Enhanced NAPL Removal and Mixing with Engineered Chaotic Advection

yufei wang¹, Yufei Wang¹, Daniel Fernàndez-Garcia¹, Guillem Sole-Mari¹, and Paula Rodríguez-Escales¹

¹Universitat Politècnica de Catalunya

November 21, 2022

Abstract

Aquifer remediation with in situ soil washing techniques and enhanced oil removal typically involve the injection of liquid solutions into the geological formation to displace and mobilize non-aqueous phase liquids (NAPLs). The efficiency of these systems is oftentimes low because the displacing fluid bypasses large quantities of NAPL due to the inherent complexity of a heterogeneous natural system. Here, chaotic advection generated by a rotating periodic injection pulse is proposed as a method to enhance NAPL removal and mixing. To evaluate the method, we perform two-phase flow simulations in multiple realizations of random permeability fields with different correlation structures and connectivity between injection and extraction wells embedded in a five-spot pattern. Results show that chaotic advection can significantly improve removal efficiency and mixing depending on several controlling factors. Chaotic advection effects are more significant under unfavorable conditions, i.e., when injection and extraction wells are well-connected through preferential channels, permeabilities are highly heterogeneous, and/or the mobility ratio between the wetting and the non-wetting fluid is larger than one. Removal efficiency reaches its maximum value when the Kubo number is close to one, i.e., when the saturation front travels one range of the permeability field in an injection pulse. These effects can develop in just a few cycles. However, removal efficiency should undergo first an early stage with detrimental effects in order to maximize removal in the long term. Chaotic advection not only enhances NAPL removal and mixing, but also reduces the uncertainty, making the system more reliable and less dependent on heterogeneity.

Enhanced NAPL Removal and Mixing with Engineered Chaotic Advection

1

2

Yufei Wang^{1,2} Daniel Fernàndez-Garcia^{1,2} Guillem Sole-Mari^{3,1,2} Paula Rodríguez-Escales^{1,2} 3

4	¹ Dept. of Civil and Environmental Engineering. Universitat Politècnica de Catalunya, Jordi Girona 1-3, 08034 Barcelona,
5	Spain
6	² Associated Unit: Hydrogeology Group (UPC-CSIC)
7	³ Earth and Environmental Sciences Area, Geochemistry Department, Lawrence Berkeley National Laboratory, 1 Cyclotron
8	Rd, Berkeley, CA 94720, USA

9	Key Points:
10	• Chaotic advection enhances non-aqueous phase liquid removal and mixing in hetero-
11	geneous porous media while reducing uncertainty
12	• Removal efficiency increases to a maximum value when the injection pulse satisfies
13	that the Kubo number is close to one
14	• The relative benefit of chaotic advection is substantially larger in conditions of unfa-
15	vorable aquifer heterogeneity and connectivity

Corresponding author: Yufei Wang, yufei.wang@upc.edu

16 Abstract

Aquifer remediation with in situ soil washing techniques and enhanced oil removal 17 typically involve the injection of liquid solutions into the geological formation to displace 18 and mobilize non-aqueous phase liquids (NAPLs). The efficiency of these systems is often-19 times low because the displacing fluid bypasses large quantities of NAPL due to the inherent 20 complexity of a heterogeneous natural system. Here, chaotic advection generated by a rotat-21 ing periodic injection pulse is proposed as a method to enhance NAPL removal and mixing. 22 To evaluate the method, we perform two-phase flow simulations in multiple realizations of 23 random permeability fields with different correlation structures and connectivity between injection and extraction wells embedded in a five-spot pattern. Results show that chaotic ad-25 vection can significantly improve removal efficiency and mixing depending on several con-26 trolling factors. Chaotic advection effects are more significant under unfavorable conditions, 27 i.e., when injection and extraction wells are well-connected through preferential channels, 28 permeabilities are highly heterogeneous, and/or the mobility ratio between the wetting and 29 the non-wetting fluid is larger than one. Removal efficiency reaches its maximum value when 30 the Kubo number is close to one, i.e., when the saturation front travels one range of the per-31 meability field in an injection pulse. These effects can develop in just a few cycles. However, 32 removal efficiency should undergo first an early stage with detrimental effects in order to 33 maximize removal in the long term. Chaotic advection not only enhances NAPL removal and 34 mixing, but also reduces the uncertainty, making the system more reliable and less dependent 35 on heterogeneity. 36

37 **1 Introduction**

Non-aqueous phase liquid (NAPL) removal from complex geological formations is 38 of great interest in aquifer remediation with in situ soil washing techniques [Huling and 39 Weaver, 1991; NRC, 2005]. These techniques typically involve the injection of liquid so-40 lutions into the geological formation to displace and mobilize the target NAPL contami-41 nant source (e.g., hydrocarbons, chlorinated solvents, mineral oil and other products from 42 chemical industry) towards extraction wells [Rao et al., 1997; Martel et al., 2004; Smalley 43 et al., 2009; Davies et al., 2014; Jackson, 2014; Jin et al., 2017; Welkenhuysen et al., 2017]. 44 The injected fluid can be water or liquid solutions with cosolvents (e.g., hydroxypropyl- β -45 cyclodextrin), surfactants (e.g., sodium dihexyl sulfosuccinate) or polymers (e.g., partially 46 hydrolyzed polyacrylamid), which are meant to improve sweeping and flushing by either re-47 ducing capillary trapping, increasing dissolution or reducing water mobility [Falta et al., 48 1999; Mccray and Brusseau, 1999; Dugan et al., 2003; Yousefvand and Jafari, 2015; Ja-49 vanbakht et al., 2017]. An early removal of NAPL can substantially eliminate the main con-50 taminant source, improving the overall cleanup efficiency at later stages of the remediation 51 process [Huling and Weaver, 1991; Soga et al., 2004]. 52

NAPL removal is more efficient in relatively homogeneous porous media and low mo-53 bility ratios [Fayers and Hewett, 1992; Soga et al., 2004; Smalley et al., 2009; Stroo et al., 54 2012]. However, aquifer heterogeneity (the spatial distribution of permeability) often ex-55 hibits well-organized high permeability geological structures that concentrate the flow in the 56 form of preferential channels [de Marsily, 1985; Western et al., 2001; Zheng and Gorelick, 57 2003; Knudby and Carrera, 2005; Le Borgne and Gouze, 2008; Fernàndez-Garcia et al., 58 2010; Bianchi et al., 2011a; Renard and Allard, 2013a; Essaid et al., 2015; Nicolaides et al., 59 2015]. In a multiphase flow problem, these channels will cause the displacing fluid to by-60 pass large quantities of the NAPL in place, resulting in a significant reduction of the sweep-61 ing efficiencies and the mixing between the wetting and the non-wetting phase [Pruess and 62 Tsang, 1990; Wan et al., 1996; Glass et al., 1998; Amundsen et al., 1999; Bertels et al., 2001; 63 Rangel-German and Kovscek, 2006; Arshadi et al., 2017, 2018; Kim et al., 2019]. The lat-64 ter is crucial for chemical flooding with surfactants and cosolvents during the remediation of 65 an aquifer contaminated with NAPL, as dissolution directly depends on the contact between

⁶⁷ liquid phases. In the petroleum industry, channeling will cause huge losses in recovered oil ⁶⁸ and monetary income [*Craig et al.*, 1957; *Craig*, 1971; *Fayers and Hewett*, 1992; *Paez Yanez* ⁶⁹ *et al.*, 2007]. The injection of fluids in a multiphase system is also important for studying the ⁷⁰ sequestration of anthropogenic CO_2 in deep saline aquifers [e.g., *Bachu*, 2000; *Bolster et al.*, ⁷¹ 2009; *Vilarrasa et al.*, 2010; *Saaltink et al.*, 2013], which constitutes an interesting alterna-⁷² tive for reducing greenhouse gas emissions to the atmosphere.

Engineered chaotic advection has been demonstrated to be an efficient technique for 73 enhancing in situ groundwater remediation technologies [e.g., Zhang et al., 2009; Mays and 74 Neupauer, 2012; Trefry et al., 2012a; Lester et al., 2013; Neupauer et al., 2014; Rodríguez-75 Escales et al., 2017; Di Dato et al., 2018]. This technique generates chaotic advection by 76 means of time-dependent water injection and extraction systems creating erratic transport 77 paths that enhance mixing by folding and stretching the solute plume. Results in this area 78 have shown that applying chaotic advection increases considerably the contact area between 79 the contaminants and the injected treatment solutions during in situ remediation of contam-80 inated groundwater, promoting the degradation of toxic compounds, including emerging 81 organic contaminants [Ottino, 1990; Ottino et al., 1994; Bagtzoglou and Oates, 2007; Luo et al., 2008; Rodríguez-Escales et al., 2017; Libera et al., 2017]. There are different ways to 83 generate chaotic advection in the subsoil. One of them is by using a rotated potential mixing 84 flow (RPM). This form of chaotic advection can be generated by assemblies of several dipole 85 injection/extraction wells operating in a plane with the same flow rate [Metcalfe et al., 2006; 86 Lester et al., 2009, 2010; Metcalfe et al., 2010; Trefry et al., 2012b]. A necessary condition 87 to generate chaos is the transient crossing of streamlines [Lester et al., 2009], which is ful-88 filled with this technique. The simplest RPM-generating sequence consists in activating a 89 dipole injection/extraction well for a certain time, then rotating the active dipole around the 90 origin by an angle, and repeating periodically. However, the RPM flow is not the only way to 91 generate chaotic advection in groundwater polluted sites. Many other kinds of well networks 92 and stirring protocols have been proposed in literature with the objective of enhancing mix-93 ing of solutions in groundwater by inducing chaotic advection [e.g., Bagtzoglou and Oates, 94 2007; Zhang et al., 2009; Piscopo et al., 2013; Neupauer et al., 2014]. 95

Although some authors Falta et al. [1999] have already observed that altering the flow 96 direction can improve NAPL remediation, there are no works aimed at evaluating the im-97 provement of chaotic advection during in situ NAPL remediation. The study of this problem 98 requires the simulation of a multiphase flow system, which at least should describe the move-99 ment of the wetting (injected fluid) and the non-wetting fluid (NAPL) through the porous 100 medium [Abriola and Pinder, 1985; Sleep and Sykes, 1989; Celia et al., 2015]. Under some 101 simplifying conditions, the governing equation of saturation of immiscible fluids resembles 102 the advection-dispersion equation (ADE) in porous media [Sleep and Sykes, 1993a,b; Bol-103 ster et al., 2009]. However, the advective and the dispersive terms in the saturation equation 104 are nonlinear functions of saturation, making impossible a direct extrapolation of the results 105 obtained in solute transport. 106

Motivated by this similarity and the success of an engineered sequence of injections 107 and extractions in contaminant transport, this paper explores the use of chaotic advection in 108 two-phase flow systems. More specifically, we evaluate the effect of chaotic advection in the 109 removal of NAPL using a five-spot injection-extraction well pattern in random heterogeneous 110 porous media with different correlation structures and connectivities. To achieve this goal 111 we have evaluated different synthetic cases where different scenarios of chaotic advection 112 have been tested in two-phase flow systems. We have defined performance metrics to analyze 113 the impact of chaotic flows on NAPL remediation, its effect on connectivity, and the role of 114 heterogeneity and mobility ratio in chaotic flow configuration. 115

