

# Are remote sensing evapotranspiration models reliable across South American climates and ecosystems?

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## Abstract

Many remote sensing-based evapotranspiration (RSBET) algorithms have been proposed in the past decades and evaluated using flux tower data, mainly over North America and Europe. Model evaluation across South America has been done locally or using only a single algorithm at a time. Here, we provide the first evaluation of multiple RSBET models, at a daily scale, across a wide variety of biomes, climate zones, and land uses in South America. We used meteorological data from 25 flux towers to force four remote sensing based ET models: Priestley & Taylor Jet Propulsion Laboratory (PT-JPL), Global Land Evaporation Amsterdam Model (GLEAM), Penman-Monteith Mu model (PM-MOD), and Penman-Monteith Nagler model (PM-VI). ET was predicted satisfactorily by all four models, with correlations consistently higher ( $R^2 > 0.6$ ) for GLEAM and PT-JPL, and PM-MOD and PM-VI presenting overall better responses in terms of PBIAS (-10

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## Key Points:

- Four remote sensing *ET* models were evaluated using 25 flux towers from across South America
- GLEAM and PT-JPL provided a significantly greater number of daily outputs
- Comparisons with flux tower-based *ET* showed that GLEAM and PT-JPL produced higher correlations whereas *RMSE* was similar for all models
- Performance of all models is reduced in dry environments

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## Abstract

Many remote sensing-based evapotranspiration (RSBET) algorithms have been proposed in the past decades and evaluated using flux tower data, mainly over North America and Europe. Model evaluation across South America has been done locally or using only a single algorithm at a time. Here, we provide the first evaluation of multiple RSBET models, at a daily scale, across a wide variety of biomes, climate zones, and land uses in South America. We used meteorological data from 25 flux towers to force four remote sensing based *ET* models: Priestley–Taylor Jet Propulsion Laboratory (PT-JPL), Global Land Evaporation Amsterdam Model (GLEAM), Penman–Monteith Mu model (PM-MOD), and Penman–Monteith Nagler model (PM-VI). *ET* was predicted satisfactorily by all four models, with correlations consistently higher ( $R^2 > 0.6$ ) for GLEAM and PT-JPL, and PM-MOD and PM-VI presenting overall better responses in terms of PBIAS ( $-10 < PBIAS < 10\%$ ). As for PM-VI, this outcome is expected, given that the model requires calibration with local data. Model skill seems to be unrelated to land-use but instead presented some dependency on biome and climate, with the models producing the best results for wet to moderately wet environments. Our findings show the suitability of individual models for a number of combinations of land cover types, biomes, and climates. At the same time, no model outperformed the other for all conditions, and all models presented poor skills for sites in certain conditions, which emphasizes the need of adapting individual algorithms to take into account intrinsic characteristics of climates and ecosystems in South America.

## 1 Introduction

Land evaporation, or evapotranspiration (*ET*), is the phenomenon by which water is converted from a liquid into its vapor phase over land. It plays a significant role in the modulation of global climate feedbacks being a key driver of the Earth's carbon, energy, and water cycles at local, regional, and global scales [Cao *et al.*, 2010; Tong *et al.*, 2017; Khosa *et al.*, 2019; Valle Júnior *et al.*, 2020; de Oliveira *et al.*, 2021]. *In situ ET* measurements can be obtained from micro-meteorological methods (e.g., eddy covariance, scintillometry, or Bowen ratio method) and those derived from the soil water balance (e.g., directly using lysimeters, or from changes in profile soil moisture content obtained gravimetrically, from neutron probes, or capacitance-based soil water monitoring equipment). Besides, plant physiological techniques such as sap flow methods, provide direct estimates of transpiration [Verhoef and Campbell, 2006; Allen *et al.*, 2011; Fisher *et al.*, 2011], but only the micro-meteorological methods provide *ET* data at the field to landscape (e.g., scintillometry) scale. Over the past three decades, eddy covariance systems have become the state-of-the-art and standard *in situ* method to quantify land surface energy and mass fluxes for different types of ecosystems [Restrepo-Coupe *et al.*, 2013; Rodrigues *et al.*, 2016; Campos *et al.*, 2019; Wang *et al.*, 2020]. However, these techniques estimate fluxes for areas of relatively limited spatial dimensions ( $\sim 1 \text{ km}^2$ ) depending on the heterogeneity of the landscape, and they are affected by specific local conditions, such as the occurrence of advection across sharp contrasts in vegetation and/or irrigation conditions, and those caused by topographic features, like cold air drainage for sloping terrain [Allen *et al.*, 2011; Mutti *et al.*, 2019; Mauder *et al.*, 2020; Rahimzadegan and Janani, 2019; Mauder *et al.*, 2020; Rwasoka *et al.*, 2011].

During the 1990s and 2000s, remote sensing based *ET* (RSBET) algorithms, using information from visible, near-infrared, and thermal infrared bands, were developed, such as the Surface Energy Balance Algorithms for Land (SEBAL, [Bastiaanssen *et al.*, 1998]), Simplified Surface Energy Balance Index (S-SEBI, Roerink *et al.* [2000]), Surface Balance Energy System (SEBS, Su [2002]), Simplified Surface Energy Balance (SSEB, Senay *et al.* [2007]), and Two-Source Energy Balance Model (TSEB, Norman *et al.* [1995]; Kustas and Norman [1999]). These algorithms were developed for sub-regional applications, with a focus on irrigation or water resources management. Over South America, their predictive skills have been assessed quite extensively, mostly for irrigated cropland [Teixeira *et al.*, 2009; Paiva *et al.*, 2011; Poblete-Echeverría and Ortega-Farias, 2012; Bezerra *et al.*, 2013, 2015;



Olivera-Guerra *et al.*, 2017; Lopes *et al.*, 2019; Mutti *et al.*, 2019]. Studies show that these models perform well when compared to field observations of *ET* [e.g. Poblete-Echeverría and Ortega-Farias, 2012; Teixeira *et al.*, 2009].

Since the late 2000s, algorithms such as PT-JPL [Fisher *et al.*, 2008], PM-MOD [Mu *et al.*, 2007, 2011], and GLEAM [Miralles *et al.*, 2011; Martens *et al.*, 2017] focused on the use of satellite-derived observations to create spatially coherent global *ET* estimates [Fisher *et al.*, 2017]. PT-JPL is at the core of the ECOSTRESS mission [Fisher *et al.*, 2020], while PM-MOD is central to the global terrestrial MODIS *ET* product (MOD16). GLEAM is used for the annual State of the Climate report since 2015 [e.g. Blunden and Arndt, 2020].

Using flux tower data, previous studies conducted in South America evaluated GLEAM and MOD16 [Ruhoff *et al.*, 2013; Moreira *et al.*, 2019; Paca *et al.*, 2019]. However, these studies validated off-the-shelf *ET* datasets generated by these models, not the models themselves. Because such *ET* products are not produced using a common dataset of meteorological variables, a comparative evaluation cannot be made in terms of model structure. Rather, different model skills would be partially linked with the quality of the inputs. A multi-site tropical study, over several continents, validating the PT-JPL model at a regional scale on a monthly basis was presented by Fisher *et al.* [2009]. However, to the best of our knowledge, studies assessing the daily predictive skills have only been conducted at the local scale [Teixeira *et al.*, 2009, 2013; Miranda *et al.*, 2017; Oliveira *et al.*, 2018; Souza *et al.*, 2019].

A major challenge to verify the results of these methods is the scarcity of ground-based observations, due to the uneven spatio-temporal distribution of the *ET* monitoring efforts. As a result, remote sensing *ET* methods are typically evaluated or parameterized using sites located only in North America, Europe [e.g., Ershadi *et al.*, 2014; McCabe *et al.*, 2016; Michel *et al.*, 2016; Xu *et al.*, 2019], Australia [Martens *et al.*, 2016] and East Asia [Jang *et al.*, 2013; Chang *et al.*, 2018; Khan *et al.*, 2018; Li *et al.*, 2019]. For example, Mu *et al.* [2011] proposed improvements to the PM-MOD *ET* global algorithm [Mu *et al.*, 2007], based on comparisons with *ET* measurements from 46 AmeriFlux sites, 45 of them located in USA and Canada. Martens *et al.* [2017] evaluated the GLEAM algorithm with 91 worldwide FLUXNET sites; however, ~65 were located in the USA and in Europe. Therefore, these models might not satisfactorily represent *ET* in sparsely sampled regions with very different climate conditions such as South America, despite this continent representing ca. 12% of the total Earth's terrestrial area.

South America spans two hemispheres, and four major climate zones, from the equator to sub-Antarctic regions, which makes it a geographically unique continent [Goymier, 2017; Trajano, 2019]. This continent hosts biomes ranging from tropical to deciduous forests, that are most sensitive to climate variability [Seddon *et al.*, 2016]. Also, five out of six of the terrestrial biomes not included in satellite-based *ET* algorithm evaluations at a global scale are found in South America (see Section 2.1). Thus, the evaluation of RSBET methods for South America offers an opportunity to reduce the current research gap, in particular at large spatial scales.

FLUXNET provides a common framework for the verification of *ET* algorithms. Nevertheless, the available sites in the FLUXNET2015 database are not evenly distributed around the world [Pastorello *et al.*, 2020]. Validating global models in South America is challenging, mainly because the data from ~90% of its FLUXNET registered sites are not readily available to the scientific community; less than 50% of South American AmeriFlux sites are available for direct access. Additionally, flux towers in woody savannas and evergreen broad-leaf forests account for nearly 65% of all Latin American FLUXNET sites while some of the biomes are not properly represented [Villarreal and Vargas, 2021].

The identification of scientific gaps and the proposed improvements are considered a priority for the future development of *ET* assessment methods from remote sensing [Fisher *et al.*, 2017]. Some of them include merging different *ET*-estimation methods, and the iden-

tification of their sources of uncertainty [Fisher *et al.*, 2017; Zhang *et al.*, 2017; Paca *et al.*, 2019]. Indeed, despite the recent developments of remote sensing *ET* methods, there are still challenges concerning the refinement of those algorithms to remedy the lack of information on specific surface characteristics and fluxes of undersampled climate zones and vegetation types, such as fractional vegetation cover and net radiation, which are a substantial source of uncertainty in global satellite-based *ET* estimates [Ferguson *et al.*, 2010; Vinukollu *et al.*, 2011; Badgley *et al.*, 2015].

Here, we evaluated the predictive skills of four satellite-based *ET* models, designed for regional and continental scale applications, over South America. The main question we seek to answer is whether such models can be applied consistently to reliably capture *ET* in South America. Specific research questions include: (i) are the models capable of correctly estimating *ET* and its components? (ii) are the models predictive skills affected by climate, land cover type or biome?

## 2 Study area, data, and methods

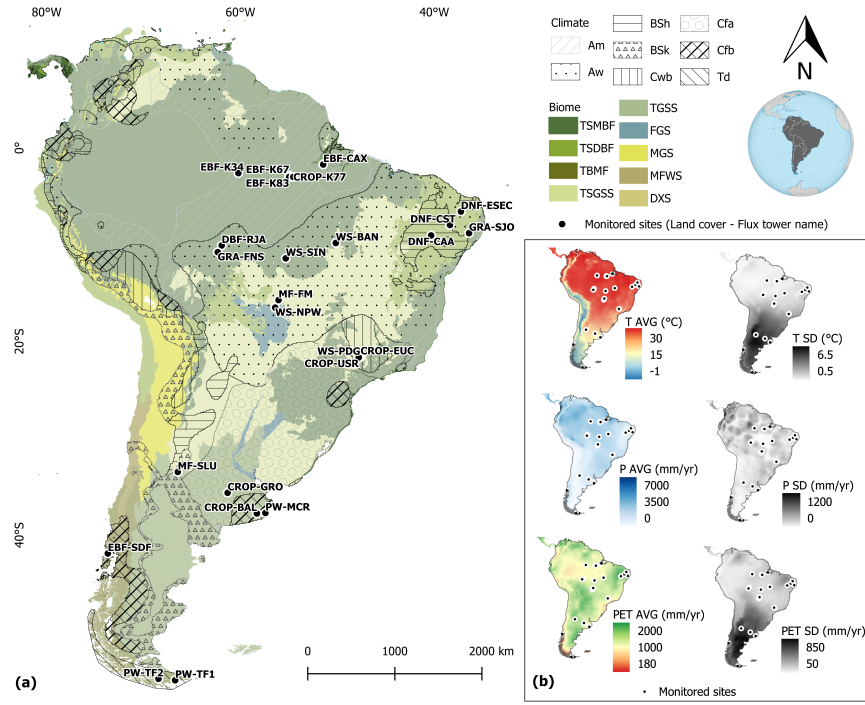
### 2.1 South American biomes, flux tower-based *ET* and meteorological data

The study area encompasses five biomes (Table S1 in the Supporting Material – SM): Tropical & Subtropical Moist Broadleaf Forests (TSMBF); Flooded Grasslands & Savannas (FGS); Tropical & Subtropical Grasslands, Savannas & Shrublands (TSGSS); Tropical & Subtropical Dry Broadleaf Forests (TSDBF) and Temperate Broadleaf & Mixed Forests (TBMF) [Olson *et al.*, 2001].

