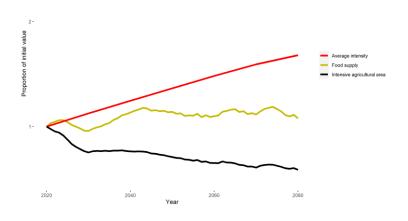


**Fig. S10:** Mapped results for the year 2080 in each scenario focusing on forest vegetation types. The mapped categories represent the dominant type within each cell. 'Mixed' contains forest and non-forest vegetation and land uses.









**Fig. S11:** Demand levels (a), supply as proportion of demand (b), land use intensity, food supply and intensive area (c), numbers of agents within amalgamated AFTs (d) and transitions between broad land use types (e) throughout SSP2-RCP 8.5 results.

# **TRACE document**

This is a TRACE document ("TRAnsparent and Comprehensive model Evaludation") which provides supporting evidence that our model presented in:

Brown et al. (in review), Agent-based modelling of alternative futures in the British land use system,

was thoughtfully designed, correctly implemented, thoroughly tested, well understood, and appropriately used for its intended purpose.

The rationale of this document follows:

Schmolke A, Thorbek P, DeAngelis DL, Grimm V. 2010. Ecological modelling supporting environmental decision making: a strategy for the future. *Trends in Ecology and Evolution* 25: 479-486.

and uses the updated standard terminology and document structure in:

Grimm V, Augusiak J, Focks A, Frank B, Gabsi F, Johnston ASA, Kułakowska K, Liu C, Martin BT, Meli M, Radchuk V, Schmolke A, Thorbek P, Railsback SF. 2014. Towards better modelling and decision support: documenting model development, testing, and analysis using TRACE. *Ecological Modelling* 

and

Augusiak J, Van den Brink PJ, Grimm V. 2014. Merging validation and evaluation of ecological models to 'evaluation': a review of terminology and a practical approach. *Ecological Modelling*.

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### **1 Problem formulation**

**This TRACE element provides supporting information on**: The decision-making context in which the model will be used; the types of model clients or stakeholders addressed; a precise specification of the question(s) that should be answered with the model, including a specification of necessary model outputs; and a statement of the domain of applicability of the model, including the extent of acceptable extrapolations.

#### Summary:

*CRAFTY-GB* is a model of the British land system (including Great Britain but excluding Northern Ireland), with the following primary and subsidiary objectives:

- To allow exploration of British land system change under a wide range of climatic and socio-economic scenarios, by representing:
  - Different sectors within the land system, including agriculture, forestry, urban, conservation and other major forms of management;
  - Different intensities of management within these systems;
  - A diverse, scenario-consistent set of socio-economic conditions affecting land management;
  - Human decision-making at individual and social levels within the land system, in terms of management and demands for different food types and ecosystem services;
  - Ecosystem service provision across a range of regulating, provisioning and cultural services.

CRAFTY-GB is an agent-based model of the British land system based on a broad range of available land system data and operating at 1km<sup>2</sup> resolution. The model is an application of the CRAFTY agent-based modelling framework (Murray-Rust et al. 2014). It is intended for use in exploring land use change under divergent climatic and socio-economic scenarios. It is primarily intended for use by scientific researchers working on issues connected with future land use change. The model is not predictive and is not intended or able to reveal likely outcomes of particular interventions, and so is not for direct use in policy formulation. It can, however, provide broad contextual information to support policy decisions, particularly with respect to interactions between land use sectors and objectives and the effects of human decision-making within the land system.

The questions that the model is intended to answer are:

- How might the British land system develop under specified climatic and socioeconomic scenario conditions?
- How might human decision-making affect outcomes within those scenarios?
- To what extent do outcomes depend on these climatic, socio-economic and behavioural conditions?
- How does the British land system affect, and how is it affected by, international food production?
- How are different ecosystem services and different land system objectives affected by simulated outcomes?

In order to answer these questions, the model produces a range of outcomes. Key amongst these are:

- Maps and timelines of land use/management, land cover, and ecosystem service provision
- Supply levels and associated valuations for all simulated ecosystem services

The domain of applicability is Great Britain, and modelling can be conducted for any temporal extent during which necessary input data are available (currently 2020-2080) – CRAFTY-GB itself involves no extrapolations beyond these data. Therefore, the model represents an open-ended 'virtual laboratory' in which simulation experiments can be run on the basis of suitable input data and the assumptions and design features of the model.

# 2 Model description

**This TRACE element provides supporting information on**: The model. Provide a detailed written model description. For individual/agent-based and other simulation models, the ODD protocol is recommended as standard format. For complex submodels it should include concise explanations of the underlying rationale. Model users should learn what the model is, how it works, and what guided its design.

### **Summary:**

## An ODD protocol for the CRAFTY-GB model is presented below.

### Introduction: technical overview

CRAFTY-GB is an application of the CRAFTY agent-based modelling framework (Murray-Rust et al. 2014), which is an Open Source framework built on reusable software components, and is an independent piece of software written in Java. Interactions between components (agents, cells, regions etc.) is specified using interfaces that enable users to create their own configuration of model components. For example, the agent interface specifies that agents have a unique ID, have a current competitiveness and, among other things, belong to a certain Agent Functional Type (AFT; Arneth, Brown, and Rounsevell 2014). As with other model components, a user may implement new agent types as long as they fulfil the contract of the interface.

To remove the need for high-level programming among model users, the CRAFTY framework and CRAFTY-GB itself can be configured and setup to run through the use of XML files. This is a form of declarative specification – the XML files declare which objects should take part in a simulation, and they are then passed over to a scheduling system. Each XML file defines one or more entities within the simulation, and will typically include other files for subcomponents. Model configuration is based on the principles below:

- A Scenario file encodes overall parameters of the simulation the number of time steps (years) over which it will run, an ID for the simulation, the means of accessing input data and the required outputs (such as videos, images and tables).
- A World file defines the regions that comprise the simulated world.
- Each Region file specifies:
  - The coordinates and capital levels of the cells in the Region, typically using a CSV or ASCII raster file

- The Allocation, Competition and Demand models used within the region, often using CSV files to specify time-varying quantities (e.g. changes in capitals and demand)
- A set of agents and their properties, making use of CSV files as necessary.
- Various land use raster data to protect or overload externally modelled land use changes such as urbanisation and protected areas.

In each of these cases, the files also specify the Java classes to be used along with their parameters, allowing users to incorporate their own code in the model.

In contrast to the declarative approach taken to configuration, CRAFTY (and hence CRAFTY-GB) uses a fixed schedule that encodes the flow of operations. To further enhance transparency of model behaviour, CRAFTY includes numeric and graphic displays for model variables. Spatially explicit outputs are also made available and include agent type, capital levels, competitiveness scores and supply of services. Any of these displays can be used to create animations of the model's behaviour over time.

# 1. Purpose

CRAFTY-GB is an application of the 'Competition for Resources between Agent Functional Types' (CRAFTY) model framework, which was designed to allow land use changes to be modelled across large spatial extents. The specific purpose of CRAFTY-GB is to allow exploration of British land system change under a wide range of climatic and socioeconomic scenarios, as outlined in Section 1 (Problem Formulation) above. The model allows the adoption of different land uses, variations in the intensity of land uses, diversification into multifunctional land uses, land abandonment and competition for land.

# 2. Entities, state variables, and scales

*Spatial units* CRAFTY-GB is based on a grid of *cells* at 1km<sup>2</sup> resolution. Each cell has defined levels of a range of *capitals*, which describe the availability of particular social, environmental or economic resources. Cells can be grouped into independent or semi-independent *regions*, but these are not applied in the default setup. A non-spatial population is assumed to exist and to generate demands for *services*, such as food, timber and access to nature. These demands are defined exogenously. Each cell may be managed by a single land use *agent*.

*Agents* Land managers are explicitly represented as agents in CRAFTY-GB (institutional agents can exist as well, and are described in Holzhauer, Brown, and Rounsevell 2019). Land management agents have functional and behavioural components to describe their forms of land management and decision-making. Agents are able to leverage the *capitals* available in a *cell* to provide a range of *services*. Each agent has a production function that maps capital levels onto service provision levels (see Sub-section **Error! Reference source not found.**, Production). An agent's *competitiveness* according to a given level of service provision can be calculated based on societal demands, supply levels and marginal benefit functions that define the economic and non-economic value of service production given the supply-demand difference at that point in the simulation.

Agents have several attributes that directly affect land use change, the two most fundamental being abandonment ("giving up") and competition ("giving-in") thresholds. If an agent's competitiveness falls below their giving up threshold, which defines the minimum return an agent is willing to accept from a cell, it abandons the cell, which then becomes available to other agents. If an agent that does not currently own a cell has a competitiveness greater than an incumbent agent's, and if the difference is larger than the incumbent's giving-in threshold,

the incumbent relinquishes its cell to the competitor – having been, in effect, 'bought out'. An agent searching for land can therefore only take over unmanaged (abandoned) cells, or those on which it can outcompete the existing agent. These processes are mediated by an abandonment probability that determines the likelihood of an agent abandoning their cell at any particular timestep, and search abilities that determine the number and order of cells that are searched for competition at each timestep (Table 1).

Agents are drawn from a typology that defines general characteristics of agents, and which is based on the Agent Functional Types (AFT) approach (see sub-section 4). As well as defining extant agents, the typology allows for new agents to be created, and for the comparison of productivity, benefit and other characteristics of "typical" agents of the type. These "Potential Agents" are used within the allocation process to represent agents who are attempting to find some land to manage, or to analyse the optimum type of agent to manage a given cell. Finally, individual agents of a given type need not be identical – all of the agent's characteristics can be drawn from user-definable distributions to provide within-type heterogeneity. See Table 1 for a complete list of agent variables.

| Variable   | Description  |  |  |  |  |
|--|--|--|--|--|--|
| Typological variables (allowing for random individual level variation) |  |  |  |  |  |
| Competition (giving-in)<br>threshold                                   | If a competing agent's competitiveness is greater than the incumbent agent's by a value larger than the giving-in threshold then the incumbent agent relinquishes that cell to the competitor.   |  |  |  |  |
| Abandonment (giving-up)<br>threshold                                   | If an agent's competitiveness falls below its giving-up threshold (defines the minimum return an agent is willing to accept from a cell) it needs to abandon its cell(s) (with giving-up probability).   |  |  |  |  |
| Abandonment (giving-up)<br>probability                                 | Probability for giving up in case the agent's competitiveness falls below the giving-up threshold  |  |  |  |  |
| Optimal production   | Amount of produced service in case of optimal conditions (all relevant capitals maximised)   |  |  |  |  |
| Capital sensitivities  | Sensitivities of production to capital values  |  |  |  |  |
| Production model   | Component responsible for calculation of service provision   |  |  |  |  |
| Search ability   | Comprising three parameters: the number of search iterations an agent type<br>can undertake per timestep, the number of cells considered for competition<br>during each search iteration, and the order (random or ranked) in which<br>those cells are competed for. |  |  |  |  |
| Social networks  | Comprising two parameters: the size and the effect of neighbourhoods within which agents of the same type benefit one another's capital, production or competitiveness levels.   |  |  |  |  |
| Individual variables (do not e   | xist at typological level)   |  |  |  |  |
| Competitiveness  | Denotes the agent's current competitiveness value (calculated in-model)  |  |  |  |  |

 Table 1: Variables of agents

*Environment*. CRAFTY-GB represents the British land system. Within this land system, heterogeneity is represented by capitals (economic, social, financial, manufactured, human and natural) that describe the locational attributes of each cell.

*Scales*. CRAFTY-GB covers the British land system at 1 km resolution. A time step in CRAFTY-GB represents a year by default, but this is not fixed and can vary to match the timescale of land use change decisions.