116 2 Methods

117

2.1 Chaotic Advection System

We evaluate an engineered chaotic advection system designed for extracting NAPL in 118 a water-wet heterogeneous porous medium. Injection and extraction wells are organized to 119 form a canonical five-spot pattern [Satkin and Bedient, 1988; Juanes and Lie, 2008]. This 120 organization of wells is frequently used for extracting NAPL in contaminated sites [Nico-121 *laides et al.*, 2015], and oil in petroleum fields [*Craig*, 1971] as well as for carbon storage 122 associated with CO₂ enhanced oil recovery (EOR) in residual oil zones [Ren and Duncan, 123 2019]. In the five-spot pattern, injection and extraction wells are uniformly distributed in 124 such a way that each extraction well is surrounded by 4 injection wells, and an injection well 125 is surrounded by 4 extraction wells (Figure 1). The wetting fluid is injected through the in-126 jection wells to displace the non-wetting fluid (NAPL) towards the central extraction well. 127

Let us denote the injection rate of the wetting fluid associated with the *j*th injection well of a given five-spot pattern as $Q_{w_j}^i(t)$, $j = \{1, 2, 3, 4\}$. Chaotic advection is generated by periodically fluctuating the injection rates $Q_{w_j}^i(t)$ in such a way that each injection well is out of phase with the others. In order to simplify the fluctuation system and reduce the number of parameters, we consider that injection rates follow a rectangular wave function with a time period *T*. This can be formulated by using the rectangular function f_i ,

$$Q_{w_i}^i(t) = Q f_j(t,T), \qquad 0 \le t < T,$$
(1)

$$f_i(t,T) = H\left(t - (j-1)T/4\right) - H\left(t - jT/4\right),\tag{2}$$

together with the statement of periodicity,

$$Q_{w_j}^i(t) = Q_{w_j}^i(t-T), \qquad t \ge T,$$
(3)

where $H(\cdot)$ is the Heaviside function. Note that the rectangular function $f_i(t,T)$ is equal to 135 1 in the time interval [jT/4 - T/4, jT/4] and zero otherwise. That is to say that the pulse 136 duration τ is equal to T/4. Q is a constant value that specifies the injection rate of the wet-137 ting fluid when the well is active $(f_j = 1)$. In short, the chaotic system set-up has two main 138 features: (1) the injection rate is constant and equal to Q for a time interval $\tau = T/4$; and 139 (2) each injection well is periodically activated with a period T. Thus, each injection period 140 is divided into 4 equal subintervals of duration T/4. In each subinterval, only one injection 141 well is active. That is, we first only activate the injection well 1 during the first time subinter-142 val, while keeping the other injection wells deactivated. Then, we deactivate injection well 1 143 and only activate injection well 2 in the second time subinterval, and so on (see Figure 1). 144

2.2 Two-Phase Flow Model

145

We consider the movement of two immiscible liquids in a horizontal two-dimensional heterogeneous aquifer. Mass transfer (e.g., volatilization and dissolution) between the two liquid phases is assumed negligible. The governing equations used to simulate the two-phase flow system are determined by the mass conservation equation of the two liquids and the generalized Darcy's law. Assuming that the porous medium and the fluids are incompressible (constant porosity and fluid densities), we have the following coupled system of partial differential equations in two dimensions,

$$\phi b \frac{\partial S_w}{\partial t} = \nabla \cdot (\kappa \lambda_w b \nabla p_w) + \sum_{j=1}^{n_i} Q_{w_j}^i(t) \delta(\mathbf{x} - \mathbf{x}_j^i) - \sum_{j=1}^{n_e} Q_{w_j}^e(t) \delta(\mathbf{x} - \mathbf{x}_j^e), \tag{4}$$



Figure 1: General five-spot arrangement of extraction and injection wells; the triangle symbols refer to the injection wells, whereas the circle symbols refer to the extraction wells; the region shown in solid lines is the domain of the synthetic test case TC1, and the region in dashed lines is the domain of the synthetic test case TC2.

$$\phi b \frac{\partial S_{nw}}{\partial t} = \nabla \cdot (\kappa \lambda_{nw} b \nabla p_{nw}) - \sum_{j=1}^{n_e} Q^e_{nw_j}(t) \delta(\mathbf{x} - \mathbf{x}^e_j),$$
(5)

where ϕ is the porosity, b is the aquifer thickness, $\mathbf{x} = (x, y)^t$, S_w and S_{nw} are the saturations 153 of the wetting and the non-wetting fluids, $\delta(\cdot)$ is the Dirac delta function, n_i is the number 154 of injection wells, n_e is the number of extraction wells, \mathbf{x}_i^i is the position of the *j*th injection 155 well, \mathbf{x}_{i}^{e} is the position of the *j*th extraction well, κ is the intrinsic permeability, λ_{w} and λ_{nw} 156 are the mobilities of the wetting and the non-wetting fluids, p_w and p_{nw} are the pressures of 157 the wetting and the non-wetting fluids, $Q_{w_i}^i$ is the injection rate of the wetting fluid at the *j*th 158 injection well, and $Q_{w_i}^e$ and $Q_{nw_i}^e$ are the extraction rates of the wetting and the non-wetting 159 fluids at the jth extraction well. The total extraction rate associated with the jth extraction 160 well is $Q_{t_i}^e = Q_{w_i}^e + Q_{nw_i}^e$. Fluid mobility is defined as the ratio of the relative permeability 161 to the viscosity of the fluid, 162

$$\lambda_w = \frac{\kappa_{rw}}{\mu_w}, \qquad \lambda_{nw} = \frac{\kappa_{rnw}}{\mu_{nw}}, \tag{6}$$

where κ_{rw} and κ_{rnw} are the relative permeabilities of the wetting and the non-wetting fluids, and μ_w and μ_{nw} are the viscosities of the wetting and the non-wetting fluids. The relative permeabilities of the non-wetting κ_{rnw} and wetting κ_{rw} phases are only functions of water

saturation and they are described by the Corey correlation model,

$$\kappa_{rnw} = \kappa_{rnwm} (1 - S_e)^{n_{nw}},\tag{7}$$

$$\kappa_{rw} = \kappa_{rwm} S_e^{\ n_w},\tag{8}$$

where κ_{rwm} , n_w , κ_{rnwm} and n_{nw} are the scaling parameters of the relative permeability

 $_{168}$ curves, and S_e is the effective saturation of the wetting fluid defined as

$$S_{e} = \frac{S_{w} - S_{wr}}{1 - S_{wr} - S_{nwr}},$$
(9)

- where S_{wr} and S_{nwr} are the residual saturations of the wetting and the nonwetting phases,
- respectively. The extraction rate of the wetting and the non-wetting phase is determined by the fractional flow function $f_w(S_w)$ through

$$Q_{w_j}^e = Q_{t_j}^e f_w(S_w), \qquad Q_{nw_j}^e = Q_{t_j}^e \left(1 - f_w(S_w)\right), \tag{10}$$

172 where

$$f_w(S_w) = \frac{\lambda_w}{\lambda_w + \lambda_{nw}}.$$
(11)

173

The difference between the two fluid pressures defines the capillary pressure,

$$p_c = p_{nw} - p_w, \tag{12}$$

¹⁷⁴ which is determined by the saturation-capillary pressure relationship or retention curve (see

equation (16)). The system only considers the presence of two liquids and therefore the sum

of saturations is equal to one, i.e., $S_w + S_{nw} = 1$.

177 **2.3 Aquifer Heterogeneity**

The intrinsic permeability and the retention curve are considered to vary in space. The natural log of the intrinsic permeability field, denoted as $Y(\mathbf{x}) = \ln \kappa(\mathbf{x})$, is considered to follow a stationary multi-Gaussian random distribution characterized by an exponential semivariogram model of variance contribution σ_V^2 , defined as

$$\gamma(\mathbf{h}) = \sigma_Y^2 \left(1 - \exp\left(-3|\mathbf{h}'|\right)\right),\tag{13}$$

where **h** is the separation vector between two points of the aquifer, and \mathbf{h}' is the separation vector obtained by orienting the correlation structure along the coordinates and scaling the

ranges to unitary values according to

$$\begin{pmatrix} h'_x \\ h'_y \end{pmatrix} = \begin{pmatrix} \cos\theta & \sin\theta \\ -\sin\theta & \cos\theta \end{pmatrix} \begin{pmatrix} a_{\max}^{-1} & 0 \\ 0 & a_{\min}^{-1} \end{pmatrix} \begin{pmatrix} h_x \\ h_y \end{pmatrix},$$
(14)

where a_{max} and a_{min} are the maximum and minimum ranges in the principal directions. The

maximum correlation direction is oriented θ degrees counterclockwise from the positive x

axis. The randomness of $Y(\mathbf{x})$ is transferred to the retention curve through the Leverett's

function $J(S_e)$ [Leverett, 1939, 1941] that scales the capillary pressure via interfacial ten-

sion, porosity and intrinsic permeability [Brown, 1951; Demond and Roberts, 1991; Lie,

¹⁹⁰ 2014]. The Leverrett's function is an invariant property written as

$$J(S_e) = \frac{p_c}{\gamma \cos \alpha} \sqrt{\frac{\kappa}{\phi}},\tag{15}$$

where γ is interfacial tension, and α is the contact angle. From this, assuming that the saturation-

capillary pressure relation follows the *Brooks and Corey* [1966] model, the retention curve is

assumed to vary as a function of the intrinsic permeability and the effective saturation by

$$p_c(S_e,\kappa) = p_L S_e^{-1/\lambda} \sqrt{\frac{\kappa_g}{\kappa}}, \qquad 0 < S_e \le 1,$$
(16)

where p_L is the characteristic Leverett entry pressure, and κ_g is the geometric mean of permeability.

196 **2.4 Fluid Mobility Ratio**

The displacement of NAPL during injection not only depends on aquifer heterogeneity but also on the fluid properties [*Nicolaides et al.*, 2015]. The mobility ratio M is the mobility of the injection fluid divided by that of the non-wetting fluid it is displacing,

$$M = \frac{\kappa_{rw}^0 \mu_{nw}}{\mu_w \kappa_{rnw}^0}.$$
(17)

To estimate the mobility ratio, in accordance with *Craig* [1971], the relative permeability of the wetting fluid is defined with the average wetting fluid saturation \overline{S}_{wBT} behind the displacing front at breakthrough (denoted as κ_{rw}^0), and the relative permeability of the non-wetting fluid is determined by the non-wetting fluid saturation ahead of the displacing front (denoted as κ_{rnw}^0), i.e., the initial saturation of NAPL. \overline{S}_{wBT} is obtained by laying a tangent line to the fractional flow curve $f_w(S_w)$ from S_{wr} and extrapolating this tangent line to $f_w = 1.0$.