We used daily meteorological data from 25 flux tower sites located across various South American biomes and land cover types to verify the predictive skill of the selected RSBET models (Figure 1a, Table S3 in SM). The time period considered for analysis was determined by the available time-series for each site (Figure S1 in SM). Further information about each biome is provided in SM. Ten sites are from FLUXNET [Pastorello *et al.*, 2020], AmeriFlux networks [Novick *et al.*, 2018] and Large-Scale Biosphere-Atmosphere Experiment in the Amazon (LBA) project [Saleska *et al.*, 2013], while the remaining data were obtained from the respective principal investigators. The spatial patterns of mean annual precipitation (*P*), air temperature (*T*), and potential evapotranspiration (*PET*) show that selected sites encompass a wide variety of climates (Figure 1b).

As we are interested in assessing models, instead of using the EC-measured latent heat flux, *LE*, to represent *ET*, we derived *LE* from the other energy balance fluxes, i.e.  $LE = Rn - G - H$  [Twine *et al.*, 2000; Wilson *et al.*, 2002; Stoy *et al.*, 2013; Fisher *et al.*, 2020], where *Rn* is the net radiation, *G* is the soil heat flux, and *H* is the sensible heat flux. The closure of the energy budget is rarely observed with flux tower measurements [Wilson *et al.*, 2002; Foken, 2008]. Usually, the available energy ( $Rn - G$ ) is greater than ( $LE + H$ ). The imbalances in the surface energy budget, reported here as an energy balance ratio, EBR (i.e.  $(LE + H)/(Rn - G)$ ), range from 0.73 to 1.16 (mean ~0.90) (Table S2, SM). It is paramount that only high-quality data were used to run and assess the models. We computed daily EBR for each site and excluded days with  $EBR < 0.75$  or  $> 1.25$ . Daily averages of meteorological variables were calculated from 30-min or hourly data only when at least 80% of the records per day were available. To obtain daytime and nighttime inputs for the MOD16 model (PM-MOD in this paper), we considered only days with a minimum of twenty 30-min daytime records and twenty during the night. As in Mu *et al.* [2011], the shortwave incoming radiation ( $Rgs \downarrow$ ) was used to distinguish between daytime ( $Rgs \downarrow > 10 \text{ W m}^{-2}$ ) and nighttime ( $Rgs \downarrow < 10 \text{ W m}^{-2}$ ). Regarding the fluxes, we used quality checked raw data that had not been gap-filled.

The quality control procedure described above was not adopted for the SDF, TF1, and TF2 towers (see Figure 1a). At those sites, horizontal advection plays an important role due to extreme weather variations throughout the year [Levy *et al.*, 2020], such that the energy



**Figure 1.** (a) Location of flux tower sites. Land cover types are indicated prior to tower names in the map: Croplands (CROP), Deciduous Needleleaf Forest (DNF), Evergreen Broadleaf Forest (EBF), Grasslands (GRA), Mixed Forest (MF), Permanent Wetland (PW), and Woody Savanna (WS); Biome types [Olson *et al.*, 2001] are indicated by shades of green, yellow and blue on the map (see legend): Tropical & Subtropical Moist Broadleaf Forests (TSMBF); Tropical & Subtropical Dry Broadleaf Forests (TSDBF); Temperate Broadleaf & Mixed Forests (TBMF); Tropical & Subtropical Grasslands, Savannas & Shrublands (TSGSS); Temperate Grasslands, Savannas & Shrublands (TGSS); Flooded Grasslands & Savannas (FGS); Montane Grasslands & Shrublands (MGS); Mediterranean Forests, Woodlands & Scrub (MFWS); Deserts & Xeric Shrublands (DXS); Climates across South America from selected representative sites are indicated by patterns on the map (see legend): Tropical savanna (Aw), Tropical monsoon (Am), Hot semi-arid (BSh), Cold semi-arid (BSk), Humid subtropical (Cfa), Temperate oceanic (Cfb), Dry-winter subtropical highland (Cwb), Polar Tundra (Td) [Peel *et al.*, 2007]. (b) Gridded annual average (AVG) and standard deviation (SD) for temperature ( $T$ ), rainfall ( $P$ ), and potential evapotranspiration ( $PET$ ) across South America and the monitored sites [Harris *et al.*, 2020].

balance closure cannot be diagnosed by EBR, as described above. For instance, the SDF zone is known as an anticyclone pathway between the Pacific and Atlantic oceans, and TF1 and TF2 are located in the extreme southern parts of Patagonia, a region characterized by strong winds. Thus, for those sites, we used  $ET$  derived from measured  $LE$ .

## 2.2 Remote sensing-based vegetation indices

The required vegetation indices (VI) to run the  $ET$  models are the Normalized Vegetation Index ( $NDVI$ ) and Enhanced Vegetation Index ( $EVI$ ). Vegetation Optical Depth ( $VOD$ ) is used in GLEAM.  $NDVI$  and  $EVI$  were derived from the 16-day Level 3 Global product of the MODerate Resolution Imaging Spectroradiometer (MODIS), aboard the Terra and Aqua satellites [Huete *et al.*, 2002]. We used both MODIS VI products, i.e. MOD13Q1 (Terra) and MYD13Q1 (Aqua), at 250 m resolution, to derive 8-day composites of  $NDVI$  and  $EVI$ .

VOD was extracted from the product described in *Moesinger et al.* [2020]. *Fisher et al.* [2008] used the Soil Adjusted Vegetation Index (SAVI) instead of *EVI* because the former does not require the blue reflectance (0.45–0.51  $\mu\text{m}$ ), however, the authors recognize that both indices are very similar. As we are interested in assessing the *ET* models rather than the products resulting from different forcing data, we used *EVI* in Fisher’s model (PT-JPL). Leaf area index (*LAI*) and other vegetation-related variables (e.g., fraction of Absorbed Photosynthetically Active Radiation,  $f_{PAR}$ ) are handled differently in each model. For example, in PT-JPL, *LAI* is obtained from total fractional vegetation cover, whereas in PM-MOD the 1-km MODIS *LAI* (MOD15) product is adopted. The original procedures to obtain those variables were not changed here. The following treatment was applied to the MODIS-derived data. “Good quality” pixels were selected, based on the quality assurance (QA) flags. Next, an autoregressive model was applied to fill in the gaps [Akaike, 1969]. Finally, we implemented a temporal filter to improve the  $f_{PAR}$  and *LAI* time series to reproduce precisely all pre-processing steps of the standard PM-MOD algorithm [Mu et al., 2011]. Filtering of  $f_{PAR}$  and *LAI* allowed for the correction of underestimated values (abrupt and unrealistic drops in the time series) that mostly originate from cloud contamination effects which were not correctly identified in the quality control fields.

## 2.3 Summary of remote sensing-based *ET* models

### 2.3.1 GLEAM

GLEAM is a semi-empirical/process-based model that estimates the total evaporative flux and its components. In this study, version 3 of the algorithm is used [Martens et al., 2017]. The main aspects of the model are described briefly, while for details we refer to *Martens et al.* [2017] and *Miralles et al.* [2011]. The model calculates potential evaporation for four sub-grid land cover fractions: (1) open water, (2) low vegetation, (3) tall vegetation, and (4) bare soil using the *Priestley and Taylor* [1972] equation. For tall and low vegetation cover fractions, potential transpiration is constrained using an empirical evaporative stress factor which is calculated as a function of soil moisture at root-zone depth and microwave *VOD* as described in *Martens et al.* [2017]. *VOD* is a microwave parameter closely linked to vegetation water content [Liu et al., 2013] and in GLEAM it is used to represent phenological changes in vegetation. The soil moisture in the root-zone is calculated with a multi-layer water-balance model forced by precipitation and satellite surface soil moisture retrievals. For bare soil, the evaporative stress factor is calculated as a function of surface soil moisture only whereas for open water evaporation no stress factor is applied. For the tall vegetation cover fraction, rainfall interception loss is estimated with Gash’s analytical model [Gash, 1979; *Miralles et al.*, 2010]. The *ET* is then calculated as the sum of low and tall vegetation transpiration, rainfall interception loss, bare soil evaporation, and open-water evaporation with each weighted by the respective fraction.

### 2.3.2 PT-JPL

The global *ET* model proposed by *Fisher et al.* [2008] is based on the *Priestley and Taylor* equation for potential *ET* (*PET*), which is partitioned into actual plant transpiration, soil evaporation, and interception evaporation, i.e.  $E_{trans} + E_{soil} + E_{int}$ . To reduce potential *ET* to actual *ET*, the PT-JPL model applies ecophysiological constraints based on land surface information such as vegetation properties and humidity/vapor pressure deficit (*VPD*). *Fisher et al.* [2008] used *NDVI* and *SAVI* as a proxy for plant physiological status. We used *EVI* because it provides a better indication of green vegetation cover than *NDVI*, as acknowledged by *Fisher et al.* [2008]. The model partitions available energy using four plant-related constraints: *LAI*, green canopy fraction, plant temperature, and plant moisture. Similar to PM-MOD (see next subsection), vegetation cover, canopy wetness, etc. determine how the available energy is partitioned among the *ET* terms. A unique aspect related to the plant temperature constraint is the determination of an optimal temperature,  $T_{opt}$

[Potter *et al.*, 1993], which corresponds to an optimal stomatal conductance. The latter co-determines  $E_{trans}$ .

### 2.3.3 PM-MOD

The MOD16 ET model (PM-MOD) is based on the Penman-Monteith equation to produce a daily global ET product summing up daytime and nighttime ET [Mu *et al.*, 2011]. In this model, total  $ET$  is partitioned into  $E_{soil}$ ,  $E_{int}$ , and  $E_{trans}$ . To compute  $E_{soil}$ , PM-MOD uses potential soil evaporation and a soil moisture constraint function based on  $VPD$  and air relative humidity ( $RH$ ) [Fisher *et al.*, 2008]. The evaporation of the water intercepted by the canopy,  $E_{int}$ , is calculated using the relevant equations from a revised version of the Biome-BGC model [Thornton, 1998]. The PM-MOD assumes that  $E_{int}$  occurs when the vegetation is covered with water, i.e. when the water cover fraction ( $f_{wet}$ )  $> 0$ , which is constrained by  $RH$  [Mu *et al.*, 2011]. In the PM-MOD model  $f_{wet}$  is calculated as in the PT-JPL model:  $f_{wet}$  is set to 0 if  $RH < 70\%$  and  $f_{wet} = RH^4$  if  $70 < RH < 100\%$  [Running *et al.*, 2019]. The PM-MOD model is designed to allow  $E_{trans}$  to occur during daytime and nighttime, by adding constraints to stomatal conductance for  $VPD$  and minimum temperature, and ignoring constraints relating to high air temperature [Running *et al.*, 2019]. The partitioning of available energy into soil or interception evaporation is based on vegetation cover ( $Fc$ ), which is assumed to be equal to the  $f_{PAR}$  from the MODIS product MOD15A2 [Mu *et al.*, 2011]. Although this method is based on the PM equation, PM-MOD does neither require wind speed nor soil moisture data for the parameterization of aerodynamic and surface resistance. Further details about PM-MOD can be found in Mu *et al.* [2011] and Running *et al.* [2019]. Note that some updates have been implemented in PM-MOD since Mu *et al.* [2011], which can be found in Running *et al.* [2019]. These were also considered here in the implementation of PM-MOD.

### 2.3.4 PM-VI

This model relies upon the hypothesis that  $ET$  is mostly controlled by specific dominant processes, such as transpiration and photosynthesis, hence a good correlation between such processes and  $ET$  is necessary for good model performance [Nagler *et al.*, 2007]. There are several formulations to estimate  $ET$  from VIs [Nagler *et al.*, 2005, 2009]. In this study, we selected the algorithm proposed by Nagler *et al.* [2013], which estimates  $ET$  using the reference crop evapotranspiration,  $ET_o$ , from the FAO-56 Penman-Monteith (PM) equation [Allen *et al.*, 1998], and a crop coefficient,  $K_{cVI}$ , derived from a vegetation index.  $K_{cVI}$  can be calculated in different ways [e.g., Nagler *et al.*, 2005, 2013]. Following Nouri *et al.* [2016] and Oliveira *et al.* [2015],  $K_{cVI}$  was calculated as:

$$K_{cVI} = a \left( 1 - e^{-b \times EVI} \right) - c \quad (1)$$

where  $a$ ,  $b$  and  $c$  are fitted coefficients. We used a parameter optimization tool based on a genetic algorithm to optimise the coefficients to estimate ET values close to the measured ones [Oliveira *et al.*, 2015]. The fitting procedure minimizes the objective function ( $OF$ ) given by the sum of squared differences between tower-based  $ET$  ( $ET_{obs}$ ) and  $ET$  estimates from the models ( $ET_{sim}$ ) at time  $i$ :

$$OF = \sum_{i=1}^n [ET_{obs}(i) - ET_{sim}]^2 \quad (2)$$

This model, herein referred to as PM-VI, has frequently been employed to estimate ET at local and regional scales [Oliveira *et al.*, 2015; Nouri *et al.*, 2016; Jarchow *et al.*, 2017]. Although obtaining  $ET_o$  requires a considerable amount of meteorological variables, the PM-VI implementation is easier and has a lower computational cost compared to other mod-



els. Unlike the other three models, PM-VI requires the calibration of the fitting coefficients, which can be a major issue for regions where  $ET$  and  $VI$  are poorly correlated or when correlations change over time [Chong *et al.*, 1993]. To calibrate the fitting coefficients, we randomly selected 20% of the available data at each site and used the remaining 80% to validate the model.