# 3. Process overview and scheduling

At each modelled timestep, the level of service production achieved by an agent is given a benefit value via a function that relates production levels to unmet demand. Agents compete for land based on these benefit values, and this competition is affected by individual or typological behaviour, as defined above. Table 2 gives an overview of the CRAFTY-GB simulation schedule.

| Table 2. Basic simulation schedule for CRAFTY-GB. | Table | 2. Basic | simulation | schedule | for CRAF | TY-GB. |
|---|-------|----------|------------|----------|----------|--------|
|---|-------|----------|------------|----------|----------|--------|

| 2. Dasie sinulation schedule for CICALTTOD.  |
|--|
| Timestep   |
| 1. Read in masks that constrain land use changes in this timestep (e.g. Urban mask)                |
|  |
| <b>2.</b> For each agent $\in$ Agents  |
| a. Update competitiveness based on residual demand   |
| b. If competitiveness < giving-up threshold, draw random number on [0,1]                           |
| and compare against giving-up probability. If lower, abandon cell                                  |
| <b>3.</b> For each region $\in$ Regions  |
| allocate-land:   |
| a. Allocate most competitive agent type to unoccupied cells, if consistent with                    |
| giving-up threshold and masks  |
| b. For each agent type $t \in$ Agent Types, undertake <i>n</i> search iterations of <i>m</i> cells |
| c. For every searched cell, calculate t's competitiveness  |
| d. If t's competitiveness > (cell owner's competitiveness + cell owner's giving-                   |
| in threshold), and if permitted by masking rules, owner relinquishes cell                          |
| e. Agent of type t takes cell over, with parameters drawn randomly from                            |
| typological ranges, if used.   |
| <b>4.</b> For each agent $\in$ Agents  |
| Update supply of services produced   |
|  |
| <b>5.</b> (For each region $\in$ Regions   |
| Update supply and unmet demand)  |

Figure 1 shows the flow of operation within each tick (or timestep). Each timestep starts by updating the decision-making context for land use agents – the levels of demand, capitals and any restrictions related for example to protected areas. Updates are made to the levels of demand across each region, and levels of capitals within each cell. These are loaded from external files, either as direct values or as functions to be sampled from on a yearly basis. Next, the land use agents respond and adapt to this altered context:

- First, each agent updates its level of supply, based on current capital levels. The total supply of each service is then calculated.
- Next, each agent's competitiveness is calculated on the basis of the difference between total supply and demand, and the valuation per unit unmet demand of each service.
- Any agents who give up are removed from the model.
- The active allocation procedure now runs, allowing new agents to take over unmanaged land and allowing other land transitions to take place.

Once all of the land use agents have been updated, final accounting is carried out, such as calculating total supply and demand, creating output files, displaying model state and creating videos.

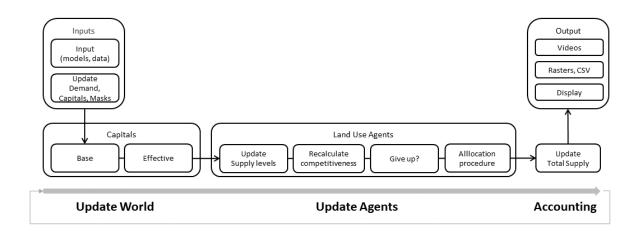


Figure 1. CRAFTY-GB flow diagram. This represents a single timestep for a single region.

## 4. Design concepts

**Basic principles.** The concept of Agent Functional Types is used to group land-use agents by their productive and decision-making characteristics. This typology is intended to allow generalisations of the attributes (traits) of individual actors in order to simplify model development and application, and to provide a transparent representation of agent decisional processes and behaviour. The AFT concept derives from a direct analogy with the use of Plant Functional Types (PFTs) in Dynamic Vegetation Models (Arneth, Brown, and Rounsevell 2014).

CRAFTY-GB inherits a number of design criteria from the CRAFTY framework on which it is based. These are:

- 1) The model must be able to run at 1km<sup>2</sup> resolution across national extent. This requirement holds for runtime costs, complexity, and the availability of data to parameterise and calibrate the model.
- 2) The model should take into account a wide range of societal demands for ecosystem services, including those that have no direct financial value.
- 3) The model must be able to represent multifunctional land use, and be responsive to the trade-offs between provision of various services.
- 4) The model should be able to represent the diversity of human behaviour that determines land management.
- 5) The model should be easy to refine and extend. This includes incorporating different data on services, capitals, land uses and agents, as well as adding complexity and variation to individual agents.

In CRAFTY-GB, these are extended to cover the purposes set out in Section 1 of this TRACE document.

The decision making submodel (see sub-section **Error! Reference source not found.**) acknowledges the existence of different modes of decision making like habits, heuristics and rule-based behaviour, and deliberative decision making. Decision are triggered by certain environmental or individual conditions or changes thereof which are checked every time step of the simulation. Table 3 provides an overview of the main assumptions that guided the

CRAFTY framework development, and which therefore underpin the operation of CRAFTY-GB.

| Model assumption   | Details  | Justification   |
|--|--|---|
| Potential productivity<br>of land can be<br>represented by a range<br>of capitals  | Capitals representing natural<br>productivity (for any good or service<br>such as a specific food or timber crop)<br>and any anthropogenic effects on<br>productivity (such as availability of<br>finance or infrastructure) can be used<br>as a basis for the description of<br>ecosystem services.   | Well-established method of characterising<br>and modelling land systems (Boumans et<br>al. 2002; Scoones 1998; Harrison et al.<br>2013; Pedde et al. 2019).   |
| Production of services<br>by land managers can<br>be described by a<br>function dependent<br>upon access to capitals<br>and productive<br>abilities. | The ability of land managers to<br>produce services is dependent on the<br>underlying productivity and attributes<br>of the land, expressed via capitals<br>(above) and their individual or<br>typological productive ability, which<br>may depend upon a number of<br>personal characteristics and<br>behavioural factors. (The form of the<br>production function is not set, but a<br>Cobb-Douglas function is used by<br>default). | An established method that allows for<br>production levels to vary according to<br>context and agent characteristics (Douglas<br>1976; Fulginiti and Perrin 1998; Martin<br>and Mitra 2001).  |
| The competitiveness of<br>land managers depends<br>upon demand for<br>specific services.   | Pre-determined demands exist for<br>ecosystem goods and services, and<br>land managers compete to satisfy these<br>demands (where not satisfied by<br>imports). Land managers are more<br>competitive when they can produce<br>greater (total) quantities of services for<br>which there is unmet demand.  | Demands for services are known to be<br>expressed via the economic value of<br>service production, and, in the absence of<br>behavioural factors, land use is driven<br>primarily by economics. Partly, decisions<br>are made on grounds of non-monetary (or<br>indirectly monetary) demands – e.g. for<br>green space - and CRAFTY is designed to<br>be capable of handling these, where they<br>can be parameterised. No fixed<br>assumption about the relationship<br>between unmet (residual) demand and<br>utility values (competitiveness) is made. |
| Land managers can be<br>classified into Agent<br>Functional Types<br>according to their<br>behaviours and<br>functions.                              | The management practices and<br>behaviours of land managers allows<br>them to be classified into a typology<br>analogous to the Plant Functional<br>Types used in Dynamic Global<br>Vegetation Models, increasing<br>modelling efficiency.   | The use of types increases computational<br>efficiency by providing a description of<br>land management and human behaviour<br>at a level of abstraction that decreases the<br>need for empirical parameterisation but<br>retains the characteristics most important<br>to large-scale land use change (Arneth,<br>Brown, & Rounsevell, 2014).  |
| Three mechanisms of land use change.   | Land use (or ownership) changes<br>when agents abandon land, take over<br>unmanaged land, or take over<br>managed land from the current owner.   | Analogous to main forms of land use<br>change in the real world.  |
| Each cell is managed by a single agent   | Multiple ownership of cells is not supported   | The scale of application is not defined<br>and so can be set to the appropriate scale<br>of land holdings in any particular case<br>(the minimum size of holding that is of   |

**Table 3**: Design assumptions made in CRAFTY-GB

|  |  | interest to the modeller). Agents may be<br>permitted to manage multiple cells. In<br>CRAFTY-GB, a 1 km <sup>2</sup> resolution is<br>selected as representative of typical<br>British land holdings.  |
|--|--|--|
| Agents have a fixed set<br>of potential actions  | The set of potential actions an agent<br>may select in decision making<br>processes and perform afterwards<br>needs to be defined and assigned<br>beforehand.  | The evolution of potential actions during<br>the time span of simulation can be<br>emulated by defining them beforehand<br>and by their dependence on evolving<br>capital and demand levels, which can in<br>turn be affected by other model<br>components.  |
| Wide range of land-use<br>relevant behaviour can<br>be represented by<br>'giving-in' and<br>'giving-up' thresholds | Range of personal characteristics and<br>behaviours known to affect land use<br>decisions can be often abstracted in<br>two values giving (relative)<br>willingness of land managers to<br>change land use or abandon land.<br>Believed to be a necessary<br>simplification for large-scale land use<br>models that adequately mimics<br>observed behaviour but can be<br>'overwritten' by more specific<br>decisions (see sections Agents and<br>Submodels, Decision Making). | Known that numerous factors affect<br>personal decision-making (e.g. Siebert,<br>Toogood, and Knierim 2006; Gorton et al.<br>2008; Brown et al. 2020; Bartkowski and<br>Bartke 2018) - too many to model or<br>parameterise. Several studies have<br>suggested that, for modelling purposes, a<br>wide range of behaviours are reducible to<br>a small number of dimensions similar to<br>those used here (Berger 2001; Polhill,<br>Gotts, and Law 2001; Siebert, Toogood,<br>and Knierim 2006; Gorton et al. 2008;<br>Murray-Rust et al. 2011). |
| Knowledge and social<br>influence flows over<br>geographical social<br>networks.                                   | Land managers are connected via<br>proximity-dependent social ties that<br>transport information, norms and<br>practices.  | Adoption of management practices<br>depends on horizontal spatial ties to<br>institutions and organisations (Brown et<br>al. 2020; Bartkowski and Bartke 2018;<br>Brown, Alexander, and Rounsevell 2018).  |
| Demand for urban land<br>is not subject to<br>competition with other<br>land uses                                  | Urban land is allocated externally to<br>the model and acts as a mask for land<br>use change within CRAFTY.  | As a relatively small but essential land<br>cover, urban land is likely to take<br>precedence and is not currently<br>modellable in the CRAFTY framework   |
| Protected Areas can be<br>represented as spatial<br>constraints on the<br>intensity of land<br>management          | Protected Areas are classified into two<br>levels and used to constrain land use<br>transitions between levels of intensity.   | No fixed rules for land use change exist in<br>most British protected areas but limits on<br>intensification are consistent with<br>objectives for environmental protection  |

*Emergence.* Emergent effects that could be observed as outcomes of experiments using CRAFTY-GB are spatially explicit changes to land ownership and management, the intensification of land uses, including mono- or multi-functional land uses, changes in productivities and yields of different land uses, and total supply levels.

*Adaptation.* Individual agents in CRAFTY-GB do not adapt their rules during a simulation run. However, the agent population adapts to changing conditions, and individual variation allows for adaptation in behavioural characteristics within types. Social interaction allows for indirect adaptation through alteration of capital values, allowing land management decisions to evolve and affect one another over time and space.

*Learning.* Agents can learn from neighbours to whom they are associated in social networks. This learning takes the form of improvements in capitals (e.g. representing knowledge), production or competitiveness, and is scaled by the degree of social networking.

*Fitness.* Agents' survival in the system depends upon their *competitiveness*, which is determined by an agent's ability to contribute to services for which unmet demands exist.

**Prediction.** CRAFTY-GB allows for contingent, explorative prediction only - i.e. it provides realisations of outcomes given the set of input conditions and model design. It does not represent an attempt to predict real-world outcomes, although model results can speak to what these real-world outcomes might be, when properly interpreted.

*Sensing.* Agents in CRAFTY-GB are aware of current demand levels and the production levels required if they are to avoid giving up their cells. They use the capital levels (attributes of a cell) to produce supply of services based on their respective production functions. Potential agents are aware of a defined number of abandoned/vacant cells that they may occupy depending upon their competitiveness. Agents are aware of the competitiveness of other agents in a region and may relinquish their cells to agents that are more competitive. Social networks allow agents to implicitly become aware of advantageous management practices used by their neighbours.

*Interaction.* Direct interactions occur between new ('potential') and existing agents that compete for cell ownership. Interactions also occur within social networks, allowing changes to production conditions to be shared.

*Stochasticity.* Agents can have individual variation in giving-up and giving-in threshold parameters, levels of service production, and the importance of each capital to service provision (each agent will have the same values throughout its lifetime, however). The allocation model includes stochasticity (representing agent-heterogeneity) as new agents consider only a limited number of cells on the grid, and the identity of these cells depends upon the random number seed being used. When giving-up probabilities are non-zero, there is stochasticity in giving-up events because the threshold is checked against a randomly drawn value.

*Collectives.* Two types of 'passive' agent collectives exist during a course of a simulation run. First is the list of agents that possess land parcels (cells) in a simulated landscape (grid), which can be global or regional in nature (covering the entire modelled land surface or some portion of it). Second is the list of potential agents that enter the system to takeover cells from existing agents (if possible) or occupy a vacant or abandoned cell on the grid. 'Active' collectives are those formed through social networks of neighbouring agents, defined by geographical proximity.

**Observation.** CRAFTY-GB can provide a range of observations and displays to help understand the model behaviour. Each of the submodels has a display, which is either numeric or graphical, showing curves for variables of note. A range of spatially explicit outputs is also available; these include maps of agent types, capital levels, competitiveness scores and supply of services. Any of these displays can be used to create videos of the model's behaviour over time. Output of a number of simulation data is possible in CSV or raster files. Table 4 gives an overview.