When M < 1 fluid displacement is said to have favorable mobility conditions. In this 207 case, for a given pressure gradient, the wetting fluid can travel at a lower velocity than the 208 non-wetting fluid, effectively pushing the NAPL towards extraction wells. On the contrary, 209 when M > 1 the wetting fluid can travel faster than the non-wetting fluid and there is a 210 tendency for the NAPL to be bypassed. During the in situ remediation of a contaminated 211 aquifer, liquid solutions with chemicals (e.g., surfactants, alkalis, or polymers) are some-212 times injected into the aquifer to improve field conditions by decreasing the mobility ratio 213 M [Huling and Weaver, 1991]. Similar strategies are often used in petroleum engineering to 214 enhanced oil recovery [Abidin et al., 2012; Raffa et al., 2016]. Here, we analyze the effect of 215 the mobility ratio on the performance of chaotic advection by changing the viscosity of the 216 wetting fluid so as to represent favorable and unfavorable mobility conditions. Thus, we con-217 sider three different mobility ratios, i.e., M = 0.5, 1.4, and 2.2. Considering that NAPL vis-218 cosity (e.g., Chlorohydrocarbons and oil products) typically ranges from 0.35 to 28 [mPa·s] 219 [Schwille, 1981; Huling and Weaver, 1991; Reid et al., 1997; Boulding, 1996], approximately 220 equivalent to a mobility ratio ranging between 0.1 and 1.8, these values cover a wide range of 221 applications. 222

223 2.5 Synthetic Test Cases

The objective of the synthetic test cases is to compare chaotic advection NAPL re-224 moval with a constant injection scheme in complex geological formations. The effect of 225 chaotic advection is studied in a wide variety of permeability fields. For this, we consider 226 two synthetic test cases, denoted as TC1 and TC2, that respectively represent two hetero-227 geneous aquifers with different correlation structure of the permeability field and hydraulic 228 connectivity between injection and extraction wells. Let us consider the general arrange-229 ment of injection and extraction wells shown in Figure 1. The distance between two adja-230 cent injection or extraction wells is 212 m and the aquifer thickness is 20 m. The aquifer sys-23 tem is assumed to be initially filled with residual water $S_{wr} = 0.2$ and a large amount of 232 NAPL. The injection-extraction system operates over 20 years. The model domain of TC1, 233 denoted as V_1 , is delimited by the inner square region shown in Figure 1. This test case in-234 volves 4 injectors and 1 central extraction well. TC1 represents a generic five-spot injection-235 extraction system embedded in an isotropic two dimensional heterogeneous $Y(\mathbf{x})$ field with 236 $a_{\text{max}} = a_{\text{min}} = 51$ m. The model domain of TC2 is defined by the dashed lines shown in Fig-237 ure 1. The domain contains 5 extraction wells and 4 injectors. TC2 represents the same five-220 spot pattern but in this case the injection-extraction system is embedded in an anisotropic heterogeneous $Y(\mathbf{x})$ field with $a_{\text{max}} = 225 \text{ m}$, $a_{\text{min}} = 22.5 \text{ m}$, and $\theta = 45^{\circ}$. The maximum 240 correlation direction is oriented along the line connecting injection and extraction wells to 241 enhance hydraulic connectivity. The maximum range is smaller than the field scale to as-242 sure the effect of permeability heterogeneity is activated; otherwise, the field is relatively 243 homogeneous. The two heterogeneous systems share the same geometric mean of the intrin-244 sic permeability, $\kappa_g = 10^{-14} \text{ m}^2$, and we explore three different degrees of heterogeneity, 245 $\sigma_v^2 = 0.1, 2$, and 6, which represent a mild, moderate, and highly heterogeneous aquifer. We 246 chose to work with a low κ_g value to test chaotic advection under adverse conditions with 247 permeability values that fluctuate between 10^{-18} and 10^{-10} m². NAPL is typically difficult 248 to recover in low permeability formations [Mackay and Cherry, 1989]. We note though that 249 the analysis is presented using dimensionless variables to make the results more general. The 250 geostatistical parameters of the $Y(\mathbf{x})$ random fields are summarized in Table 1. 251

All domain boundaries are set to no-flow conditions. The extents of V_1 and V_2 are 252 212×212 m² and 300×300 m², respectively. V_2 is larger than V_1 to allow aligning the domain 253 boundaries with the stratification in the TC2 case. Note that, otherwise, the injected fluid 254 255 would be forced to move through the stratification. Of course, some boundary effects are expected but the intend here is not to exactly reproduce a large field system but to compare 256 chaotic advection removal with a conventional scheme. The total extraction rate assigned to 257 the central well is always constant and fixed to $Q_{t_5}^e = Q$ in all cases. Chaotic advection fol-258 lows always a rectangular wave function with amplitude Q and period T. The constant injec-259

tion scheme considers that $Q_{w_j}^i = Q/4$ in TC1 and $Q_{w_j}^i = Q/2$ in TC2 for $j = \{1, 2, 3, 4\}$. In TC2, the total extraction rate of corner wells are fixed to $Q_{t_j}^e = Q/4$ for $j = \{1, 2, 3, 4\}$. Parameters adopted during the simulations are listed in Table 1. The domains of TC1 and TC2 are respectively discretized into 101 and 201 squared cells to represent the variability of the random fields. The resolution of the random fields vary between 15 and 150 cells per range, which is considered sufficient to represent the inherent spatial variability of permeability. Figure 2 shows illustrative test fields and corresponding well arrangements. To improve visualization, TC2 is rotated 45° clockwise from the *x* positive axis.

The simulation approach is as follows. For each test case, we first consider a stochastic 268 description of $Y(\mathbf{x})$ with 100 equally likely realizations characterized by $\sigma_V^2 = 2$ and M = 2.2269 to explore the range of uncertainty. Within each realization, two-phase flow simulations with 270 constant-injection and chaotic advection removal are conducted with different periods T, 271 which vary from 0.5 to 20 years. Performance metrics are then characterized by their statis-272 tical moments (mean behavior and uncertainty) and sample probability density functions 273 (PDFs). Finally, we investigate the effect of the degree of heterogeneity σ_V^2 and mobility 274 ratio M in individual realizations. The effect of σ_V^2 is analyzed by re-scaling the variance of the $Y(\mathbf{x})$ values adopted in a given realization so as to always replicate the same specific 276 heterogeneous patterns. The viscosity of the non-wetting fluid is kept constant to 13 mPa·s, 277 while the viscosity of the wetting fluid is changed from 1.0 to either 0.2 or $5.0 \text{ mPa} \cdot \text{s}$ (see 278 Table 1), which can represent, for instance, chemical flooding with polymers during NAPL 279 remediation or enhanced oil recovery with CO₂ sequestration, respectively. 280

Random fields are generated with the Sequential Gaussian Simulation method imple-281 mented in the SGSIM code [Journel and Huijbregts, 1976]. We use the open-source Matlab 282 Reservoir Simulation Toolbox (MRST)[Krogstad et al., 2015] to simulate two-phase flow 283 using the IMplicit Pressure Explicit Saturation (IMPES) algorithm [Lie, 2014; Chen et al., 284 2006; Yanosik and McCracken, 1979]. The numerical discretization of the flow solution and 285 the $Y(\mathbf{x})$ field is the same. The maximum time step for updating saturation is constraint by 286 the Courant-Friedrichs-Lewy (CFL) condition to assure that the time step for updating saturation is smaller than that for updating pressure [*Courant et al.*, 1928; *Coats*, 2003]. The 288 two-point flux approximation (TPFA) is employed to solve the pressure equation. The one-289 point upstream weighting scheme is used to avoid artificial dispersion [Forsyth and Sammon, 290 1986; Sammon, 1988; Allen, 1985]. This upstream weighting scheme has first-order spatial 291 accuracy [Sleep and Sykes, 1993a,b]. 292

2.6 Performance Metrics

293

We define two different performance metrics to evaluate the relative efficiency of the proposed chaotic advection system involved in a five-spot pattern: the removal efficiency and the saturation distribution index. The removal efficiency RE(t) measures the volume of NAPL recovered at time *t* relative to the initial volume of NAPL in the aquifer, and the distribution index DI(t) quantifies the degree of uniformity of the wetting fluid saturation distribution at time *t* [*Le Borgne et al.*, 2010; *Nicolaides et al.*, 2015]. The formal definition of these metrics can be written as,

$$RE(t) = \frac{1}{V_{nwi}} \int_0^t Q_{nw_5}(t) dt,$$
 (18)

$$DI(t) = 1 - \frac{\sigma^2(t)}{\sigma_{\max}^2(t)},\tag{19}$$

where Q_{nw_5} is the NAPL extraction rate obtained at the central well of the five-spot pattern

(see Figure 1), V_{nwi} is the initial volume of NAPL in the V_1 -domain,



Figure 2: Test fields and corresponding well arrangements; the first panel of the figure shows one realization of the isotropic random field used in the synthetic test case TC1, and the second panel shows one realization of the anisotropic random field used in the synthetic test case TC2 (the domain has been rotated 45° clockwise from the *x* positive axis to improve visualization).

Parameters	TC1	TC2	References
(n_e, n_i) [-]	(1, 4)	(5, 4)	Craig [1971]
<i>V</i> [m ³]	$212 \times 212 \times 20$	$300 \times 300 \times 20$	-
$n_x \times n_y \times n_z[-]$	101×101	201×201	-
$\kappa_g [\mathrm{m}^2]$	10^{-14}	10^{-14}	Mackay and Cherry [1989]; Wu et al. [1994]
σ_Y^2 [-]	0.1 / 2.0 / 6.0	0.1 / 2.0 / 6.0	Craig [1971]; Dillard et al. [1997]
(a_{\max}, a_{\min}) [m]	(51, 51)	(225, 22.5)	Kitanidis [1997]
θ [degrees]	0°	45°	-
ϕ [-]	0.2	0.2	Wu et al. [1994]
$Q [10^{-3} \text{m}^3/\text{s}]$	0.228	0.228	-
(ρ_{nw}, ρ_w) [kg/m ³]	(898, 981)	(898, 981)	Schwille [1981]
µ _{nw} [mPa∙ s]	13.0	13.0	Schwille [1981]
$\mu_w \text{ [mPa· s]}$	0.2 / 1.0 / 5.0	0.2 / 1.0 / 5.0	Schwille [1981]
M [-]	2.2 / 1.4 / 0.5	2.2 / 1.4 / 0.5	Schwille [1981]
(S_{wr}, S_{nwr}) [-]	(0.2, 0.2)	(0.2, 0.2)	Wu et al. [1994]; Lie [2014]
$(\kappa_{rwm},\kappa_{rnwm})$ [-]	(0.2, 0.8)	(0.2, 0.8)	Craig [1971]; Lie [2014]
(n_w, n_{nw}) [-]	(2, 2)	(2, 2)	Sleep and Sykes [1993a]
p_L [Pa]	8 ^{<i>a</i>}	8	Zhong et al. [2001]; Wipfler et al. [2004]
λ[-]	0.5	0.5	Sleep and Sykes [1993a]

Table 1: Summary of the parameters adopted during the simulations for the two synthetic test cases.

^a Low entry pressure is used considering the imbibition process.

$$V_{nwi} = \int_{V_1} \phi S_{nw} (t=0) dV,$$
 (20)

 σ^2 is the variance of the S_w -distribution in the V_1 -domain, and σ^2_{max} is the maximum variance of S_w in the V_1 -domain. The variance of saturation associated with the wetting fluid is defined as

$$\sigma^{2}(t) = \frac{1}{V_{1}} \int_{V_{1}} S_{w}^{2}(t) dV - \left(\frac{1}{V_{1}} \int_{V_{1}} S_{w}(t) dV\right)^{2}.$$
(21)

Note that the performance metrics only consider the simulated values obtained in the 306 V_1 region. This intends to minimize boundary effects in TC2. The maximum variance σ_{max}^2 307 is obtained when the distribution of saturation (wetting fluid) exhibits a bimodal distribution 308 with two segregated modes. In a multiphase injection-extraction removal system, this hap-309 pens at early stages after injection, when the displacement is piston-like and the saturation of 310 the injected fluid behind the displacement front is significantly different from the saturation 311 ahead, which is close to the initial saturation. With time, driven by capillary dispersion and 312 heterogeneity (and in our case chaotic advection), these two distinct saturations will mix. In 313 an ideal case, when the saturation distribution is perfectly mixed, the saturation distribution 314 approaches a unimodal distribution with $\sigma^2 = 0$ and DI = 1. Similar metrics of mixing 315 (substituting S_w by solute concentrations) can be found in the literature of solute transport 316 in porous media [Jha et al., 2011]. Here, knowing that S_w ranges between 0 and 1, we esti-317 mated σ_{max}^2 by the following upper bound of variance [Bhatia and Davis, 2000], 318

$$\sigma_{\max}^2(t) = \frac{1}{V_1} \int_{V_1} S_w(t) dV - \left(\frac{1}{V_1} \int_{V_1} S_w(t) dV\right)^2.$$
 (22)

The chaotic advection system results are compared with a standard injection-extraction removal system characterized by constant injection rates. Whenever necessary for a better interpretation of the results, performance metrics are presented as fractional increase with respect to the constant-injection solution. For any given performance metric χ , the fractional increase of χ is determined as

$$\Delta \chi = \frac{\chi_c - \chi_r}{\chi_r},\tag{23}$$

where the subscripts c and r denote the chaotic advection and the reference constant injection rate simulation results, respectively.

