Resilient California water portfolios require infrastructure investment partnerships that are viable for all partners

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Abstract

Water scarcity is a growing problem around the world, and regions such as California are working to develop diversified, interconnected, and flexible water supply portfolios. To meet their resilient water portfolio goals, water utilities and irrigation districts will need to cooperate across scales to finance, build, and operate shared water supply infrastructure. However, planning studies to date have generally focused on partnership-level outcomes (i.e., highly aggregated mean cost-benefit analyses), while ignoring the heterogeneity of benefits, costs, and risks across the individual investing partners. This study contributes an exploratory modeling analysis that tests thousands of alternative water supply investment partnerships in the Central Valley of California, using a high-resolution simulation model to evaluate the effects of new infrastructure on individual water providers. The viability of conveyance and groundwater banking investments are as strongly shaped by partnership design choices (i.e., which water providers are participating, and how do they distribute the project's debt obligation?) as by extreme hydrologic conditions (i.e., floods and droughts). Importantly, most of the analyzed partnership structures yield highly unequal distributions of water supply and financial risks across the partners, limiting the viability of cooperative partnerships. Partnership viability is especially rare in the absence of groundwater banking facilities, or under dry hydrologic conditions, even under explicitly optimistic assumptions regarding climate change. These results emphasize the importance of high-resolution simulation models and careful partnership structure design when developing resilient water supply portfolios for institutionally complex regions confronting scarcity.

- 1 Resilient California water portfolios require infrastructure investment
- 2 partnerships that are viable for all partners
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11 Key Points

- Water portfolio planning frameworks need to account for heterogeneity across
 participating partners, not just "average" performance
- Exploratory modeling shows how most infrastructure partnerships are highly unequal in
 their water deliveries and financial risks
- Viable partnership design is shown to be especially difficult for dry hydrologic scenarios,
 or when building conveyance without storage

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21 working to develop diversified, interconnected, and flexible water supply portfolios. To meet

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Valley of California, using a high-resolution simulation model to evaluate the effects of new

29 infrastructure on individual water providers. The viability of conveyance and groundwater

30 banking investments are as strongly shaped by partnership design choices (i.e., which water

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36 conditions, even under explicitly optimistic assumptions regarding climate change. These results

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40 1 Introduction

In May 2021, California Governor Newsom announced a \$5.1 billion package for 41 "immediate drought response and long-term water resilience investments" (Office of Governor 42 Gavin Newsom, 2021). This follows the administration's Water Resilience Portfolio Initiative 43 (WRPI), an ambitious blueprint for bolstering the state's water security (Newsom et al., 2020). 44 The WRPI recommends a suite of actions to overcome challenges such as population growth, 45 groundwater overdraft, and aging infrastructure, as well as climate change, which is already 46 making droughts more frequent and severe (AghaKouchak, Cheng, Mazdiyasni, & Farahmand, 47 2014; AghaKouchak et al., 2021; Berg & Hall, 2017; Diffenbaugh, Swain, & Touma, 2015). 48 Focal point recommendations in the WRPI include (1) expanding, improving, and diversifying 49 the state's water storage and conveyance infrastructure, (2) developing flexible institutions for 50 water sharing (e.g., groundwater banking), and (3) preparing for more climate change-related 51 extreme droughts and floods. The WRPI envisions a future in which separate agencies, utilities, 52 and stakeholder groups collaboratively develop and manage a shared network of water 53 infrastructure that bridges local, regional, and statewide scales. However, at present, it is not 54 clear that planners have the tools they need to build this "cohesive, resilient 'water system of 55 56 systems' across California" (Newsom et al., 2020). In this work, we show that traditional water supply planning tools are unsuitable for the task at hand. Exploratory modeling contributes new 57 insights for designing and evaluating collaborative water investment partnerships under 58 uncertainty. These insights have broad relevance beyond California, including the entire Western 59

U.S. and other water-scarce regions around the world seeking to develop more resilient watersupply systems.

62 Water supply planning analyses generally rely on simulation models to evaluate the impacts of alternative policies and investments. However, modern supply networks in regions 63 such as California are extremely complex, both in terms of the engineered system of reservoirs, 64 canals, and groundwater recharge facilities, as well as the institutional systems of environmental 65 regulations, water rights, and groundwater banking arrangements (Escriva-Bou, Mccann, et al., 66 2020). Exacerbating these complexities, atmospheric rivers deliver a large fraction of the state's 67 annual precipitation during a few short events (Dettinger, Ralph, Das, Neiman, & Cayan, 2011; 68 Gershunov, Shulgina, Ralph, Lavers, & Rutz, 2017), introducing strong interdependencies 69 between floods and droughts. This makes it critical to resolve daily-scale dynamics (Hanak, 70 Jezdimirovic, et al., 2018; Kocis & Dahlke, 2017; Malek et al., 2022; Zeff et al., 2021), while 71 72 simultaneously multi-decadal simulations are needed to properly evaluate the impacts of longterm infrastructure investments, the slow dynamics of groundwater storage change (Manna, 73 74 Walton, Cherry, & Parker, 2019), and the deep uncertainties in climatic, economic, and 75 regulatory changes. Lastly, it is critical that planning models resolve a wide range of spatial scales and system actors in order to evaluate how water moves through statewide infrastructure 76 77 networks in response to local actions by individual water utilities, irrigation districts, and water 78 storage districts (Zeff et al., 2021) (hereafter referred to collectively as "water districts"). To 79 date, much of this complexity is beyond the reach of the state's primary water resources planning models (e.g., CalSim (California Department of Water Resources, 2017; Draper et al., 2004), 80 CalLite (Islam et al., 2011), CALVIN (Draper, Jenkins, Kirby, Lund, & Howitt, 2003)). 81 California has over \$33 billion of water-related expenditures per year, 85% of which are 82

funded by local agencies (Hanak, Chappelle, et al., 2018). Despite recent high-profile 83 commitments to water infrastructure from state and federal governments (Office of Governor 84 Gavin Newsom, 2021; The White House, 2021), individual water districts are responsible for the 85 vast majority of the funding. Water districts typically finance large capital projects through debt 86 that must be repaid over decades using water sales revenues. A key question is whether the 87 additional water gained from a project will generate sufficient revenues to cover debt payments 88 without requiring budget cuts, unpopular customer rate hikes, or even bankruptcy (Chapman & 89 Breeding, 2014; Jeffrey Hughes et al., 2014; Leurig, 2010). However, benefits can be difficult to 90 assess in light of system complexities and uncertainties as described above. Moreover, as the 91 state's water supply portfolio becomes more diverse, flexible, and inter-connected, it becomes 92 increasingly difficult for water districts to evaluate individual capital projects due to interactions 93 across the infrastructural and institutional networks (Haimes, 2018). 94

Cooperative finance and operation of infrastructure can benefit water districts through
economies of scale, reduced redundancy, and increased flexibility (Escriva-Bou, Sencan, Hanak,
& Wilkinson, 2020; Jeff Hughes & Fox, 2019; Riggs & Hughes, 2019). Larger and more diverse
coalitions may also be better positioned to raise capital and harness state and federal subsidies
(Cypher & Grinnell, 2007; Hansen, Mullin, & Riggs, 2020; Newsom et al., 2020). However,
cooperation introduces significant new complexities related to the heterogeneity of outcomes
across the participants. Capital projects that look favorable at the partnership level may yield

102 poor results for individual partners due to differences in water rights, groundwater recharge

- 103 capacities, and locations within the infrastructure network. Some partners may also bear an
- outsized share of the risk in unfavorable future scenarios (i.e., losses under climate change)
 (Gold, Reed, Trindade, & Characklis, 2019; Gorelick, Zeff, Hughes, Eskaf, & Characklis, 2019;
- Herman, Zeff, Reed, & Characklis, 2014). An additional challenge is assigning the share of
- 107 project debt to be borne by each partner; standard approaches for apportioning cost shares to
- 108 multiple beneficiaries are unsuitable when there is significant uncertainty or a large number of
- 109 potential partners (De Souza, Medellín-Azuara, Lund, & Howitt, 2011; Giglio & Wrightington,
- 110 1972). Recent water portfolio planning studies (San Joaquin River Restoration Program, 2011;
- Sunding, 2015; U.S. Bureau of Reclamation, 2017, 2020) have focused on expected
- 112 costs/benefits at highly aggregated levels, typically under minimal uncertainty, while
- vulnerability assessments under broader uncertainty have focused on individual water districts
- 114 (Groves et al., 2015; Lempert & Groves, 2010; Tariq, Lempert, Riverson, Schwartz, & Berg,
- 115 2017) or aggregate regional outcomes (Connell-Buck, Medellín-Azuara, Lund, & Madani, 2011;
- 116 Schwarz et al., 2019, 2018; Selmon et al., 2019). There is little research to date on disaggregating
- 117 costs, benefits, and risks to design robust partnerships that are broadly satisfactory to all partners
- 118 (Herman, Reed, Zeff, & Characklis, 2015; Jafino, Kwakkel, & Taebi, 2021).

In this work, we contribute an exploratory modeling framework (see detailed review by 119 120 Moallemi, Kwakkel, de Haan, & Bryan, 2020) for testing thousands of candidate infrastructure 121 partnership structures across multiple future hydrologic scenarios, using a flexible and highresolution water resources simulation model, the California Food-Energy-Water System 122 (CALFEWS (Zeff et al., 2021)). Each partnership structure is assessed in terms of aggregate 123 performance as well as its impacts on individual partners, in order to search for investments that 124 are viable for all partners across multiple plausible futures. We take an explicitly optimistic view 125 of uncertainty in this study by modeling outcomes under present-day demands, institutions, and 126 regulatory contexts. Similarly, we assume a limited degree of hydrologic uncertainty by focusing 127 on stationary hydro-climatic variability, rather than explicitly focusing on the more severe floods 128 and droughts expected with climate change (Gonzalez et al., 2018). A major aim of this 129 exploratory analysis is to show that, even under these strongly optimistic assumptions, traditional 130 planning frameworks are unlikely to produce viable investment partnerships. 131

Our results focus on the southern Central Valley, a water-stressed and agriculturally-132 productive region that relies heavily on overdrafted aquifers to meet its irrigation and drinking 133 water demands, particularly during drought, and may face severe cutbacks under the Sustainable 134 Groundwater Management Act (Faunt & Sneed, 2015; Hanak et al., 2019; Levy et al., 2021; 135 Newsom et al., 2020). Water districts are mobilizing to develop new infrastructure (e.g., canals, 136 groundwater recharge facilities) and flexible institutions (e.g., water trading, groundwater 137 banking) in order to balance supplies and demands (Escriva-Bou, Sencan, et al., 2020; Hanak, 138 Jezdimirovic, et al., 2018; Jezdimirovic, Hanak, & Escriva-Bou, 2020; Scanlon, Reedy, Faunt, 139 Pool, & Uhlman, 2016). Thus, the region is emblematic of the challenges and opportunities 140 141 facing water supply organizations throughout the Western U.S. and other water-stressed regions

142 of the world, and can advance our ability to develop resilient and sustainable water portfolios

143 capable of managing intensifying scarcity and climate-related uncertainty (AghaKouchak et al.,

144 2021; Famiglietti, 2014; Jiménez Cisneros et al., 2015; Lall et al., 2018).