## 2.4 Quantifying model reliability

The model predictive skill was visually evaluated with scatter plots of measured versus modelled  $ET$ , as well as through the coefficient of determination ( $R^2$ ), root mean square error ( $RMSE$ ), percent bias ( $PBIAS$ ), concordance correlation coefficient ( $\rho$ ), slope ( $m$ ), and intercept ( $b$ ) of the linear regression. The data used in the analysis were filtered for rainy days ( $P > 0.5$  mm). Our analysis proceeded from a general (no distinction among sites) to a site-by-site and group level analysis, i.e. per biome, climate, or land use. As the number of flux towers, and record length for each tower, within the groups was different, a sampling procedure was adopted to compute the per-group validation metrics: (i) skill metrics for each group were calculated using samples from each tower within the group. The sample size  $N$  was defined as half of the record length of the shortest available tower record within the corresponding group; (ii) the samples were taken by randomly sampling the pool of available data within each tower dataset; (iii) this procedure was repeated 1000 times to get the mean and standard deviation ( $SD$ ) of each metric per group. To establish a relationship between model predictive skill and water availability at individual tower sites, we obtained the aridity index ( $AI = P/ET_o$ ) from the global dataset provided by Trabucco and Zomer [2019]. For many tower sites, the available meteorological data (even from nearby meteorological stations) were not sufficient to provide a reliable  $AI$ ; hence the choice for a global dataset.

## 3 Results

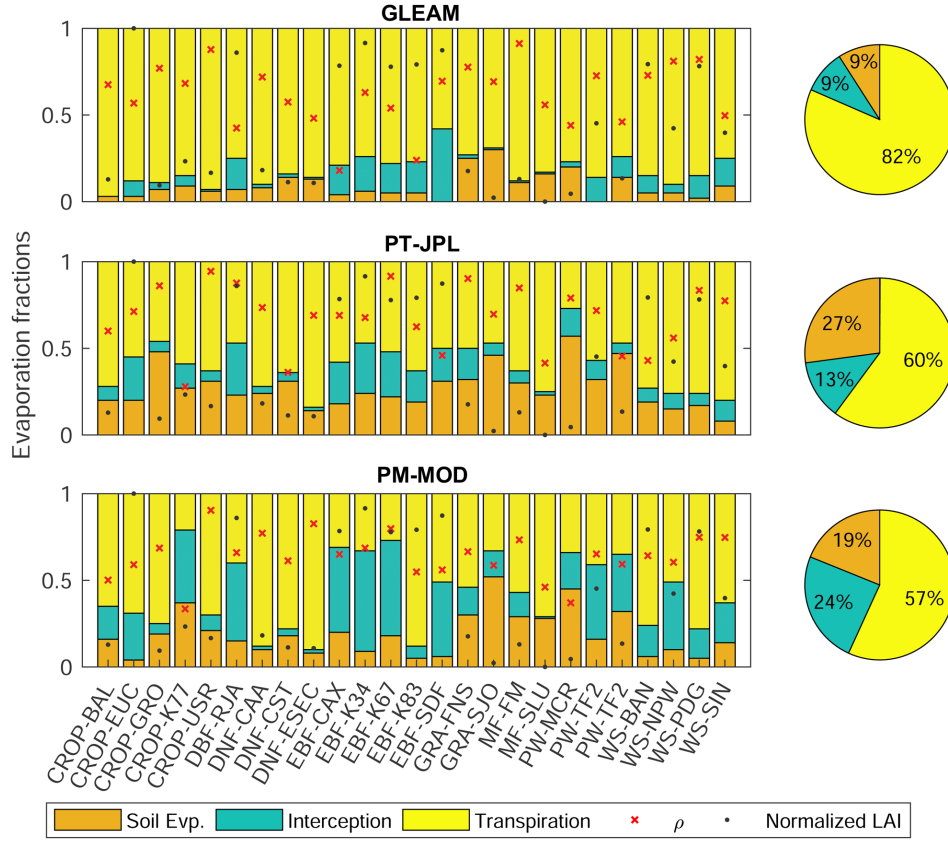
### 3.1 ET partitioning

Partitioning of  $ET$  among the three components ( $E_{soil}$ ,  $E_{int}$ ,  $E_{trans}$ ) exhibited more variation for the PT-JPL and PM-MOD models. On average,  $E_{trans}$  accounted for 60% (PT-JPL) and 56% (PM-MOD) of  $ET$  but, across sites, it presented a smaller range (30% to 85%) for PT-JPL than for PM-MOD (20 to 90%) (Figure 2). GLEAM  $E_{trans}$  accounted for 82% of  $ET$  on average, varying between 60% and 95% across sites. Average interception across sites reached 9% (GLEAM), 13% (PT-JPL), and 24% (PM-MOD) of total  $ET$ .  $E_{int}$  fractions range were similar for GLEAM and PT-JPL ( $SD \approx 9\%$ ), whereas PM-MOD  $E_{int}$  varied more among sites ( $SD = 18\%$ ).  $E_{int}$  was often correlated with  $LAI$ , especially for the GLEAM estimates ( $R^2 = 0.57$ , Figure S2 in SM). PT-JPL  $E_{soil}$  estimates exceeded the other models, particularly for sites with low  $LAI$  values (e.g., ESEC, CST, and USR).

### 3.2 Overall model skills

Since each model requires a different input dataset (Table S3, SM), the data available to run and validate each model varied. GLEAM and PT-JPL provided a significantly greater number of daily outputs: 7301 (GLEAM), 7277 (PT-JPL), 5905 (PM-MOD), and 6638 (PM-VI). The complete data set was used to produce scatter plots of  $ET$  records and model simulations for each location (See Figures S4-S7 in SM). To allow a fair analysis, the results shown in the main text were obtained using data from days that were common across models, resulting in 4718 data points.

To illustrate the relative contribution of each site to the scatter plots in Figure 3, we display the regression lines (light grey lines) between model and tower-based  $ET$  for each tower site, and the mean metrics across individual sites. In general,  $ET$  was reasonably predicted by all models, as suggested by the relatively low spread of most points in the scatter plots, many regression lines close to the 1:1 line, mean root mean square error ( $RMSE$ ) be-



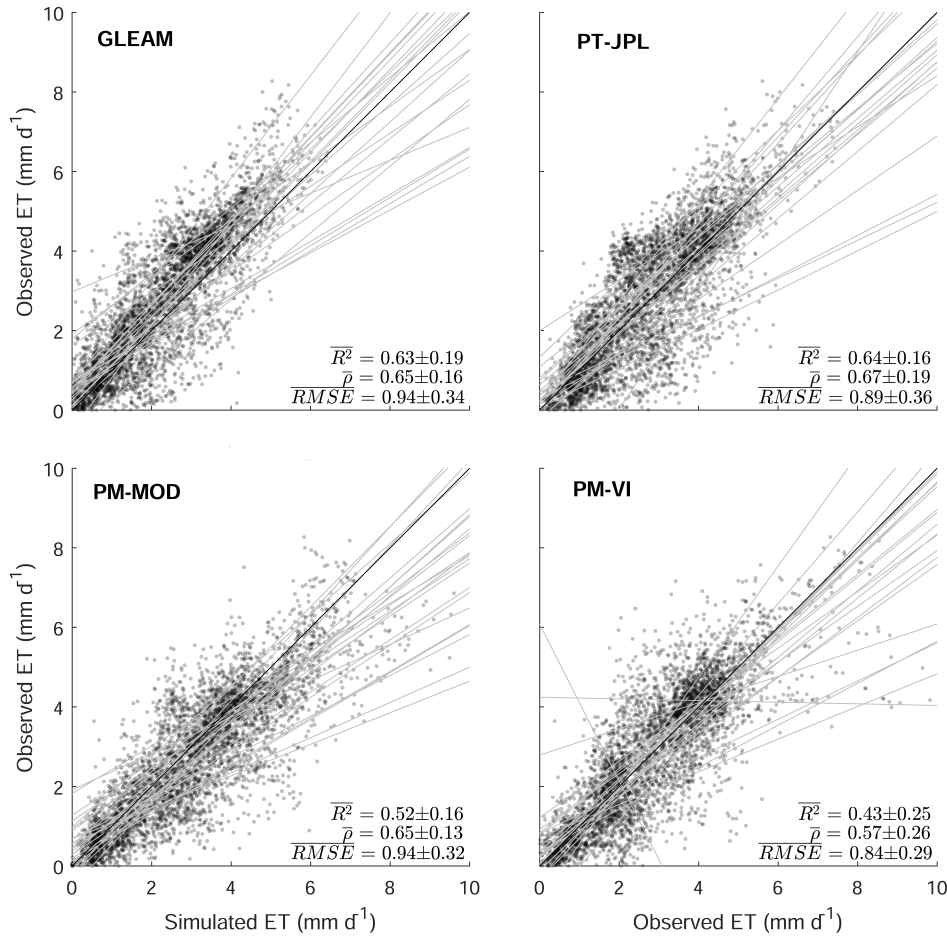
**Figure 2.** Evaporation fractions estimated by the models at each site (stacked bars) and average partitioning of land evaporation per model (pie diagram). Black dots:  $LAI$  scaled between 0 and 1 based on the minimum and maximum values of  $LAI$ . Red  $\times$ : the concordance correlation coefficient.

low  $1 \text{ mm d}^{-1}$  and mean concordance correlation coefficient,  $\bar{\rho}$ , mostly above 0.65 (Figure 3). Nevertheless there is some spread for a few sites, e.g., in the PT-JPL scatter plot that displays a few sites with large bias despite strong overall correlation and  $\rho$ .

The models slightly overestimate  $ET$  as suggested by higher density of points below the 1:1 line, except for GLEAM, which slightly underestimates. Correlations were similar between GLEAM and PT-JPL, with an average value of  $\sim 0.65$  and the highest values at individual sites reaching close to 0.9, as indicated by the standard deviations (0.19 and 0.18, respectively). From Figure 3, it becomes evident that, despite the relatively lower spread of points for PM-VI, this model presented a less consistent performance across towers, as suggested by the contrasting slopes presented by the regression lines in that plot; hence the lower average determination coefficient ( $\bar{R}^2$ ) and  $\bar{\rho}$ . For complementary information, see Figure S3 in SM.

### 3.3 Model skills per biome, land use, and climate

Figure 4 presents  $\rho$ ,  $RMSE$ ,  $PBIAS$ , and  $R^2$  for each model across six biomes, eight land use types, and seven climate classes in South America. Error bars are shown for all metrics, and they represent the standard deviation resulting from the resampling procedure outlined in 2.4. Note that the analysis about the FGS and TBMF biomes are based on one and three towers, respectively. For most biomes,  $RMSE$  and  $R^2$  did not significantly diverge. In



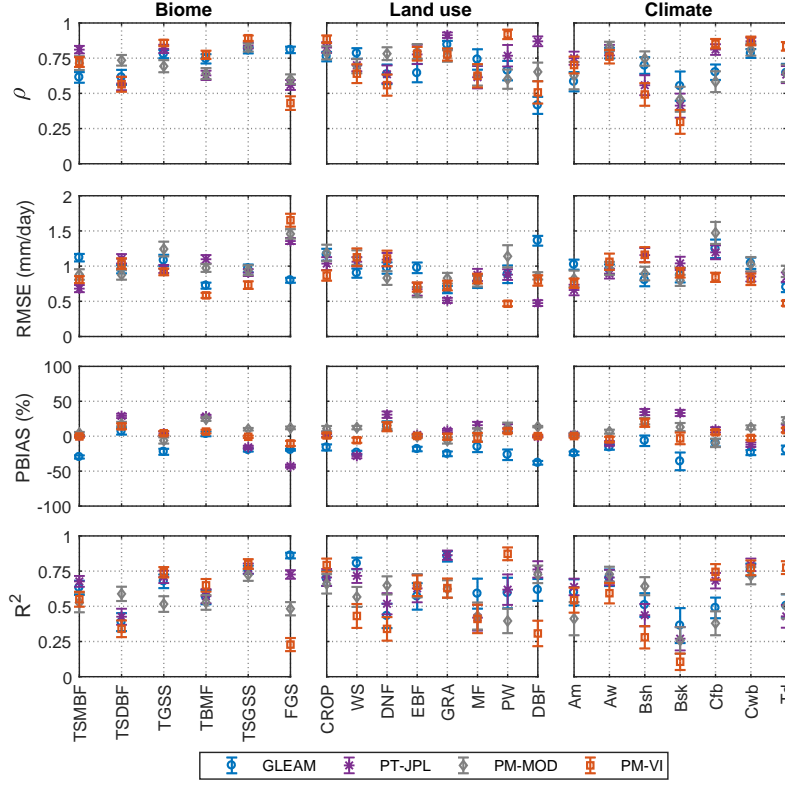
**Figure 3.** Scatter plots of observed vs. simulated daily evapotranspiration at all flux tower sites, for each model. The light grey lines show the regression slope of individual sites. The coefficient of determination ( $R^2$ ), root mean square error ( $RMSE$ ) and percent bias ( $PBIAS$ ) were averaged across towers and are displayed on the plots ( $N = 4,718$ ).

general, TSGSS showed the best overall metrics for all models, while PM-VI in FGS (NPW site) presented the poorest ( $\rho < 0.5$ ,  $RMSE > 1.5 \text{ mm d}^{-1}$ , and  $R^2 < 0.25$ ). Model performance across towers within each biome did not vary much, as suggested by the relatively low range of the error bars for all metrics.