326 2.7 Dimensionless Variables

To facilitate the interpretation, we present the results in terms of dimensionless vari-327 ables. In statistical physics, the Kubo number Ku is a dimensionless measure of the correla-328 tion time of the fluctuations typically used for analyzing the behavior of moving particles in 329 turbulent, random or chaotic velocity fields [Kubo, 1963; Mazzino, 1997; Castiglione, 2000; Vlad et al., 2001]. The Kubo number has been also used to study solute transport in tempo-331 rally fluctuating flow through randomly heterogeneous porous media [Dentz and Carrera, 332 2005; De Dreuzy et al., 2012]. In solute transport, the Kubo number compares the average 333 travel distance of a particle with the integral scale of $Y(\mathbf{x})$. Here, based on this, we define the 334 following Kubo number in two-phase flow systems subject to periodic injection pulses, 335

$$Ku = \frac{v_f \tau}{\ell},\tag{24}$$

where τ is the pulse duration, v_f is the mean velocity of the saturation front, and ℓ is a measure of the correlation scale of $Y(\mathbf{x})$. The mean velocity of the saturation front is estimated by the breakthrough time of the saturation of the wetting fluid at the central extraction well of the five-spot pattern under constant injection conditions, denoted as t_{BT} ,

$$v_f \approx \frac{L}{\langle t_{BT} \rangle},$$
 (25)

where L is the separation distance between injection and extraction wells, and $\langle \cdot \rangle$ denotes 340 the ensemble average of the Monte Carlo simulations. Note that this way the Kubo number 341 directly includes the breakthrough time, which is known to control the efficiency of NAPL 342 removal in real applications. Here, we choose to use the range of the $Y(\mathbf{x})$ field as a mea-343 sure of correlation because it provides an indication of the average extent of low/high per-344 meability zones. Since in our chaotic removal setup the injected fluid moves half of the time 345 along and transverse to the direction of stratification in an average sense, we have used $\ell =$ 346 $(a_{\text{max}} + a_{\text{min}})/2$ to estimate the Kubo number. We have also normalized the time by 347

$$t^* = \frac{v_f t}{\ell}.$$
(26)

The number of cycles completed after a time *t* during chaotic advection removal is defined as N = t/T. Dividing (26) by (24) and knowing that $T = 4\tau$, we have

$$N = \frac{t^*}{4\mathrm{Ku}}.$$
(27)

In the appendix we show that the governing equations of the two-phase flow system considered can be written in a dimensionless form that directly depends on the Kubo number and these dimensionless variables.

In a given realization of the random field, we expect the hydraulic connectivity be-353 tween injection and extraction wells to control the efficiency of the injection-extraction sys-354 tem; high connectivity can generate fast flow pathways between injection and extraction 355 wells leading to early breakthrough times [Silliman and Wright, 1988; Labolle and Fogg, 356 2001; Bianchi et al., 2011b; Renard and Allard, 2013b; Edery et al., 2014]. Several indi-357 cators of connectivity have been proposed in subsurface hydrology for flow and contaminant transport in aqueous phase systems [Sánchez-Vila et al., 1999; Fernàndez-Garcia et al., 359 2002; Knudby and Carrera, 2005; Trinchero et al., 2008]. Here, following Fernàndez-Garcia 360 et al. [2010], we define an indicator of point-to-point connectivity for multiphase flow sys-361 tems as, 362

$$CS = \frac{t_0}{t_{BT}},\tag{28}$$

where t_0 is an expected or reference value of the saturation breakthrough time. Injection and extraction wells are well connected in terms of saturation displacement when CS > 1, since the observed breakthrough time is more rapid than that its expected reference value. The larger the CS value, the better connection exists between injection and extraction wells, and one should expect geological bodies of high permeability connecting injection and extraction wells. We chose to measure t_0 by the expected value of the breakthrough time of saturation $\langle t_{BT} \rangle$ obtained in isotropic random fields under constant injection conditions.

370 3 Results and Discussions

371

3.1 Mean Behavior: The Role of the Fluctuation Period

The ensemble averages of the fractional increase of the removal efficiency $\langle \Delta RE \rangle$ and 372 distribution index $\langle \Delta DI \rangle$ are shown in Figure 3 as a function of the Kubo number for different removal times. Here, we only employ chaotic periods smaller than the total recovery 374 time to assure that chaotic advection is active, otherwise the flow is effectively steady. Re-375 sults show that $\langle \Delta RE \rangle$ increases to a maximum value when a specific pulse duration τ of a 376 periodically applied rotating injection pulse yields a Kubo number close to one, i.e., $Ku \approx 1$. 377 This result is somehow analogous to the effect of temporal flow fluctuations on solute trans-378 port (single phase). Dentz and Carrera [2005] and De Dreuzy et al. [2012] found that the 379 effective transverse dispersion coefficient of a solute plume is maximized when Ku = 1. This suggests that in a two-phase flow system, when Ku = 1, advective transport and temporal 381 fluctuations are synchronized to improve NAPL displacement towards the extraction well, 382 most likely due to an enhancement of transverse dispersion of the displacing fluid saturation. 383 In practice, this means that the saturation front should travel one range of the $Y(\mathbf{x})$ field (an 384 average extent of low/high permeability zones) in an injection pulse to maximize NAPL re-385 moval. When $Ku \ll 1$, the frequency of temporal fluctuations is too high to properly sample 386 the permeability field. When Ku>1, chaotic advection generates a highly nonuniform partial 387 sweep of the porous medium. 388

Remarkably, a more pronounced peak is observed in the anisotropic case, which suggests that chaotic removal works best under unfavorable field conditions, i.e., when some of the injectors are potentially correlated with the central well. In the anisotropic case we obtain a maximum fractional increase close to 12%, which is six times larger than that of the isotropic case. In particular, the pulse duration for maximizing removal efficiency in TC1 and TC2 are $\tau \approx \langle t_{BT} \rangle / 5$ and $\tau \approx \langle t_{BT} \rangle / 2$, respectively. It is logical to think that the pulse duration should be smaller than the breakthrough time, otherwise the wetting fluid can gain access to the extraction well in the first injection of the fluctuation cycle. Following this line



Figure 3: Ensemble average of the fractional increase of the removal efficiency and distribution index as a function of the Kubo number for different removal times in isotropic (first panel) and anisotropic (second panel) random permeability fields with $\sigma_Y^2 = 2$ and M = 2.2.

of thought, it is not surprising that $\langle \Delta RE \rangle$ rapidly declines after passing through a maximum when the pulse duration, and therefore the Kubo number, becomes too large ($\langle \Delta RE \rangle$ can reach negative values in the isotropic case).

The ensemble average of the fractional increase of the distribution index $\langle \Delta DI \rangle$ is also 400 shown in Figure 3. Chaotic advection is also demonstrated to enhance mixing. However, a 401 clear peak is only observed in the isotropic case after a long time (when $t^* > 15$) and at Kubo 402 numbers slightly larger than 1 (between 1.5 and 3). Probably, the peak cannot be seen in the 403 other cases because the maximum Kubo number available is relatively small. These results 404 highlights that maximum mixing does not necessarily imply maximum removal, most likely 405 because the wetting fluid can only effectively displace the non-wetting fluid when $S_w \gg S_{wr}$ 406 due to the non-linear nature of the relative permeability. 407

Figure 4 depicts the temporal evolution of $\langle \Delta RE \rangle$ and $\langle \Delta DI \rangle$ for different Kubo num-408 bers. Chaotic effects on $\langle \Delta RE \rangle$ require a certain time to develop during which sometimes it 409 exhibits a valley, then reaches a maximum, and after this it slowly declines with time. These 410 features are more intense for anisotropic fields. The valley displays negative values decreas-411 ing with the Kubo number. In general, the time needed to reach the valley and the peak is 412 relatively smaller in anisotropic fields, meaning that chaotic advection effects develop faster 413 in well-connected permeability fields. From (27), we have that $N = t^*/4$, which means that 414 the number of cycles require to reach maximum removal is only about 2 and 3 cycles in the 415 anisotropic and isotropic case, respectively. The valley seems to take place in the first cy-416 cle. From a practical point of view, it is important to recognize that these results suggest 417 that in order to maximize removal in the long term one should undergo first an early stage 418 with detrimental effects in removal efficiency. The temporal evolution of $\langle \Delta DI \rangle$ is similar 419 to $\langle \Delta RE \rangle$, but in this case the valley and the peak take place at different times and with less 420 intensity. In fact, the valley becomes only apparent when Ku > 1. 421



Figure 4: Temporal evolution of the ensemble average of the fractional increase of removal efficiency and distribution index for different Kubo numbers in isotropic (first panel) and anisotropic (second panel) random permeability fields with $\sigma_Y^2 = 2$ and M = 2.2.

3.2 Uncertainty in Chaotic Advection NAPL Removal

422

The inherent complexity of a heterogeneous geological system typically produces large 423 uncertainties in the efficiency of NAPL removal in field applications. For instance, it is well 424 known that DNAPL removal efficiencies of current remediation techniques are limited and 425 highly variable when moving from laboratory to field scale [Soga et al., 2004]. In this sec-426 tion, we demonstrate that chaotic advection removal not only improves performance metrics, 427 but also reduces uncertainty, making the application of in situ removal techniques more re-428 liable and less sensitive to the underlying heterogeneity of the permeability field. To show 429 this, Figure 5 presents the coefficients of variation of the removal efficiency CV_{RE} and distri-430 bution index CV_{DI} as a function of the Kubo number for different removal times. For com-431 parison purposes, the horizontal dashed lines shown in the figures indicate the corresponding 432 coefficient of variation obtained with a constant injection scheme. In general, the coefficient 433 of variation of RE is one order of magnitude larger than that of DI. Results demonstrate 434 that chaotic removal can significantly reduce the uncertainty of removal efficiency and dis-435 tribution index relative to a constant injection scheme. This effect is more pronounced in the 436 anisotropic case with unfavorable conditions. In this case, the coefficients of variation of RE 437 and DI are respectively reduced from 0.15 to 0.12 and from 0.023 to 0.008. The uncertainty 438 in removal efficiency exhibits its minimum value when $\langle \Delta RE \rangle$ is maximum (Ku \approx 1). This 439 suggests that the increase in removal efficiency and distribution index due to chaotic advec-440 tion always goes along with a reduction of their uncertainty. 441

The temporal evolution of the coefficient of variation of performance metrics is depicted in Figure 6 for different Kubo numbers. For comparison, the dashed lines correspond to the constant injection scheme. In general, the uncertainty of *RE* and *DI* exhibits large fluctuations at early times, which ultimately vanish to approach a well-defined asymptotic value at large times. The effects of chaotic advection are more pronounced in the anisotropic case with similar overall behavior. The time needed to reach an asymptotic value in removal



Figure 5: Coefficient of variation of the removal efficiency and distribution index as a function of the Kubo number for different removal times in isotropic (first panel) and anisotropic (second panel) random permeability fields with $\sigma_Y^2 = 2$ and M = 2.2.

efficiency strongly depends on the Kubo number. When Ku ≈ 1 , CV_{RE} approaches the asymptotic value more rapidly than in other cases, i.e., in less than one range of the $Y(\mathbf{x})$ field. The coefficient of variation of the distribution index, at late times, decreases with the Kubo number. This illustrates that low-frequency fluctuations can further reduce the uncertainty of mixing but at the expenses of removal efficiency and its uncertainty.