145 **2 Methods**

146 2.1 Overview

147 The Friant-Kern Canal is a major water conveyance system in the southeastern Central Valley. As part of the Central Valley Project, it diverts San Joaquin River water from Lake 148 Millerton to the "Friant contractors" to the south (Figure 1). The canal also conveys water from 149 other local rivers to a wider array of water districts and groundwater banks willing to take 150 surplus deliveries during high-flow periods. However, widespread subsidence caused by 151 groundwater overdraft has damaged the canal and reduced its capacity by 60% along critical 152 153 reaches (Faunt & Sneed, 2015; Friant Water Authority, 2019). Water districts relying on the canal, especially the Friant contractors, are advocating for its rehabilitation. Simultaneously, 154 many water districts are working to build more groundwater recharge and banking facilities 155 (Dahlke et al., 2018; Hanak, Jezdimirovic, Escriva-Bou, & Ayres, 2020; Jezdimirovic et al., 156 2020; Scanlon et al., 2016). Building conveyance and groundwater recharge simultaneously 157 could yield synergistic benefits when the additional deliveries due to canal expansion, which 158 primarily occur during high-flow periods, exceed immediate irrigation demands and require 159 storage until drier periods (Alam, Gebremichael, Li, Dozier, & Lettenmaier, 2020; Hanak, 160 Jezdimirovic, et al., 2018; Kocis & Dahlke, 2017; Wendt, van Loon, Scanlon, & Hannah, 2021). 161 Local farms, water districts, and politicians have lobbied for external support, but subsidies are 162 unlikely to fund the projects in full (Whisnand, 2021). Thus, water districts that stand to benefit 163 will need to collaborate to raise the remaining capital, creating the challenge of defining, 164 designing, and evaluating these investment partnerships. 165

In this study, an exploratory ensemble modeling approach is used to evaluate thousands 166 of plausible partnership structures. Each candidate partnership structure contains a subset of 41 167 water districts (Tiers 1-3 in Figure 1) that could potentially benefit from rehabilitating the canal 168 and/or adding a new groundwater bank along the canal. Our exploratory sampling design assigns 169 170 to each partner an ownership share that controls its access to capacity in the new infrastructure and its share of debt payments. Each candidate partnership structure is applied to three different 171 capital projects (canal rehabilitation, groundwater bank development, and both). Each of the 172 capital projects are evaluated using three different 50-year hydrologic scenarios (wet, average, 173 and dry) that are selected from a larger stochastic streamflow ensemble that captures the 174 system's internal hydro-climatic variability. All partnership structure, capital project, and 175 hydrologic scenario combinations are simulated using the CALFEWS model (Zeff et al., 2021). 176 Water portfolio investments are assessed using two key performance metrics. First is the 177 captured water gain in billions of liters per year, GL/year (or thousands of acre-feet per year, 178 kAF/year). This metric measures the expected "new water" delivered by the investment, defined 179 as the difference in average annual water deliveries to partner districts with and without the new 180 infrastructure under a particular hydrologic scenario. Each capital project represents a large up-181 front investment that requires annual debt payments, with each partner's share of debt equal to its 182 ownership share. The second measure of performance is the cost of gains in dollars per million 183

- liters, \$/ML, (or dollars per acre-foot, \$/AF), defined as the annual debt payment divided by the 184
- captured water gain. Effective partnerships will reliably provide their partners with significant 185 186 water gains at relatively low cost.
 - Madera Canal uis Reservoir Millerton Lake Pacheco San Joaquin Tunnel Pine Flat River Lake Friant-Kern Canal Kings California Lake River Aqueduct Kaweah Kaweah R Lake Tule River Surface reservoir Las Success Perillas Natural river Pumping Canal Lake Plant **Urban turnout** sabella Tier 1 district Kern Cross River **Tier 2 district** Valley Canal **Tier 3 district** Other district Arvin-Edison Canal Friant contractors dmonston Map tiles by Stamen Design, umping CC BY 3.0 -- Map data (C) Plant OpenStreetMap contributors
- Figure 1. Geography of water supply in the southern Central Valley. Five major reservoirs store 188 runoff from the Sierra Nevada mountains in the east and release it into the region's major rivers, 189 where much of the flow is withdrawn by water districts. Millerton Lake diverts San Joaquin 190 River water into the Madera Canal and Friant-Kern Canal (FKC) as part of the Central Valley 191 Project (CVP). Water districts receiving CVP contract water from the FKC are shown as "Friant 192 Contractors". The CVP and State Water Project also pump water from the Sacramento-San 193 Joaquin River Delta to San Luis Reservoir, where it is routed via a series of pumps and canals to 194
- water districts in the valley and urban districts along the coast. Water districts are grouped into 195
- three tiers based on the results of the experiment, with Tier 1 districts having the highest 196
- potential to benefit from new infrastructure and Tier 3 districts having the lowest potential. Note 197
- that three coastal urban districts from the experiment are not shown (one Tier 2 and two Tier 3). 198
- 2.2 CALFEWS simulation model 199
- The California Food-Energy-Water System (CALFEWS) is an open-source, 200
- Python/Cython-based model for simulating the movement of water supplies within California, 201
- 202 with a focus on the Central Valley (Zeff et al., 2021). CALFEWS operates across multiple
- scales, from statewide representation of major inter-basin transfer projects to distributed local 203



representation of district-level conjunctive surface and groundwater management. The model has
been found to reproduce historical reservoir storages, canal flows, surface water deliveries, and
groundwater banking accounts with a high degree of fidelity.

CALFEWS has three major advantages when compared to more commonly-used water 207 supply models in California (e.g., CalSim (California Department of Water Resources, 2017; 208 Draper et al., 2004), CalLite (Islam et al., 2011), CALVIN (Draper et al., 2003)). Firstly, it 209 models dynamics at a finer spatio-temporal resolution (water district representation, daily time 210 step), while still allowing for simulations to be run at a statewide scale over many decades. This 211 gives CALFEWS an unprecedented ability to track the impacts of district-level decision-making 212 on statewide water supply projects, and the impacts of short-lived high-flow periods (e.g., 213 atmospheric rivers) on long-lived infrastructure investments. Secondly, CALFEWS uses a rules-214 based representation of system dynamics, in contrast to the mathematical programming 215 techniques used by the aforementioned models. This rules-based approach can more flexibly 216 represent system complexities such as adaptive operations, environmental regulations, 217 groundwater banking arrangements, and distributed district-level decision-making. Lastly, 218 219 ensembles of CALFEWS simulations can be dispatched in parallel on high-performance computing infrastructure, enabling high-throughput exploratory modeling approaches (see 220

221 Section 2.5).

222 2.3 Infrastructure project alternatives

The first infrastructure project considered in this study is the rehabilitation of the Friant-223 Kern Canal. Widespread groundwater overdraft in the region has damaged the canal via 224 subsidence (Faunt & Sneed, 2015). In certain sections near the Tule River, canal capacity has 225 been reduced by almost 60% (Friant Water Authority, 2019; U.S. Bureau of Reclamation, 2020). 226 Water districts and government representatives have recently been negotiating a partnership to 227 rehabilitate the canal. Although the new design specifications are uncertain, we assume that the 228 entire length of the canal will be returned to its original design capacity and that the cost borne 229 by the districts involved in the partnership will be \$50 million. This is based on a recent funding 230 agreement for the project, in which the Friant contractors have agreed to pay \$50 million out of 231 an estimated \$500 million total (Whisnand, 2021). The remainder will come from federal and 232 state funding sources as well as legal settlement agreements associated with environmental 233 damages and land subsidence. The partnership design efforts in this study are focused on the \$50 234 million share currently allocated to the Friant contractors. This is an unusually favorable case 235 study for infrastructure investment partnerships due to the unusually high level of outside 236 funding available; for other capital projects with lower subsidy levels, viability will generally be 237 more difficult to achieve. 238

The second capital project is a jointly managed groundwater bank along the Friant-Kern Canal in the vicinity of the Tule River. This project is hypothetical and is not based on any particular existing or planned groundwater bank, but water districts throughout the region have been investing in new recharge facilities and banking relationships (Escriva-Bou, Sencan, et al., 2020; Hanak et al., 2020). The partnership's share of the total capital cost (i.e., after any subsidies) is assumed to be \$50 million, in line with cost estimates of other large groundwater banks currently under development (Jezdimirovic et al., 2020). There are three main parameters

- controlling the function of groundwater recharge and recovery facilities in the CALFEWS model
- 247 (Zeff et al., 2021): the baseline recharge capacity, the baseline recovery capacity, and the storage
- volume of infiltration basins. The two "baseline" values refer to initial capacities at the
 beginning of the recharge or recovery season; both capacities will decrease over the season with
- extended use. The uncertainty bounds for these parameters are set based on the ranges of pre-
- existing groundwater banks in the region: infiltration capacity between 0 and 1.5 GL/day (1.2)
- kAF/day, infiltration pond storage volume between 0 and 1.5 GL (1.2 kAF), and recovery
- capacity between 0 and 0.9 GL/day (0.7 kAF/day). For comparison, the upper limits of these
- ranges are 50%, 50%, and 88% of the estimated parameters for the largest groundwater bank in

the region, the Kern Water Bank (Kern Water Bank Authority, 2021).

Access to both pieces of infrastructure is restricted to the set of water districts investing in the partnership. Additionally, each partner is assigned priority access to a fraction of the new capacity that is proportional to its ownership share in the project (e.g., a district with a 20% ownership share will have priority access over 20% of the new capacity). If a district is not using its priority capacity at any given time, access is opened up to the broader set of partners. Note that the canal rehabilitation project does not impact non-partners' access to the existing capacity. It only restricts access to the expanded capacity at the top of the canal.

Both capital projects are assumed to be financed with revenue bonds that require equal-263 sized annual payments over a 50-year period with 3% interest. This is a conservative assumption 264 (in the sense of not over-estimating costs) because recent revenue bonds issued by California 265 water districts have generally carried between 2.5-3.5% interest with maturities of 25 years or 266 fewer (California State Treasurer's Office, 2020), and a 50-year maturity bond would carry an 267 additional premium in practice. Under these assumptions, partnerships would make annual debt 268 payments of approximately \$1.943 million for either canal rehabilitation or groundwater bank 269 development, and \$3.887 million for both projects. Each partner's share of the overall debt 270 payment is proportional to its ownership share. 271

272 2.4 Hydrologic scenarios

The hydrologic scenarios used in this work are generated using the California and West 273 Coast Power System (CAPOW), an open-sourced, Python-based model for simulating the 274 275 operation of the U.S. west coast bulk electric power system. CAPOW has a major focus on the impact of hydrometeorological variables (streamflow, temperature, wind speed, insolation) on 276 277 system reliability and pricing (Su, Kern, Denaro, et al., 2020). As such, a major component of the model is its stochastic engine, which uses a hybrid vector autoregressive-bootstrapping 278 279 approach to generate daily synthetic records of hydrometeorological variables at many locations 280 across the west coast. These synthetic records are shown to maintain historical correlation patterns across space (i.e., northern vs. southern California) and time (i.e., intra- and interannual 281 autocorrelation), while overcoming the limitations of the limited length of historical hydro-282 climatic observations. Although CAPOW is trained on historical data, it enables the generation 283 of statistical replicate time series of synthetic hydro-climatic scenarios, better capturing a wide 284 variety of plausible futures as well as extremes beyond the limited number of observed rare 285

events in the modern historical record. This makes it a valuable tool for planning over the
medium term (e.g., initiating capital investments in the next decade), where interannual
variability is a major contributor to overall hydrologic uncertainty (Doss-Gollin, Farnham,
Steinschneider, & Lall, 2019; Lehner et al., 2020; Su, Kern, Reed, & Characklis, 2020).