The central panels in Figure 4 provide evidence for the high variability of model predictive skills across different land uses (LU), which suggest that: (i) no model outperforms the others for all LU types, (ii) each model has intrinsic and in some cases exclusive characteristic that makes it more suitable for certain LU. Only for croplands (CROP) we found similar metrics among models ( $\rho \approx 0.8$ ,  $0.8 < RMSE < 1.2 \text{ mm d}^{-1}$ ,  $-20\% < PBIAS < 10\%$ ,  $0.6 < R^2 < 0.8$ ). Conversely, for most LU, the metrics variation is remarkable (e.g., DBF:  $0.4 < \rho < 0.9$ ,  $-50\% < PBIAS < 10\%$ ,  $0.25 < R^2 < 0.80$ ). On average, each model has the best skills for two LU; e.g., ET prediction for GRA and DBF was best with PT-JPL ( $\rho \approx 0.9$ ,  $RMSE \approx 0.5 \text{ mm d}^{-1}$ ,  $PBIAS \approx 0\%$ ,  $R^2 > 0.75$ ) whereas PM-VI presented similar skills for estimation of ET for CROP and PW. Likewise, model skill is related to the climate type. The analysis of  $\rho$  and  $R^2$  over semi-arid regions (BSk and BSh) indicates a relatively poor skill of all models (except PM-MOD for BSh climate). This is in contrast to the



overall good performance over more humid environments (e.g., Aw and Cwb). The greatest divergence among model performances was found for the Polar Tundra (Td) climate zone, for which PM-VI presented the highest  $\rho$  and  $R^2$  (both  $> 0.75$ ), lowest  $RMSE$  ( $\sim 0.5 \text{ mm d}^{-1}$ ) and  $PBIAS$  ( $< 10\%$ ).

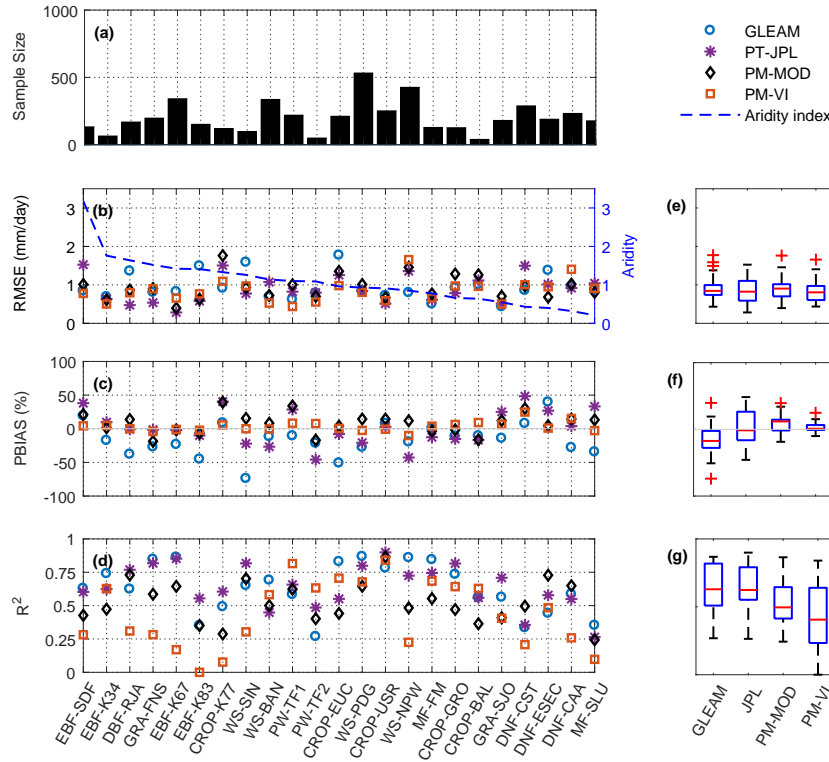


**Figure 4.** Model performance per biome, land use and climate. The error bars represent the standard deviation of the metrics within each class. Biome types: Temperate Grasslands, Savannas & Shrublands (TGSS); Tropical & Subtropical Grasslands, Savannas & Shrublands (TSGSS); Tropical & Subtropical Dry Broadleaf Forests (TSDBF); Tropical & Subtropical Moist Broadleaf Forests (TSMBF); Flooded Grasslands & Savannas (FGS); Mediterranean Forests, Woodlands & Scrub (MFWS). Land use types: Cropland (CROP); Woodland Savanna (WS); Deciduous Needleleaf Forest (DNF); Evergreen Broadleaf Forest (EBF); Open Shrubland (OSH); Mixed Forest (MF); Permanent Wetland (PW); Deciduous Broadleaf Forest (DBF). Climate Zones: Temperate oceanic (Cfb); Tropical savanna (Aw), Tropical monsoon (Am), Hot semi-arid (BSh), Cold semi-arid (BSk), Humid subtropical (Cfa), Dry-winter subtropical highland (Cwb), Polar Tundra (Td).

### 3.4 Individual sites

In this section, we explore the model performance at individual towers. Model skills for all individual sites are depicted in Figure 5. Sites with  $N < 30$  are not discussed here but are considered in the scatter plots shown in the SM (Figures S4-S7). To facilitate the comparison of our results with previous analyses using the same models, only three statistics are shown in Figure 5:  $RMSE$ ,  $PBIAS$ , and  $R^2$ . Other metrics are displayed in the scatter plots

in Figures S4-S7 in the SM. In Figure 5, the metrics for the various towers are displayed in order of increasing aridity (varying from  $\sim 3$  to 0, left to right), as suggested by the AI (see section 2.4). In general, there is a good agreement between the PM-based models in terms of  $RMSE$  and  $PBIAS$ . Despite the oscillations in statistical metrics among sites, especially for PM-VI, there is a general tendency of decreasing  $R^2$  as aridity increases, which is accompanied by an increase in  $PBIAS$ . Conversely,  $RMSE$  does not seem to be affected by aridity; however, the absence of a downward trend in  $RMSE$  actually suggests a higher relative error as  $ET$  decreases. In terms of individual metrics,  $RMSE$  values varied between  $\sim 0.5$  and  $\sim 1.5$  mm d $^{-1}$  for all models, with  $RMSE < 1$  mm d $^{-1}$  for most sites. The boxplots show that  $RMSE$  variation is similar among models, except for PT-JPL which presents the lowest  $RMSE$  (e.g., K67). Figure 5 shows that  $PBIAS$  for PM-VI varies around zero across sites, which is expected given the model requires calibration with local data. However, based on  $R^2$ , it is apparent that this model's skill is quite limited for  $AI > \sim 1.2$ . In general, the PT-based models showed larger biases, with PT-JPL and GLEAM consistently overestimating and underestimating  $ET$ , respectively. In terms of  $R^2$ , the PT-models ranked better than the PM-models for more than  $\sim 50\%$  of the towers.



**Figure 5.** Comparison of statistics of the models in estimating evapotranspiration ( $ET$ ). (a) Sample size ( $N$ ) used to compute the statistics; (b)  $RMSE$  = Root Mean Square Error; (c) Percent Bias ( $PBIAS$ ); (d)  $R^2$  = coefficient of determination. A summary of each model's statistics is depicted in the boxplots: (e)  $RMSE$ ; (f)  $PBIAS$ ; (g)  $R^2$ . Flux towers are arranged according to the aridity index (with aridity increasing from left to right).

## 4 Discussion

We conducted the first multi-remote sensing ET model analysis in South America (SA) using a common set of forcing and validation data located on flux tower sites across a diverse range of land covers, climates, and biomes. Forcing data include both *in situ* (e.g., temperature and radiation) and remote sensing data, mainly related to vegetation (e.g., *LAI* and *EVI*). Many of these sites are not yet available in flux network databases, including sites with land cover (deciduous needle-leaf forests, DNF), a biome (FGS), and two climate types (polar tundra, hot semi-arid) that have not been previously assessed in other regional studies on the performance of satellite-based ET models. Moreover, some classes included here were considered for validation of individual models only (e.g., semi-arid and tropical climate types, TSDBF biome, etc).

Generally, model predictive skill over SA resembles what has been reported for other continents, including satisfactory values of coefficient of determination ( $R^2 > 0.6$ ) of the models (except PM-VI) for most validation sites, and consistently better results for the GLEAM and PT-JPL models, with *RMSE* ranging from  $\sim 0.5$  to  $1.5 \text{ mm d}^{-1}$ . Also, in accordance with previous analysis, GLEAM and PT-JPL presented somewhat higher *RMSE* than PM-MOD, and the performance of all models decreased with increasing aridity [McCabe *et al.*, 2016; Michel *et al.*, 2016]. Nonetheless, the general analysis (Section 3.2) indicates that all models can be used reliably over most of the environmental conditions in SA covered in our study. The analysis across towers and groups (i.e., biome, land use type and climate, section 3.3, Figure 4) identified considerable differences in terms of model skill.

Our results agree with previous studies from [Ershadi *et al.*, 2014; McCabe *et al.*, 2016; Michel *et al.*, 2016; Miralles *et al.*, 2016] who applied PM-MOD, GLEAM (except Ershadi *et al.* [2014]) and PT-JPL to sites located in Africa, Asia, Australia, Europe and Middle East and reported that PM-MOD showed, for most sites, lower correlations with measured ET compared to GLEAM and PT-JPL. Unlike previous analysis, our study agrees with Michel *et al.* [2016] in the sense that model skill seems to be unrelated to land cover. Michel *et al.* [2016] also reported a wide variation of  $R^2$  (0.2–0.8) and *RMSE* ( $0.8\text{--}2 \text{ mm d}^{-1}$ ), for different sites under mixed forests. Conversely, contrasting results between our results and previous studies were found for woodland savanna. While we found  $0.5 < R^2 < 0.8$  and  $0.7 < RMSE < 1.5 \text{ mm d}^{-1}$ , Michel *et al.* [2016] reported  $R^2 < 0.2$  and  $1 < RMSE < 3 \text{ mm d}^{-1}$ .

Overall, our group-wise analysis based on climate agrees with previous studies. For example, the poor model skill found here for the cold semi-arid (Bsk) climate ( $0.1 < R^2 < 0.5$ ) resembles that found by Michel *et al.* [2016] and McCabe *et al.* [2016] for several sites in the United States. While aridity could have played a role here, it could also be caused by the fact that semi-arid sites can often only support sparse canopies. Such canopies present challenges when it comes to the description of aerodynamic transfer for example and radiation partitioning (see e.g. Verhoef and Allen [2000]). Our findings also show a poor to moderate model skill for ET predictions for sites located in the Cfb climate zone, with PM-MOD having the worst performance. Conversely, PM-MOD presented the best predictive skill for the BSh climate, according to most metrics.

Besides the three RSBET models commonly assessed (GLEAM, PT-JPL, and PM-MOD), our analysis included the PM-VI model, which has been validated mostly for cropland or riparian ecosystems [e.g., Nagler *et al.*, 2005, 2009, 2013; Jarchow *et al.*, 2017]. Here, we tested PM-VI for a much wider variety of biomes, climates and land uses, and found a poor predictive skill for several sites with  $AI > 1.2$  (e.g. K67, K77, K83) or  $AI < 0.8$  (e.g., CAA and SLU), even though the model accounts for a site-specific calibration. Considering the good results obtained for  $\sim 50\%$  of the towers and the fact that, compared to the other models, PM-VI has a much simpler implementation, this model does have potential as long as sufficient data are available for calibration or, at least, validation. However, the need for local calibration is a hurdle for its implementation for most regions that are unsampled;

therefore future studies are necessary to investigate which factors are most relevant in the determination of the model fitting coefficients, and to provide distributed reference values for its coefficients (e.g., based on land use dynamics).

We were able to identify a number of probable causes for poor model performance at individual sites, including (i) patch-scale heterogeneities; (ii) “mixed pixels”, i.e. mixed response of different vegetation types within a pixel; (iii) time-lag between  $ET_{obs}$  and  $EVI$ ; (iv) model sensitivity to individual inputs; (v) low correlation between  $ET$  and vegetation indices (see Section 3.0 in the SM for more details). Although we did not verify this in our study, we do not dismiss the possibility that known uncertainties in the estimation of site-specific vegetation characteristics (e.g.,  $f_{PAR}$  and leaf stomatal conductance in the PM-MOD; Ershadi *et al.* [2014]) are also causes of lower model performance. In our study, we used soil heat flux ( $G$ ) which is generally measured below ground (usually at 5–20 cm deep) using soil heat flux plates. It could be argued that not correcting  $G$  for the heat storage between the plate and the soil surface could lead to sub-optimal estimates of  $ET$  when  $LE$  is calculated as the residual of the energy balance, especially for towers where the soil is bare or covered by sparse vegetation, where  $G$  can be relatively large. This, in turn, could lead to the conclusion that the models are performing worse than is actually the case. Although desirable, correcting  $G$  for heat storage is rarely possible due to data unavailability (few sites only measure soil moisture and temperature, which are required to estimate soil heat capacity, and heat storage using the calorimetric method). Moreover, at daily scales and for most sites,  $G$  is either negligible in SA (summer or winter, when the amount of heat stored during the day roughly equals that lost during the night) or represents a minor portion only (spring and autumn) of the energy balance. As detailed and discussed in Section S3.0 and Figure S8 in SM, it is highly unlikely that neglecting such corrections will have affected the results.