The probability density function (PDF) of RE and DI provides a broader description 453 of the ensemble of realizations. This is shown in Figure 7 for a removal time of 10 years. We compare the results obtained by using a constant injection with those obtained by using 455 chaotic advection with optimal Kubo number. The PDFs were estimated through an iterative 456 optimal kernel density estimator [Engel et al., 1994] to minimize spurious statistical fluctu-457 ations. As expected from our previous results, the central tendency of the PDFs is shifted 458 towards larger removal efficiencies and distribution indexes due to chaotic advection. In a 459 constant injection scheme, removal efficiency exhibits a wide distribution with a long tail as-460 sociated with relatively large removal efficiencies. In contrast, NAPL removal with chaotic 461 advection yields a more symmetric and narrower distribution of the removal efficiency with 462 high probabilities centered at relatively large quantities. 463

3.3 Impact of Chaotic Advection on Connectivity

464

The hydraulic connection between injection and extraction wells depends on the spe-465 cific spatial patterns that the permeability field displays in a given realization. Within each 466 realization, the measure of connectivity CS presented in section 2.7 quantifies the presence 467 of preferential channels or high permeability regions between injection wells and the central 468 extraction well [Trinchero et al., 2008; Fernàndez-Garcia et al., 2010]. Within this context, 469 in this section we analyze the dependence between chaotic removal and connectivity. For 470 this, we present in Figure 8 the conditional expectation of the removal efficiency and distri-471 bution index (and corresponding fractional increase) relative to the connectivity indicator, 472



Figure 6: Temporal evolution of the coefficient of variation of the removal efficiency and distribution index as a function of the Kubo number for different removal times in isotropic (first panel) and anisotropic (second panel) random permeability fields with $\sigma_Y^2 = 2$ and M = 2.2.



Figure 7: Comparison of the estimated probability density functions of the removal efficiency and distribution index obtained with chaotic advection (Kubo numbers associated with maximum performance) and constant injection removal for anisotropic fields with $\sigma_Y^2=2$ and M=2.2 after 10 years of operation.



Figure 8: Removal efficiency and distribution index (and their corresponding fractional increase) as a function of point-to-point connectivity in two-phase flow systems for different removal times and injection-extraction method in isotropic and anisotropic random fields with $\sigma_Y^2 = 2$ and M = 2.2.

i.e., $\langle RE|CS \rangle$ and $\langle DI|CS \rangle$, for different removal times. Here, the pulse duration is chosen to satisfy maximum performance.

Results agree with field observations in that removal efficiency decreases with con-475 nectivity. The inefficiency of NAPL removal is mainly attributed to the inability of the dis-476 placing fluid to sweep the NAPL trapped in low permeability regions, which is bypassed 477 when the injection and extraction wells are well connected [De Dreuzy et al., 2012; De Bar-478 ros et al., 2013; Edery et al., 2014]. As expected, the anisotropic random fields reflect large 479 CS values (Figure 2). TC2 exhibits elongated lenses of high/low permeabilities between the 480 injectors and the central extraction well. Looking at both the removal efficiency and its frac-481 tional increase we see that even though the removal efficiency decreases with connectivity, 482 the fractional increase due to chaotic advection becomes important with increasing CS. That 483 is, the fractional increase of the removal efficiency increases with unfavorable connectivity 484 conditions, meaning that chaotic removal works best in the worst case scenario. Hence, to 485 some extent, results suggest that chaotic advection can partially overcome channeling effects 486 during NAPL removal. 487

Connectivity affects the distribution index (mixing of saturations) in a similar way. 488 Recalling that the distribution index is a measure of mixing, results indicate that strong con-489 nectivity patterns in a heterogeneous aquifer tend to preclude the occurrence of mixing. It is 490 logical to think that well-connected fields will concentrate the wetting fluid in small regions, 491 making it difficult for mixing to occur [De Barros et al., 2013]. This is equally true for both 492 injection modes, but one can easily appreciate that chaotic removal renders the system less 493 dependent on connectivity, because the chaotic advection can partially break fast flow paths. 494 Similar effects have been reported in solute transport [De Dreuzy et al., 2012]. DI decreases 495 with CS but at a smaller rate during chaotic advection removal. 496

3.4 The roles of the Degree of Heterogeneity and Mobility Ratio

497

In this section we evaluate the roles that heterogeneity and mobility ratio have on re-498 moval efficiency and mixing. The effect of σ_V^2 is analyzed by re-scaling the variance of the 499 $Y(\mathbf{x})$ values adopted in a given realization. A similar approach was used by [Neupauer et al., 500 2014]. This way we make sure that we always deal with the same heterogeneous pattern, i.e., 501 the same organization of permeability values. The $Y(\mathbf{x})$ fields used are shown in Figure 2. 502 We also changed the viscosity of the wetting fluid to analyze the effect of the mobility ratio 503 on chaotic removal. The mobility ratio is changed from M = 2.2 to M = 1.4 and M = 0.5. Figure 9 shows the fractional increase of removal efficiency obtained in the isotropic and 505 anisotropic case after 20 years as a function of the Kubo number for different σ_V^2 and M val-506 ues. The general behavior follows our previous results; maximum removal close or slightly 507 larger than Ku ≈ 1 followed by a rapid decline. The location of the peak slightly depends 508 on the mobility ratio, indicating that unfavorable displacement (M > 1) may require slightly 509 smaller frequencies of chaotic fluctuations. However, the important point here is to realize 510 that chaotic removal in two-phase flow systems is significantly affected by two competing 511 factors: mobility ratio and heterogeneity. When the degree of heterogeneity is not significant 512 $(\sigma_V^2 < 2)$, the unfavorable displacement caused by high mobility ratios controls removal 513 efficiency, i.e., ΔRE increases with the mobility ratio. When heterogeneity is important 514 $(\sigma_V^2 > 2)$, channeling controls the displacing process regardless of the mobility ratio. We 515 note also that chaotic advection is more effective under unfavorable conditions of heterogene-516 ity, i.e., high σ_v^2 in well-connected anisotropic fields. This is because removal efficiency is 517 typically small in highly heterogeneous systems due to channeling, thus leaving a large op-518 portunity for improvement. In this case, for this realization of the random field, we obtain 519 a fractional increase larger than 20% at peak values. In this context, we note that Neupauer 520 et al. [2014] also found that chaotic advection in solute transport (single phase) is most ad-521 vantageous in highly heterogeneous fields with large σ_V^2 . 522



Figure 9: Fractional increase of the removal efficiency as a function of the Kubo number for different degrees of heterogeneity and mobility ratios in one realization of the isotropic (first panel) and anisotropic (second panel) random field.

To visually illustrate the benefits of chaotic advection removal, Figures 10 and 11 of-523 fer the map of the wetting fluid saturation after 20 years of operation in the isotropic and 524 anisotropic permeability fields shown in Figure 2 for M = 2.2 and $\sigma_Y^2 = 0.1, 2$, and 6. Con-525 stant injection results are compared with those corresponding to the chaotic sequences that 526 yielded maximum removal (RE) and maximum mixing (DI). We can easily see that chaotic 527 removal with optimal fluctuations can significantly outperform the constant injection scheme. 528 This is particularly notable when field conditions are unfavorable, i.e., large σ_V^2 and injection-529 extraction wells oriented along the principal correlation direction. Note for instance that even 530 though severe stratification (anisotropic case with σ_v^2 =6) controls the distribution of satura-531 tion in a constant injection scheme, the application of chaotic advection can largely palliate 532 this shortcoming. This figure also illustrates the dichotomy between maximizing removal ef-533 ficiency or mixing. Results have shown that these two conditions occur at different chaotic 534 periods and removal times. In practice, the use of one or another will depend on the project 535 objectives. For instance, at early stages of remediation one may favor removal efficiency, but 536 at late times, when liquid solutions with cosolvents, surfactants or polymers are meant to be 537 used to alter fluid properties and improve performance, one may wish to promote mixing 538 [Huling and Weaver, 1991; Rao et al., 1997; Neupauer et al., 2014]. 539



Figure 10: Spatial distribution of the wetting fluid saturation after 20 years of operation in one realization of the isotropic random field for different degrees of heterogeneity and injection-extraction method with M=2.2.



Figure 11: Spatial distribution of the wetting fluid saturation after 20 years of operation in one realization of the anisotropic random field for different degrees of heterogeneity and injection-extraction method with M=2.2.

As an illustrative example of the method, we finally compare in Figure 12 the temporal 540 evolution of the NAPL flow rate recovered at the central well of the five-spot pattern pro-541 duced by chaotic advection with that of the constant injection scheme obtained in a given 542 realization of the anisotropic permeability field for σ_Y^2 =6 and M = 2.2 (worst case sce-543 nario). Once the non-wetting fluid breaks through the extraction well, the fractional flow 544 rate of NAPL rapidly declines, generating a long tail of poor removal efficiencies with time. 545 Instead, enhanced NAPL removal with chaotic advection produces a more persistent NAPL 546 extraction rate sequence characterized by important NAPL removal spikes. The net result in 547 this case is a relative increase in removal efficiency of 22%. 548



Figure 12: Comparison of temporal evolution of the NAPL flow rate recovered at the central well of the five-spot pattern produced by chaotic advection with that of the constant injection scheme obtained with the anisotropic permeability field for $\sigma_Y^2 = 6$ and M = 2.2.