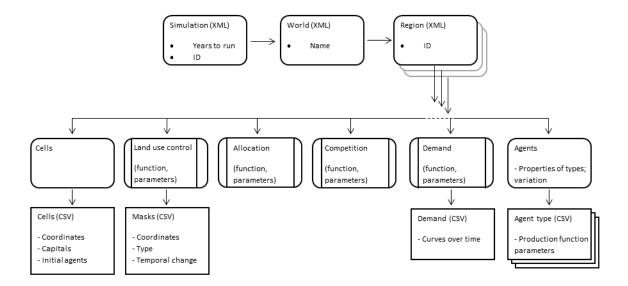
| Data              | CSV          | Raster       | Agg.CSV      | GUI          | Video        |
|-------------------|--------------|--------------|--------------|--------------|--------------|
| Agent ID          | $\checkmark$ | -            | -            | ×            | -            |
| LandUseIndex      | $\checkmark$ | $\checkmark$ | $\checkmark$ | ~            | $\checkmark$ |
| Capital levels    | $\checkmark$ | $\checkmark$ | ×            | $\checkmark$ | $\checkmark$ |
| Service Demand    | -            | -            | $\checkmark$ | ~            | $\checkmark$ |
| Service Supply    | $\checkmark$ | $\checkmark$ | $\checkmark$ | ~            | $\checkmark$ |
| Productivity      | $\checkmark$ | $\checkmark$ | ×            | $\checkmark$ | $\checkmark$ |
| Service Product.  | $\checkmark$ | ×            | ×            | ~            | x            |
| Competitiveness   | $\checkmark$ | $\checkmark$ | ×            | ~            | ~            |
| Giving In Thresh. | $\checkmark$ | x            | ×            | ~            | x            |
| Volatility        | ×            | $\checkmark$ | ×            | ×            | ×            |
| TakeOvers         | ×            | -            | $\checkmark$ | ×            | ×            |
| Performed Actions | $\checkmark$ | $\checkmark$ | ×            | ×            | x            |

Table 4: CRAFTY output matrix

## 5. Initialization

Initialisation proceeds through a set of interlinked XML and CSV files to allow the model's configuration by non-programmers. XML files define basic simulation parameters and provide properties for the initialisation of model components coded as Java objects, while CSV files provide data when there are many values required. The approach is highly flexible and extendable.

CRAFTY-GB initialises by reading the Scenario.xml file and follows the links therein to the configuration of outputs and the world configuration, which in turn contains links to regions and these to their model components like agent types, the competition model being used, or the allocation model. A cell.csv file includes the coordinates and capital levels of the cells in a region, the initial allocation of agents on these cells, and agent properties that are applied when these agents are initialised. Figure 2 gives an overview of a standard setting of XML and CSV files.



**Figure 2:** Overview of model configuration, showing relationships between files and what each file provides.

### 6. Input data

### Capital levels.

Capitals are divided into human, social, manufactured, financial and natural capitals, with natural capital further divided into yields or suitabilities for arable, pastoral and forest land uses or species. For CRAFTY-GB, social, human, financial and manufactured capitals were derived from UK-SSP projections of eight socio-economic indicators from (Merkle et al. 2022) (Table 5). Natural capitals were created in two distinct steps. Forest suitabilities were modelled using the Ecological Site Classification (ESC) model originally developed by (Pyatt, 1995) and since used frequently in forestry modelling for the UK (Forest Research 2021). This model uses data on accumulated temperature, continentality, wind risk, moisture deficit, soil moisture regime, and soil nutrient regime to predict biophysical suitability and associated potential yield class (timber growth) for a range of tree species. In the scenarios, these data were derived from UK-specific RCPs (Robinson et al. 2022) (Table 6).

To project land suitability for arable and pastoral land a General Additive Model (GAM; Hastie & Tibshirani, 1990) was produced to link land cover classes from Land Cover Map 2015 (Rowland et al., 2017) to UK-RCP covariates. Land Cover target class 3 (arable) and 4 (improved Grassland) were used as the training maps for arable and improved grassland, respectively, whilst semi-natural grassland was trained on LCM target classes 5-7, 9 and 10 (neutral, calcareous and acid grassland; heather; and heather grassland). UK-RCP-derived bioclimatic variables for growing degree days (GDD), minimum and maximum temperature, and soil moisture deficit (SMD) and surplus (SMS) were used as covariates, following Pearson et al. (2004). Urban areas were masked out in advance of model training. The baseline map of arable suitability was further processed to take into consideration changes in agricultural yields through time as modelled by the IMPRESSIONS European integrated assessment model (Harrison et al., 2019) and these augmented arable layers were used as a capital layer within CRAFTY-GB. The two grassland suitability maps were used directly as capital layers within CRAFTY-GB.

**Table 5:** Description of socio-economic capitals. For each of the capitals, individual values per area and time slice were formed as means between two indicator variables and subsequently normalised to [0,1]. Values between decades were interpolated. Full details of the indicator variables underlying the socio-economic capitals are given in (Merkle et al. 2022). Natural capital, split into 11 suitabilities, is described in 6.

| Capital      | Indicator Variables   | Linear Rescaling Thresholds |                    |                       |                     |                           |
|--------------|---|-----------------------------|--------------------|-----------------------|---------------------|---------------------------|
|              |   | Very Low<br>[0 ; 0.2]       | Low<br>[0.2 ; 0.4] | Medium<br>[0.4 ; 0.6] | High<br>[0.6 ; 0.8] | Very<br>High<br>[0.8 ; 1] |
| Social       | Income quintile ratio (S80/S20)   | 60;25                       | 25;10              | 10;5                  | 5;2                 | 2;1                       |
|              | Proportion of people who agree to<br>"people around here are willing to<br>help their neighbours" | 0;30                        | 30;50              | 50 ; 70               | 70;90               | 90;100                    |
| Human        | Life expectancy at birth  | 30;50                       | 50;60              | 60;70                 | 70;80               | 80;110                    |
|              | Proportion of people aged $25 - 64$ with tertiary education                                       | 0;10                        | 10;20              | 20;30                 | 30;45               | 45;80                     |
| Financial    | Household Income per capita [EUR PPS]   | 0;5                         | 5;10               | 10;25                 | 25;50               | 50;80                     |
|              | Proportion of people who agree to "I can save any amount of my income"                            | 0;20                        | 20;30              | 30;40                 | 40;50               | 50;100                    |
| Manufactured | Gross Fixed Capital Formation per<br>Area [mEUR/km <sup>2</sup> ]                                 | 0;0.75                      | 0.75 ; 1.25        | 1.25 ; 3              | 3;10                | 10;500                    |
|              | Average of total speed-weighted<br>road length [Speed-weighted<br>km/km <sup>2</sup> ]            | 0;0.1                       | 0.1 ; 0.2          | 0.2 ; 0.3             | 0.3 ; 1             | 1;4                       |

**Table 6:** Description of suitabilities comprising natural capital. All are normalised to a [0,1] scale at baseline and are linked to empirical production values through supply normalisation (described below). Abbreviations are as follows: GAM - General additive model; LCM – Land Cover Map: IAP – Integrated Assessment Platform; ESC – Ecological Site Classification.

| Suitability   | Explanation   | Source/reference  |
|---|---|---|
| Arable suitability                                      | GAM-projected arable suitability index (0 to 1) based on<br>relationship between bioclimatic covariates and LCM target class 3,<br>modified by changes in arable yields from IMPRESSIONS<br>integrated model. | GAMs (Hastie and<br>Tibshirani, 1990)<br>LCM 2015 (Rowland et<br>al., 2017) |
| Improved grassland suitability                          | GAM-projected semi-natural grassland suitability (0-1 index) based<br>on relationship between bioclimatic covariates and LCM target class<br>4.   | IMPRESSIONS IAP<br>(Harrison et al., 2019)<br>Biophysical covariates        |
| Semi-natural<br>grassland suitability                   | GAM-projected semi-natural grassland suitability (0-1 index) based<br>on relationship between bioclimatic covariates and LCM target<br>classes 5-7,9 and 10.  | Pearson et al. (2002).  |
| Natural: Short<br>Rotation Coppice<br>(SRC) suitability | ESC modelling: Willow yield   | ESC (Forest Research 2021)  |
| Natural: Agro-forestry tree suitability                 | ESC modelling: Sycamore yield   |   |
| Natural: Non-native conifer suitability                 | ESC modelling: Sitka spruce yield   |   |
| Natural: Non-native broadleaf suitability               | ESC modelling: Beech yield  |   |
| Natural: Native conifer suitability                     | ESC modelling: Scots pine yield   |   |
| Natural: Native<br>broadleaf suitability                | ESC modelling: Sessile Oak yield  |   |
| Natural: Native<br>broadleaf suitability                | ESC modelling: Silver Birch yield   |   |
| Natural: General tree species suitability               | ESC modelling: Combination of all other yields  |   |

# **Protected** areas

Protected areas belonging to 11 different types of national and international designation and to 5 different private land-owning organisations (NGOs) were included in the model (Table 7). Each protected area was first categorised into IUCN Protected Area Management Categories according to the existing categorisation of the IUCN National Committee United Kingdom (2012) or, where no existing categorisation was found, according to landowners' stated objectives. Two broad levels of protected area emerged from this classification: IUCN category IV and V areas where many forms of land use are found (all of the officially designated protected areas in the UK), and IUCN category II areas where land use is more tightly controlled (most of the NGO-owned protected areas). We therefore adopted two forms of constraint within the protected areas, with all land use except the most intensive being permitted in the first group, and no land use change except to the most extensive or conservation management permitted in the second. We also prevented land use change on areas classified as water, bare rock, coastal sediment and marsh in the baseline land use map. Institutions were used to enforce land use protections, and were represented as having complete power and knowledge.

**Table 7:** Types of protected area included in the model, their equivalent IUCN ranking (taken from IUCN National Committee United Kingdom (2012) or determined based on management objectives), data sources and the modelled constraint each type of protected area places on land use change.

| Type of protected area                        | IUCN<br>category | Data source   | Effect in<br>CRAFTY-GB |
|---|------------------|---|------------------------|
| International                                 |                  |   |                        |
| Biosphere Reserves                            | IV               | (UNESCO 2017)   | Not intensive          |
| Ramsar site                                   | IV               | (JNCC 2020)   |                        |
| Special Area of Conservation<br>(SAC)         | IV               |   |                        |
| Special Protection Area (SPA)                 | IV               |   |                        |
| National                                      |                  |   |                        |
| Area of Outstanding Natural<br>Beauty (AONB)  | V                | (Natural England, 2020a; Natural Resources Wales, 2021a)  | Not intensive          |
| Site of Special Scientific<br>Interest (SSSI) | IV               | (Natural England, 2021c; Natural Resources Wales, 2020; SNH, 2020)                                |                        |
| Heritage Coast (HC)                           | V                | (Natural England, 2017; Natural Resources Wales, 2017a)   |                        |
| Local Nature Reserve (LNR)                    | IV               | (Natural England, 2021a; Natural Resources<br>Wales, 2018; Scottish Government, 2020a)            |                        |
| National Nature Reserve<br>(NNR)              | IV               | (Natural England, 2021b; Natural Resources<br>Wales, 2021b; Scottish Government, 2020b)           |                        |
| National Park (NP)                            | V                | (Natural England, 2020c; Natural Resources<br>Wales, 2017b; Scottish Government, 2021a,<br>2021b) |                        |
| National Scenic Area (NSA)                    | V                | (Scottish Government, 2021c)  |                        |
| NGOs  |                  |   |                        |
| John Muir Trust (JMT)                         | Π                | JMT, personal communication   | No Change              |
| National Trust / National                     | V                | (National Trust, 2021; National Trust for   |                        |
| Trust for Scotland (NT/NTS)                   |                  | Scotland, 2015)   |                        |
| RSPB  | Π                | (RSPB, 2021)  |                        |
| Scottish Wildlife Trust                       | II               | (Scottish Wildlife Trust, 2016)   |                        |
| Other NGO                                     | Π                | Trees for Life, personal communication  |                        |

### Land uses (agent types).

CRAFTY-GB includes a range of agent types designed to capture the main forms of land use in Great Britain, including gradations of intensity and multi-functionality. Agent types were divided between arable land uses (intensive arable for food, intensive arable for fodder, sustainable arable and extensive arable), pastoral land uses (intensive pastoral, extensive pastoral, very extensive pastoral), forest land uses (productive native conifer, productive non-native broadleaf, productive non-native broadleaf, multifunctional mixed woodland and native woodland for conservation), and combined classes (bioenergy and agroforestry) (Table 8). Variation in ecosystem service provision within these classes allows them to represent a continuous range of forms of land management rather than arbitrarily distinct groups.