There are, however, some issues worth mentioning here. Cause (vi), for instance, is a major issue for PM-VI, as expected because the model is highly dependent on VI dynamics (see Section 2.4) [Nagler *et al.*, 2005]. Regarding cause (iv), the superior performance of the PT models over PM-MOD at most sites is probably linked to uncertainties resulting from the estimation of aerodynamic resistance [Ershadi *et al.*, 2014]. In PM-MOD, the aerodynamic and surface resistances of each  $ET$  component (soil, interception, and transpiration) are parametrized based on biome-specific values of leaf-scale boundary layer conductance, for example [Mu *et al.*, 2011]. Compared to the previous version of PM-MOD [Mu *et al.*, 2007], this new approach resulted in a perceptible improvement only for cropland and deciduous broadleaf forest flux tower sites, whereas for other land uses no meaningful change was reported [Ershadi *et al.*, 2015]. Conversely, PT models are highly dependent on  $R_n$  (causes iv and v); hence they often fail in dry environments (see metrics for  $AI < \sim 0.6$  in Figure 5) where  $ET$  seasonality is dictated by  $P$  and not radiation, or in regions with low  $R_n$  (e.g., TF2). Poor model responses at K77 (cropland, Figure S10 in SM) were attributed to causes (i) and (ii), as remnants of forest and shrubs were identified within the tower footprint and within MODIS pixel. VI products with higher resolution than MODIS exist and have been used to estimate  $ET$  [e.g., Aragon *et al.*, 2018; Fisher *et al.*, 2020]; thus offering a possible solution for causes (i) and (ii). Time lag between  $ET$  and  $EVI$  (cause iii) was identified at EUC, where  $EVI$  followed the decline of  $ET$  after  $\sim 1$ – $2$  months of interval.

Remote sensing based  $ET$  partitioning is expected to present some divergences from ground based measurements. This is the case especially for  $E_{soil}$ , because of the difficulty to obtain remote sensing information on soil characteristics that drive  $E_{soil}$ , such as soil temperature and moisture [Talsma *et al.*, 2018a,b], in particular at high vegetation cover fractions. Globally, transpiration has been reported to account for 57–90% of global  $ET$ , based on *in situ* data and model outputs [Jasechko *et al.*, 2013; Wei *et al.*, 2017; Paschalis *et al.*, 2018]. Although these are global estimates, we expected  $E_{trans}$  to be the largest  $ET$  component also in SA due to its prevailing tropical climate and corresponding vegetation types. Our results show that this was indeed the case for GLEAM with an  $E_{trans}/ET$  ratio of  $\sim 80\%$ , and for PT-JPL and PM-MOD with values of 57 and 60%, respectively. Nonetheless, based

on our findings, model predictive skill in estimating total  $ET$  is not necessarily associated with its ability to partition  $ET$  accurately.

Concomitantly, inconsistencies in  $ET$  partitioning do not necessarily translate into inaccurate model estimates of total  $ET$ ; this depends on the modelling approach. On the one hand, if total  $ET$  results from the sum of  $ET$  components independently, then an under- or overestimation of  $ET$  components can reduce the overall model skill, or reasonable  $ET$  estimates can be achieved as the consequence of an occasional compensation of errors in  $E_{trans}$ ,  $E_{soil}$  and  $E_{int}$ . On the other hand, if the  $ET$  partitioning is derived from the estimate of a proxy value for total  $ET$ , such as available energy (as in PM-MOD and PT-JPL), the  $ET$  partitioning is unlikely to influence the total  $ET$  estimates. Still, good estimates of  $ET$  components are important to differentiate the roles of vegetation and soil, i.e., how they contribute to vertical soil water fluxes and changes in profile soil water content. Reliable knowledge of the distribution between  $E_{soil}$  and  $E_{trans}$  is also important when this information is used in hydrological models to calculate other water balance components, such as runoff.

Ground-based  $ET$  partitioning data are generally not widely available; this also goes for most land cover types included in this study. We compared the models' outputs with field experiment studies that measured one or more  $ET$  components either at the same sites as those used here or within the same region (Table 1).  $ET$  partitioning values derived from GLEAM seem to be more consistent with ground-based information available for tropical rainforests, croplands and grasslands than for wetlands, and mixed and deciduous needle-leaf forests (Table 1). As for PT-JPL; its  $ET$  partitioning fits reasonably well with observations made for both tropical rain- and dry forests. Note that PT-JPL (as well as PM-MOD) constrain  $E_{trans}$  based on  $f_{wet}$ ; hence, compared to GLEAM, transpiration will be lower under high  $RH$  in the model but  $ET$  can be high due to water availability in the soil and intercepted rainfall. Nonetheless, the overall predictive skill of PT-JPL was satisfactory at such sites (Figure 5 and Figure S6 in SM). Regarding PM-MOD, the main inconsistency is the  $E_{inter}$  for tropical forests (Table 1). Despite the wide variability in  $E_{trans}/ET$  among models, their overall predictive skill was satisfactory, that is, not associated with their capability to correctly estimate each  $ET$  component individually (see SM for further discussion). No model was able to consistently capture the  $ET$  partitioning across all sites correctly, which is expected given the uncertainty of each  $ET$  component and the climate and land-cover variability in SA. However, the joint estimates of all models covered totally or partially all field-derived evidence on  $ET$  partitioning. This suggests that continental  $ET$  estimates for understudied regions, such as the SA, would benefit from merging  $ET$  outputs from models that are based on different methods [e.g., Paca *et al.*, 2019].

Despite our efforts to gather as much tower data as possible, with the goal of having a common data set for all models, we faced several limitations including: differences in lengths of observational time series across towers (up to 3 years), as well as lack in overlap of these time series; uneven distribution of towers across groups (e.g., biomes); and, finally, South American geographical features that were not considered in this study (e.g., MGS biome or desert climate type, BWk). Thus, it was not possible to assess, for all towers, model responses during all seasons. Nonetheless, the fact that our dataset encompasses a wide variety of climates enabled us to evaluate, to a reasonable extent, model responses for contrasting seasons and fill in the gaps flagged up in the literature, such as the absence, in a similar analysis, of towers in the tropical climate zone pointed out by McCabe *et al.* [2016].

## 5 Conclusion

Our results show that, in general,  $ET$  can be reasonably well predicted by all four models, despite an overall tendency of overestimation by PT-JPL and PM-MOD, and underestimation by GLEAM. Similar to results from other continents, model predictive skill in South America decreases as aridity increases. Our analysis emphasizes the need of improving model  $ET$  partitioning, although the link between flawed  $ET$  partitioning and poor

**Table 1.** Comparison of evaporation fractions for several land uses between this study and field-based estimates. FE = field estimates. Land covers that present field data from the same modeling sites or same geographical region are indicated with '\*’.

LULC	FE	$E_{trans}$ (%)			$E_{soil}$ (%)			$E_{int}$ (%)			References		
		GLEAM	PT-JPL	PM-MOD	FE	GLEAM	PT-JPL	PM-MOD	FE	GLEAM		PT-JPL	PM-MOD
EBF*	80–84	74–79	47–63	31–88	NA	4–6	18–24	5–20	15–25	17–20	18–29	7–58	Leopoldo et al. [1995]; Shuttleworth and Pereira [1988]
DNF*	50–81	84–94	64–84	78–90	NA	8–14	14–24	8–18	10	1–2	2–5	2–4	Gaj et al. [2016]; Sun et al. [2019]; de Queiroz et al. [2020]
CROP*	NA	93	63	70	20–4	6	31	21	10	1	6	9	Denmead et al. [1997]; Cabral et al. [2012]
CROP*	85	88	55	69	NA	3	20	4	13	9	25	27	Cabral et al. [2010]
WS*	NA	86	76	78	NA	2	17	5	8	13	7	17	Cabral et al. [2015]
GRA	50–78	69–73	47–49	33–54	NA	25–30	32–46	30–52	NA	1–2	7–18	15–16	Ferretti et al. [2003]; Sutanto et al. [2012]; Wang et al. [2014]
MF	36–74	82–88	63–75	58–71	19	11–16	23–30	28–29	NA	1	2–7	1–14	[Aron et al., 2020; Paul-Limoges et al., 2020]
PW	33–38	73–86	28–57	34–41	NA	0–20	32–57	16–54	NA	3–14	6–16	21–43	Zhang et al. [2018]



model skill is not evident based on our results. Having reliable *ET* partitioning coefficients as part of the FLUXNET-type datasets would be immensely valuable in this respect, but unfortunately such data are difficult to obtain, as they require labour-intensive and expensive methods (such as sapflow gauges and lysimeters), that also present problems with regards to upscaling from plot to field-scale. Correlations are consistently higher for GLEAM and PT-JPL, with  $R^2 > 0.5$  for most sites, whereas PM-MOD and PM-VI presented better performances in terms of PBIAS ( $-10 < PBIAS < 10\%$  for most sites). As for PM-VI, this outcome is expected, given the model requires calibration with local data.

The model skill seems to be unrelated to land cover type as we found a wide variability of metric values within the same class and across models. Conversely, a clear relationship between model skill and climate was noticed, with poor responses occurring in semi-arid regions whereas an overall good performance was found for more humid environments. Except for the FGS biome, we found that skill across models was mostly similar within the same biome.

Despite the relatively high number of towers (compared to previous global analyses that used a similar amount of sites), gathering a balanced amount of data and uniform distribution of towers across different biomes and climate zones across the whole continent was challenging. Thus, we emphasize the importance of expanding the flux tower network in South America as well as the formation of bilateral collaboration for future contributions. Previous studies [e.g. Michel *et al.*, 2016; McCabe *et al.*, 2016] have expressed the need of extending the evaluation of RSBET models to uncharted biomes and climate conditions. Our analysis fills this gap by assessing the reliability of four RSBET models over South America; we provide benchmarking metrics that can serve the improvement of ET models for better capturing ET over this continent.

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## Supporting Information for

### “Are remote sensing evapotranspiration models reliable across South American climates and ecosystems?”

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## Contents

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## S1.0 Study area – Biomes

The tropical & subtropical moist broadleaf forests, i.e. TSMBF, cover approximately 50% of the South American territory. This biome is mainly characterized by a warm climate and high rainfall rates. ET in this region is responsible for generating ~30% of the atmospheric moisture that precipitates in the Amazon basin [Eltahir and Bras, 1994]. The dry seasonal forests, i.e. TSDBF, are located mostly in the center (Bolivia) and east (Brazilian semi-arid region) of South America; here mean rainfall variability is temporally and spatially high, with pronounced dry seasons that can extend up to 10 months. The biome that encompasses grasslands, savannas, and shrublands, i.e. TSGSS is the second largest biome in South America. This biome has a high concentration of endemic species whose habitats have experienced exceptional losses, therefore many areas of this biome are part of the global biodiversity hotspots for conservation [Myers *et al.*, 2000]. The wetlands, i.e. FBS, host a vast diversity of aquatic and palustrine vegetation [Junk *et al.*, 2006a,b] and provide crucial ecosystem services to the Pantanal (the largest FGS's ecoregion) [Costanza *et al.*, 1997]. FBS are completely surrounded by grasslands, savannas & shrublands, and dry forests. The temperate forests, i.e. TMBF, are located in southern Chile and Argentina. This region contains the Central Chile biodiversity hotspot [Myers *et al.*, 2000]. Mean annual temperature decreases from north to south and from low to high altitudes in the Andes. Within this biome, peatland ecosystems cover most of the area (440,000 km<sup>2</sup>, Arroyo *et al.* [2005]) in the southernmost part of South America and along the Pacific coast of Chile.

## S2.0 Gap filling of meteorological data

Relative humidity ( $RH$ ) and vapor pressure ( $e_a$ ) records are the most common missing variables among the tower data. Missing  $e_a$  values were filled using  $RH$  and saturated air vapor pressure ( $e_{sat}$ ) calculated from Tetens' equation [Tetens, 1930]. Missing  $RH$  values

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were filled accordingly, given that  $e_a$  and temperature ( $T$ ) are available at that day or 30-min interval. Days without  $T$  records were discarded. Missing atmospheric pressure ( $P_{atm}$ ) values were estimated from ground elevation at the tower sites using equation 7 from the FAO-56 manual [Allen *et al.*, 1998].

### S3.0 Results and Discussion

#### *ET* partitioning

Greater insight into model-based *ET* partitioning can be gained by comparing our results with other estimates from the literature. Previous model-based estimates using the Gash model [Gash *et al.*, 1995; Valente *et al.*, 1997] for USR (sugar cane) and PDG (woodland savanna) showed that  $E_{int}$  accounts for  $\sim 10\%$  of *ET* [Cabral *et al.*, 2012, 2015]. For USR, all three models limited  $E_{int}/ET < 10\%$ ; with PM-MOD offering a mean estimate ( $\sim 9\%$ ) closest to the estimates cited above, and GLEAM presenting the lowest value of  $E_{int}/ET$  ( $< 2\%$ ). While GLEAM estimated  $E_{soil}/ET \approx 5\%$ , PM-MOD and PT-JPL estimates (20 and 30%, respectively) agreed better with the 20–40% reported by [Denmead *et al.*, 1997] for two sugar cane fields in Australia. For the EUC site, interception estimates from GLEAM ( $\sim 10\%$ ) are much closer to the  $E_{int}/ET \approx 13\%$  reported by Cabral *et al.* [2010], when compared to  $\sim 25\%$  estimated by PT-JPL and PM-MOD. For PDG, PT-JPL produced an interception ratio (7% of *ET*) closest to the  $E_{int}/ET = 8\%$  reported by Cabral *et al.* [2015] when compared to GLEAM ( $E_{int}/ET = 13\%$ ) and PM-MOD ( $E_{int}/ET = 17\%$ ). The transpiration fractions ( $E_{trans}/ET$ ) simulated in this study for the sites in dry regions were mostly higher ( $> 75\%$ ) than the range (50–80%) found in several studies by Jasechko *et al.* [2013] (Table 1 in the main text). However,  $E_{trans}/ET > 75\%$  for the Brazilian semi-arid areas are reasonable during the rainy season, when about 70% of the annual *ET* occurs [Mutti *et al.*, 2019; Marques *et al.*, 2020].