For the present work, the CAPOW stochastic engine is used to generate daily synthetic 290 full natural flow records at fifteen major water supply reservoirs in California. These full natural 291 flow records can be used to statistically reconstruct more detailed time series of streamflows, 292 gains, and snowpacks (Zeff et al., 2021). We generate 101 alternative 50-year time series, which 293 span a wide range of hydrologic outcomes across extreme quantiles of interest. Figure 2 and 294 Supporting Information (SI) Figure S1 compare each 50-year synthetic record with the 111-year 295 historical record in terms of 1-, 2-, 4-, and 8-year flow duration curves of full natural flow across 296 the major South San Joaquin River Basin (SSJRB) reservoirs (Millerton, Pine Flat, Kaweah, 297 Success, and Isabella), and the major Sacramento River Basin (SRB) reservoirs (Shasta, 298 Oroville, New Bullards Bar, and Folsom), respectively. These curves demonstrate that the 299 stochastic engine can accurately represent regional flows in terms averages, extremes, and 300 persistence of multi-year wet and dry periods, while also providing a wider range of extremes for 301

302 risk assessment.

These 101 scenarios are then sorted in terms of total 50-year full natural flow across the 303 SSJRB reservoirs. The time series with the highest, median, and lowest flows are selected to 304 represent "wet", "average", and "dry" hydrologic scenarios, respectively. The "average" scenario 305 is found to track the historical flow duration curve well in the SSJRB (Figure 2), while the "wet" 306 and "dry" scenarios are significantly wetter and drier, especially with respect to normal and wet 307 periods (e.g., mid- to high-range exceedance). With respect to the most extreme droughts (e.g., 308 low exceedance), the "average" scenario is found to be quite extreme, while the "dry" scenario is 309 similar to the historical record. The "wet" and "dry" scenarios are less extreme in the context of 310 the SRB (SI Figure S1), and the "average" scenario appears to be wetter than average. This is 311 due to the imperfect correlation in hydrologic conditions across the state; the wettest years in 312 northern California are not necessarily the wettest years in the southern Central Valley, and vice 313 versa. The SSJRB reservoirs are used for scenario selection because they have the most direct 314 influence on the Friant-Kern Canal. 315

316 2.5 Exploratory modeling framework

This study employs a random sampling framework to develop 3,000 plausible candidate 317 partnership structures. The sampling design requires a three-step process to generate each 318 319 partnership structure. First, the size of the partnership *np* is randomly drawn from a Poisson distribution with a mean of 8 partners, excluding zero. This distribution, which creates 320 321 partnership structures ranging from approximately one to twenty partners, was selected to strike a balance between relatively dense sampling at smaller partnership sizes (which are more likely 322 to be viable) and broader exploration of larger partnership structures. Second, np partners are 323 randomly selected without replacement from the set of 41 candidate water districts. Third, the 324 ownership shares of the *np* partners are sampled from a uniform distribution and normalized to 325 sum to one. 326



Figure 2. Comparison of synthetic and historical full natural flow (FNF) for the Tulare Lake

Basin. Panels (a-d) show the total full natural flows for the five major reservoirs of the South

330 San Joaquin River Basin (Millerton, Pine Flat, Kaweah, Success, and Isabella), under the

historical record and the wet, average, and dry synthetic scenarios. Panels (e-h) show the full

natural flow duration curves for each time series over 1-, 2-, 4-, and 8-year periods.

Each of the 3,000 sampled partnership structures is combined with each of three capital 333 projects: canal rehabilitation, groundwater bank development, and both. Three groundwater 334 recharge and recovery parameters are randomly sampled from uniform distributions across their 335 uncertainty bounds (Section 2.3). This results in 9,000 partnership structure/capital project 336 combinations, to which we add the "Friant-16" canal rehabilitation partnership representing 337 business-as-usual planning (see Section 3.4), and the baseline case with no new infrastructure. 338 Lastly, each of these combinations is combined with each of the three stochastic hydrologic 339 scenarios: dry, average, and wet. In total, this represents 27,006 partnership structure/capital 340 project/hydrologic scenario combinations. 341

Each of these 27,006 combinations is evaluated using the CALFEWS model. Simulations are dispatched in parallel across 800 cores using the Longleaf Cluster at the University of North Carolina at Chapel Hill. Results such as reservoir releases, canal flows, water contract deliveries, and groundwater banking balances for each simulation are stored using the hdf5 file format. Each simulation is evaluated by comparing its performance to the performance of the baseline

347 no-infrastructure case under the same hydrologic scenario.

348 **3 Results**

349 3.1 Evaluating new infrastructure investments along the Friant-Kern Canal

350 The candidate partnerships explored in this study display a wide range of partnershiplevel performance (Figures 3a-b). Five percent deliver at least 62 GL/year (50 kAF/year) of 351 captured water gains, with a maximum of 118 GL/year (96 kAF/year). For context, Lake 352 Millerton has a capacity of 642 GL (521 kAF), and 1 GL (1 kAF) is enough irrigate roughly 353 0.82-2.2 km² (250-667 acres) of crops in the region (University of California Agriculture and 354 Natural Resources, 2021). Five percent of candidate partnerships cost less than \$45/ML 355 (\$56/AF), with a minimum of \$30/ML (\$36/AF). These results would be competitive with other 356 water supply projects under consideration throughout the Central Valley (Jezdimirovic et al., 357 2020). However, most candidate partnerships perform more modestly, with a median gain of 21 358 GL/year (17 kAF/year) and a median cost of \$120/ML (\$147/AF). This cost of gains, which 359 includes only the investment's debt payments and not the additional costs of procuring and 360 transporting the water itself, would represent a significant increase above typical rates of \$32-361 154/ML (\$40-190/AF) charged by water districts for irrigation deliveries in the region. Worse 362 still, 9% of candidate partnerships yield negative captured water gains, representing investments 363 that actually reduce partners' water deliveries on average. This occurs when new infrastructure 364 triggers unpredictable dynamics within the water system that allow non-partners to benefit over 365 partners. The cost of gains for these partnerships is effectively infinite, and more broadly 13% of 366 partnerships have very high costs over \$1,000/ML (\$1,233/AF). These capital investments 367 represent a serious financial risk if the future water gains are insufficient to allow partners to sell 368 enough water to pay off their debt, even under our explicitly optimistic assumptions in this study 369 that neglect the broader extremes possible under climate change. Thus, near term capital 370 investments can have very complex and potentially severe downside risks for partners. 371



372

Figure 3. Distribution of performance for thousands of candidate infrastructure investment 373 partnerships. (a) Distribution of partnership-level captured water gains. (b) Distribution of 374 partnership-level cost of gains. (c) Distribution of captured water gains for individual water 375 districts. (d) Distribution of cost of gains for individual water districts. Distributions show the 376 variability of results across all candidate partnership structures, capital projects, and hydrologic 377 scenarios. For Panels (b) and (d), all costs over \$1,000/ML (\$1,233/AF) are consolidated into 378 "1000+". Water districts are grouped into three tiers based on the results of the experiment, with 379 Tier 1 districts having the highest potential to benefit from new infrastructure and Tier 3 districts 380 having the lowest potential. 381

Decomposing these results by hydrologic scenario, the costs of gains are generally 382 highest in the dry scenario (Figure 4) because the size of debt payments is independent of how 383 much water ends up being available. However, costs in the wet scenario are similar to the 384 average scenario, suggesting an important asymmetry: the downside risk in a drier future is 385 larger than the upside risk in a wetter future. Decomposing variability by capital project, we find 386 similar cost distributions for canal rehabilitation projects vs. joint canal-groundwater banking 387 projects. However, the latter risks underperformance in dry scenarios if insufficient water is 388 available to warrant the capital cost of both projects. Groundwater bank-only partnerships 389 generally perform well in the wet scenario but much poorer in average and dry scenarios. 390

Although the project type and hydrologic scenario do impact performance in important ways, substantial variability remains after accounting for these factors (Figure 4). This variability is attributed to the partnership structure itself (i.e., which water districts are included as partners, and what ownership share does each partner take?). For example, the subset of candidate partnerships that expand the canal under the wet hydrologic scenario (Figure 4a) are identical other than their partnership structures, yet experience widely varying partnership-level costs ranging from \$32/ML to over \$1,000/ML (\$39-1,233/AF). For candidate partnerships with

- 398 groundwater banking (either alone or in combination with canal expansion), three additional
- capacity factors also contribute to the variability of outcomes (e.g., soil infiltration rate).
- 400 However, these appear to play a minor role compared to partnership structure. Thus, the impacts
- 401 of water portfolio investments can depend critically on the subset of water districts in the
- 402 partnership and their relative ownership shares. This represents a major challenge for standard
- infrastructure planning approaches that focus primarily on physical design parameters such as
- 404 capacity (perhaps in combination with climate-related vulnerability analyses), while neglecting
- 405 contractual design factors such as infrastructure access and ownership.



Figure 4. Distribution of performance (partnership-level cost of gains) for candidate partnerships
across different capital projects and hydrologic scenarios. Panels (a-c) involve a canal expansion,

409 (d-f) involve a groundwater bank, and (g-i) involve both projects. Panels (a/d/g) use the wet

- 410 hydrologic scenario, (b/e/h) use the average scenario, and (c/f/i) use the dry scenario. All costs
- 411 over \$1,000/ML (\$1,233/AF) are consolidated into "1000+".
- 412 3.2 Potential benefits are highly heterogeneous across water districts

Next, we disaggregate partnership-level performance into the captured water gains 413 (Figure 3c) and costs of gains (Figure 3d) experienced by each of the 41 water districts in the 414 exploratory experiment. Each district's results include only the subset of candidate partnership 415 structures for which it is a participant. Most districts display a wide range of captured water 416 gains (e.g., from <0 to >20 GL/year), but with outcomes concentrated in a narrower band (e.g., 417 >90% of partnerships between 0-10 GL/year). The range of costs experienced by each district is 418 even wider, with all districts except for one spanning from near-zero to over \$1,000/ML. Despite 419 the high variability within each district, there are clear distributional differences as well. Water 420 districts can be grouped into three roughly-equal sized tiers based on their median cost of water 421 gains: Tier 1 under \$100/ML (\$123/AF), Tier 2 under \$240/ML (\$296/AF), and Tier 3 over 422 \$240/ML. Tier 1 districts have the highest potential to benefit from partnering in these 423 investments, while Tier 3 districts have the lowest. These distributional differences provide 424

425 crucial information for the partnership negotiation process, but would be unavailable using426 traditional aggregate cost-benefit analyses.