549 4 Conclusions

We have proposed and evaluated the use of chaotic advection to enhance non-aqueous 550 phase liquid removal during in situ soil washing remediation techniques and enhanced oil 551 recovery in complex geological formations. Chaotic advection is generated through the ap-552 plication of a rotating periodic injection pulse in a five-spot injection-extraction pattern. To 553 evaluate the method, we have performed two-phase flow simulations in multiple realizations 554 of randomly heterogeneous permeability fields with different correlation structures and con-555 nectivity structures between injection and extraction wells. Performance metrics include 556 removal efficiency and the distribution index of saturation (mixing). We have shown that 557 chaotic advection can significantly improve removal efficiency and mixing. The performance 558 of the method depends on the Kubo number, the connectivity between injection and extrac-559 tion wells, the degree of heterogeneity and the mobility ratio. The most important findings 560 are listed as follows: 561

Chaotic advection improves NAPL removal efficiency and the mixing between the
 wetting and non-wetting phases. This is relatively more pronounced when the perme ability field displays unfavorable conditions, i.e., when the injection and extraction
 wells are well-connected to each other through preferential channels, the permeability
 field is highly heterogeneous, and/or the mobility ratio between the wetting and the

567 568	non-wetting fluid is larger than one (unfavorable displacement). This makes chaotic advection extremely useful in worst-case field applications.
569	2. Removal efficiency increases to a maximum value when the pulse duration of a peri-
570	odically applied rotating injection pulse satisfies that the Kubo number is close to one,
571	i.e., when the saturation front travels one range of the permeability field (an average
572	extent of low/high permeability zones) per injection pulse. This maximum value, in
573	an ensemble average sense, is around 12% in unfavorable conditions of the permeabil-
574	ity field with $\sigma_Y^2 = 2$.
575	3. Chaotic advection can fully develop maximum strength after a few injection cycles.
576	However, results have shown that removal efficiency should first undergo an early
577	stage with detrimental effects in order to maximize removal in the long term.
578	4. The application of chaotic advection not only enhances non-aqueous phase liquid re-
579	moval, but also reduces its uncertainty, making the removal system more reliable and
580	less dependent on heterogeneity. Again, this reduction is higher in unfavorable condi-
581	tions of the permeability field.
582 583 584 585 586 587	5. Maximum removal efficiency and mixing occur at different chaotic periods and removal times. In practice, the use of one or another will depend on the project objectives. For instance, at early stages of remediation one may favor removal efficiency, but at late times one may wish to promote mixing when liquid solutions with cosolvents, surfactants or polymers are meant to be used to alter fluid properties and improve performance.
588	6. The mobility ratio and the degree of heterogeneity are two competing factors con-
589	trolling the performance of chaotic advection in two-phase flow systems. When σ_Y^2 is
590	large the performance of chaotic advection is mainly controlled by heterogeneity. On
591	the contrary, when σ_Y^2 is small, the mobility ratio controls the overall behavior of the
592	system.
593	These results encourage the application of chaotic advection in multiphase flow prob-

lems. In this context, we note that the application of chaotic advection in mathphase now prob sites and/or enhanced oil removal comes at almost no additional cost since the removal system consisting of several wells is not necessarily modified with respect to standard practices, except for the way the technology (delivering of fluids) is put into practice. Moreover, this method can be easily implemented with other techniques to further improve removal efficiencies. The enhancement of mixing between phases can favor for instance surfactant dissolution and mobilization of NAPLs, chemical flooding or CO_2 sequestration through enhanced oil removal.

A Governing Equations in Dimensionless Form

In this appendix we show that the governing equations of the two-phase flow system can be written in a dimensionless form. The appendix shows that the Kubo number and the dimensionless variables used in our analysis arise naturally from the mass conservation equations. These quantities have therefore an important role in analyzing chaotic advection removal systems, defining the controlling parameters as well as the characteristic scales of the problem (in space and time). Let us define the following dimensionless variables,

$$x^* = \frac{x}{\ell}, \qquad y^* = \frac{y}{\ell}, \qquad t^* = \frac{v_f t}{\ell}.$$
 (A.1)

The Darcy flux and pressure of the wetting and non-wetting fluid are written in dimensionless form as

$$q_w^* = \frac{q_w}{v_f \phi}, \qquad q_{nw}^* = \frac{q_{nw}}{v_f \phi}, \qquad p_w^* = \frac{p_w \kappa_g}{\mu_w v_f \phi \ell}, \qquad p_{nw}^* = \frac{p_{nw} \kappa_g}{\mu_w v_f \phi \ell}. \tag{A.2}$$

With these definitions, Darcy's law is expressed as

$$q_w^* = -\kappa_{rw} \exp\left(Y'\right) \nabla^* p_w^*,\tag{A.3}$$

$$q_{nw}^* = -\frac{\gamma \kappa_{rnw}}{M} \exp\left(Y'\right) \nabla^* p_{nw}^*,\tag{A.4}$$

where $\nabla^* = [\partial/\partial x^*, \partial/\partial y^*], \gamma = \kappa_{rw}^0 / \kappa_{rnw}^0$, and Y' is the deviation of the natural log of the

intrinsic permeability from the mean, i.e., $Y' = Y - \langle Y \rangle$. By construction, Y' mainly depends on the degree of heterogeneity σ_Y^2 , the two-point statistics (variogram), and the hydraulic

connectivity CS. The injection and extraction flow rates are written in dimensionless form as

$$Q_{w_{j}}^{i*} = \frac{Q_{w_{j}}^{i}}{v_{f}\phi b\ell}, \qquad Q_{w_{j}}^{e*} = \frac{Q_{w_{j}}^{e}}{v_{f}\phi b\ell}, \qquad Q_{nw_{j}}^{e*} = \frac{Q_{nw_{j}}^{e}}{v_{f}\phi b\ell}, \qquad Q^{*} = \frac{Q}{v_{f}\phi b\ell}.$$
 (A.5)

Substituting (A.1), (A.2) and (A.5) into (4) and (5), we have

$$\frac{\partial S_{w}}{\partial t^{*}} = -\nabla^{*} \cdot q_{w}^{*} + \sum_{j=1}^{n_{i}} Q_{w_{j}}^{i*}(t^{*})\delta(\mathbf{x}^{*} - \mathbf{x}_{j}^{i*}) - \sum_{j=1}^{n_{e}} Q_{w_{j}}^{e*}(t^{*})\delta(\mathbf{x}^{*} - \mathbf{x}_{j}^{e*}), \tag{A.6}$$

$$\frac{\partial S_{nw}}{\partial t^*} = -\nabla^* \cdot q_{nw}^* - \sum_{j=1}^{n_e} Q_{nw_j}^{e*}(t^*) \delta(\mathbf{x}^* - \mathbf{x}_j^{e*}).$$
(A.7)

⁶¹⁷ Knowing by the properties of the Heaviside function that

$$H(t - (j - 1)T/4) - H(t - jT/4) = H(t^* - (j - 1)Ku) - H(t^* - jKu),$$
(A.8)

618 the periodic wave function is written as

$$Q_{w_i}^{i*}(t^*) = Q^* f_j(t^*, \operatorname{Ku}), \qquad 0 \le t^* < 4\operatorname{Ku},$$
 (A.9)

$$Q_{w_i}^{i*}(t^*) = Q_{w_i}^{i*}(t^* - 4\mathrm{Ku}), \qquad t^* \ge 4\mathrm{Ku},$$
 (A.10)

where we see that the parameters controlling the governing equations can be reduced to few

dimensionless parameters. The Kubo number Ku plays a central role.

621 Acknowledgments

This work was partially supported by the European Commission, through project MAR-SOLUT (grant H2020-MSCA-ITN-2018); by the Spanish Ministry of Economy and Competitiveness, through project MONOPOLIOS (RTI 2018-101990-B-100, MINECO/FEDER);

and by the Catalan Agency for Management of University and Research Grants through FI

611

616

⁶²⁵ 2017 (EMC/2199/2017). The MRST model and simulation data are available on Zenodo,

with doi: 10.5281/zenodo.4095594.

628 References

- Abidin, A., T. Puspasari, and W. Nugroho (2012), Polymers for Enhanced Oil Recovery 629 Technology, Procedia Chemistry, doi:10.1016/j.proche.2012.06.002. 630 Abriola, L. M., and G. F. Pinder (1985), A Multiphase Approach to the Modeling of Porous 631 Media Contamination by Organic Compounds: 1. Equation Development, Water Re-632 sources Research, doi:10.1029/WR021i001p00011. 633 Allen, M. B. (1985), Numerical modelling of multiphase flow in porous media, Advances in 634 Water Resources, doi:10.1016/0309-1708(85)90062-4. 635 Amundsen, H., G. Wagner, U. Oxaal, P. Meakin, J. Feder, and T. Jøssang (1999), Slow two-636 phase flow in artificial fractures: Experiments and simulations, Water Resources Research, 637 doi:10.1029/1999WR900147. 638 Arshadi, M., A. Zolfaghari, M. Piri, G. A. Al-Muntasheri, and M. Sayed (2017), The ef-639 fect of deformation on two-phase flow through proppant-packed fractured shale sam-640 ples: A micro-scale experimental investigation, Advances in Water Resources, doi: 641 10.1016/j.advwatres.2017.04.022. 642 Arshadi, M., M. Khishvand, A. Aghaei, M. Piri, and G. A. Al-Muntasheri (2018), Pore-Scale 643 Experimental Investigation of Two-Phase Flow Through Fractured Porous Media, Water Resources Research, doi:10.1029/2018WR022540. 645 Bachu, S. (2000), Sequestration of co2 in geological media: criteria and approach for site 646 selection in response to climate change, Energy Conversion and Management, 41(9), 953 647 - 970, doi:https://doi.org/10.1016/S0196-8904(99)00149-1. 648 Bagtzoglou, A. C., and P. M. Oates (2007), Chaotic advection and enhanced ground-649 water remediation, Journal of Materials in Civil Engineering, doi:10.1061/(ASCE) 650 0899-1561(2007)19:1(75). 651 Bertels, S. P., D. A. DiCarlo, and M. J. Blunt (2001), Measurement of aperture distri-652 bution, capillary pressure, relative permeability, and in situ saturation in a rock frac-653 ture using computed tomography scanning, Water Resources Research, doi:10.1029/ 654 2000WR900316. Bhatia, R., and C. Davis (2000), A better bound on the variance, American Mathematical 656 Monthly, doi:10.2307/2589180. 657 Bianchi, M., C. Zheng, C. Wilson, G. R. Tick, G. Liu, and S. M. Gorelick (2011a), Spatial 658 connectivity in a highly heterogeneous aquifer: From cores to preferential flow paths, Wa-659 ter Resources Research, 47(5), doi:10.1029/2009WR008966. 660 Bianchi, M., C. Zheng, C. Wilson, G. R. Tick, G. Liu, and S. M. Gorelick (2011b), Spatial 661 connectivity in a highly heterogeneous aquifer: From cores to preferential flow paths, Wa-662 ter Resources Research, 47(5). 663 Bolster, D., M. Dentz, and J. Carrera (2009), Effective two-phase flow in heterogeneous 664 media under temporal pressure fluctuations, Water Resources Research, 45(5), doi: 665 10.1029/2008WR007460. 666 Boulding, J. R. (1996), EPA environmental assessment sourcebook, CRC Press. 667 Brooks, R. H., and A. T. Corey (1966), Properties of Porous Media Affecting Fluid Flow, 668 doi:10.3758/s13423-011-0165-y. 669 Brown, H. W. (1951), Capillary Pressure Investigations, Journal of Petroleum Technology, 670 doi:10.2118/951067-G. 671 Castiglione, P. (2000), Diffusion coefficients as function of Kubo number in random fields, 672 Journal of Physics A: Mathematical and General, doi:10.1088/0305-4470/33/10/302. 673 Celia, M. A., S. Bachu, J. M. Nordbotten, and K. W. Bandilla (2015), Status of CO2 storage 674 in deep saline aquifers with emphasis on modeling approaches and practical simulations, 675 doi:10.1002/2015WR017609. 676 Chen, Z., G. Huan, and Y. Ma (2006), Computational methods for multiphase flows in porous 677 media (Vol. 2). 678 Coats, K. H. (2003), IMPES stability: Selection of stable timesteps, SPE Journal, doi:10. 679 2118/84924-PA. 680
 - -25-