All models found practically negligible values of  $E_{int}$  ( $< 2\%$  of *ET*) in three Brazilian semi-arid sites (CAA, CST, and ESEC), despite the fact that previous studies showed that interception loss for native seasonally dry forests (as CAA and CST) accounts for  $\sim 10\%$  of *ET* [de Queiroz *et al.*, 2020]. Regarding *ET* partitioning from croplands (K77, USR, EUC, GRO, BAL), we found  $E_{trans}/ET$  ranging from 85–93% (GLEAM), 46–63% (PT-JPL), and 22–76% (PM-MOD). Hence, the partitioning simulated by GLEAM agrees with previous studies showing that  $E_{trans}$  tends to be high in croplands (70–90% of *ET*), even under low

*LAI* conditions [Zhou *et al.*, 2016; Wei *et al.*, 2017; Stoy *et al.*, 2019]. Despite the wide variability of  $E_{trans}/ET$  among models, the overall predictive skill was satisfactory, thus not associated with their capability to correctly estimate each  $ET$  component individually.

### Impact of correcting $G$

The impact of the correction of  $G$  in the energy balance, thus in the estimation of the observed  $ET$  ( $= Rn - G - H$ ), is addressed here. Most Fluxnet-type datasets do not offer soil moisture and soil temperature data, hence correction of  $G$  is possible only for a reduced number of validation tower sites. At daily scales, in particular under densely vegetated areas,  $G$  is often negligible. The grass sites included in this study lack the data required for correcting  $G$ . Therefore, we select the GRO tower (soybean) to assess the role of  $G$  in the model metrics. In Figure S8, we show the metrics when tower  $ET$  is calculated using  $G$  directly measured by soil heat flux plates below the surface (raw  $G$ , black dots) and when the corrected  $G$  is used ( $G$  surface, red dots), i.e. when heat storage above the plate has been taken into account. As shown in Fig S9, no significant change was noted in  $R^2$ ,  $PBIAS$  or  $RMSE$ ; and in most cases the correction caused a deterioration in the metrics. The largest changes were:

- GLEAM:  $R^2$  decreased from 0.79 to 0.74;
- PM-MOD:  $PBIAS$  increased from -4.59 to 2.07%
- PT-JPL:  $RMSE$  increased from 0.83 to 0.9 mm d<sup>-1</sup>

### Diagnosis

Some factors, e.g. those relating to tower location and the surrounding environment, can also affect model performances. A likely cause for model-observation mismatch, for instance at site K77, could be the heterogeneity of the fluxes measured by the towers [Bai *et al.*, 2015]. Previous studies have shown that land cover heterogeneity may induce some variations in the footprint [Chen *et al.*, 2009]. In the case of the K77 site, the footprint is estimated to extend up to 200 m away from the mast, based on crop and flux sensor heights; however, high resolution images of the K77 surroundings reveal a certain heterogeneity (remnant of forest and shrubs) within a 130-m radius around the tower. Depending on wind speed and direction, tower measurements for this site could therefore be affected by non-crop fluxes.



Aside from patch-scale heterogeneities, there is also the possibility that the 250-m MODIS pixel is returning a mixed response of different types of vegetation within it (sub-pixel spatial heterogeneity), which has been shown to compromise model performance [Fisher *et al.*, 2008; Mu *et al.*, 2011; Nagler *et al.*, 2005; McCabe *et al.*, 2016; Fisher *et al.*, 2020]. “Mixed” pixels of croplands have been shown to be problematic, especially when the VI adopted is *EVI* [e.g., Wardlow *et al.*, 2007]. In this regard, *NDVI* might present a certain advantage as it has been shown to better capture the changes in the green biomass level, making it more sensitive to sudden changes among crops in the senescence phase, whereas *EVI* would be more suitable for mapping crops at the peak phase [Embry and Nothnagel, 1994; Wardlow *et al.*, 2007].

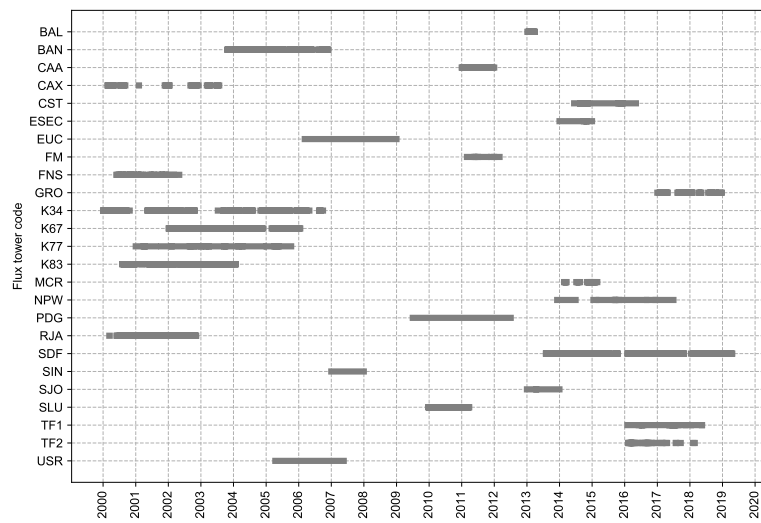
Therefore, the mismatch between  $ET_{sim}$  and  $ET_{obs}$  for K77 site are probably linked to the VI adopted here. An alternative for future work would be to use even higher spatial resolution VNIR/*NDVI* data than the 250 m MODIS data. Examples include 10-m Sentinel-2 *NDVI* or 3-m Planet *NDVI*. For instance, Aragon *et al.* [2018] used 3-m Planet *NDVI* to produce ultra high resolution *ET* using PT-JPL. Another promising way forward is the 70-m ECOSTRESS *ET* data (e.g., Figure S11), which enables a closer alignment to eddy covariance footprints [Fisher *et al.*, 2020]. The land cover representativity issue is even worse in the case of GLEAM that uses microwave *VOD* data for vegetation phenology, a much coarser product. As for the EUC site (~500-m footprint), the plantation border is at a sufficient distance from the tower (~2 km), compared to the surface heterogeneities encountered for the other sites, and occupies multiple MODIS pixels. Hence none of the problems mentioned above are reasonable explanations for the model overprediction at that site in the dry period. Moreover, *EVI* data from MODIS captured the vegetation variability during transition in terms of amplitude but we noticed a lag between  $ET_{obs}$  and *EVI* (Figure S9). A similar pattern was reported by [Fisher *et al.*, 2008] while validating PT-JPL against  $ET_{obs}$  for a savanna tower.

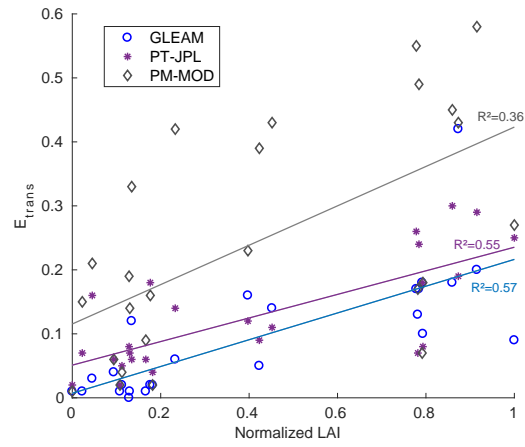
In the case of PT-JPL model estimates at the EUC site, the model’s higher sensitivity to *Rn* than to *RH* seems to be a plausible explanation for the model performance (Figure 5; Figure S9). In the PT-JPL model, *RH* affects *ET* indirectly, through the soil moisture constraint and interception loss. Conversely, influence of *Rn* is indirect: a fraction of it is used to compute all three *ET* components and it affects the temperature which, in turn, is used to calculate the plant temperature constraint [Fisher *et al.*, 2008]. Moreover, *Rn* has a direct effect on *ET* as it controls the potential *ET* in the PT equation. Both the soil moisture and

the plant temperature constraints are positively correlated with modelled  $ET$  and they have been shown to be the most sensitive parameters in the PT-JPL model, potentially resulting in  $\sim 20\%$  of model uncertainty [García *et al.*, 2013]. Although the  $RH$ -dependent constraint presented a larger variability at the EUC site, it tended to follow  $ET_{obs}$  variability whereas the  $Rn$ -indirectly-dependent constraint remained constant at its maximum value (one) during the months during which overprediction occurred.

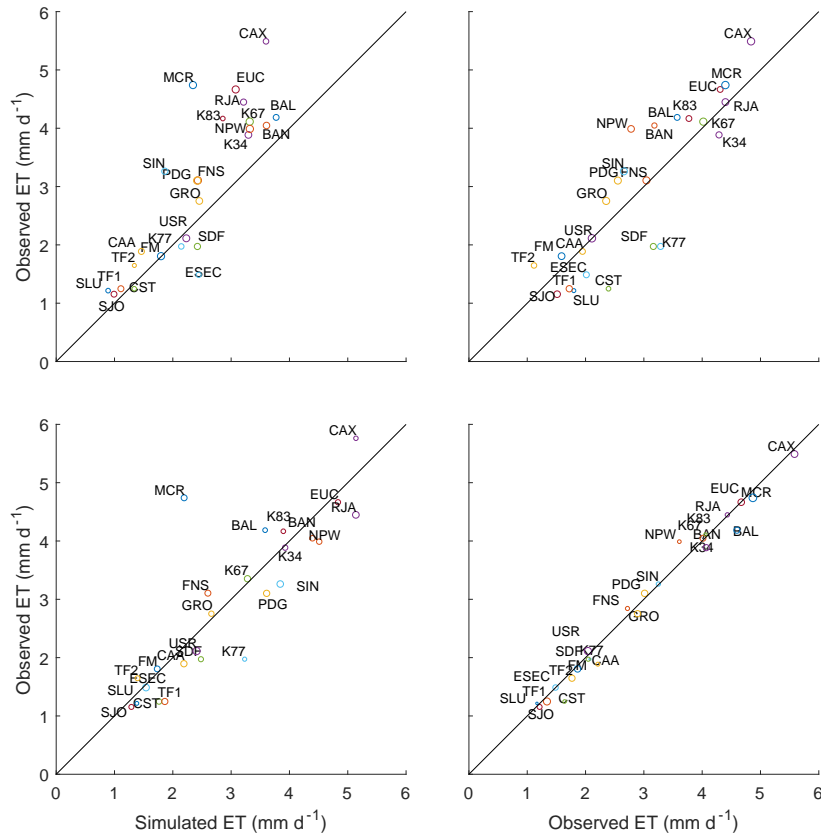
The performance of the PM-based models observed here might have been affected by the uncertainties in the estimation of site-specific properties, as we used the leaf conductances (as well as other parameters) from the Biome Properties Look-Up Table in [Mu *et al.*, 2011], instead of calibrating them for the selected sites. As noted by Ershadi *et al.* [2015], the parameterisation seems to be crucial for PM modes. Therefore, a clear path for improving PM-based models over South America is to include its towers for calibrating the biome specific parameters (e.g., potential stomatal conductance, surface conductance etc) in order to make them more representative and account for their inter-continental variability. Moreover, Mu *et al.* [2011] also acknowledged that values of  $LAI/f_{PAR}$  from MODIS may introduce a bias in  $ET$  estimates, which may explain the systematic errors at some sites (e.g., CAA, FM, K83).

Regarding the PM-VI model, we found it to be the model with the largest variability ( $0 < R^2 < 0.85$ ,  $0.15 < m < 1.25$ ,  $0 < \rho < 0.98$ ) in performance among the selected towers. Because this model consists of a combination of  $ET_o$  and an  $EVI$ -based crop coefficient, we would expect it to perform best at crop sites. Although that was true for some sites (EUC and USR), PM-VI had the poorest performances at most sites with an aridity index  $> 1.2$ . Moreover, the model was capable of estimating  $ET$  for non-crop areas, with good predictive skills found at sites with moist forest (K34), mixed forest (FM), woodland savanna (PDG), and two permanent wetland sites (TF1 and TF2). It is interesting to note the contrasting behaviour of the PM-based models. While the PM-MOD is much more complex than PM-VI, the latter outperforms the former at some sites (e.g., TF1 and TF2). This is most likely caused by the fact that PM-MOD has not been calibrated for the flux towers considered here, unlike PM-VI.