Allocation of the initial distribution of land uses was based on the 2015 Land Cover Map (LCM2015) produced by the UK Centre for Ecology & Hydrology (2016) (Rowland et al., 2017) and the National Forest Inventory (NFI) 2010-2015 (Forestry Commission 2021). Further datasets were used to define the extent and location of specific land uses, and full details are given in Table 8. Urban areas were derived from land cover data at the baseline (LCM 2015) and then projected in the scenarios by an independent urban model (described in detail in Merkle et al., in review). This model created 1km gridded urban surface projections through a newly developed urban allocation algorithm based on a neighbourhood density function, SSP-specific sprawl parameter settings, and SSP-specific land exclusions of protected areas and flood risk areas. Land not otherwise used was modelled as unmanaged.

In some cases, input data were incomplete and had to be further processed before being used. This was true of some coastal areas and islands (particularly estuaries and the Shetland isles). Data gaps in Shetland were filled using a regression model using topographic variables (i.e. elevation, slope, and aspect) trained upon the data of the nearest Orkney island. Gaps in coastal areas were filled using nearest-neighbour values. We also used 5x5 moving average interpolation to smooth hard boundaries between administrative units in the capitals. Finally, where scenario input data for 2020 were not consistent with baseline data, those data series were normalised by the equivalent baseline values.

**Table 8:** Allocation of initial distribution of land uses. Levels of intensity are assigned discretely in terms of agent types, but modelled continuously across these types according to availability and usage of agricultural inputs and production levels.

| Land use (agent)                    | Notes  | Initial allocation  |
|-------------------------------------|--|---|
|                                     |  |   |
| <b>Intensive arable</b> (food)      | Farmers managing intensively for crop production for food.   | Allocated to LCM2015 aggregate class 'Arable'. Food and fodder types distributed randomly within that according to  |
| Intensive<br>arable (fodder)        | Farmers managing intensively for crop production for livestock fodder, ultimately producing meat and milk.   | (modelled) baseline demand levels to provide the required amount of each  |
| Sustainable arable                  | Farmers managing organically or otherwise less intensively for<br>crop production for food   | Allocated to LCM2015 aggregate class 'Arable' to give an area coverage equal to the 2015 area of organic arable in the UK (as reported by (DEFRA 2016a), with specific cells chosen spatially randomly  |
| Extensive arable                    | Farmers managing with few inputs for limited crop production<br>for food; equivalent to subsistence production where capitals are<br>very low  | Allocated to the LCM2015 aggregate class 'Arable' cells within the lowest 10% of modelled suitability for arable crops  |
| Intensive pastoral                  | Farmers managing intensively for livestock   | LCM2015 Improved grassland  |
| Extensive pastoral                  | Farmers managing extensively for livestock on semi-natural grassland   | LCM2015 Semi-natural grassland  |
| Very extensive<br>pastoral          | Minimal management involving some grazing  | LCM2015 Mountain, heath, bog and LCM2015 Semi-<br>natural grassland (Fen Marsh Swamp)   |
| Bioenergy                           | Dedicated production of Short Rotation Coppice / Miscanthus  | Assigned to LCM2015 aggregate class 'Arable' to cover the 2015 extent of arable bioenergy land (DEFRA 2016b), assigned to locations of Energy Crops Scheme (Tranche 2) agreements 2013-2015 (Natural England 2020b)                                 |
| Agroforestry                        | Farmers practicing silvo-pastoral or silvo-arable forms of<br>agroforestry, combining trees with either grazing or crops, for<br>timber, crop and livestock production.  | NFI 'low-density' class when otherwise unassigned.  |
| Productive non-<br>native conifer   | Production-focused forest managers with non-native conifer<br>plantations. Primary objective is softwood timber production.  | LCM2015 Coniferous woodland class, sub-divided by NFI<br>Conifer class, located where modelled suitability is higher<br>for non-native than for native species  |
| Productive non-<br>native broadleaf | Production focused forest managers with non-native broadleaf<br>plantations. (Not currently common, but felt to have potential<br>importance in the future). Primary objective is hardwood timber<br>production. | LCM2015 Broadleaf woodland, sub-divided by NFI broadleaf class, located where modelled suitability is higher for non-native than for native species   |
| Productive native conifer           | Production focused forest managers with native conifer<br>plantations. Primary objective is softwood timber production.  | LCM2015 Coniferous woodland, sub-divided by NFI<br>Conifer class, located where modelled suitability is higher<br>for native than for non-native species  |
| Productive native<br>broadleaf      | Production focused forest managers with native broadleaf plantations. Primary objective is hardwood timber production.   | LCM2015 Broadleaf woodland, sub-divided by NFI broadleaf class, located where modelled suitability is higher for native than for non-native species.  |
| Multifunctional<br>mixed woodland   | Forest managers with mixed woodlands and multiple objectives practising low-intensity management   | LCM2015 Broadleaf or Coniferous woodland, subdivided by NFI mixed classes   |
| Native woodland<br>(conservation)   | Conservation focused forest managers. Primary objective is to conserve biodiversity.   | LCM2015 Broadleaf or coniferous woodland, excluding NFI classes indicating active management or no forest cover, and located where modelled broadleaf suitability is within the lowest 50% or modelled conifer suitability is within the lowest 10% |
| Urban                               | Urban and industrial areas   | Modelled separately   |
| Unmanaged                           | Represents areas with minimal to no management, often where<br>biophysical conditions preclude significant productivity e.g.<br>high montane or deep peat areas  | Unassigned cells  |

### Services and demand levels.

A range of provisioning, regulating and cultural ecosystem services and other indicators (e.g. biodiversity, employment) of relevance to the UK-SSP scenarios were modelled. These services are defined in Table 9, and their provision by different agent types based on capital levels is given in Brown et al. (2022). In this implementation, the relative calibration of service provision is approximate and largely assumption-based, though informed by empirical or modelled evidence where possible.

| Services               | Details   |
|------------------------|---|
| Food crops             | Crops for human consumption                                 |
| Fodder crops           | Crops for consumption by ruminant and monogastric livestock |
| Grass-fed meat         | Red meat produced in pastoral systems                       |
| Grass-fed milk         | Milk produced in pastoral systems                           |
| Bioenergy fuel         | Bioenergy crops; short rotation coppice & miscanthus        |
| Softwood               | Softwood (conifer) timber                                   |
| Hardwood               | Hardwood (broadleaf) timber                                 |
| Biodiversity           | Biological diversity  |
| Landscape diversity    | Diversity of landscape elements                             |
| Carbon sequestration   | Quantity of carbon sequestered (above & below ground)       |
| Recreation             | Recreation potential  |
| Flood Regulation       | Land ability to store water                                 |
| Employment             | Potential for employment associated with land management    |
| Sustainable production | Abstract service providing sustainability in agriculture    |

 Table 9: Goods and services modelled in CRAFTY-GB.

In modelling production of crops and livestock products, we assume divisions between crop production for direct human consumption, crop production for livestock consumption, and grass-fed livestock production. We assume that pastoralist agents produce grass-fed milk (intensive pastoral only) and red meat, while 'arable for fodder' agents effectively produce crop-fed red and white meat, and milk. Monogastrics are gramivores, so are fed only from cropland. Evidence for production levels includes an existing application of the CRAFTY framework to Scotland (Burton et al., in prep), and literature evidence on ecosystem services provision in different land use types (Burton et al., 2018; Rolo et al., 2021).

Non-food demands were taken from the stakeholder-defined scenarios, and are described in (Merkle et al. 2022). Demand levels for foods were derived from the LandSyMM (Land System Modular Model; www.landsymm.earth) global modelling framework (Rabin et al., 2020). Within LandSyMM, the dynamic global vegetation model LPJ-GUESS simulates physiological, demographic, and disturbance processes for a variety of plant functional types (Smith et al., 2014), while the land use model PLUM simulates land use and management based on global trade and cell-level (0.5°) productivity (Alexander et al., 2018). Food demand was calculated from scenario projections of country-level population and gross domestic product (GDP), using the historical relationship of per capita GDP to consumption of each of six crop types – C3 cereals, C4 cereals, rice, oil crops, pulses, and starchy roots – plus ruminant and monogastric livestock. Separate demand levels were calculated for food crops for human consumption and for feed for monogastric livestock and ruminant livestock not raised on pasture. Both types of demand account for crops used for processing, seed stocks, and for losses sustained during the production process. Demands were also adjusted to take account of imports and exports, as calculated by PLUM.

In the case of CRAFTY-GB, the total food production of the UK simulated by LandSyMM was taken as the national demand (i.e. aggregated from the 0.5° grid that LandSyMM uses). Because the simulated LandSyMM baseline (representing the year 2020) is not based on land cover data,

while the baseline land allocation of CRAFTY-GB is, all LandSyMM demands were normalised relative to their 2020 values, giving a continuous series of annual changes in demand levels as proportions of 2020 demand.

First the domestic production of feed and food crops was calculated. Food crops scale with the production of agents in CRAFTY, from a baseline quantity of 35.65 Mt of crops (an average of 771 tonnes for each of the 46,252 purely arable agents in CRAFTY-GB at the baseline, including subsequent losses, processing and seeds etc.). Feed crops were converted to livestock products through product-specific Feed Conversion Ratios taken from (Alexander et al., 2016). Monogastrics are fed exclusively on these feed crops (including those imported), meaning that the demands for Mt of pork, poultry and eggs could be immediately converted into demands for Mt of feed crops. Ruminant livestock (according to demands for Beef, Mutton, Goat and milk) were similarly converted, and the remaining available feed crops were assigned proportionally to them. Leftover demand for these livestock products was converted to a pasture demand by scaling from the baseline, and for comparison by using an additional pasture food conversion ratio.

# Services and demand levels

We use combinations of the Representative Concentration Pathways (RCP) climate scenarios (van Vuuren et al., 2011) and Shared Socioeconomic Pathways (SSP) socio-economic scenarios (O'Neill et al. 2017). A combined set of these scenarios was specified for the British context through a combination of stakeholder engagement and computational or statistical modelling.

The SSPs were specified for the UK as described in Pedde et al. (2021), (Harmáčková et al. 2022) and (Merkle et al. 2022). These substantial extensions of the global SSPs provide detailed narratives of social, economic and political developments across the UK until 2100. The narratives integrate national stakeholder knowledge on locally-relevant drivers and indicators with higher level information from the European and global SSPs. These narratives were simplified and converted into model parameterisations, and SSPs were put in a global context through LandSyMM global land system modelling to provide consistency with the broader SSP framework and to account for the UK's international trade.

Climatic conditions are taken from the CHESS-SCAPE data set, which provides several climate variables at 1 km<sup>2</sup> spatial resolution and several temporal resolutions, from daily to decadal. CHESS-SCAPE is derived from the 12 km<sup>2</sup> resolution UKCP18 regional predictions for the UK. UKCP18 regional predictions were obtained by running a perturbed parameter ensemble of a regional climate model (RCM), nested within a global climate model (GCM) for RCP8.5 (Murphy et al., 2018). CHESS-SCAPE was derived from this regional data set by: (i) downscaling from 12 km<sup>2</sup> to 1 km<sup>2</sup> using a modified version of the CHESS methodology (Emma L. Robinson, Blyth, Clark, Finch, et al. 2017); (ii) bias-correcting to observed historical climate using the CHESS-met dataset (Emma L. Robinson, Blyth, Clark, Comyn-Platt, et al. 2017); and (iii) time-shifting and pattern scaling to provide RCPs 2.6, 4.5 and 6.0, using members of the CMIP5 ensemble to define target trajectories of global temperature change (Taylor et al., 2012). Full details can be found in (E. L. Robinson et al. 2022). The highest temporal resolution of CHESS-SCAPE is daily. From these were calculated 20-year meanmonthly climatologies, at a 10-year time-step, giving spatially and temporally explicit values for several climate variables for the UK, including temperature and precipitation. The climate variables were used to calculate Penman-Monteith potential evapotranspiration with interception correction (PETI), following the method of Robinson et al. (2017). This is potential evapotranspiration calculated for a short grass, with a correction applied on rain days to account for the greater efficiency of evaporation of water from the canopy surface before it can reach the soil. The air temperature was used to calculate growing degree days (GDD), which is a count of the number of days for which mean air temperature was greater than 5°C. The air temperature, precipitation, PETI and GDD were then used as inputs to the crop, grassland and forest modelling to produce annual scenario-specific capital values.