681	Courant, R., K. Friedrichs, and H. Lewy (1928), On the partial difference equations of mathematical physics. <i>Math. Ann.</i>
683	Craig, F. F. (1971), The Reservoir Engineering Aspects of Waterflooding, in <i>Climate Change</i>
684	2013 - The Physical Science Basis, doi:10.1017/CBO9781107415324.004.
685	Craig, F. F., J. Sanderlin, D. Moore, and T. Geffen (1957), A laboratory study of gravity seg-
686	regation in frontal drives, Aime.
687	Davies, R. J., S. Almond, R. S. Ward, R. B. Jackson, C. Adams, F. Worrall, L. G. Herring-
688	shaw, J. G. Gluyas, and M. A. Whitehead (2014), Oil and gas wells and their integrity:
689	Implications for shale and unconventional resource exploitation, doi:10.1016/j.marpetgeo. 2014 03 001
601	De Barros F. D. Fernàndez-Garcia D. Bolster, and X. Sanchez-Vila (2013). A risk-based
692	probabilistic framework to estimate the endpoint of remediation. Concentration rebound
693	by rate-limited mass transfer, <i>Water Resources Research</i> , 49(4), 1929–1942.
694	De Dreuzy, J. R., J. Carrera, M. Dentz, and T. Le Borgne (2012), Asymptotic dispersion for
695	two-dimensional highly heterogeneous permeability fields under temporally fluctuating
696	flow, Water Resources Research, 48(1), doi:10.1029/2011WR011129.
697	de Marsily, G. (1985), Flow and transport in fractured rocks: connectivity and scale effect, in
698	International symposium on the hydrogeology of rocks of low permeability. International
699	Association of Hydrogeologists, Iucson, AZ (USA), pp. 267–277.
700	Demond, A. H., and P. V. Roberts (1991), Effect of interfacial forces on two-phase capillary
701	pressure—saturation relationships, <i>Water Resources Research</i> , doi:10.1029/90WR02408.
702	Dentz, M., and J. Carrera (2005), Effective solute transport in temporally fluctuating flow
703	through heterogeneous media, <i>Water Resources Research</i> , 41(8).
704	Di Dato, M., F. P. de Barros, A. Fiori, and A. Bellin (2018), Improving the Efficiency
705	of 5-D Hydrogeological Mixers: Dilution Ennancement via Coupled Engineering-
706 707	10.1002/2017WR022116.
708	Dillard, L. A., H. I. Essaid, and W. N. Herkelrath (1997), Multiphase flow modeling of a
709	crude-oil spill site with a bimodal permeability distribution, Water Resources Research,
710	doi:10.1029/97WR00857.
711	Dugan, P. J., J. E. McCray, and G. D. Thyne (2003), Influence of a solubility-enhancing agent
712	(cyclodextrin) on NAPL-water partition coefficients, with implications for partitioning
713	tracer tests, Water Resources Research, doi:10.1029/2002WR001672.
714	Edery, Y., A. Guadagnini, H. Scher, and B. Berkowitz (2014), Origins of anomalous trans-
715	port in heterogeneous media: Structural and dynamic controls, Water Resources Research,
716	50(2), 1490–1505.
717	Engel, J., E. Herrmann, and T. Gasser (1994), An iterative bandwidth selector for kernel esti-
718	mation of densities and their derivatives, Journal of Nonparametric Statistics, 4(1), 21–34,
719	doi:10.1080/10485259408832598.
720	Essaid, H. I., B. A. Bekins, and I. M. Cozzarelli (2015), Organic contaminant transport and
721	fate in the subsurface: Evolution of knowledge and understanding, <i>Water Resources Re</i> -
722	search, doi:10.1002/2015WR017121.
723	Falta, R. W., C. M. Lee, S. E. Brame, E. Roeder, J. T. Coates, C. Wright, A. L. Wood, and
724	C. G. Enfield (1999), Field test of high molecular weight alcohol flushing for subsur-
725	face nonaqueous phase liquid remediation, <i>Water Resources Research</i> , doi:10.1029/
720	Favers F I and T A Hewett (1992) A review of current trends in petroleum reservoir de-
728	scription and assessment of the impacts on oil recovery. Advances in Water Resources.
729	doi:10.1016/0309-1708(92)90002-J.
730	Fernàndez-Garcia, D., P. Trinchero, and X. Sanchez-Vila (2010). Conditional stochas-
731	tic mapping of transport connectivity, <i>Water Resources Research</i> , 46(10), doi:10.1029/
732	2009WR008533.
733	Fernàndez-Garcia, D., X. Sánchez-Vila, and T. H. Illangasekare (2002), Convergent-flow
734	tracer tests in heterogeneous media: combined experimental-numerical analysis for deter-

735 736	mination of equivalent transport parameters, <i>Journal of Contaminant Hydrology</i> , 57(1), 129 – 145, doi:https://doi.org/10.1016/S0169-7722(01)00214-5.
737	Forsyth, P. A., and P. H. Sammon (1986), Practical considerations for adaptive implicit
738	methods in reservoir simulation, <i>Journal of Computational Physics</i> , doi:10.1016/
739	Class P. J. M. J. Nichell and J. Verrington (1008). A modified invesion percelation model
740	for low-capillary number immiscible displacements in horizontal rough-walled frac-
741	tures: Influence of local in-plane curvature. <i>Water Resources Research</i> . doi:10.1029/
743	98WR02224.
744	Huling, S. G., and J. W. Weaver (1991), Ground water issue: Dense non-aqueous phase liq-
745	uids, Tech. rep.
746	Jackson, R. B. (2014), The integrity of oil and gas wells, doi:10.1073/pnas.1410786111.
747	Javanbakht, G., M. Arshadi, T. Qin, and L. Goual (2017), Micro-scale displacement of napl
748 749	by surfactant and microemulsion in heterogeneous porous media, Advances in water re- sources, 105, 173–187.
750	Jha, B., L. Cueto-Felgueroso, and R. Juanes (2011), Quantifying mixing in viscously un-
751 752	stable porous media flows, <i>Physical Review E - Statistical, Nonlinear, and Soft Matter Physics</i> , doi:10.1103/PhysRevE.84.066312.
753	Jin, L., S. Hawthorne, J. Sorensen, L. Pekot, B. Kurz, S. Smith, L. Heebink, V. Herde-
754	gen, N. Bosshart, J. Torres, C. Dalkhaa, K. Peterson, C. Gorecki, E. Steadman, and
755	J. Harju (2017), Advancing CO2 enhanced oil recovery and storage in unconventional oil
756	play—Experimental studies on Bakken shales, <i>Applied Energy</i> , doi:10.1016/j.apenergy.
757	2017.10.004.
758	Junnes R and K A Lie (2008) Numerical modeling of multiphase first contact missible
759	flows part 2 front-tracking/streamline simulation <i>Transport in porous media</i> 72(1) 97–
761	120.
762	Kim, M., K. Kim, W. S. Han, J. Oh, and E. Park (2019), Density-Driven Convection in a
763	Fractured Porous Media: Implications for Geological CO 2 Storage, Water Resources
764	Research, doi:10.1029/2019wr024822.
765	Kitanidis, P. K. (1997), Introduction to geostatistics: applications in hydrogeology, Cam-
766	bridge university press.
767	Knudby, C., and J. Carrera (2005), On the relationship between indicators of geostatistical,
768	flow and transport connectivity, <i>Advances in Water Resources</i> , 28(4), 405 – 421, doi:https://doi.org/10.1016/j.advances.2004.00.001
769	//doi.org/10.1016/j.advwatres.2004.09.001.
770	MRST_AD_ An open-source framework for rapid prototyping and evaluation of reservoir
772	simulation problems in Society of Petroleum Engineers - SPE Reservoir Simulation Sym-
773	posium 2015.
774	Kubo, R. (1963), Stochastic Liouville equations, <i>Journal of Mathematical Physics</i> , doi:10.
775	1063/1.1703941.
776	Labolle, E. M., and G. E. Fogg (2001), Role of molecular diffusion in contaminant migration
777 778	and recovery in an alluvial aquifer system, in <i>Dispersion in Heterogeneous Geological Formations</i> , pp. 155–179, Springer.
779	Le Borgne, T., and P. Gouze (2008), Non-Fickian dispersion in porous media: 2. Model val-
780	idation from measurements at different scales, Water Resources Research, doi:10.1029/
781	2007WR006279.
782	Le Borgne, T., M. Dentz, D. Bolster, J. Carrera, JR. De Dreuzy, and P. Davy (2010), Non-
783	fickian mixing: Temporal evolution of the scalar dissipation rate in heterogeneous porous $\frac{1}{2}$
784	media, Advances in Water Resources, 33(12), 1468–14/5.
785	Lester, D., G. Metcalle, M. Irelry, A. Ord, B. Hobbs, and M. Rudman (2009), Lagrangian
786	topology of a periodically reoriented potential now. Symmetry, optimization, and mixing,

⁷⁸⁷ *Physical Review E*, 80(3), 036,208.

788 789 790	Lester, D., M. Rudman, G. Metcalfe, M. Trefry, A. Ord, and B. Hobbs (2010), Scalar dispersion in a periodically reoriented potential flow: Acceleration via lagrangian chaos, <i>Physical Review E</i> , <i>81</i> (4), 046,319.
791 792	Lester, D. R., G. Metcalfe, and M. G. Trefry (2013), Is chaotic advection inherent to porous media flow?, <i>Physical Review Letters</i> , doi:10.1103/PhysRevLett.111.174101.
793 794	Leverett, M. (1939), Flow of Oil-water Mixtures through Unconsolidated Sands, <i>Petroleum Transactions of AIME</i> , doi:10.2118/939149-g.
795 796	Leverett, M. (1941), Capillary Behavior in Porous Solids, <i>Transactions of the AIME</i> , doi: 10.2118/941152-G.
797 798 799	Libera, A., F. P. de Barros, and A. Guadagnini (2017), Influence of pumping operational schedule on solute concentrations at a well in randomly heterogeneous aquifers, <i>Journal of Hydrology</i> , doi:10.1016/j.jhydrol.2016.12.022.
800 801	Lie, K. (2014), An introduction to reservoir simulation using MATLAB: user guide for the Matlab Reservoir Simulation Toolbox (MRST), <i>SINTEF ICT</i> .
802 803 804	Luo, J., M. Dentz, J. Carrera, and P. Kitanidis (2008), Effective reaction parameters for mixing controlled reactions in heterogeneous media, <i>Water Resources Research</i> , doi: 10.1029/2006WR005658.
805 806 807 808 809 810	 Mackay, D. M., and J. A. Cherry (1989), Groundwater contamination: pump-and-treat remediation, <i>Environmental Science & Technology</i>, 23(6), 630–636, doi:10.1021/es00064a001. Martel, R., A. Hébert, R. Lefebvre, P. Gélinas, and U. Gabriel (2004), Displacement and sweep efficiencies in a dnapl recovery test using micellar and polymer solutions injected in a five-spot pattern, <i>Journal of Contaminant Hydrology</i>, 75(1), 1 – 29, doi: https://doi.org/10.1016/j.jconhyd.2004.03.007.
811 812	Mays, D. C., and R. M. Neupauer (2012), Plume spreading in groundwater by stretching and folding, <i>Water Resources Research</i> , doi:10.1029/2011WR011567.
813 814 815	Mazzino, A. (1997), Effective correlation times in turbulent scalar transport, <i>Physical Review</i> <i>E</i> - <i>Statistical Physics, Plasmas, Fluids, and Related Interdisciplinary Topics</i> , doi:10.1103/ PhysRevE.56.5500.
816 817 818 819 820	 Mccray, J. E., and M. L. Brusseau (1999), Cyclodextrin-enhanced in situ flushing of multiple-component immiscible organic liquid contamination at the field scale: Analysis of dissolution behavior, <i>Environmental Science and Technology</i>, doi:10.1021/es980117b. Metcalfe, G., M. Rudman, A. Brydon, L. Graham, and R. Hamilton (2006), Composing chaos: An experimental and numerical study of an open duct mixing flow, <i>AIChE Journal</i>, 52(1), 9–28.
821 822 823 824 825	 Metcalfe, G., D. Lester, A. Ord, P. Kulkarni, M. Rudman, M. Trefry, B. Hobbs, K. Regenaur-Lieb, and J. Morris (2010), An experimental and theoretical study of the mixing character-istics of a periodically reoriented irrotational flow, <i>Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences</i>, 368(1918), 2147–2162.
826 827 828	Neupauer, R. M., J. D. Meiss, and D. C. Mays (2014), Chaotic advection and reaction dur- ing engineered injection and extraction in heterogeneous porous media, <i>Water Resources Research</i> , doi:10.1002/2013WR014057.
829 830 831	Nicolaides, C., B. Jha, L. Cueto-Felgueroso, and R. Juanes (2015), Impact of viscous fin- gering and permeability heterogeneity on fluid mixing in porous media, <i>Water Resources</i> <i>Research</i> , doi:10.1002/2014WR015811.
832 833	NRC (2005), Contaminants in the Subsurface: Source Zone Assessment and Remediation, Washington, DC: The National Academies Press., doi:10.17226/11146.
834 835	 Otuno, J. (1990), Mixing, Chaotic Advection, And Turbulence, Annual Review of Fluid Mechanics, doi:10.1146/annurev.fluid.22.1.207. Ottino, L.M., S.C. Japa, and V.S. Chakravarthy (1004). From Devralde's stratehing and
836 837	folding to mixing studies using horseshoe maps, <i>Physics of Fluids</i> , doi:10.1063/1.868309.
838 839 840 841	kinson (2007), New Attempt in Improving Sweep Efficiency at the Mature Koluel Kaike and Piedra Clavada Waterflooding Projects of the S. Jorge Basin in Argentina, doi: 10.2118/107923-MS.