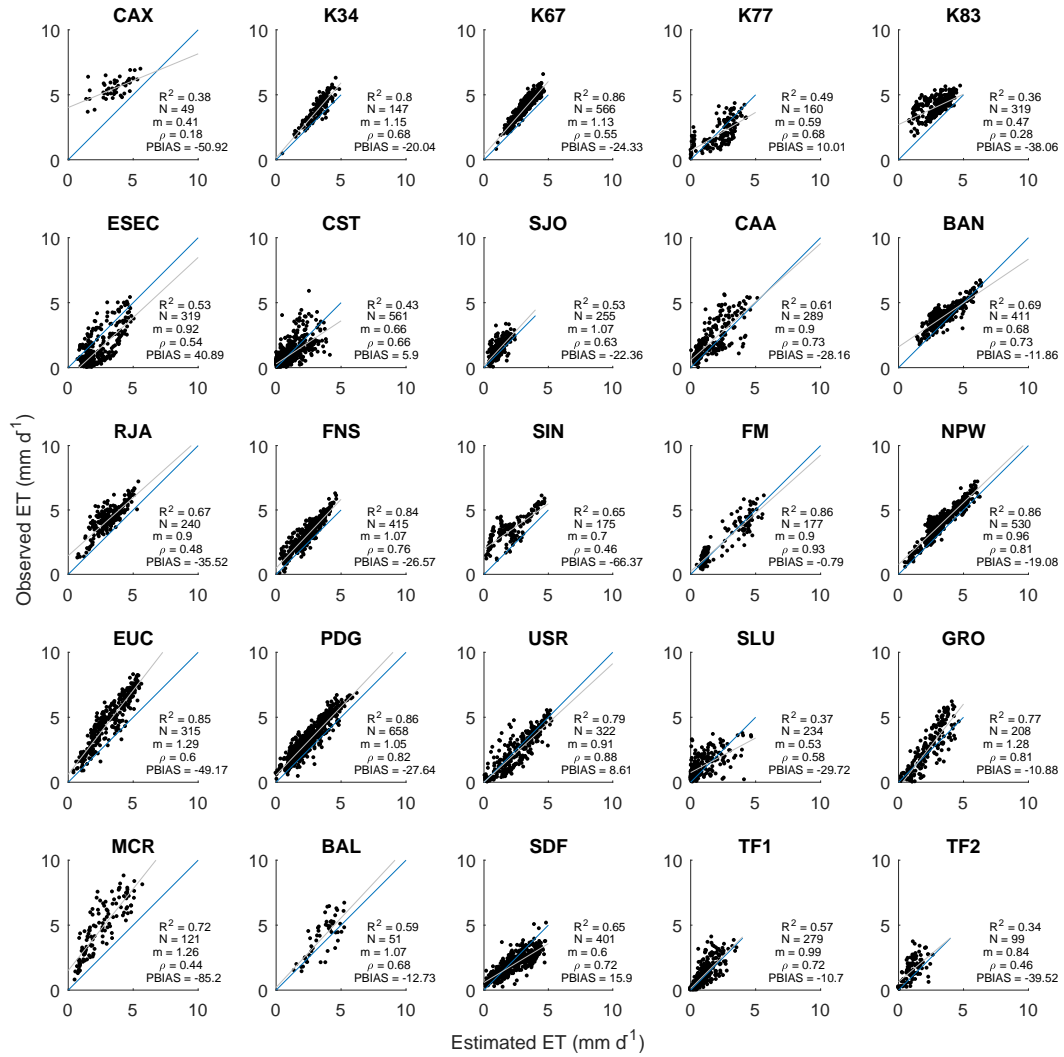
**Figures S1-10****Figure 1.** Latent heat flux data availability at individual flux towers.



**Figure 2.** Relationship between  $E_{int}$  and  $LAI$  for GLEAM, PT-JPL and PM-MOD.

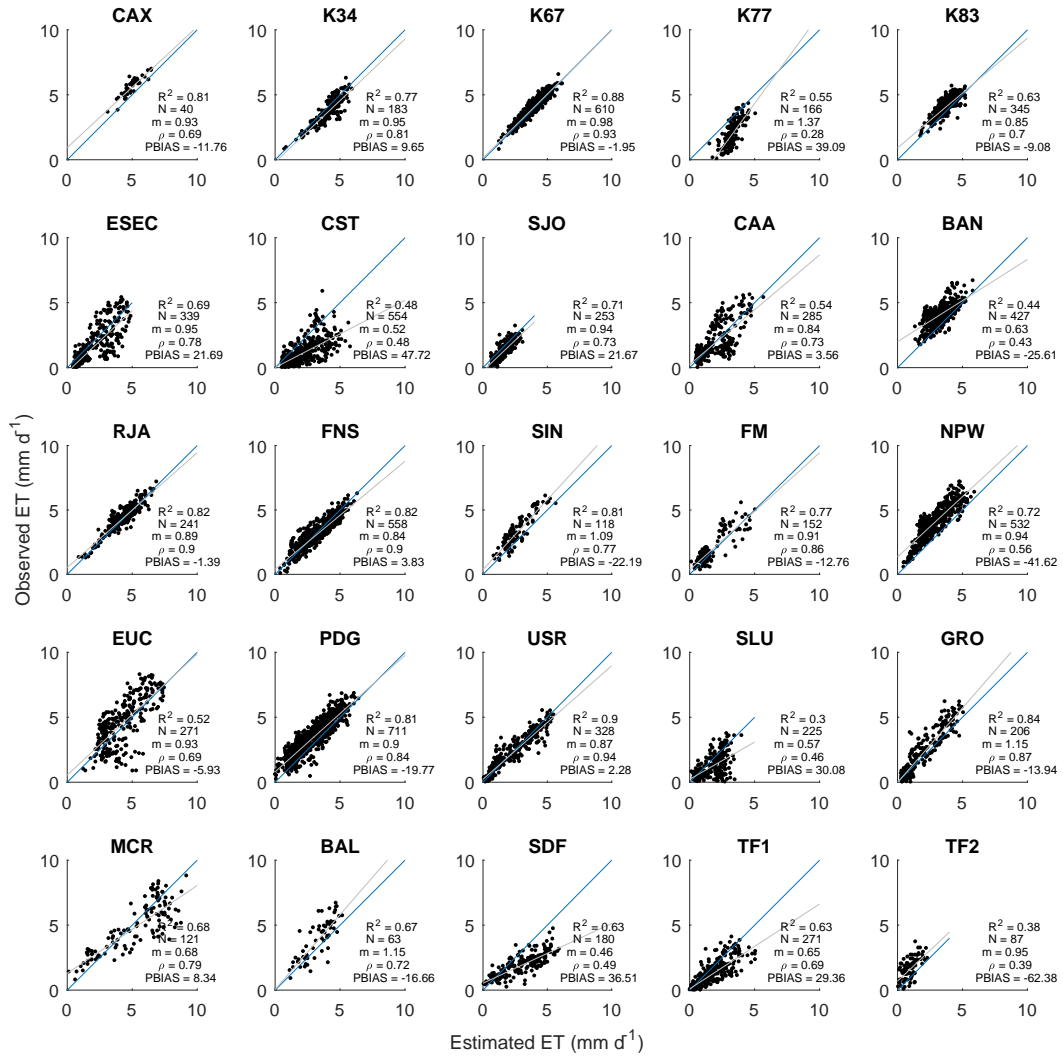


**Figure 3.** Comparison of mean observed and simulated  $ET$ . Circle sizes are proportional to individual model  $R^2$ .

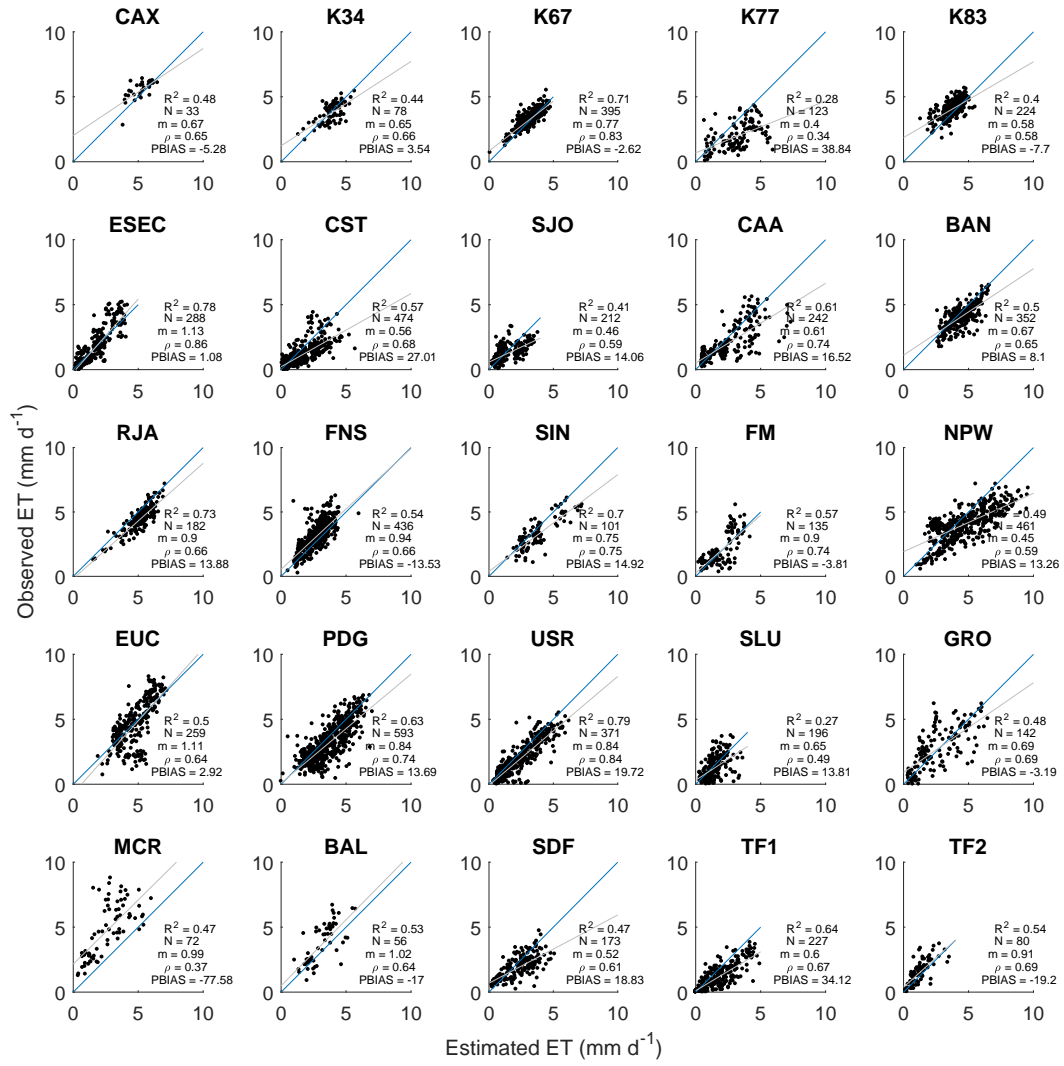


**Figure 4.** Comparison between observed and simulated ET from GLEAM. N = sample size;  $R^2$  = determination coefficient, m = slope of the least squares regression line,  $\rho$  = concordance correlation coefficient, PBIAS = percent bias.