## 7. Submodels

Allocation Model. Land ownership within CRAFTY-GB changes according to three different mechanisms, which simulate both individual and collective aspects of land use dynamics. Firstly, agents may leave the model owing to a competitiveness score that falls below an agent's giving-up threshold. Secondly, when land is unmanaged, due to abandonment or lack of managers, it can be taken over by a newly created agent. By default, the set of potential agents is evaluated to determine their competitiveness score on each unmanaged cell  $(c_{a,i})$ . The agents are sampled such that the probability of an agent of type *a* attempting to take over a cell scales with its competitiveness on a cell with 'perfect' capital levels;  $P(a) \propto c_{ai}^{\gamma}$ , where  $\gamma=0$  gives a random selection and  $\gamma \rightarrow \infty$  tends towards optimal selection. For more general land use transitions, an allocation procedure runs between existing and potential agents to determine ownership changes. This can include direct competition, where incoming agents attempt to take over existing cells; such an attempt succeeds where new agent has a competitiveness on the cell greater than or equal to the existing agent's competitiveness plus giving in threshold:  $c_{new} \ge$  $c_{curr} + giving_up_{curr}$ . Different allocation models are possible, however, and can be used to explore the relationships between human behaviour and local or global optimality. Once an agent is located, we assume it does not change location, due to the large costs involved.

*Production function.* Each agent has a production function, which maps capital levels onto service provision:

$$(1) P_S = F_A(C_i)$$

There is no limitation on the form of this function, but here a Cobb-Douglas style function is used to combine optimal production levels  $(o_s)$  with dependence on each capital to give service productivity:

(2) 
$$p_s = o_s \prod_C c_i^{\lambda_C}$$
;

where  $\lambda_c$  is a weighting factor specific to capital *c*.

**Population, Services, Demand and Utility.** We assume that there is a population present in any given region with a level of demand for services D. At the same time, there is a supply of these services from within the region, and the difference between the two is the residual (or unmet) demand, R. The marginal utility of production (i.e., the utility attributed to the production of one additional unit of a service) is a function of this residual demand:

$$(3) \qquad m_s = u_s(r_s);$$

where  $m_s$  is the marginal utility,  $u_s$  is a function that describes the utility of production, and  $r_s$  is the residual demand, for service s. The form of the function  $u_s$  is linear by default. For a given bundle of service provision, an agents' competitiveness (or utility) is given by:

$$(4) U_S = \sum_S p_S m_S$$

# **3** Data evaluation

**This TRACE element provides supporting information on**: The quality and sources of numerical and qualitative data used to parameterize the model, both directly and inversely via calibration, and of the observed patterns that were used to design the overall model structure. This critical evaluation will allow model users to assess the scope and the uncertainty of the data and knowledge on which the model is based.

### **Summary:**

CRAFTY-GB makes use of a range of datasets from different sources. These are summarized here, along with pre-existing evaluation exercises of those datasets. Model development did not involve any additional evaluation of data, and only a small amount of calibration to data (as described below). Model structure was based on a conceptual design (see Section 4) rather than patterns in data.

Data used in CRAFTY-GB are summarized in Table 10, along with their sources and any known evaluation exercises. In most cases, these data formed direct input to the model. Some calibration to data was carried out by running the model without any baseline land use data (i.e. from an empty map) and comparing the resultant numbers and distributions of agents with those contained in the baseline land use data (as in Brown, Seo, and Rounsevell 2019). This comparison was used to check the parameterization of agent types, with some adjustments made to ensure that parameters were not unrealistic in their effects. No agreement target was used because real-world land use patterns are long-term products of numerous factors and processes not contained in the model, but movement towards observed land use distributions was interpreted as improvement. Four rounds of this calibration exercise were carried out, before the modelling team agreed that parameter values had no obvious inconsistencies with the data.

**Table 10**: Input data for CRAFTY-GB, their sources and details of evaluation. Input data that are purely assumption-based are not described here.

| Data type          | Data<br>coverage                | Specific variables  | Source   | Evaluation  |
|--------------------|---------------------------------|---|--|---|
| CAPITALS           | Social capital                  | Income quintile ratio<br>(S80/S20)  | OECD, 2011   | Data subject to detailed evaluation<br>by OECD and EU member states<br>(UK in this case) (EUROSTAT<br>2013; OECD 2013)  |
|                    |                                 | Proportion of people who<br>agree to "people around<br>here are willing to help their<br>neighbours"          | UKHLS, 2015  | Data subject to detailed evaluation described in (Lynn and Knies 2016)  |
|                    | Human capital                   | Life expectancy at birth  | ONS, 2018  | Data subject to detailed evaluation described in (ONS 2017)   |
|                    |                                 | Proportion of people aged 25 – 64 with tertiary education   | Eurostat, 2019   | Data subject to standardised<br>EUROSTAT evaluation procedures<br>(EUROSTAT 2018, 2022)   |
|                    | Financial<br>capital            | Household Income per<br>capita [EUR PPS]  | ONS, 2017  | Data subject to detailed evaluation described in (ONS 2022)   |
|                    |                                 | Proportion of people who<br>agree to "I can save any<br>amount of my income"                                  | UKHLS, 2017  | Data subject to detailed evaluation described in (Lynn and Knies 2016)  |
|                    | Manufactured capital            | Gross Fixed Capital<br>Formation per Area<br>[mEUR/km <sup>2</sup> ]  | Eurostat, 2017   | Data subject to standardised<br>EUROSTAT evaluation procedures<br>(EUROSTAT 2018, 2022)   |
|                    |                                 | Average of total speed-<br>weighted road length<br>[Speed-weighted km/km <sup>2</sup> ]                       | GRIP, 2015   | Data subject to validation as described in (Meijer et al. 2018)   |
|                    | Natural<br>capitals             | Arable suitability  | Developed for this model; method described in Brown et al. (2022)        | Statistical evaluation as described in Brown et al. (2022)  |
|                    |                                 | Improved grassland suitability  | Developed for this model; method described in Brown et al. (2022)        | Statistical evaluation as described in Brown et al. (2022)  |
|                    |                                 | Semi-natural grassland suitability  | Developed for this model; method described in Brown et al. (2022)        | Statistical evaluation as described in Brown et al. (2022)  |
|                    |                                 | Tree species suitabilities  | ESC (Forest Research 2021)   | No single evaluation or validation<br>protocol; piecemeal evaluation and<br>improvement over 20 years of<br>model usage (Forest Research<br>2021; Pyatt 1995) |
| LAND COVER         | UK land cover                   | Land cover classes, locations   | (UK Centre for Ecology &<br>Hydrology 2016)                              | Validated as described by UK<br>Centre for Ecology & Hydrology<br>(2016)  |
|                    | National<br>Forest<br>Inventory | Forest classes, locations   | (Forestry Commission 2021)   | Validated primarily through on-the-<br>ground surveys, as described by<br>Forestry Commission (2021)  |
|                    | Crop type areas                 | Total area of organic crops,<br>total area of arable<br>bioenergy   | (DEFRA 2016a, 2016b)   | Largely survey-based, evaluated as described in by DEFRA (2016a, 2016b)   |
|                    | Energy crop locations           | Locations of Energy Crops<br>Scheme (Tranche 2)<br>agreements 2013-2015                                       | Natural England (2020b)  | Based on direct records of scheme uptake locations  |
| PROTECTED<br>AREAS | Protected area locations        | Biosphere Reserves<br>Ramsar site, Special Area of<br>Conservation (SAC),<br>Special Protection Area<br>(SPA) | (UNESCO 2017)<br>(JNCC 2020)   | All Protected Area data are direct<br>records of area boundaries<br>(shapefiles) and therefore were not<br>evaluated  |
|                    |                                 | Area of Outstanding<br>Natural Beauty (AONB)  | (Natural England, 2020a; Natural<br>Resources Wales, 2021a)              |   |
|                    |                                 | Site of Special Scientific<br>Interest (SSSI)   | (Natural England, 2021c; Natural<br>Resources Wales, 2020; SNH,<br>2020) |   |
|                    |                                 | Heritage Coast (HC)   | (Natural England, 2017; Natural<br>Resources Wales, 2017a)               |   |
|                    |                                 |   | 22   |   |

|  | Local Nature Reserve<br>(LNR)   | (Natural England, 2021a; Natural<br>Resources Wales, 2018; Scottish<br>Government, 2020a)  |   |
|--|---|--|---|
|  | National Nature Reserve<br>(NNR)  | (Natural England, 2021b; Natural<br>Resources Wales, 2021b; Scottish<br>Government, 2020b)   |   |
|  | National Park (NP)  | (Natural England, 2020c; Natural<br>Resources Wales, 2017b; Scottish<br>Government, 2021a, 2021b)  |   |
|  | National Scenic Area<br>(NSA)   | (Scottish Government, 2021c)   |   |
|  | John Muir Trust (JMT)   | JMT, personal communication  |   |
|  | National Trust / National<br>Trust for Scotland<br>(NT/NTS)                                     | (National Trust, 2021; National<br>Trust for Scotland, 2015)   |   |
|  | RSPB  | (RSPB, 2021)   |   |
|  | Scottish Wildlife Trust   | (Scottish Wildlife Trust, 2016)  |   |
|  | Other NGO   | Trees for Life, personal communication   |   |
| Social<br>network<br>extent                | Radius of influence<br>between agents   | Brown et al. (Calum Brown,<br>Alexander, and Rounsevell 2018)  | Based on values generated and<br>evaluated in the analysis of Brown<br>et al. (Brown, Alexander, and<br>Rounsevell 2018)  |
| Food<br>production                         | Crop production   | Derived from the LandSymm global<br>model (Rabin et al. 2020)  | Extensively evaluated as described<br>in (Rabin et al. 2020; Alexander et<br>al. 2018)  |
|  | Livestock production  | Derived from the LandSymm global<br>model (Rabin et al. 2020) and feed<br>conversion ratios of (Alexander et<br>al. 2016)  | Extensively evaluated as described<br>in (Rabin et al. 2020; Alexander et<br>al. 2018, 2016)  |
| Ecosystem<br>service                       | Timber & fuel production  | Dependent on natural capitals, as described above  | Evaluation as described under natural capitals in this table  |
| provision                                  | Biodiversity, carbon,<br>recreation, flood regulation<br>& employment provision                 | Qualitatively based on literature findings   | No direct evaluation, but<br>comparison and interpretation of<br>literature values (e.g. Burton et al.<br>2018; Rolo et al. 2021)   |
| Food demands                               | Crop & livestock product demands  | Derived from the LandSymm<br>global model (Rabin et al. 2020)<br>and feed conversion ratios of<br>(Alexander et al. 2016)  | Extensively evaluated as described<br>in (Rabin et al. 2020; Alexander et<br>al. 2018, 2016)  |
| All scenario-<br>dependent<br>model inputs | Capital levels, behaviours,<br>production and demand<br>levels, ecosystem service<br>valuations | Stakeholder-derived, as described in<br>Pedde et al. (2021), (Harmáčková et<br>al. 2022) and (Merkle et al. 2022)  | Participatory stakeholder<br>evaluation; see Pedde et al. (2021),<br>(Harmáčková et al. 2022) and<br>(Merkle et al. 2022)   |
|  | network<br>extent<br>Food<br>production<br>Ecosystem<br>service<br>provision<br>Food demands    | (LNR)National Nature Reserve<br>(NNR)National Park (NP)National Scenic Area<br>(NSA)<br>John Muir Trust (JMT)<br>National Trust for Scotland<br>(NT/NTS)<br>RSPB<br>Scottish Wildlife Trust<br>Other NGOSocial<br>network<br>extentFood<br>productionEcosystem<br>service<br>provisionFood demandsFood demandsCrop & livestock product<br>extentCrop & livestock product<br>mandsFood demandsCrop & livestock product<br>model inputsCrop & livestock productCrop & livestock product<br>extendCrop & livestock product<br>mands | (LNR)Resources Wales, 2018; Scottish<br>Government, 2020a)National Nature Reserve<br>(NNR)(Natural England, 2021b; Natural<br>Resources Wales, 2011b; Scottish<br>Government, 2020b)National Park (NP)(Natural England, 2020c; Natural<br>Resources Wales, 2017b; Scottish<br>Government, 2021a)National Scenic Area<br>(NSA)(Scottish Government, 2021c)John Muir Trust (JMT)JMT, personal communication<br>(National Trust for Scotland<br>(NT/NTS)RSPB<br>Scottish Wildlife Trust(Scottish Government, 2021c)National Trust / National<br>Trust for Scotland<br>(NT/NTS)(RSPB, 2021)Scotish Wildlife Trust<br>Other NGO(Scottish Wildlife Trust, 2016)Social<br>network<br>extentRadius of influence<br>between agentsFood<br>productionCrop productionDerived from the LandSymm global<br>model (Rabin et al. 2020)Livestock production<br>service<br>provisionDirivestock production<br>Biodiversity, carbon,<br>recreation, flood regulation<br>& employment provisionDerived from the LandSymm global<br>model (Rabin et al. 2020) and feed<br>conversion ratios of (Alexander et<br>al. 2016)Food demandsCrop & livestock product<br>demandsDerived from the LandSymm<br>global model (Rabin et al. 2020)<br>and feed conversion ratios of<br>(Alexander et al. 2016)Food demandsCrop & livestock product<br>demandsDerived from the LandSymm<br>global model (Rabin et al. 2020)<br>and feed conversion ratios of<br>(Alexander et al. 2016)Food demandsCrop & livestock product<br>demandsDerived from the LandSymm<br>global model (Rabin et al. 2020)<br>and feed conversion ratios of<br>(Alexander et al. 2016) |

# 4 Conceptual model evaluation

**This TRACE element provides supporting information on**: The simplifying assumptions underlying a model's design, both with regard to empirical knowledge and general, basic principles. This critical evaluation allows model users to understand that model design was not ad hoc but based on carefully scrutinized considerations.