Piscopo, A. N., R. M. Neupauer, and D. C. Mays (2013), Engineered injection and extraction 842 to enhance reaction for improved in situ remediation, Water Resources Research, doi:10. 843 1002/wrcr.20209. 844 Pruess, K., and Y. W. Tsang (1990), On two-phase relative permeability and capillary 845 pressure of rough-walled rock fractures, Water Resources Research, doi:10.1029/ 846 WR026i009p01915. 847 Raffa, P., A. A. Broekhuis, and F. Picchioni (2016), Polymeric surfactants for enhanced oil 848 recovery: A review, doi:10.1016/j.petrol.2016.07.007. 849 Rangel-German, E. R., and A. R. Kovscek (2006), A micromodel investigation of two-850 phase matrix-fracture transfer mechanisms, Water Resources Research, doi:10.1029/ 851 2004WR003918. 852 Rao, P. S. C., M. D. Annable, R. K. Sillan, D. Dai, K. Hatfield, W. D. Graham, A. L. Wood, 853 and C. G. Enfield (1997), Field-scale evaluation of in situ cosolvent flushing for enhanced 854 aquifer remediation, Water resources research, 33(12), 2673-2686. 855 Reid, R. C., J. M. Prausnitz, and T. K. Sherwood (1997), The properties of gases and liquids, 856 4th ed. ed., McGraw-Hill, New York. 857 Ren, B., and I. Duncan (2019), Modeling oil saturation evolution in residual oil zones: Im-858 plications for CO2 EOR and sequestration, Journal of Petroleum Science and Engineering, 859 doi:10.1016/j.petrol.2019.02.072. Renard, P., and D. Allard (2013a), Connectivity metrics for subsurface flow and transport, 861 Advances in Water Resources, 51, 168 - 196, doi:https://doi.org/10.1016/j.advwatres. 862 2011.12.001, 35th Year Anniversary Issue. 863 Renard, P., and D. Allard (2013b), Connectivity metrics for subsurface flow and transport, 864 Advances in Water Resources, 51, 168–196. Rodríguez-Escales, P., D. Fernàndez-Garcia, J. Drechsel, A. Folch, and X. Sanchez-Vila 866 (2017), Improving degradation of emerging organic compounds by applying chaotic ad-867 vection in Managed Aquifer Recharge in randomly heterogeneous porous media, Water 868 Resources Research, doi:10.1002/2016WR020333. 869 Saaltink, M. W., V. Vilarrasa, F. De Gaspari, O. Silva, J. Carrera, and T. S. Rötting (2013), A method for incorporating equilibrium chemical reactions into multiphase flow models 871 for co2 storage, Advances in Water Resources, 62, 431 - 441, doi:https://doi.org/10.1016/j. 872 advwatres.2013.09.013, computational Methods in Geologic CO2 Sequestration. 873 Sammon, P. H. (1988), Analysis of upstream differencing, SPE Reservoir Engineering (Soci-874 ety of Petroleum Engineers), doi:10.2118/14045-PA. 875 Satkin, R. L., and P. B. Bedient (1988), Effectiveness of various aquifer restoration schemes 876 under variable hydrogeologic conditions, Groundwater, 26(4), 488–498. 877 Schwille, F. (1981), Groundwater pollution in porous media by fluids immiscible with water, 878 in Studies in Environmental Science, doi:10.1016/S0166-1116(08)71937-X. 879 Silliman, S., and A. Wright (1988), Stochastic analysis of paths of high hydraulic conductivity in porous media, Water Resources Research, 24(11), 1901–1910. 881 Sleep, B. E., and J. F. Sykes (1989), Modeling the transport of volatile organics in variably 882 saturated media, Water Resources Research, doi:10.1029/WR025i001p00081. 883 Sleep, B. E., and J. F. Sykes (1993a), Compositional simulation of groundwater contamination by organic compounds: 1. Model development and verification, Water Resources Research, doi:10.1029/93WR00283. 886 Sleep, B. E., and J. F. Sykes (1993b), Compositional simulation of groundwater contam-887 ination by organic compounds: 2. Model applications, Water Resources Research, doi: 888 10.1029/93WR00284. 889 Smalley, P., A. Ross, C. Brown, T. Moulds, and M. Smith (2009), Reservoir Technical Lim-890 its: A Framework for Maximizing Recovery From Oil Fields, SPE Reservoir Evaluation & 891 Engineering, doi:10.2118/109555-PA. 892 Soga, K., J. Page, and T. Illangasekare (2004), A review of napl source zone remediation 893 efficiency and the mass flux approach, Journal of Hazardous Materials, 110(1), 13 - 27, 894 doi:https://doi.org/10.1016/j.jhazmat.2004.02.034. 895

896 897	Stroo, H. F., A. Leeson, J. A. Marqusee, P. C. Johnson, C. H. Ward, M. C. Kavanaugh, T. C. Sale, C. J. Newell, K. D. Pennell, C. A. Lebrón, and M. Unger (2012), Chlorinated ethene source remediation: Lessons learned <i>Environmental Science & Technology</i> 46(12)
898	6438–6447, doi:10.1021/es204714w, pMID: 22558915.
900	Sánchez-Vila, X., P. M. Meier, and J. Carrera (1999), Pumping tests in heterogeneous
901	aquifers: An analytical study of what can be obtained from their interpretation using ja- cob's method. <i>Water Resources Research</i> 35(4), 943–952, doi:10.1029/1999WR900007
002	Trefry M G D R Lester G Metcalfe A Ord and K Regenauer-Lieb (2012a) Toward en-
904	hanced subsurface intervention methods using chaotic advection. <i>Journal of Contaminant</i>
905	Hydrology, 127(1), 15 - 29, doi:https://doi.org/10.1016/i.iconhyd.2011.04.006, gO10:
906	Groundwater Quality Management in a Rapidly Changing World.
907	Trefry, M. G., D. R. Lester, G. Metcalfe, A. Ord, and K. Regenauer-Lieb (2012b), Toward en-
908	hanced subsurface intervention methods using chaotic advection, <i>Journal of Contaminant</i> Hydrology, 127(1-4), 15–29
909	Trinchero P. Y. Sánchez Vila and D. Fernàndez Garcia (2008). Point to point connectiv
910	ity, an abstract concept or a key issue for risk assessment studies?, <i>Advances in Water Re-</i>
912	<i>sources</i> , <i>31</i> (12), 1/42 – 1/53, doi:https://doi.org/10.1016/j.advwatres.2008.09.001.
913	Vilarrasa, V., D. Bolster, S. Olivella, and J. Carrera (2010), Coupled hydromechanical mod-
914	eling of co2 sequestration in deep saline aquifers, <i>International Journal of Greenhouse</i>
915	of the EGU General Assembly 2000
916	Vlad M. E. Spineapu, I. Micguigh and P. Balacou (2001). Diffusion in biased turbulance
917	Physical Review F 63(6) 066 304
010	Wan J T K Tokunaga C F Tsang and G S Bodyarsson (1996) Improved glass micro-
920	model methods for studies of flow and transport in fractured porous media. <i>Water Re-</i>
921	sources Research, doi:10.1029/96WR00755.
922	Welkenhuysen, K., J. Rupert, T. Compernolle, A. Ramirez, R. Swennen, and K. Piessens
923	(2017), Considering economic and geological uncertainty in the simulation of realistic
924	investment decisions for CO2-EOR projects in the North Sea, Applied Energy, doi:10.
925	1016/j.apenergy.2016.10.105.
926	Western, A. W., G. Blöschl, and R. B. Grayson (2001), Toward capturing hydrologically sig-
927	nificant connectivity in spatial patterns, Water Resources Research, 37(1), 83–97, doi:
928	10.1029/2000WR900241.
929	Wipfler, E. L., M. I. Van Dijke, and S. E. Van Der Zee (2004), Three-phase flow analysis of
930	dense nonaqueous phase liquid infiltration in horizontally layered porous media, <i>Water</i>
931	Resources Research, doi:10.1029/2003 w R002948.
932	wu, Y. P. S. Huyakorn, and N. S. Park (1994), A vertical equilibrium model for assessing
933 934	Resources Research, doi:10.1029/93WR03412.
935	Yanosik, J., and T. McCracken (1979), A Nine-Point, Finite-Difference Reservoir Simulator
936 937	for Realistic Prediction of Adverse Mobility Ratio Displacements, <i>Society of Petroleum</i> <i>Engineers Journal</i> , doi:10.2118/5734-PA.
938	Yousefvand, H., and A. Jafari (2015), Enhanced oil recovery using polymer/nanosilica, Pro-
939	cedia Materials Science, 11, 565 - 570, doi:https://doi.org/10.1016/j.mspro.2015.11.068,
940	5th International Biennial Conference on Ultrafine Grained and Nanostructured Materials,
941	UFGNSM15.
942	Zhang, P., S. L. Devries, A. Dathe, and A. C. Bagtzoglou (2009), Enhanced mixing and
943	plume containment in porous media under time-dependent oscillatory flow, Environmental
944	Science and Technology, doi:10.1021/es900854r.
945	Zheng, C., and S. M. Gorelick (2003), Analysis of solute transport in flow fields influenced
946	by preterential flowpaths at the decimeter scale, <i>Groundwater</i> , $41(2)$, $142-155$, doi:10.
947	$\frac{1111}{3} \frac{1}{43} - \frac{1}{3} \frac{1}{43} \frac{1}{3} \frac{1}{$

Zhong, L., A. Mayer, and R. J. Glass (2001), Visualization of surfactant-enhanced nonaque ous phase liquid mobilization and solubilization in a two-dimensional micromodel, *Water*

⁹⁵⁰ *Resources Research*, doi:10.1029/2000WR900300.