427 Decomposing the district-level performance into separate components for each capital project and hydrologic scenario, we find costs to be highly variable across all combinations, 428 highlighting the importance of partnership structure on district-level outcomes (Figure 5). Tier 1 429 districts generally have much lower costs than Tier 2-3 for canal-only projects, while the 430 distributions are more uniform across tiers for bank-only projects. Joint canal-groundwater 431 banking projects exhibit intermediate behavior; Tier 1 districts generally have lower costs than 432 Tier 2, which have lower costs than Tier 3, but these differences are more gradual than in the 433 canal-only partnerships. With regards to hydrology, the costs of gains are significantly higher on 434 average in the dry scenario than the average and wet scenarios, but this effect varies across 435 districts, suggesting that some districts will experience more climate-related risk than others. We 436 reiterate that our experimental design is expressly optimistic with regards to climate uncertainty, 437 considering extreme realizations of stationary stochastic variability but not anthropogenically 438 forced climate non-stationarity. Thus, under climate change, the downside climate-related risks 439 are likely even greater than represented here. 440

The geographic distribution of the tiers provides useful context (Figure 1). Tier 1 441 stretches from the Kaweah River to the Kern River along the western side of the Friant-Kern 442 Canal, before wrapping around the other major canals to the south. These districts' proximity to 443 the confluences of major canals and rivers allows them to receive water from multiple sources, 444 increasing their chances of capitalizing on the investment. Additionally, many of the Tier 1 445 districts overlie highly suitable soils for groundwater recharge (O'Geen et al., 2015) (SI Figure 446 S2), which has allowed them to build dedicated recharge facilities within their boundaries (Alam 447 et al., 2020; Scanlon et al., 2016; Wendt et al., 2021). These facilities increase the benefits from 448 canal expansion by storing more surplus flows during winter/spring months when agricultural 449 demands are limited. Districts without significant groundwater recharge capacity are more 450 limited in how much water they can receive during these periods. This also helps to explain why 451 the benefits from groundwater bank development are more uniformly distributed across the 452 districts (Figure 5d-f) because these facilities level the playing field by allowing districts with 453 unsuitable soils to store groundwater off-site. Lastly, Figure 1 shows that the water districts with 454 the highest potential to gain from new infrastructure are not necessarily the most "obvious" ones. 455 In this case, we find a number of Friant contractors (the group of districts currently negotiating to 456 invest in the project (Whisnand, 2021)) in Tiers 2-3, while a number of non-contractors are in 457 Tier 1. This demonstrates the value of the exploratory modeling approach to partnership design, 458 which allows a much broader array of suitable alternatives to be discovered. 459



Figure 5. Distribution of performance (cost of gains for individual districts) for candidate

- 462 partnerships across different capital projects and hydrologic scenarios. Panels (a-c) involve a
- 463 canal expansion, (d-f) involve a groundwater bank, and (g-i) involve both projects. Panels (a/d/g)
- 464 use the wet hydrologic scenario, (b/e/h) use the average scenario, and (c/f/i) use the dry scenario.
- 465 All costs over 1,000/ML (1,233/AF) are consolidated into "1000+".
- 466 3.3 Heterogeneity limits the viability of partnerships
- 467 Results thus far have shown how the benefits of water portfolio partnerships can vary
 468 widely based on the type of infrastructure, the hydrologic scenario, and the partnership structure

469 itself. We now demonstrate how this range of district-level outcomes can threaten investment viability. The partnerships in this work are assumed to be voluntary, meaning all partners have 470 agreed to the terms of cooperation. This implies that each partner believes they will be better off 471 with the partnership than without it (Giglio & Wrightington, 1972). In practice, coalition-472 building is complex due to power and information asymmetries, incentive misalignments, and 473 relationships between participants (BenDor & Scheffran, 2019; Hansen et al., 2020; Lubell, 474 Blomquist, & Beutler, 2020; Madani, 2010; Read, Madani, & Inanloo, 2014). However, we 475 adopt the following simple definition: an infrastructure partnership is considered viable under a 476 particular hydrologic scenario if all partners pay less than \$200/ML (\$247/AF) for their share of 477 captured water gains. This is an optimistically conservative definition, in the sense of not 478 erroneously labelling viable partnerships as non-viable, because water districts in the region 479 typically charge their customers \$32-154/ML (\$40-190/AF) for irrigation water (University of 480 California Agriculture and Natural Resources, 2021). Therefore, \$200/ML in new project debt, 481 on top of additional water procurement costs, would represent a significant and likely 482

483 unacceptable increase.

484 Despite our optimistically conservative viability definition, only 8% of candidate partnerships we have explored are viable (Figure 6). In the preponderance of cases, the worst-off 485 partner pays significantly more than the partnership-level cost. In 61% of cases, performance is 486 487 considered viable at the partnership level (i.e., overall cost is <\$200/ML), but non-viable when 488 impacts are disaggregated (i.e., at least one partner pays >\$200/ML). This suggests that in a majority of cases, the limiting factor for project viability is not the overall volume of captured 489 water gains, but rather the uneven distribution of these gains across the partnership in light of 490 each partner's share of project debt. Traditional planning approaches based on aggregate impacts 491 are incapable of uncovering these distributional issues, and thus risk leading to cooperative 492 capital investments that are harmful to a subset of partners. 493

Decomposing viability by hydrologic scenario also delivers valuable insights (Figure 6 494 inset). In the wet hydrologic scenario, 15% of candidate partnerships are found to be viable, 495 compared to 9% in the average scenario and 1% in the dry scenario. This suggests that if the 496 future is drier than the past, investment partnerships will be more likely to fail for at least one 497 partner. California has already begun getting hotter and drier under anthropogenic climate 498 change (Gonzalez et al., 2018), which presents a major obstacle to meeting the state's 499 collaborative portfolio investment goals. Decomposing the results by capital project shows that 500 the canal expansion alone is very unlikely to be viable (1%). Groundwater bank development 501 improves the chances of viability, either in isolation (7%) or in combination with canal 502 expansion (17%). It is instructive to compare these results to Figure 4, which shows that canal 503 expansion projects and joint canal-groundwater banking projects have roughly similar odds of 504 505 achieving viability at the partnership level (i.e., aggregate costs <\$200/ML). Thus, although canal expansion appears promising in aggregate, it tends to distribute captured water gains more 506 507 unevenly and thus is less likely to satisfy all of the participating districts (see also Figure 5).



Figure 6. Viability of candidate infrastructure investment partnerships. Each simulated

510 partnership is plotted according to the cost of gains for the partnership as a whole vs. the worst-

off partner. The project type and hydrologic scenario used for each simulation are represented by

512 marker type and color, respectively. Viable partnerships (those with costs < 200/ML (247/AF)

513 for the worst-off partner) are represented with black outlines and higher opacity. All costs over

\$1,000/ML (\$1,233/AF) are consolidated into "1000+". Inset shows the viability of candidate
 partnership structures under each combination of capital project and hydrologic scenario,

515 partnership structures under each combination of capital project and hydrologic sc

represented by color as well as the percentage printed in each square.

517 3.4 Comparing alternative partnerships

We now explore these issues in more detail through a comparison of three alternative 518 partnership structures. First, we consider a partnership between 16 Friant contractors ("Friant-519 16") for canal expansion only. In contrast to the randomly sampled partnerships from the 520 exploratory study, Friant-16 is deliberately constructed to represent the baseline performance 521 522 under business-as-usual planning because the Friant contractors are currently negotiating to establish such a partnership (Whisnand, 2021). The ownership shares for Friant-16 are assumed 523 to be proportional to historical deliveries of CVP-Friant contract water (Congressional Research 524 Service, 2007). 525

In addition, we consider two high-performing partnerships from the exploratory study: 526 one with eight partners ("Alt-8") and one with three ("Alt-3"). Both are joint canal-groundwater 527 banking investments. The selection procedure for choosing Alt-8 and Alt-3 is as follows. First, 528 the set of 27,000 simulations is filtered down to include only those that meet each of the 529 following criteria: (1) the partnership is viable, i.e., the cost of gains for the worst-off partner is 530 less than \$200/ML; (2) the partnership-wide captured water gain is greater than 55 GL/year (45 531 kAF/year); and (3) the total captured water gain for all non-partners in the region is greater than 532 zero. These thresholds are set *a posteriori* by iteratively increasing the constraints until an elite 533 subset of solutions remain. In practice, the thresholds could be defined by decision-makers based 534

- on their particular context and preferences. Note that Friant-16 fails on all three criteria under the
- average hydrologic scenario: its worst-off partner cost is over \$1000/ML, its total captured water
- gain for the partnership is 46 GL/year, and its captured water gain for non-partners is -2 GL/year
 (i.e., the new infrastructure reduces deliveries to regional non-partner districts). Each of these
- (i.e., the new infrastructure reduces deliveries to regional non-partner districts). Each of thesecriteria eliminates a significant subset of simulations (SI Figures S3-S4). After filtering by the
- 540 three criteria, only 103 simulations remain: 51 from the wet scenario and 52 from the average
- 541 scenario, while no partnerships remain from the dry scenario. SI Figure S5 shows how these
- 542 simulations vary along multiple performance metrics. From the 52 candidate partnerships that
- 543 perform satisfactorily in the average scenario, Alt-3 and Alt-8 are selected manually as examples
- that perform well across all metrics. Alt-3 is chosen as a representative small successful
- 545 partnership, while Alt-8 is chosen to from among the larger partnerships, which may be 546 preferable due to political and financial concerns.



Figure 7. Comparison of three alternative infrastructure investment partnerships at aggregate
scale. For each partnership and each hydrologic scenario, Panel (a) shows the total captured
water gain for the partnership, Panel (b) shows the average captured water gain per partner, and
Panel (c) shows the average cost of gains for the partnership. Results are represented by color as
well as the number printed in each square.

At the partnership level, Alt-3 is found to deliver the highest total captured water gains, at 553 73 GL/year (59 kAF/year) in the average scenario (Figure 7a). Alt-8 delivers 71 GL/year (57 554 kAF/year) of gains, while Friant-16 delivers significantly less at 46 GL/year (38 kAF/year). 555 Gains in the wet scenario are slightly lower for Friant-16 and slightly higher for the other two 556 partnerships, while dry scenario gains are roughly 50% lower across all partnerships. Because 557 Alt-3 only has three partners, it has significantly higher gains on a per-partner basis (Figure 7b). 558 559 However, after accounting for the larger debt shares borne by each district in the smaller partnerships and the additional capital expense required for Alt-3 and Alt-8 to develop a 560 groundwater bank, the three partnerships are found to have similar costs of gains (Figure 7c): 561

\$43-50/ML (\$53-62/AF) in the wet scenario, \$42-53/ML (\$52-68/AF) in the average scenario,
and \$84-97/ML (\$104-119/AF) in the dry scenario.