### Summary:

Conceptual model evaluation has taken the following main forms:

- The fundamental conceptualization of the system, represented by the CRAFTY framework, has been described in detail and justified on the basis of empirical knowledge and basic principles, in a number of publications. It is evaluated for the case of CRAFTY-GB in this document.
- The widespread application of the CRAFTY framework has provided a number of tests of conceptual model utility.
- The relevance of the conceptual model to the problem that CRAFTY-GB addresses is also evaluated here.
- The conceptual division of British land uses into agent functional types has been evaluated with respect to underlying habitat classes.

A conceptual model evaluation is presented in Table 3 above, detailing the assumptions embedded in CRAFTY-GB as well as justifications for them on the basis of qualitative and quantitative information. This is a general justification, and not specific to the particular purpose of CRAFTY-GB. As such, it complements existing conceptual model evaluations of the CRAFTY framework given in (Brown, Brown, and Rounsevell 2016; Rounsevell, Robinson, and Murray-Rust 2012; Arneth, Brown, and Rounsevell 2014; Murray-Rust et al. 2014).

The relevance and value of this conceptual design has been assessed in various ways in publications applying the CRAFTY framework. CRAFTY has been applied in a number of theoretical or abstracted case studies (Murray-Rust et al. 2014; Arneth, Brown, and Rounsevell 2014; Brown et al. 2014; Brown, Holzhauer, and Metzger 2018; Synes et al. 2019; Urban et al. 2021; Holzhauer, Brown, and Rounsevell 2019), and to real-world studies in Yunnan Province, China (Synes et al. 2016), Sweden (Blanco, Holzhauer, et al. 2017; Blanco, Brown, et al. 2017), Scotland (Burton et al., in prep), Europe (Brown, Seo, and Rounsevell 2019; Brown, Holman, and Rounsevell 2021), Brazil (Millington et al. 2021) as well as Great Britain (Brown et al., in review). Not all of these applications contain formal conceptual model evaluations, but each speaks to the relevance of the model to particular research questions, and the fit of model assumptions to knowledge about real-world systems.

In the specific case of CRAFTY-GB, conceptual model evaluation has focused on the model purpose (Section 1), the main objective being **to allow exploration of British land system change under a wide range of climatic and socio-economic scenarios**. The ability of the conceptual model to capture scenario characteristics is therefore of paramount importance, and is described in detail in Brown et al. (in review). Table 11 below characterizes the fit between conceptual model design and scenario conditions. While the model is not able to represent every aspect of every scenario, it is felt to provide good coverage across a greater range of scenario conditions than existing models founded on more restrictive conceptual designs (see Brown et al. (in review) for further discussion of this).

| Scenario   | Description   | Conceptual model fit  |   |   |  |  |
|--|---|---|---|---|--|--|
|  |   | Behaviour   | Capitals  | Ecosystem service<br>demands and<br>valuations  | Production   |  |
| SSP1 -<br>Sustainability<br>SSP2 – Middle<br>of the Road | UK-SSP1 shows the UK transitioning to a fully<br>functional circular economy as society quickly<br>becomes more egalitarian leading to healthier<br>lifestyles, improved well-being, sustainable use<br>of natural resources, and more stable and fair<br>international relations. It represents a sustainable<br>and co-operative society with a low carbon<br>economy and high capacity to adapt to climate<br>change.<br>UK-SSP2 is a world in which strong public-<br>private partnerships enable moderate economic | The presence of<br>spatially-defined<br>social networks<br>allows the model to<br>be tailored to social<br>conditions and<br>distinct impacts in<br>each scenario.<br>Land-use-specific<br>decision-making,<br>along with | The presence of<br>socio-economic as<br>well as natural<br>capitals allows the<br>model to respond<br>to the full range of<br>scenario<br>conditions, which<br>directly affect<br>ecosystem service<br>provision. This is<br>a key advantage<br>over models that<br>consider only<br>certain influences<br>on production<br>(e.g. climatic or<br>economic).<br>The ability to<br>mask protected,<br>urban or other<br>areas also allows<br>for direct<br>interventions in<br>simulated land use<br>change<br>independently of<br>capital dynamics | CRAFTY-GB is<br>designed to<br>incorporate a<br>representative<br>range of<br>ecosystem<br>services. The<br>flexible nature of<br>ecosystem service<br>demand levels and<br>valuation<br>functions mean<br>that model<br>responses can<br>reflect scenario-<br>specific<br>preferences for<br>different services<br>and means of<br>providing benefits<br>in return for<br>service provision. | The dependence of<br>production or provision<br>levels on a full range of<br>scenario characteristics<br>(expressed through<br>behaviours, capitals, and<br>demands) as well as any<br>directly-modelled policy<br>interventions (e.g.<br>support for particular<br>land uses) means that<br>production, in principle,<br>varies in line with<br>scenario storylines,<br>rather than being a semi-<br>independent outcome of<br>biophysical conditions.<br>Production of multiple<br>services allows trade-<br>offs at individual and<br>higher levels to be<br>assessed, and ensures<br>that full impacts of land<br>use changes can be<br>accounted for, if<br>ecosystem services are<br>representative. |  |
|  | growth but inequalities persist. It represents a<br>highly regulated society that continues to rely on<br>fossil fuels, but with gradual increases in<br>renewable energy resulting in intermediate<br>adaptation and mitigation challenges.  | individual-level<br>randomness in<br>production and<br>agent behaviour,<br>allows the nature<br>and rate of land use  |   |   |  |  |
| SSP3 –<br>Regional<br>rivalry                            | The dystopian scenario, UK-SSP3, shows how<br>increasing social and economic barriers may<br>trigger international tensions, nationalisation in<br>key economic sectors, job losses and, eventually<br>a highly fragmented society with the UK<br>breaking apart. It represents a society where<br>rivalry between regions and barriers to trade<br>entrench reliance on fossil fuels and limit<br>capacity to adapt to climate change.   | change to reflect<br>individual, social<br>and political<br>scenario<br>components.<br>Individual-level<br>randomness in<br>production and<br>agent behaviour   |   |   |  |  |
| SSP4 -<br>Inequality                                     | UK-SSP4 shows how a society dominated by<br>business and political elites may lead to<br>increasing inequalities by curtailing welfare<br>policies and excluding the majority of a<br>disengaged population. The business and<br>political elite facilitate low carbon economies but<br>large differences in income across segments of<br>UK society limits the adaptive capacity of the<br>masses.   | allows  |   |   |  |  |
| SSP5 – Fossil-<br>fuelled<br>development                 | UK-SSP5 shows the UK transitioning to a highly<br>individualistic society where the majority<br>become wealthier through the exploitation of<br>natural resources combined with high economic<br>growth. It represents a technologically advanced<br>world with a strong economy that is heavily<br>dependent on fossil fuels, but with a high<br>capacity to adapt to the impacts of climate<br>change.  |   |   |   |  |  |

**Table 11**: Descriptions of scenarios represented by CRAFTY-GB and the conceptual elements of relevance to each scenario.

A final evaluation exercise with relevance to conceptual model design focused on the division of British land uses into agent functional types. This division was evaluated with respect to underlying habitat classes, and so is not only conceptual but also partly practical in nature. This evaluation is described in detail in Brown et al. (in review), and summarized here. First, the extent of agreement between baseline land uses and the underlying input data (Land Cover Map 2015) was assessed. This showed good general agreement between AFTs and LCM land-cover classes, although with large variations across individual grid cells. Second, the extent of EUNIS ecosystem types (European Environment Agency, 2019) within each land use type was examined. Because they are derived from different sources, these two maps were not expected to align closely but to reveal the basic ability of the land use typology to capture ecosystem characteristics. Nevertheless, results showed good agreement between classes in each dataset,

but again with large variation within types (as expected, given the range of characteristics allowable in each CRAFTY-GB class). Third, land use types were compared to the 'CEH Land Cover® *Plus*: Pesticides v2.0' and 'CEH Land Cover® *Plus*: Fertilisers 2010-2015 (England)' datasets. These datasets report annual application intensity per km<sup>2</sup> grid cell of 162 ingredients for pesticides and nitrogen (N), phosphorus (P), and potassium (K) for fertilisers. Once more there was large and expected variation of application levels within individual cells assigned to each land use class, and good agreement of average application levels with land use types. Intensive agricultural AFTs showed the highest application intensities of both pesticides and fertilisers, while the application is substantially lower in extensive AFTs (both arable and pastoral). While this evaluation was not used to calibrate model parameters, it provided some indication that the conceptual design of agent functional types was suitable for capturing land use and ecosystem variations in Britain.

# **5** Implementation verification

This TRACE element provides supporting information on: (1) whether the computer code implementing the model has been thoroughly tested for programming errors, (2) whether the implemented model performs as indicated by the model description, and (3) how the software has been designed and documented to provide necessary usability tools (interfaces, automation of experiments, etc.) and to facilitate future installation, modification, and maintenance.

## **Summary:**

The core code of the CRAFTY framework, used in CRAFTY-GB, has been thoroughly tested, using a combination of unit tests, debugging and sense checks on outputs. Model outputs were also iteratively evaluated during model development to ensure that performance was as expected, and the model description compared to model functioning. The software has been designed with a range of usability tools, including a graphical user interface that updates 'live' as the model runs, an online interface to explore model outputs, and an open-access, documented code base.

Unit tests were used in the development of the CRAFTY framework, with thorough checks also made on model implementation and performance (Murray-Rust et al., 2014). This ensured that the shared code base is sound for all applications of the framework. In the case of CRAFTY-GB, model performance was also assessed by comparing expected and realized outcomes across a range of parameterisations, though the primary aim here was to check on input data and calibration (as described in Sections 3 and 4 above). This also ensured performance in-line with the model description.

Substantial effort has been put into model usability. By default, CRAFTY provides facilities to graphically control and monitor model parameters, processes and outputs, as well as a range of file types and contents to capture model results. CRAFTY-GB adopts these facilities and therefore can provide a range of observations and displays to help understand model behaviour. Each of the submodels has a display, which is either numeric or graphical, showing curves for variables of note. A range of spatially explicit outputs is also available; these include maps of agent types, capital levels, competitiveness scores and supply of services. Any of these displays can be used to create videos of the model's behaviour over time (see Table 4, above).

The CRAFTY framework code is open-access and documented through ODD protocols as well as informal written descriptions. Installation and usage instructions are also provided. CRAFTY-GB is available through an online interface, where a model description is available

to help users interact with pre-generated outputs visualized in a range of figures. All of these usability tools are available via <u>https://landchange.earth/CRAFTY</u>.

# 6 Model output verification

**This TRACE element provides supporting information on**: (1) how well model output matches observations and (2) how much calibration and effects of environmental drivers were involved in obtaining good fits of model output and data.

### Summary:

Model output verification is of limited relevance for CRAFTY-GB because the model is designed to explore future conditions and relevant observations in historical or present-day settings are unavailable. A limited exercise to compare 'naïve' model runs against land cover data has been performed, but calibration was minimal. CRAFTY-GB reproduces data used in its development (e.g. in terms of stable baseline land uses and service supply levels), but has not been tested against independent data (Section 8).

As described in Section 3, some calibration to data was carried out by running the model without any baseline land use data (i.e. from an empty map) and comparing the resultant numbers and distributions of agents with those contained in the baseline land use data (as in Brown, Seo, and Rounsevell 2019). This comparison was used to check the parameterization of agent types, with some adjustments made to ensure that parameters were not unrealistic in their effects. No agreement target was used because real-world land use patterns are long-term products of numerous factors and processes not contained in the model, but movement towards observed land use distributions was interpreted as improvement. Four rounds of this calibration exercise were carried out, before the modelling team agreed that parameter values had no obvious inconsistencies with the data.