564 When performance is disaggregated, however, significant differences emerge (Figure 8). District-level captured water gains for each partnership are quite heterogeneous, with some 565 districts receiving less than average and others receiving more. This is not necessarily a problem; 566 as long as each district's captured water gain is proportional to its ownership share and thus its 567 share of project debt, then the costs of gains across the partnership will be uniform. However, 568 this is found to be very uncommon. The costs of gains for Friant-16 are especially 569 heterogeneous, spanning from under \$23/ML to over \$1,000/ML. This means that some districts 570 receive more than their "fair share" of gains based on their ownership share, and others receive 571 far less than their fair share (in fact, some districts' gains can be negative). The other two 572 partnerships experience smaller but meaningful levels of cost heterogeneity. Interestingly, the 573 costs for Alt-8 are much more heterogeneous under the dry scenario than the wet and average 574 scenarios. This suggests that climate-related risk is unevenly held across the partnership, which 575 is vital information for partners to have when planning major investments under uncertainty. If 576 577 the future turns out to be drier than the historical record, then planning studies based on historical records are likely to underestimate not only the overall cost of gains from the capital project but 578 579 also the level of inequality in how these costs are distributed.

District tier is generally indicative of performance across the three partnerships, but 580 imperfectly so. The highest-performing partnership, Alt-3, is composed entirely of Tier 1 581 districts. In Alt-8, Tier 1 districts have the lowest costs in general, followed by Tiers 2-3. This 582 disparity is exacerbated in the dry scenario, where Tiers 2-3 bear almost all of the negative 583 impacts. Friant-16 displays weaker clustering, with districts from all three tiers present across the 584 low- to mid-ranges of the cost spectrum. However, the highest costs are paid by districts in Tiers 585 2-3. Overall, these results are consistent with the district-level variability in Figures 3d and 5. It 586 is better, in general, to be Tier 1 than Tier 3, but Tier 3 districts can do very well in well-587 designed partnerships and Tier 1 districts can do very poorly in poorly-designed partnerships. 588

589 Lastly, an important factor in the success of Alt-3 is simply the size of the partnership. All else equal, smaller coalitions are more likely to remain viable because they have fewer 590 partners to satisfy. Smaller partnerships also deliver more water to each partner on average, 591 which is beneficial as districts attempt to maximize their capture of surface water supplies to 592 avoid fallowing under the Sustainable Groundwater Management Act (Hanak et al., 2020). On 593 the other hand, many districts are incapable of accepting very large quantities of water during 594 short-lived high-flow events, and thus may prefer a smaller ownership share within a larger 595 partnership. Additionally, larger and more diverse partnerships will be more capable of 596 harnessing subsidies to lower costs (Cypher & Grinnell, 2007; Hansen et al., 2020; Newsom et 597 598 al., 2020).



Figure 8. Comparison of three alternative infrastructure investment partnerships. Panels (a/c/e) 600 601 show the captured water gains for each district in the Friant-16, Alt-8, and Alt-3 partnerships, respectively, while Panels (b/d/f) show the cost of gains for each district in these partnerships. 602 Each panel is split into three layers showing performance on the wet, average, and dry 603 hydrologic scenarios. Each district is represented by a circle, with color and size representing the 604 district's tier and its ownership share in the project, respectively. Within each layer, districts are 605 arranged by ownership share from smallest (top) to largest (bottom). The vertical blue, yellow, 606 and red lines represent partnership-level averages under the wet, average, and dry scenarios. All 607 costs over \$1,000/ML (\$1,233/AF) are consolidated into "1000+". 608

609 4 Discussion

610 4.1 Policy implications

This work has a number of important policy implications for regions working to adapt to 611 water scarcity through collaborative infrastructure investments. First, these results caution 612 against the use of highly aggregated models and mean cost-benefit performance assessments that 613 fail to resolve specific multi-actor dynamics within complex water supply systems. Traditional 614 capital investment planning frameworks tacitly assume that all partners will benefit equally from 615 joint infrastructure investments, or that benefits will be distributed according to historical usage 616 patterns. Our results show that investment partnership outcomes are often highly heterogeneous 617 across participating water districts, highlighting the importance of disaggregation to ensure that 618 investments provide benefits not only to the "average" partner, but to all partners. In the case of 619

620 the planned Friant-Kern Canal rehabilitation project, results suggest that the business-as-usual

621 partnership (the Friant contractors) may not be the ideal set of partners, and a wider set of

622 regional water districts should be considered for participation. More broadly, this work

highlights that contractual details regarding the operation and ownership of shared infrastructure

- are crucial design elements on par with physical design parameters such as conveyance capacity.
- 625 Neglecting these details (a current standard practice in planning studies) can cause large errors in
- 626 predicted performance.

Our results also emphasize the interconnectedness of the individual components within 627 institutionally complex water supply systems. Multiple capital investments under consideration 628 should be evaluated concurrently based on their interactive and cumulative effects rather than in 629 isolation. Moreover, the bundling of multiple components into a joint portfolio of investments 630 can yield synergistic benefits. For example, pairing canal expansion with storage infrastructure 631 such as groundwater recharge facilities can improve the value of conveyance by increasing local 632 capacity to store surplus flows from the canal. Groundwater banking can also widen the set of 633 water districts willing to invest in conveyance by providing a mechanism for districts with poor 634 local recharge capacity to store their water elsewhere. More broadly, these synergistic effects 635 support California's vision of a flexible infrastructure network that encourages coordination and 636 637 cooperation across scales.

However, this interconnectedness also amplifies the challenge of accurately evaluating 638 capital projects within complex supply networks. For example, the local capacity for 639 groundwater recharge (both in-district recharge and out-of-district banking) has a large impact on 640 overall project performance, but information about groundwater recharge capacity across the 641 region is widely dispersed and, in some cases, non-existent. Moreover, these capacities are 642 evolving quickly as water districts adapt to new requirements under the Sustainable Groundwater 643 Management Act (Hanak et al., 2020). This makes it difficult to accurately represent 644 groundwater recharge within planning models and therefore to evaluate candidate capital 645 projects. More broadly, this highlights the challenge of modeling an increasingly complex and 646 interconnected system with data that is siloed and diffuse. The state is working to improve data 647 availability following the Open and Transparent Water Data Act of 2016 (California Department 648 of Water Resources, 2021), but more work is needed to inform planning efforts under the WRPI. 649 This is a ubiquitous challenge for water resources systems globally. 650

651 Lastly, our results provide a stark picture of the impacts of a drier climate on water infrastructure investments. Even under expressly optimistic assumptions (e.g., a conservative 652 viability threshold, a capital project with unusually high subsidies, and a limited range of 653 hydrologic uncertainty that does not explicitly include climate warming), the vast majority of 654 candidate partnerships are not viable under the dry scenario. Moreover, the downside risks are 655 often borne unequally, with a subset of partners bearing the bulk of ill effects in unfavorable 656 futures. This work thus provides a warning that poorly designed regional water infrastructure 657 investment partnerships may provide marginal supply resiliency benefits if the future turns out to 658 be drier than the past. Simultaneously, long-lived debt obligations from new investments 659 (combined with existing obligations from past investments) can pose serious financial risks for 660 water districts and their customers if benefits turn out to be lower than expected (Chapman & 661

Breeding, 2014; Jeffrey Hughes et al., 2014). Planning under the WRPI should consider not only
water supply resilience, but also financial resilience for the organizations tasked with providing
affordable water.

665 4.2 Future directions

Future work will extend this framework through more advanced solution-generation 666 techniques (e.g., multi-objective evolutionary algorithms (Majer et al., 2014; Nicklow et al., 667 2010; Reed, Hadka, Herman, Kasprzyk, & Kollat, 2013)) and a broader exploration of robustness 668 under climatic, economic, and regulatory uncertainties (Kasprzyk, Nataraj, Reed, & Lempert, 669 2013; Lempert, Groves, Popper, & Bankes, 2006; Moallemi, Zare, et al., 2020). Additionally, 670 future work should consider whether supply and financial portfolios in water-scarce regions can 671 be made more resilient using infrastructure real options and adaptive pathways (Fletcher, 672 Lickley, & Strzepek, 2019; Haasnoot, Kwakkel, Walker, & ter Maat, 2013; Herman, Quinn, 673 Steinschneider, Giuliani, & Fletcher, 2020), flexible partnership design (Gorelick et al., 2019), or 674 novel financial tools such as environmental impact bonds or index insurance contracts (Brand et 675 al., 2021; Maestro, Barnett, Coble, Garrido, & Bielza, 2016; Zeff & Characklis, 2013). Lastly, 676 this work has focused primarily on equality at the water district level (i.e., whether costs are 677 equally distributed across partners) as opposed to equity and justice (Jafino et al., 2021; Osman 678 & Faust, 2021) (i.e., whether different water districts and their customers have differing access to 679 water and differing ability to pay for infrastructure, and how these differences intersect with 680 economic and political power, racial injustice, and responsibility for historical groundwater 681 overdraft and subsidence) (Dobbin & Lubell, 2021; Fernandez-Bou et al., 2021; Pauloo et al., 682 2020). Explicit consideration of these issues in direct co-production with disadvantaged 683 communities (Lemos et al., 2018; Minkler, Vásquez, Tajik, & Petersen, 2008) will be an 684 685 important extension of this study in light of the WRPI's stated goal of alleviating the growing water affordability challenge. 686

687 **5.** Conclusions

Population growth, groundwater overdraft, and climate change represent an 688 unprecedented challenge to water security in California, the Western US, and other water-689 stressed regions around the world. The Water Resilience Portfolio Initiative provides a vision for 690 bolstering the state's water supplies through an interconnected, collaborative, and flexible *water* 691 system of systems (Newsom et al., 2020), and represents not only a roadmap for California, but 692 also a potential template for other regions looking to develop their own resilient water supply 693 portfolios. For such a vision to work, individual water providers within the broader system will 694 need to collaborate in financing and building new shared infrastructure. However, traditional 695 planning frameworks based on highly aggregated models and mean cost-benefit estimates are ill-696 equipped to evaluate multi-party investment partnerships due to the significant complexities and 697 uncertainties within the distributed supply network. In this paper, we demonstrate the need to 698 evaluate partnerships at the level of individual water providers, and the challenge of designing 699 partnerships that can provide acceptable water supply benefits to each partner relative to its share 700 of project debt. 701

702 We leverage a high-resolution water supply simulation model within a parallelized exploratory modeling framework in order to explore alternative partnership structures, capital 703 projects, and hydrologic scenarios at an unprecedented scale. Even under explicitly optimistic 704 assumptions regarding climate change and other uncertainties, we find that vast majority of 705 alternative partnership structures tend to deliver water supply benefits and financial risks that are 706 highly uneven, threatening the viability of these cooperative investments. Designing viable 707 partnerships is especially challenging under drier hydrologic conditions, when insufficient 708 surplus water is available to warrant the investment in additional conveyance and storage 709 infrastructure. Additionally, our results demonstrate the synergy between conveyance and 710 storage for capturing surplus water during peak flow events; partnerships may be more likely to 711 succeed if they can combine multiple pieces of infrastructure that interact positively. 712 Importantly, however, partnership design choices (i.e., which water providers are participating, 713 and how do they distribute the project's debt obligation?) may be even more important to the 714 success of a partnership than the future hydrology or the type of infrastructure. As a whole, this 715 research investigates several under-studied challenges in the evaluation and planning of new 716 717 infrastructure investment partnerships within complex water supply networks. Confronting these challenges will be crucial if California and other regions are to achieve their resilient water 718

719 portfolio goals.

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727 **Open research**

- All data and code for this project, including figure generation, will be made available on GitHub
- and Zenodo prior to publication.

730 **References**

- AghaKouchak, A., Cheng, L., Mazdiyasni, O., & Farahmand, A. (2014). Global warming and
 changes in risk of concurrent climate extremes: Insights from the 2014 California drought.
 Geophysical Research Letters, 41, 8847–8852. https://doi.org/10.1002/2014GL062308
- AghaKouchak, A., Mirchi, A., Madani, K., Di Baldassarre, G., Nazemi, A., Alborzi, A., ...
 Wanders, N. (2021). Anthropogenic Drought: Definition, Challenges, and Opportunities. *Reviews of Geophysics*, 59, e2019RG000683. https://doi.org/10.1029/2019RG000683
- Alam, S., Gebremichael, M., Li, R., Dozier, J., & Lettenmaier, D. P. (2020). Can Managed
 Aquifer Recharge Mitigate the Groundwater Overdraft in California's Central Valley?
- 739 *Water Resources Research*, *56*(8). https://doi.org/10.1029/2020WR027244
- 740 BenDor, T. K., & Scheffran, J. (2019). Agent-Based Modeling of Environmental Conflict and

- 741 *Cooperation*. Boca Raton, FL: Taylor & Francis.
- Berg, N., & Hall, A. (2017). Anthropogenic warming impacts on California snowpack during
 drought. *Geophysical Research Letters*, 44, 2511–2518.
 https://doi.org/10.1002/2016GL072104
- 745 Brand, M. W., Quesnel, K., Saksa, P., Ulibarri, N., Bomblies, A., Mandle, L., ... Gibbons, J. P.
- 746 (2021). Environmental Impact Bonds: a common framework and looking ahead.
- *Environmental Research: Infrastructure and Sustainability*, *1*(2), 023001.
- 748 https://doi.org/10.1088/2634-4505/ac0b2c
- 749 California Department of Water Resources. (2017). *CalSim 3.0 Draft Report*. Sacramento, CA.
- California Department of Water Resources. (2021). AB 1755: Open and Transparent Water Data
 Platform for California. Retrieved from https://water.ca.gov/ab1755
- California State Treasurer's Office. (2020). All Issuance CY2019 on 12-22-20. Retrieved from
 https://www.treasurer.ca.gov/cdiac/datafile/2019.xls
- 754 Chapman, T. A., & Breeding, J. M. (2014). U.S. Public Finance Waterworks, Sanitary Sewer,
- and Drainage Utility Systems: Methodology And Assumptions. Retrieved from
 https://www.spratings.com/documents/20184/908554/US_PF_Event_RFCRndTblsJan2015
 Article1/30d125eb-1066-4730-8ab1-f2cd6a6d6f9a
- 758 Congressional Research Service. (2007). San Joaquin River Restoration Settlement Act.
- Connell-Buck, C. R., Medellín-Azuara, J., Lund, J. R., & Madani, K. (2011). Adapting
 California's water system to warm vs. dry climates. *Climatic Change*, *109*(SUPPL. 1), 133–
 149. https://doi.org/10.1007/s10584-011-0302-7
- Cypher, T., & Grinnell, C. (2007). Governments Working Together. A Citizen's Guide to Joint
 Powers Agreements. Retrieved from
- 764 https://sgf.senate.ca.gov/sites/sgf.senate.ca.gov/files/GWTFinalversion2.pdf
- Dahlke, H. E., LaHue, G. T., Mautner, M. R. L., Murphy, N. P., Patterson, N. K., Waterhouse,
 H., ... Foglia, L. (2018). Managed Aquifer Recharge as a Tool to Enhance Sustainable
 Groundwater Management in California. In *Advances in Chemical Pollution*,
- *Environmental Management and Protection* (Vol. 3, pp. 215–275).
- 769 https://doi.org/10.1016/bs.apmp.2018.07.003
- De Souza, S., Medellín-Azuara, J., Lund, J. R., & Howitt, R. E. (2011). *Beneficiary Pays Analysis of Water Recycling Projects*. Retrieved from
- http://www.waterboards.ca.gov/water issues/programs/grants loans/water recycling/docs/e
- con tskfrce/beneficiarypays.pdf
- Dettinger, M. D., Ralph, F. M., Das, T., Neiman, P. J., & Cayan, D. R. (2011). Atmospheric
 Rivers, Floods and the Water Resources of California. *Water*, *3*(2), 445–478.
 https://doi.org/10.3390/w3020445
- Diffenbaugh, N. S., Swain, D. L., & Touma, D. (2015). Anthropogenic warming has increased
 drought risk in California. *Proceedings of the National Academy of Sciences*, *112*(13),
 3931–3936. https://doi.org/10.1073/pnas.1422385112
- 780 Dobbin, K. B., & Lubell, M. (2021). Collaborative Governance and Environmental Justice:

781 782	Disadvantaged Community Representation in California Sustainable Groundwater Management. <i>Policy Studies Journal</i> , 49(2), 562–590. https://doi.org/10.1111/psj.12375
783	Doss-Gollin, J., Farnham, D. J., Steinschneider, S., & Lall, U. (2019). Robust Adaptation to
784	Multiscale Climate Variability. <i>Earth's Future</i> , 7(7), 734–747.
785	https://doi.org/10.1029/2019EF001154
786	Draper, A. J., Jenkins, M. W., Kirby, K. W., Lund, J. R., & Howitt, R. E. (2003). Economic-
787	engineering optimization for California water management. <i>Journal of Water Resources</i>
788	<i>Planning and Management</i> , 129(3), 155–164. https://doi.org/10.1061/(ASCE)0733-
789	9496(2003)129:3(155)
790 791 792 793	 Draper, A. J., Munévar, A., Arora, S. K., Reyes, E., Parker, N. L., Chung, F. I., & Peterson, L. E. (2004). CalSim: Generalized model for reservoir system analysis. <i>Journal of Water Resources Planning and Management</i>, 130(6), 480–489. https://doi.org/10.1061/(ASCE)0733-9496(2004)130:6(480)
794 795 796 797	 Escriva-Bou, A., Mccann, H., Hanak, E., Lund, J., Gray, B., Blanco, E., Tweet, A. (2020). Water Accounting in Western US, Australia, and Spain: Comparative Analysis. <i>Journal of Water Resources Planning and Management</i>, 146(3), 04020004. https://doi.org/10.1061/(ASCE)WR.1943-5452.0001157
798	Escriva-Bou, A., Sencan, G., Hanak, E., & Wilkinson, R. (2020). <i>Water Partnerships between</i>
799	<i>Cities and Farms in Southern California and the San Joaquin Valley</i> . San Francisco, CA.
800	Retrieved from https://www.ppic.org/wp-content/uploads/water-partnerships-between-
801	cities-and-farms-in-southern-california-and-the-san-joaquin-valley.pdf
802	Famiglietti, J. S. (2014). The global groundwater crisis. <i>Nature Climate Change</i> , <i>4</i> (11), 945–948.
803	https://doi.org/10.1038/nclimate2425
804	Faunt, C. C., & Sneed, M. (2015). Water availability and subsidence in California's Central
805	Valley. San Francisco Estuary and Watershed Science, 13(3), 0–8.
806	https://doi.org/10.15447/sfews.2015v13iss3art4
807	Fernandez-Bou, A. S., Ortiz-Partida, J. P., Dobbin, K. B., Flores-Landeros, H., Bernacchi, L. A.,
808	& Medellín-Azuara, J. (2021). Underrepresented, understudied, underserved: Gaps and
809	opportunities for advancing justice in disadvantaged communities. <i>Environmental Science</i>
810	and Policy, 122(April), 92–100. https://doi.org/10.1016/j.envsci.2021.04.014
811	Fletcher, S., Lickley, M., & Strzepek, K. (2019). Learning about climate change uncertainty
812	enables flexible water infrastructure planning. <i>Nature Communications</i> , 10(1), 1–11.
813	https://doi.org/10.1038/s41467-019-09677-x
814	Friant Water Authority. (2019). Subsidence: A critical challenge to Friant-Kern Canal water
815	deliveries. Retrieved from
816	https://static1.squarespace.com/static/58c2eccc15d5db46200ea426/t/5df2e69ea705f61846a2
817	58bd/1576199845717/FWA_Subsidence_Challenge_V3_web.pdf
818	Gershunov, A., Shulgina, T., Ralph, F. M., Lavers, D. A., & Rutz, J. J. (2017). Assessing the
819	climate-scale variability of atmospheric rivers affecting western North America.
820	<i>Geophysical Research Letters</i> , 44(15), 7900–7908. https://doi.org/10.1002/2017GL074175
821	Giglio, R. J., & Wrightington, R. (1972). Methods for apportioning costs among participants in

- regional systems. *Water Resources Research*, 8(5), 1133–1144.
- Gold, D. F., Reed, P. M., Trindade, B. C., & Characklis, G. W. (2019). Identifying Actionable
 Compromises: Navigating Multi-City Robustness Conflicts to Discover Cooperative Safe
 Operating Spaces for Regional Water Supply Portfolios. *Water Resources Research*, 55, 1–
- 826 27. https://doi.org/10.1029/2019WR025462
- Gonzalez, P., Garfin, G. M., Breshears, D. D., Brooks, K. M., Brown, H. E., Elias, E. H., ...
- Udall, B. H. (2018). Southwest. In D. R. Reidmiller, C. W. Avery, D. R. Easterling, K. E.
- 829 Kunkel, K. L. M. Lewis, T. K. Maycock, & B. C. Stewart (Eds.), Impacts, Risks, and
- 830 *Adaptation in the United States: Fourth National Climate Assessment, Volume II* (pp. 1101–
- 831 1184). Washington, DC, USA: U.S. Global Change Research Program.
- 832 https://doi.org/10.7930/NCA4.2018.CH25
- Gorelick, D. E., Zeff, H. B., Hughes, J., Eskaf, S., & Characklis, G. W. (2019). Exploring
 Treatment and Capacity-Sharing Agreements Between Water Utilities. *Journal American Water Works Association*, *111*(9), 26–40. https://doi.org/10.1002/awwa.1359
- Big Groves, D. G., Bloom, E., Lempert, R. J., Fischbach, J. R., Nevills, J., & Goshi, B. (2015).
- Barring and Management, 141(7), 05014008. https://doi.org/10.1061/(ASCE)WR.19435452.0000471
- Haasnoot, M., Kwakkel, J. H., Walker, W. E., & ter Maat, J. (2013). Dynamic adaptive policy
 pathways: A method for crafting robust decisions for a deeply uncertain world. *Global Environmental Change*. https://doi.org/10.1016/j.gloenvcha.2012.12.006
- Haimes, Y. Y. (2018). *Modeling and managing interdependent complex systems of systems*. John
 Wiley & Sons, Inc.
- Hanak, E., Chappelle, C., Gray, B., McCann, H., Ajami, N., Baerenklau, K., ... Mitchell, D.
 (2018). *California's Water: Paying for Water*.
- Hanak, E., Escriva-Bou, A., Gray, B., Green, S., Harter, T., Jezdimirovic, J., ... Seavy, N.
 (2019). *Water and the Future of the San Joaquin Valley*. San Francisco, CA. Retrieved from https://www.ppic.org/wp-content/uploads/water-and-the-future-of-the-san-joaquin-valleyfebruary-2019.pdf
- Hanak, E., Jezdimirovic, J., Escriva-Bou, A., & Ayres, A. (2020). A Review of Groundwater
 Sustainability Plans in the San Joaquin Valley (Public comments submitted to the
- 853 *California Department of Water Resources).* San Francisco, CA. Retrieved from
- https://www.ppic.org/wp-content/uploads/ppic-review-of-groundwater-sustainability-plans in-the-san-joaquin-valley.pdf
- Hanak, E., Jezdimirovic, J., Green, S., Escriva-Bou, A., Bostic, D., & Mccann, H. (2018).
 Replenishing Groundwater in the San Joaquin Valley. Sacramento, CA. Retrieved from https://www.ppic.org/wp-content/uploads/r-0417ehr.pdf
- Hansen, K., Mullin, M., & Riggs, E. K. (2020). Collaboration Risk and the Choice to
- Consolidate Local Government Services. *Perspectives on Public Management and Governance*, 3(3), 223–238. https://doi.org/10.1093/ppmgov/gvz017
- 862 Herman, J. D., Quinn, J. D., Steinschneider, S., Giuliani, M., & Fletcher, S. (2020). Climate