# 7 Model analysis

**This TRACE element provides supporting information on**: (1) how sensitive model output is to changes in model parameters (sensitivity analysis), and (2) how well the emergence of model output has been understood.

### **Summary:**

Several analyses have been run on the CRAFTY framework that have relevance to CRAFTY-GB. However, CRAFTY-GB run-times are relatively long (e.g. approx. 9 hours on a consumer workstation), which limits the scope for rigorous sensitivity analyses. Model sensitivity and output emergence has been analysed qualitatively and to a limited extent, as described below, including in terms of model stochasticity.

Sensitivity analyses of the CRAFTY modelling framework underlying the CRAFTY-GB model show that model results are particularly sensitive to capital levels and demand values, with less sensitivity to parameters controlling agent behavior (e.g. Brown, Holzhauer, and Metzger 2018; Murray-Rust et al. 2014). An exception is the level of multifunctional production by agents (i.e. levels of production of more than one ecosystem service), which can have a large effect on the balance among land use classes. In the case of CRAFTY-GB, model sensitivity was not formally assessed, but we made explorative changes and used these for informal evaluation in modelling

group. The scenario analysis presented in Brown et al. (in review) was itself a form of sensitivity analysis, and was used to understand main driving factors responsible for model outcomes.

# 8 Model output corroboration

This TRACE element provides supporting information on: How model predictions compare to independent data and patterns that were not used, and preferably not even known, while the model was developed, parameterized, and verified. By documenting model output corroboration, model users learn about evidence which, in addition to model output verification, indicates that the model is structurally realistic so that its predictions can be trusted to some degree.

### Summary:

Formal model output corroboration has not been performed for CRAFTY-GB. The model does not produce predictions, and exploratory outputs are for a range of future scenario-based conditions. The relevance of the model to these future conditions is defined in the previous sections of this document. Qualitative comparisons to relevant observed outcomes are presented below.

CRAFTY-GB's primary purpose is exploratory modelling of future scenario effects on the British land system. As such, it is impossible to verify that model outputs accurately reflect outcomes in those scenarios unless and until one actually occurs. Corroboration of alternative outcomes is possible in principle, for example in historical conditions or (qualitatively) in terms of responses to single drivers. However, these latter options are partially precluded by the absence of sufficiently detailed data. Historical data do not provide comparable, high-resolution time series of land uses, ecosystem service supply or demand levels, without which CRAFTY-GB results cannot be generated and/or assessed. Observations of effects related to single drivers are unavailable due to the concurrent actions of multiple drivers in reality. Nevertheless, checking model outputs qualitatively against independent information is possible and can be informative. Several such comparisons are presented below to enable readers to draw their own conclusions about model reliability.

- Changing demands; it is known that increasing demand for particular services does generate increased production in the land system, and that this tends to occur in more productive areas as we find. However, there are also various ways that production can increase; for example intensification often follows from demand increases (and extensification from demand decreases) both of which occur in our model results.
- Food production is generally prioritized in reality, and food supply approximately equals demand. Our valuation of ES is arbitrary but has the equivalent outcome.
- At a basic level, cross-sectoral trade-offs are a major feature of the land system, and can be explored here.
- Consolidation of productive areas and abandonment or change in marginal areas are strong patterns in British land use that are also replicated in the model.
- Low capitals produce inefficient, changeable land systems (SSP3)

#### References

- Alexander, Peter, Calum Brown, Almut Arneth, John Finnigan, and Mark D. A. Rounsevell. 2016. "Human Appropriation of Land for Food: The Role of Diet." *Global Environmental Change: Human and Policy Dimensions* 41 (November): 88–98.
- Alexander, Peter, Sam Rabin, Peter Anthoni, Roslyn Henry, Thomas A. M. Pugh, Mark D. A. Rounsevell, and Almut Arneth. 2018. "Adaptation of Global Land Use and Management Intensity to Changes in Climate and Atmospheric Carbon Dioxide." *Global Change Biology*, March. https://doi.org/10.1111/gcb.14110.
- Arneth, A., C. Brown, and M. D. A. Rounsevell. 2014. "Global Models of Human Decision-Making for Land-Based Mitigation and Adaptation Assessment." *Nature Climate Change* 4 (7): 550–57.
- Bartkowski, Bartosz, and Stephan Bartke. 2018. "Leverage Points for Governing Agricultural Soils: A Review of Empirical Studies of European Farmers' Decision-Making." *Sustainability: Science Practice and Policy* 10 (9): 3179.
- Berger, T. 2001. "Agent-Based Spatial Models Applied to Agriculture: A Simulation Tool for Technology Diffusion, Resource Use Changes and Policy Analysis." *Agricultural Economics* 25 (2–3): 245–60.
- Blanco, Victor, Calum Brown, Sascha Holzhauer, Gregor Vulturius, and Mark D. A. Rounsevell. 2017. "The Importance of Socio-Ecological System Dynamics in Understanding Adaptation to Global Change in the Forestry Sector." *Journal of Environmental Management* 196: 36–47.
- Blanco, Victor, Sascha Holzhauer, Calum Brown, Fredrik Lagergren, Gregor Vulturius, Mats Lindeskog, and Mark D. A. A. Rounsevell. 2017. "The Effect of Forest Owner Decision-Making, Climatic Change and Societal Demands on Land-Use Change and Ecosystem Service Provision in Sweden." *Ecosystem Services* 23 (December 2016): 174–208.
- Boumans, Roelof, Robert Costanza, Joshua Farley, Matthew A. Wilson, Rosimeiry Portela, Jan Rotmans, Ferdinando Villa, and Monica Grasso. 2002. "Modeling the Dynamics of the Integrated Earth System and the Value of Global Ecosystem Services Using the GUMBO Model." *Ecological Economics: The Journal of the International Society for Ecological Economics* 41 (3): 529–60.
- Brown, C., S. Holzhauer, and M. J. Metzger. 2018. "Land Managers' Behaviours Modulate Pathways to Visions of Future Land Systems." *Regional Environmental Change*. https://link.springer.com/article/10.1007/s10113-016-0999-y.
- Brown, Calum, Peter Alexander, and Mark Rounsevell. 2018. "Empirical Evidence for the Diffusion of Knowledge in Land Use Change." *Journal of Land Use Science* 13 (3): 269–83.
- Brown, Calum, Ken Brown, and Mark Rounsevell. 2016. "A Philosophical Case for Process-Based Modelling of Land Use Change." *Modeling Earth Systems and Environment* 2 (2): 50.
- Brown, Calum, Ian Holman, and Mark Rounsevell. 2021. "How Modelling Paradigms Affect Simulated Future Land Use Change." *Earth System Dynamics* 12: 211–31.
- Brown, Calum, Eszter Kovács, Irina Herzon, Sergio Villamayor-Tomas, Amaia Albizua, Antonia Galanaki, Ioanna Grammatikopoulou, Davy McCracken, Johanna Alkan Olsson, and Yves Zinngrebe. 2020. "Simplistic Understandings of Farmer Motivations Could Undermine the Environmental Potential of the Common Agricultural Policy." *Land Use Policy*, November, 105136.
- Brown, Calum, Dave Murray-Rust, Jasper Van Vliet, Shah Jamal Alam, Peter H. Verburg, and Mark D. Rounsevell. 2014. "Experiments in Globalisation, Food Security and Land Use Decision Making." *PloS One* 9 (12). https://doi.org/10.1371/journal.pone.0114213.

- Brown, Calum, Bumsuk Seo, and Mark Rounsevell. 2019. "Societal Breakdown as an Emergent Property of Large-Scale Behavioural Models of Land Use Change." *Earth System Dynamics Discussions*, no. May: 1–49.
- Burton, Vanessa, Darren Moseley, Calum Brown, Marc J. Metzger, and Paul Bellamy. 2018. "Reviewing the Evidence Base for the Effects of Woodland Expansion on Biodiversity and Ecosystem Services in the United Kingdom." *Forest Ecology and Management* 430 (December): 366–79.
- DEFRA. 2016a. "Organic Farming Statistics 2015." https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachme nt data/file/524093/organics-statsnotice-19may16.pdf.
  - —. 2016b. "Crops Grown For Bioenergy in England and the UK: 2015." https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachme nt data/file/578845/nonfood-statsnotice2015i-19dec16.pdf.
- Douglas, Paul H. 1976. "The Cobb-Douglas Production Function Once Again: Its History, Its Testing, and Some New Empirical Values." *Economy, Journal of Political* 84 (5): 903–16.
- European Environment Agency. 2019. "Ecosystem Types of Europe." February 7, 2019. https://www.eea.europa.eu/data-and-maps/data/ecosystem-types-of-europe-1.
- EUROSTAT. 2013. "Meeting of Providers of OECD Income Distribution Data 2.2 Comparability of OECD with Other International and National Estimates on Income Inequality and Poverty." EU. https://www.oecd.org/els/soc/2.2b%20Eurostat-EUSILC-Comparability.pdf.
  - ——. 2018. "Methodology for Data Validation 2.0 Revised Edition 2018." https://ec.europa.eu/eurostat/ramon/statmanuals/files/methodology\_for\_data\_validatio n\_v2\_0\_rev2018.pdf.
- Forest Research. 2021. "Ecological Site Classification Decision Support System (ESC-DSS)." 2021. https://www.forestresearch.gov.uk/tools-and-resources/fthr/ecological-siteclassification-decision-support-system-esc-dss/.
- Forestry Commission. 2021. "Forestry Commission Open Data." 2021. https://dataforestry.opendata.arcgis.com/search?q=national%20forest%20inventory%202016.
- Fulginiti, Lilyan E., and Richard K. Perrin. 1998. "Agricultural Productivity in Developing Countries." *Agricultural Economics* 19 (1): 45–51.
- Gorton, Matthew, Elodie Douarin, Sophia Davidova, and Laure Latruffe. 2008. "Attitudes to Agricultural Policy and Farming Futures in the Context of the 2003 CAP Reform: A Comparison of Farmers in Selected Established and New Member States." *Journal of Rural Studies* 24 (3): 322–36.
- Harmáčková, Zuzana, Simona Pedde, James M. Bullock, Ornella Dellaccio, Jennifer Dicks, George Linney, Magnus Merkle, Mark Rounsevell, Jon Stenning, and Paula A. Harrison. 2022. "Improving Regional Applicability of the UK Shared Socioeconomic Pathways Through Iterative Participatory Co-Design." https://doi.org/10.2139/ssrn.4010364.
- Harrison, Paula A., Ian P. Holman, George Cojocaru, Kasper Kok, Areti Kontogianni, Marc J. Metzger, and Marc Gramberger. 2013. "Combining Qualitative and Quantitative Understanding for Exploring Cross-Sectoral Climate Change Impacts, Adaptation and Vulnerability in Europe." *Regional Environmental Change* 13 (4): 761–80.
- Hastie, Trevor J., and Robert J. Tibshirani. 1990. *Generalized Additive Models*. Vol. 1. CRC Press.