- adaptation as a control problem: Review and perspectives on dynamic water resources
- planning under uncertainty. *Water Resources Research*, *56*, e24389.
- 865 https://doi.org/10.1029/2019wr025502
- Herman, J. D., Reed, P. M., Zeff, H. B., & Characklis, G. W. (2015). How should robustness be
 defined for water systems planning under change? *Journal of Water Resources Planning and Management*, *141*(10), 04015012. https://doi.org/10.1061/(ASCE)WR.19435452.0000509
- Herman, J. D., Zeff, H. B., Reed, P. M., & Characklis, G. W. (2014). Beyond optimality:
 Multistakeholder robustness tradeoffs for regional water portfolio planning under deep
 uncertainty. *Water Resources Research*, 50(10), 7692–7713.
 https://doi.org/10.1002/2014WR015338
- Hughes, Jeff, & Fox, R. (2019). Strengthening Utilities through Consolidation: The Financial
 Impact. Retrieved from
- http://uswateralliance.org/sites/uswateralliance.org/files/publications/Final_Utility
- 877 Consolidation Financial Impact Report_022019.pdf
- Hughes, Jeffrey, Tiger, M., Eskaf, S., Berahzer, S. I., Royster, S., Boyle, C., & Batten, D. (2014).
 Defining a Resilient Business Model for Water Utilities. Retrieved from https://efc.sog.unc.edu/sites/default/files/4366 Exec Summary 0.pdf
- Islam, N., Arora, S., Chung, F., Reyes, E., Field, R., Munévar, A., ... Chen, Z. Q. R. (2011).
 CalLite: California Central Valley Water Management Screening Model. *Journal of Water Resources Planning and Management*, *137*(1), 123–133.
 https://doi.org/10.1061/(ASCE)WR.1943-5452.0000089
- Jafino, B. A., Kwakkel, J. H., & Taebi, B. (2021). Enabling assessment of distributive justice
 through models for climate change planning: A review of recent advances and a research
 agenda. *Wiley Interdisciplinary Reviews: Climate Change*, e721.
 https://doi.org/10.1002/wcc.721
- Jezdimirovic, J., Hanak, E., & Escriva-Bou, A. (2020). What's the Plan to End Groundwater
 Overdraft in the San Joaquin Valley? Retrieved November 10, 2020, from
 https://www.ppic.org/blog/whats-the-plan-to-end-groundwater-overdraft-in-the-san joaquin-valley/
- Jiménez Cisneros, B. E., Oki, T., Arnell, N. W., Benito, G., Cogley, J. G., Döll, P., ... Nishijima,
 A. (2015). Freshwater resources. In *Climate Change 2014 Impacts, Adaptation and*
- 895 *Vulnerability: Part A: Global and Sectoral Aspects. Contribution of Working Group II to*
- the Fifth Assessment Report of the Intergovernmental Panel On Climate Change (pp. 229–
- 270). Cabridge, United Kingdom and New York, NY, USA: Cambridge University Press.
 https://doi.org/10.1017/CBO9781107415379.008
- 899 Kasprzyk, J. R., Nataraj, S., Reed, P. M., & Lempert, R. J. (2013). Many objective robust
- decision making for complex environmental systems undergoing change. *Environmental Modelling and Software*, 42, 55–71. https://doi.org/10.1016/j.envsoft.2012.12.007
- Kern Water Bank Authority. (2021). Frequently Asked Questions. Retrieved from https://www.kwb.org/faqs/
- 904 Kocis, T. N., & Dahlke, H. E. (2017). Availability of high-magnitude streamflow for

- groundwater banking in the Central Valley, California. *Environmental Research Letters*, *12*(8), 084009. https://doi.org/10.1088/1748-9326/aa7b1b
- Lall, U., Johnson, T., Colohan, P., Aghakouchak, A., Brown, C., Mccabe, G., ... Arumugam, S.
 (2018). Water. In D. R. Reidmiller, C. W. Avery, D. R. Easterling, K. E. Kunkel, K. L. M.
 Lewis, T. K. Maycock, & B. C. Stewart (Eds.), *Impacts, Risks, and Adaptation in the*
- 910 United States: Fourth National Climate Assessment (Vol. II, pp. 145–173). Washington,
- 911 DC, USA: U.S. Global Change Research Program.
- 912 https://doi.org/10.7930/NCA4.2018.CH3
- Lehner, F., Deser, C., Maher, N., Marotzke, J., Fischer, E., Brunner, L., ... Hawkins, E. (2020).
 Partitioning climate projection uncertainty with multiple Large Ensembles and CMIP5/6. *Earth System Dynamics*, 11, 491–508. https://doi.org/10.5194/esd-2019-93
- Lemos, M. C., Arnott, J. C., Ardoin, N. M., Baja, K., Bednarek, A. T., Dewulf, A., ... Wyborn,
 C. (2018, December 1). To co-produce or not to co-produce. *Nature Sustainability*. Nature
 Publishing Group. https://doi.org/10.1038/s41893-018-0191-0
- Lempert, R. J., & Groves, D. G. (2010). Identifying and evaluating robust adaptive policy
 responses to climate change for water management agencies in the American west.
 Technological Forecasting and Social Change, 77(6), 960–974.
 https://doi.org/10.1016/j.techfore.2010.04.007
- Lempert, R. J., Groves, D. G., Popper, S. W., & Bankes, S. C. (2006). A general, analytic
 method for generating robust strategies and narrative scenarios. *Management Science*,
 52(4), 514–528. https://doi.org/10.1287/mnsc.1050.0472
- Leurig, S. (2010). *The Ripple Effect: Water risk in the municipal bond market*. Boston, MA.
 Retrieved from https://www.ceres.org/resources/reports/ripple-effect-water-risk-municipal-bond-market
- Levy, Z. F., Jurgens, B. C., Burow, K. R., Voss, S. A., Faulkner, K. E., Arroyo-Lopez, J. A., &
 Fram, M. S. (2021). Critical aquifer overdraft accelerates degradation of groundwater
 quality in California's Central Valley during drought. *Geophysical Research Letters*.
 https://doi.org/10.1029/2021gl094398
- Lubell, M., Blomquist, W., & Beutler, L. (2020). Sustainable Groundwater Management in
 California: A Grand Experiment in Environmental Governance. *Society & Natural Resources*, *33*(12), 1447–1467. https://doi.org/10.1080/08941920.2020.1833617
- Madani, K. (2010). Game theory and water resources. *Journal of Hydrology*.
 https://doi.org/10.1016/j.jhydrol.2009.11.045
- Maestro, T., Barnett, B. J., Coble, K. H., Garrido, A., & Bielza, M. (2016). Drought index
 insurance for the Central Valley Project in California. *Applied Economic Perspectives and Policy*, 38(3), 521–545. https://doi.org/10.1093/aepp/ppw013
- Maier, H. R., Kapelan, Z., Kasprzyk, J., Kollat, J., Matott, L. S., Cunha, M. C., ... Reed, P. M.
 (2014). Evolutionary algorithms and other metaheuristics in water resources: Current status,
 research challenges and future directions. *Environmental Modelling and Software*, 62, 271–
 299. https://doi.org/10.1016/j.envsoft.2014.09.013
- 945 Malek, K., Reed, P., Zeff, H., Hamilton, A., Wrzesien, M., Holtzman, N., ... Pavelsky, T.

- 946 (2022). Bias Correction of Hydrologic Projections Strongly Impacts Inferred Climate
- 947 Vulnerabilities in Institutionally Complex Water Systems. *Journal of Water Resources*948 *Planning and Management*, *148*(1), 1–14. https://doi.org/10.1061/(asce)wr.1943949 5452.0001493
- Manna, F., Walton, K. M., Cherry, J. A., & Parker, B. L. (2019). Five-century record of climate
 and groundwater recharge variability in southern California. *Scientific Reports*, 9(1), 1–8.
 https://doi.org/10.1038/s41598-019-54560-w
- Minkler, M., Vásquez, V. B., Tajik, M., & Petersen, D. (2008). Promoting environmental justice
 through community-based participatory research: The role of community and partnership
 capacity. *Health Education and Behavior*, 35(1), 119–137.
 https://doi.org/10.1177/1090198106287692
- Moallemi, E. A., Kwakkel, J., de Haan, F. J., & Bryan, B. A. (2020). Exploratory modeling for
 analyzing coupled human-natural systems under uncertainty. *Global Environmental Change*, 65, 102186. https://doi.org/10.1016/j.gloenvcha.2020.102186
- Moallemi, E. A., Zare, F., Reed, P. M., Elsawah, S., Ryan, M. J., & Bryan, B. A. (2020).
 Structuring and evaluating decision support processes to enhance the robustness of complex human–natural systems. *Environmental Modelling and Software*, *123*, 104551.
 https://doi.org/10.1016/j.envsoft.2019.104551
- Newsom, G., Ross, K., Blumenfeld, J., Bosler, K. M., Bonham, C. H., Nemeth, K., ... Tatayon,
 S. (2020). *California Water Resilience Portfolio Governor's Executive Order N-10-19*.
 Sacramento, CA. Retrieved from https://waterresilience.ca.gov/wpcontent/uploads/2020/01/California-Water-Resilience-Portfolio-2019-Final2.pdf
- Nicklow, J., Reed, P., Savic, D., Dessalegne, T., Harrell, L., Chan-Hilton, A., ... Zechman, E.
 (2010). State of the Art for Genetic Algorithms and Beyond in Water Resources Planning and Management. *Journal of Water Resources Planning and Management*, *136*(4), 412– 432. https://doi.org/10.1061/ASCÊWR.1943-5452.0000053
- O'Geen, A. T., Saal, M. B. B., Dahlke, H., Doll, D., Elkins, R., Fulton, A., ... Walkinshaw, M.
 (2015). Soil suitability index identifies potential areas for groundwater banking on
 agricultural lands. *California Agriculture*, 69(2), 75–84.
 https://doi.org/10.3733/ca.v060p02p75
- 975 https://doi.org/10.3733/ca.v069n02p75
- Office of Governor Gavin Newsom. (2021). Governor Newsom Announces \$5.1 Billion Package
 for Water Infrastructure and Drought Response as Part of \$100 Billion California
- Comeback Plan. Retrieved from https://www.gov.ca.gov/2021/05/10/governor-newsom announces-5-1-billion-package-for-water-infrastructure-and-drought-response-as-part-of-
- 980 100-billion-california-comeback-plan/
- Osman, K. K., & Faust, K. M. (2021). Toward Operationalizing Equity in Water Infrastructure
 Services: Developing a Definition of Water Equity. ACS ES&T Water, 1(8), 1849–1858.
 https://doi.org/10.1021/acsestwater.1c00125
- Pauloo, R. A., Escriva-Bou, A., Dahlke, H., Fencl, A., Guillon, H., & Fogg, G. E. (2020).
 Domestic well vulnerability to drought duration and unsustainable groundwater
 management in California's Central Valley. *Environmental Research Letters*, 15, 044010.
- 987 https://doi.org/10.1088/1748-9326/ab6f10

- Read, L., Madani, K., & Inanloo, B. (2014). Optimality versus stability in water resource
 allocation. *Journal of Environmental Management*, *133*, 343–354.
 https://doi.org/10.1016/j.jenvman.2013.11.045
- Reed, P. M., Hadka, D., Herman, J. D., Kasprzyk, J. R., & Kollat, J. B. (2013). Evolutionary
 multiobjective optimization in water resources: The past, present, and future. *Advances in Water Resources*, 51, 438–456. https://doi.org/10.1016/j.advwatres.2012.01.005
- Riggs, E., & Hughes, J. (2019). Crafting Interlocal Water and Wastewater Agreements. Chapel
 Hill, NC. Retrieved from https://efc.sog.unc.edu/wp-
- 996 content/uploads/sites/1172/2021/06/Crafting20Interlocal20Agreements_Final_01.pdf
- San Joaquin River Restoration Program. (2011). Friant-Kern Canal Capacity Restoration
 Feasibility Study. Retrieved from https://www.restoresjr.net/?wpfb_dl=1916
- Scanlon, B. R., Reedy, R. C., Faunt, C. C., Pool, D., & Uhlman, K. (2016). Enhancing drought
 resilience with conjunctive use and managed aquifer recharge in California and Arizona. *Environmental Research Letters*, 11(3), 035013. https://doi.org/10.1088/17489326/11/3/035013
- Schwarz, A., Ray, P., Arnold, W., Brown, C., Wi, S., Vasquez, J., ... Andrew, J. (2019).
 Decision Scaling Evaluation of Climate Risks to the State Water Project Final Report.
 Retrieved from https://water.ca.gov/-/media/DWR-Website/Web-Pages/Programs/All Programs/Climate-Change-Program/Climate-Action-Plan/Files/CAP-III-Decision-Scaling Vulnerability-Assessment.pdf
- Schwarz, A., Ray, P., Wi, S., Brown, C., He, M., & Correa, M. (2018). *Climate Change Risk Faced by the California Central Valley Water Resource System*. Retrieved from
 https://www.energy.ca.gov/sites/default/files/2019-12/Water_CCCA4-EXT-2018 001_ada.pdf
- Selmon, M., Schwarz, A., Coombe, P., Arnold, W., Chappell, E., Correa, M., ... Andrew, J.
 (2019). *California Department of Water Resources Climate Action Plan, Phase 3: Climate Change Vulnerability Assessment.*
- Su, Y., Kern, J. D., Denaro, S., Hill, J., Reed, P., Sun, Y., ... Characklis, G. W. (2020). An open source model for quantifying risks in bulk electric power systems from spatially and temporally correlated hydrometeorological processes. *Environmental Modelling & Software*, *126*(January), 104667. https://doi.org/10.1016/j.envsoft.2020.104667
- Su, Y., Kern, J. D., Reed, P. M., & Characklis, G. W. (2020). Compound hydrometeorological
 extremes across multiple timescales drive volatility in California electricity market prices
 and emissions. *Applied Energy*, 276(July), 115541.
 https://doi.org/10.1016/j.apenergy.2020.115541
- Sunding, D. L. (2015). *CalWater Fix Economic Analysis Draft*. Retrieved from https://cawaterlibrary.net/document/cal-water-fix-economic-analysisdraft/? sft keywords=delta-conveyance
- Tariq, A., Lempert, R. J., Riverson, J., Schwartz, M., & Berg, N. (2017). A climate stress test of
 Los Angeles' water quality plans. *Climatic Change*, *144*(4), 625–639.
 https://doi.org/10.1007/s10584-017-2062-5

- The White House. (2021). FACT SHEET: The American Jobs Plan. Retrieved from
 https://www.whitehouse.gov/briefing-room/statements-releases/2021/03/31/fact-sheet-the american-jobs-plan/
- U.S. Bureau of Reclamation. (2017). North-of-the-Delta Offstream Storage Investigation: Draft
 Feasibility Study. Retrieved from https://sitesproject.org/resources/feasibility-report/
- 1034 U.S. Bureau of Reclamation. (2020). Friant-Kern Canal Middle Reach Capacity Correction
 1035 Project. Retrieved from https://www.usbr.gov/mp/docs/fkc-feasibility-report.pdf
- 1036 University of California Agriculture and Natural Resources. (2021). Kern County: Irrigation
 1037 Management & Agronomy. Retrieved from
 1038 https://cekern.ucanr.edu/Irrigation Management/
- Wendt, D. E., van Loon, A. F., Scanlon, B. R., & Hannah, D. M. (2021). Managed aquifer
 recharge as a drought mitigation strategy in heavily-stressed aquifers. *Environmental Research Letters*, 16(1). https://doi.org/10.1088/1748-9326/abcfe1
- Whisnand, C. (2021). Friant-Kern Canal repair process continues with repayment contract.
 Retrieved from https://www.recorderonline.com/news/friant-kern-canal-repair-processcontinues-with-repayment-contract/article a3946542-f53d-11eb-a9f2-b77a81e9cef4.html
- Zeff, H. B., & Characklis, G. W. (2013). Managing water utility financial risks through third party index insurance contracts. *Water Resources Research*, 49(8), 4939–4951.
 https://doi.org/10.1002/wrcr.20364
- 1048 Zeff, H. B., Hamilton, A. L., Malek, K., Herman, J. D., Cohen, J. S., Medellin-Azuara, J., ...
- 1049 Characklis, G. W. (2021). California's food-energy-water system: An open source
- simulation model of adaptive surface and groundwater management in the Central Valley.
- 1051 *Environmental Modelling and Software*, 141, 105052.
- 1052 https://doi.org/10.1016/j.envsoft.2021.105052
- 1053



Earth's Future

Supporting Information for

Resilient California water portfolios require infrastructure investment partnerships that are viable for all partners

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Contents of this file

Figures S1 to S5

Introduction

This Supporting Information contains five additional figures, S1-S5. Figure S1 and S2 support Sections 2.4 and 3.2 of the main text, respectively, while Figures S3-S5 support Section 3.4 of the main text.



Figure S1. Comparison of synthetic and historical full natural flow (FNF) for the Sacramento River Basin. Panels (a-d) show the total full natural flows for the four major reservoirs of the Sacramento River Basin (Shasta, Oroville, New Bullards Bar, and Folsom), under the historical record and the wet, average, and dry synthetic scenarios. Panels (e-h) show the full natural flow duration curves for each time series over 1-, 2-, 4-, and 8-year periods.



Figure S2. Soil suitability for groundwater recharge in the southern Central Valley. Five major reservoirs store runoff from the Sierra Nevada mountains in the east and release it into the region's major rivers, where much of the flow is withdrawn by water districts. Millerton Lake diverts San Joaquin River water into the Madera Canal and Friant-Kern Canal as part of the Central Valley Project (CVP). The CVP and State Water Project also pump water from the Sacramento-San Joaquin Delta to San Luis Reservoir, where it is routed via a series of pumps and canals to water districts in the valley and urban districts along the coast. Water districts in the region are designated with thin black outlines, and their coloring represents the suitability of soils for groundwater recharge after accounting for deep tillage, according to the Soil Agricultural Groundwater Banking Index (O'Geen et al., 2015).



Figure S3. Viability of sampled infrastructure investment partnerships with additional constraint on partnership captured water gains. Each simulated partnership is plotted according to the cost of gains for the worst-off partner vs. the captured water gains for the partnership as a whole. The project type and hydrologic scenario used for each simulation are represented by marker type and color, respectively. Viable partnerships (those with costs <\$200/ML (\$247/AF) for the worst-off partner and captured water gains >55 GL/year (45 kAF/year) for the partnership)) are represented with black outlines and higher opacity. All costs over \$1,000/ML (\$1,233/AF) are consolidated into "1000+". Inset shows the viability of candidate partnership structures under each combination of capital project and hydrologic scenario, represented by color as well as the percentage printed in each square.



Figure S4. Viability of sampled infrastructure investment partnerships with additional constraint on captured water gains for non-partners. Each simulated partnership is plotted according to the cost of gains for the worst-off partner vs. the captured water gains for the partnership as a whole. The project type and hydrologic scenario used for each simulation are represented by marker type and color, respectively. Viable partnerships (those with costs <\$200/ML (\$247/AF) for the worst-off partner and captured water gains >0 GL/year across all non-partners in the region)) are represented with black outlines and higher opacity. All costs over \$1,000/ML (\$1,233/AF) are consolidated into "1000+". Inset shows the viability of candidate partnership structures under each combination of capital project and hydrologic scenario, represented by color as well as the percentage printed in each square.



Figure S5. Performance of partnerships that meet multiple criteria. Each vertical axis represents a different performance metric. Each curve represents a simulated partnership, and its intersection with each vertical axis corresponds to its performance on that metric. Blue and orange curves correspond to the wet and average hydrologic scenarios, respectively, and color shading represents the number of partners. The Alt-3 and Alt-8 partnerships are shown in bold with dotted and dashed black emphases, respectively.

References

O'Geen, A. T., Saal, M. B. B., Dahlke, H., Doll, D., Elkins, R., Fulton, A., … Walkinshaw, M. (2015). Soil suitability index identifies potential areas for groundwater banking on agricultural lands. *California Agriculture*, *6*9(2), 75–84. https://doi.org/10.3733/ca.vo69n02p75