- Holzhauer, Sascha, Calum Brown, and Mark Rounsevell. 2019. "Modelling Dynamic Effects of Multi-Scale Institutions on Land Use Change." *Regional Environmental Change* 19 (3): 733–46.
- IUCN National Committee United Kingdom. 2012. "Putting Nature on the Map: Identifying<br/>Protected Areas in the UK."<br/>https://portals.iucn.org/library/sites/library/files/documents/2012-102.pdf.
- JNCC. 2020. "UK Protected Area Datasets for Download." 2020. https://jncc.gov.uk/ourwork/uk-protected-area-datasets-for-download/.
- Lynn, Peter, and Gundi Knies. 2016. "UNDERSTANDING SOCIETY The UK Household Longitudinal Study Waves 1-5 Quality Profile." Institute for Social and Economic Research University of Essex. https://www.understandingsociety.ac.uk/sites/default/files/downloads/documentation/ mainstage/quality-profile.pdf.
- Martin, Will, and Devashish Mitra. 2001. "Productivity Growth and Convergence in Agriculture versus Manufacturing." *Economic Development and Cultural Change* 49 (2): 403–22.
- Meijer, Johan R., Mark A. J. Huijbregts, Kees C. G. Schotten, and Aafke M. Schipper. 2018. "Global Patterns of Current and Future Road Infrastructure." *Environmental Research Letters: ERL [Web Site]* 13 (6): 064006.
- Merkle, Magnus, Ornella Dellaccio, Rob Dunford, Zuzana Harmáčková, Paula A. Harrison, J-F Mercure, Simona Pedde, et al. 2022. "Creating Quantitative Scenario Projections for the UK Shared Socioeconomic Pathways." https://doi.org/10.2139/ssrn.4006905.
- Millington, James D. A., Valeri Katerinchuk, Ramon Felipe Bicudo da Silva, Daniel de Castro Victoria, and Mateus Batistella. 2021. "Modelling Drivers of Brazilian Agricultural Change in a Telecoupled World." *Environmental Modelling & Software*, February, 105024.
- Murphy, J. M., G. R. Harris, D. M. H. Sexton, E. Kendon, P. Bett, R. Clark, and K. Yamazaki. 2018. "UKCP18 Land Projections: Science Report. Met Office." Met Office.
- Murray-Rust, D., C. Brown, J. van Vliet, S. J. Alam, D. T. Robinson, P. H. Verburg, and M. Rounsevell. 2014. "Combining Agent Functional Types, Capitals and Services to Model Land Use Dynamics." *Environmental Modelling & Software* 59 (September): 187–201.
- Murray-Rust, Dave, Nicolas Dendoncker, Terry P. Dawson, Lilibeth Acosta-Michlik, Eleni Karali, Eleonore Guillem, and Mark Rounsevell. 2011. "Conceptualising the Analysis of Socio-Ecological Systems through Ecosystem Services and Agent-Based Modelling." *Journal of Land Use Science* 6 (2–3): 83–99.
- National Trust. 2021. "National Trust Open Data." 2021. https://uk-nationaltrust.opendata.arcgis.com/.
- National Trust for Scotland. 2015. "National Trust for Scotland Property Boundaries." 2015. https://marine.gov.scot/information/national-trust-scotland-property-boundaries.
- Natural England. 2017. "Heritage Coasts (England)." 2017. https://naturalenglanddefra.opendata.arcgis.com/datasets/heritage-coasts
  - england/explore?location=52.802383%2C-2.195731%2C6.95&showTable=true.
  - —. 2020a. "Areas of Outstanding Natural Beauty (England)." https://data.gov.uk/dataset/8e3ae3b9-a827-47f1-b025-f08527a4e84e/areas-ofoutstanding-natural-beauty-england.
    - -. 2020b. "Energy Crops Scheme Agreements Tranches 1 2." https://data.gov.uk/dataset/363474ab-0d45-4dff-8857-5fcd35cdf3db/energy-crops-scheme-agreements-tranches-1-2.
    - -. 2020c. "National Parks (England)." https://data.gov.uk/dataset/334e1b27-e193-4ef5b14e-696b58bb7e95/national-parks-england.

-. 2021a. "Local Nature Reserves (England)." https://data.gov.uk/dataset/acdf4a9e-a115-41fb-bbe9-603c819aa7f7/local-nature-reserves-england. -. 2021b. "National Nature Reserves (England)." https://data.gov.uk/dataset/726484b0d14e-44a3-9621-29e79fc47bfc/national-nature-reserves-england. -. 2021c. "Sites of Special Scientific Interest (England)." 2021. https://naturalenglanddefra.opendata.arcgis.com/datasets/f10cbb4425154bfda349ccf493487a80\_0/explore?1 ocation=52.837148%2C-2.496337%2C6.94. "Heritage Coasts." Natural Resources Wales. 2017a. DataMapWales. 2017. https://datamap.gov.wales/layers/inspire-nrw:NRW HERITAGE COAST. -. 2017b. "National Parks." https://data.gov.uk/dataset/949976cb-f952-4405-9fa1bf531fdca0f5/national-parks. -. 2018. "Local Nature Reserves (LNRs)." https://data.gov.uk/dataset/c0c66de2-ef27-471f-a501-ebf2713f8649/local-nature-reserves-lnrs. "SSSIs." 2020. 2020. https://naturalresourceswales.sharefile.eu/share/view/s7097d5022294fc5b/foe8decaf112-4e5e-af93-02b2fc71ade3. 2021a. "Areas of Outstanding (AONBs)." Natural Beauty https://data.gov.uk/dataset/b40871c7-ab45-44f1-8989-47f872e4a9da/areas-ofoutstanding-natural-beauty-aonbs. -. 2021b. "National Nature Reserves (NNRs)." https://data.gov.uk/dataset/ce3bdae3cc24-4fa9-8db0-a1fc2217e995/national-nature-reserves-nnrs. OECD. 2013. "Income Distribution." https://doi.org/10.1787/data-00654-en. ONS. "Health Expectancies OMI." 2017. 2017. https://www.ons.gov.uk/peoplepopulationandcommunity/healthandsocialcare/healthan dlifeexpectancies/methodologies/healthexpectanciesqmi. "Wealth 2022. and Assets Survey OMI." 2022. https://www.ons.gov.uk/peoplepopulationandcommunity/personalandhouseholdfinanc es/debt/methodologies/wealthandassetssurveyqmi. Pearson, Richard G., Terence P. Dawson, and Canran Liu. 2004. "Modelling Species Distributions in Britain: A Hierarchical Integration of Climate and Land-Cover Data." *Ecography* 27 (3): 285–98. Pedde, Simona, Paula A. Harrison, Ian P. Holman, Gary D. Powney, Stephen Lofts, Reto Schmucki, Marc Gramberger, and James M. Bullock. 2021. "Enriching the Shared Socioeconomic Pathways to Co-Create Consistent Multi-Sector Scenarios for the UK." The Science of the Total Environment 756 (February): 143172. Pedde, Simona, Kasper Kok, Katharina Hölscher, Niki Frantzeskaki, Ian Holman, Rob Dunford, Alison Smith, and Jill Jäger. 2019. "Advancing the Use of Scenarios to Understand Society's Capacity to Achieve the 1.5 Degree Target." Global Environmental Change: Human and Policy Dimensions 56 (May): 75-85. Polhill, J. G., N. M. Gotts, and A. N. R. Law. 2001. "Imitative versus Nonimitative Strategies

- Polhill, J. G., N. M. Gotts, and A. N. R. Law. 2001. "Imitative versus Nonimitative Strategies in a Land-Use Simulation." *Cybernetics and Systems* 32 (1–2). http://www.citeulike.org/user/jamesdamillington/article/2850188.
- Pyatt, G. 1995. "An Ecological Site Classification for Forestry in Great Britain." 260. Forestry Commission Research Division. https://www.forestresearch.gov.uk/documents/4950/RIN260.pdf.
- Rabin, Sam S., Peter Alexander, Roslyn Henry, Peter Anthoni, Thomas A. M. Pugh, Mark Rounsevell, and Almut Arneth. 2020. "Impacts of Future Agricultural Change on Ecosystem Service Indicators." *Earth System Dynamics* 11 (2): 357–76.
- Robinson, E. L., C. Huntingford, V. S. Semeena, and J. M. Bullock. 2022. "CHESS-SCAPE: Future Projections of Meteorological Variables at 1 Km Resolution for the United Kingdom 1980-2080 Derived from UK Climate Projections 2018." NERC EDS Centre

*for Environmental Data Analysis.* https://doi.org/10.5285/8194b416cbee482b89e0dfbe17c5786c.

- Robinson, Emma L., Eleanor Blyth, Douglas Clark, Edward Comyn-Platt, Jon Finch, and Ali Rudd. 2017. "Climate Hydrology and Ecology Research Support System Meteorology Dataset for Great Britain (1961-2015) [CHESS-Met] v1.2." https://doi.org/10.5285/b745e7b1-626c-4ccc-ac27-56582e77b900>.
- Robinson, Emma L., Eleanor M. Blyth, Douglas B. Clark, Jon Finch, and Alison C. Rudd. 2017.
   "Trends in Atmospheric Evaporative Demand in Great Britain Using High-Resolution Meteorological Data." *Hydrology and Earth System Sciences* 21 (2): 1189–1224.
- Rolo, Victor, Jose V. Roces-Diaz, Mario Torralba, Sonja Kay, Nora Fagerholm, Stephanie Aviron, Paul Burgess, et al. 2021. "Mixtures of Forest and Agroforestry Alleviate Trade-Offs between Ecosystem Services in European Rural Landscapes." *Ecosystem Services* 50 (August): 101318.
- Rounsevell, M., D. T. Robinson, and D. Murray-Rust. 2012. "From Actors to Agents in Socio-Ecological Systems Models." *Philosophical Transactions of the Royal Society of London. Series B, Biological Sciences* 367 (1586): 259–69.
- Rowland, C. S., R. D. Morton, L. Carrasco, G. McShane, A. W. O'Neil, and C. M. Wood. 2017. "Land Cover Map 2015 (1 Km Percentage Target Class, GB). NERC Environmental Information Data Centre."
- RSPB. 2021. "RSPB Reserves." 2021. https://opendatarspb.opendata.arcgis.com/datasets/6076715cb76d4c388fa38b87db7d9d24\_0/explore?1 ocation=55.360270%2C-3.252783%2C5.99.
- Scoones, I. 1998. "Sustainable Rural Livelihoods: A Framework for Analysis." Institute of Development Studies.
- Scottish Government. 2020a. "Local Nature Reserves (Scotland)." https://data.gov.uk/dataset/ff131012-8777-42c9-a263-97cead27ddee/local-naturereserves-scotland.

—. 2020b. "National Nature Reserves (Scotland)." https://data.gov.uk/dataset/5dae8e31-3ef3-4a2e-8c6c-31068e354c83/national-nature-reserves-scotland.

- ——. 2021a. "Cairngorms National Park Designated Boundary." https://data.gov.uk/dataset/8a00dbd7-e8f2-40e0-bcba-da2067d1e386/cairngormsnational-park-designated-boundary.
  - —. 2021b. "Loch Lomond and The Trossachs National Park Designated Boundary." https://data.gov.uk/dataset/6f63d73d-c45d-4947-8ad0-2d6f52b200ff/loch-lomond-andthe-trossachs-national-park-designated-boundary.
- ———. 2021c. "National Scenic Areas." https://data.gov.uk/dataset/8d9d285a-985d-4524-90a0-3238bca9f8f8/national-scenic-areas.
- Scottish Wildlife Trust. 2016. "Our Data." Scottish Wildlife Trust: Our Data. September 19, 2016. https://scottishwildlifetrust.org.uk/our-work/our-evidence-base/our-data/.
- Siebert, Rosemarie, Mark Toogood, and Andrea Knierim. 2006. "Factors Affecting European Farmers' Participation in Biodiversity Policies." *Sociologia Ruralis* 46 (4): 318–40.
- Smith, B., D. Wårlind, A. Arneth, T. Hickler, P. Leadley, J. Siltberg, and S. Zaehle. 2014. "Implications of Incorporating N Cycling and N Limitations on Primary Production in an Individual-Based Dynamic Vegetation Model." *Biogeosciences* 11 (7): 2027–54.
- SNH. 2020. "SNH Natural Spaces Sites of Special Scientific Interest." 2020. https://gateway.snh.gov.uk/natural-spaces/dataset.jsp?dsid=SSSI.
- Synes, Nicholas W., Calum Brown, Stephen C. F. Palmer, Greta Bocedi, Patrick E. Osborne, Kevin Watts, Janet Franklin, and Justin M. J. Travis. 2019. "Coupled Land Use and Ecological Models Reveal Emergence and Feedbacks in Socio-Ecological Systems." *Ecography* 42 (4): 814–25.

- Synes, Nicholas W., Calum Brown, Kevin Watts, Steven M. White, Mark A. Gilbert, and Justin M. J. Travis. 2016. "Emerging Opportunities for Landscape Ecological Modelling." *Current Landscape Ecology Reports* 1 (4): 146–67.
- Taylor, Karl E., Ronald J. Stouffer, and Gerald A. Meehl. 2012. "An Overview of CMIP5 and the Experiment Design." *Bulletin of the American Meteorological Society* 93 (4): 485– 98.
- UK Centre for Ecology & Hydrology. 2016. "Land Cover Map 2015." 2016. https://www.ceh.ac.uk/services/land-cover-map-2015.
- UNESCO. 2017. "Biosphere Reserves around the World." 2017. http://ihpwins.unesco.org/layers/mab\_biosphere\_reserves:geonode:mab\_biosphere\_reserves.
- Urban, Mark C., Justin M. J. Travis, Damaris Zurell, Patrick L. Thompson, Nicholas W. Synes, Alice Scarpa, Pedro R. Peres-Neto, et al. 2021. "Coding for Life: Designing a Platform for Projecting and Protecting Global Biodiversity." *Bioscience*. https://doi.org/10.1093/biosci/biab099.
- Vuuren, Detlef P. van, Jae Edmonds, Mikiko Kainuma, Keywan Riahi, Allison Thomson, Kathy Hibbard, George C. Hurtt, et al. 2011. "The Representative Concentration Pathways: An Overview." *Climatic Change* 109 (1): 5.