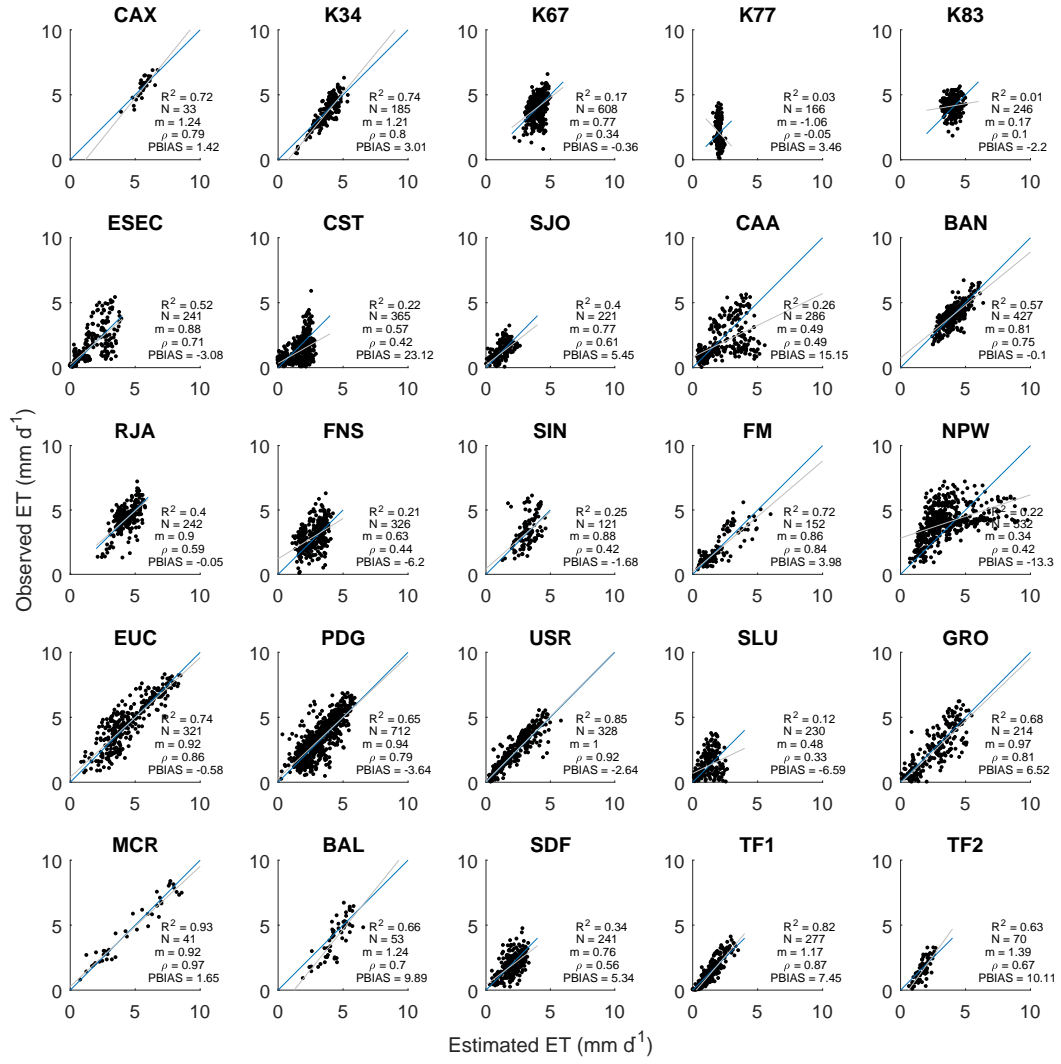




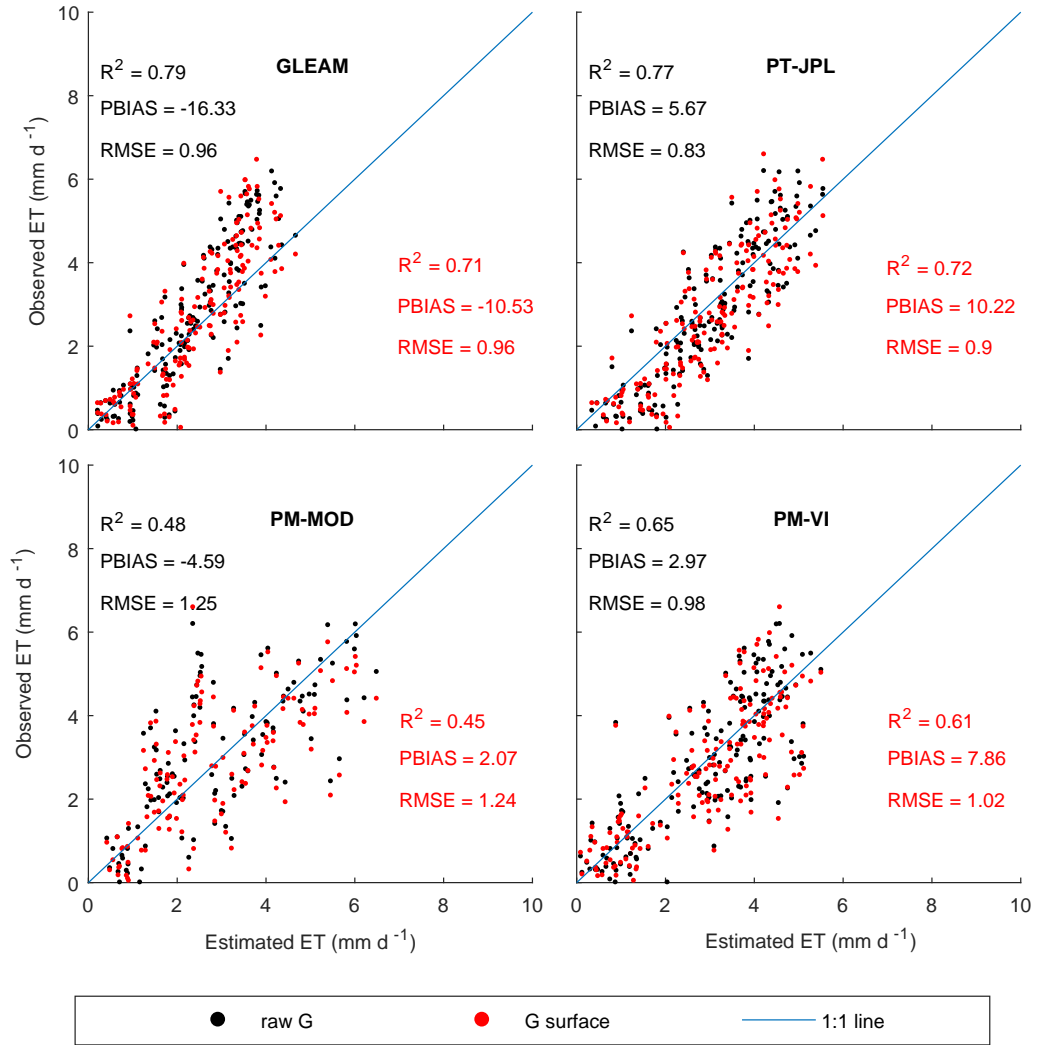
**Figure 5.** Comparison between observed and simulated ET from the PT-JPL model. N = sample size; R<sup>2</sup> = determination coefficient, m = slope of the least squares regression line, ρ = concordance correlation coefficient, PBIAS = percent bias.



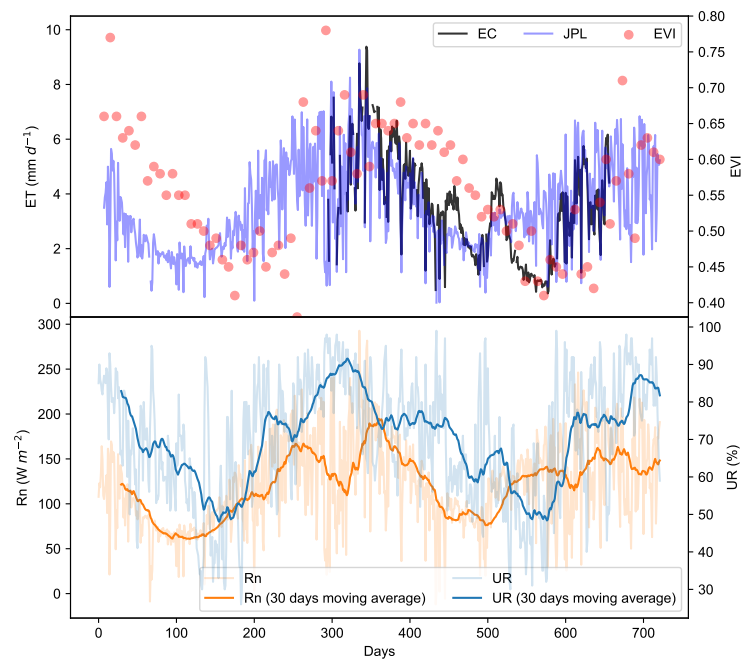
**Figure 6.** Comparison between observed and simulated ET from the PM-MOD model. N = sample size;  $R^2$  = determination coefficient, m = slope of the least squares regression line,  $\rho$  = concordance correlation coefficient, PBIAS = percent bias.



**Figure 7.** Comparison between observed and simulated ET from the PM-VI model. N = sample size;  $R^2$  = determination coefficient, m = slope of the least squares regression line,  $\rho$  = concordance correlation coefficient, PBIAS = percent bias.

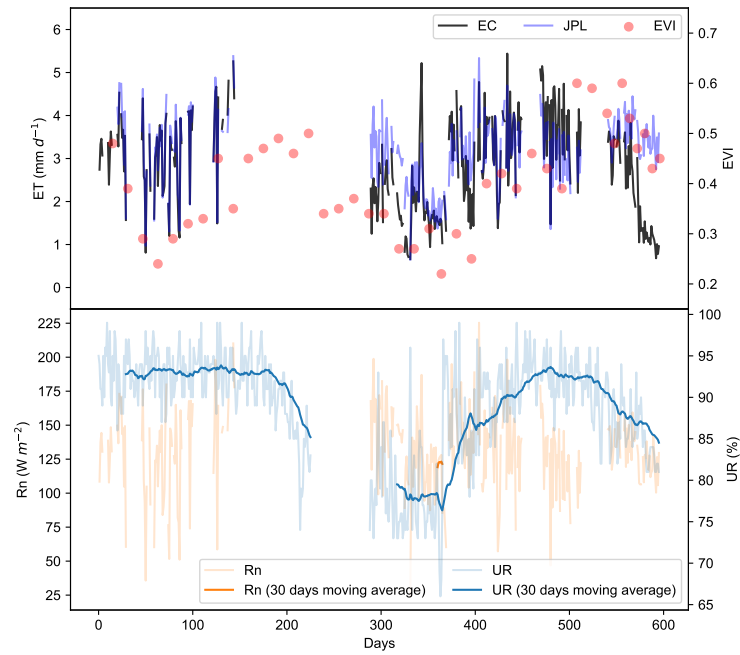


**Figure 8.** Comparison between observed and simulated *ET* with and without correcting *G*.

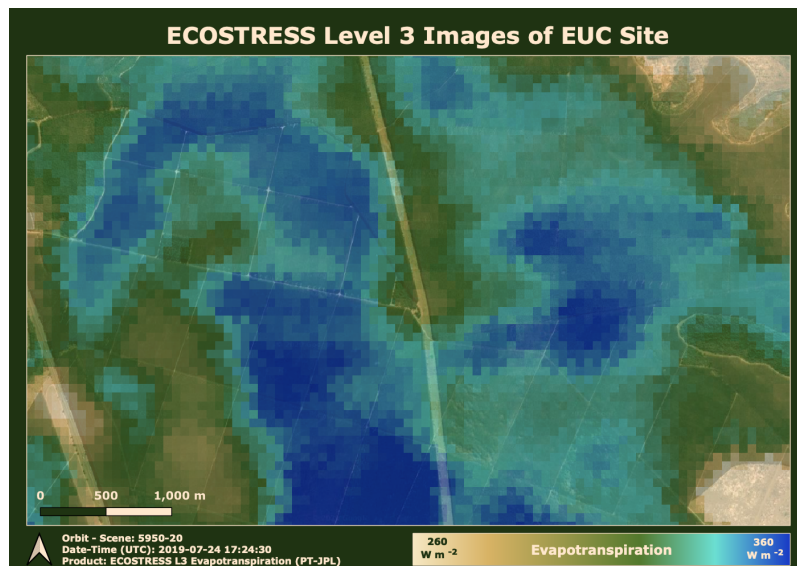


**Figure 9.** Relationship between observed ET (EC), simulated ET from PT-JPL and key meteorological variables at EUC site.





**Figure 10.** Relationship between observed ET (EC), simulated ET from PT-JPL and key meteorological variables at K77 site.



**Figure 11.** ECOSTRESS image (70-m resolution) for the EUC tower site

**Tables S1-S3****Table 1.** Description and characteristics of the biomes covered in this study.

Biome	Major ecoregion	Approximate area ( $\times 10^3 \text{ km}^2$ )	Dominant climates
Tropical & Subtropical Moist Broadleaf Forests (TSMBF)	Amazon and Atlantic rain-forests	8,750	Tropical Rainforest and Monsoon (Af and Am)
Tropical & Subtropical Dry Broadleaf Forests (TSDBF)	Caatinga and Chiquitano dry forests	1,000	Arid Steppe Hot (BSH)
Tropical & Subtropical Grasslands, Savannas & Shrublands (TSGSS)	Cerrado, Dry and Humid chaco, and Uruguayan savanna	4,250	Tropical Savanna (Aw)
Flooded Grasslands & Savannas (FGS)	Pantanal, Paraná flooded savanna, and Southern Cone Mesopotamian savanna	265	Tropical Savanna (Aw)
Temperate Broadleaf & Mixed Forests biome (TBMF)	Valdivian temperate forests and Magellanic subpolar forests	550	Temperate without dry season and warm summer (Cfb) and Polar Tundra (Td)
Temperate Grasslands, Savannas & Shrublands (TGSS)	Humid Pampas and Low Monte	576	Cold semi-arid (BSk) and Humid subtropical (Cfa)

**Table 2.** Flux towers used to validate remote sensing-based ET. Biome types: Tropical & Subtropical Moist Broadleaf Forests (TSMBF); Tropical & Subtropical Dry Broadleaf Forests (TSDBF); Temperate Broadleaf & Mixed Forests (TBMF); Tropical & Subtropical Grasslands, Savannas & Shrublands (TSGSS); Temperate Grasslands, Savannas & Shrublands (TGSS); Flooded Grasslands & Savannas (FGS). Land use classes: Evergreen Broadleaf Forest (EBF); Deciduous Broadleaf Forest (DBF); Grassland (GRA); Cropland (CROP); Woodland Savanna (WS), Mixed Forest (MF); Deciduous Needleleaf Forest (DNF); Permanent Wetland (PW). LULC = Land use/Land Cover. EBR = Energy Balance Ratio  $((LE + H)/(Rn - G))$ . NA = not available

Name	Lat	Lon	Biome/LULC	Elevation (m)	EBR	Ref.
SDF	-41.88	-73.68	TBMF/EBF	50	NA	NA
K34	-2.61	-60.21	TSMBF/EBF	90	0.86	Hutyra <i>et al.</i> [2007]
RJA	-10.08	-61.93	TSMBF/DBF	180	0.74	von Randow <i>et al.</i> [2004]
CAX	-1.72	-51.46	TSMBF/EBF	57	NA	NA
FNS	-10.77	-62.34	TSMBF/GRA	240	0.77	Hasler and Avissar [2007]
K67	-2.86	-54.96	TSMBF/EBF	194	0.97	Paca <i>et al.</i> [2019]
K83	-3.02	-54.97	TSMBF/EBF	181	0.95	Paca <i>et al.</i> [2019]
K77	-3.01	-54.54	TSMBF/CROP	101	1.16	Paca <i>et al.</i> [2019]
SIN	-11.41	-55.32	TSMBF/WS	349	0.88	Vourlitis <i>et al.</i> [2008]
BAN	-9.82	-50.16	TSGSS/WS	168	0.9	Borma <i>et al.</i> [2009]
TF1	-54.97	-66.73	TBMF/PW	40	NA	Kutzbach [2019a]
TF2	-54.83	-68.45	TBMF/PW	60	NA	Kutzbach [2019b]
EUC	-21.58	-47.6	TSGSS/CROP	710	1.02	Cabral <i>et al.</i> [2011]
PDG	-21.62	-47.63	TSGSS/WS	710	0.99	Cabral <i>et al.</i> [2015]
USR	-21.64	-47.79	TSGSS/CROP	541	0.97	Cabral <i>et al.</i> [2012]
NPW	-16.49	-56.41	FGS/WS	120	NA	Dalmagro <i>et al.</i> [2018]
FM	-15.72	-56.07	TSGSS/MF	154	0.74	Rodrigues <i>et al.</i> [2014]
MCR	-37.55	-57.3	TGSS/PW	1	NA	Tonti <i>et al.</i> [2018]
GRO	-35.62	-61.32	TGSS/CROP	80	NA	NA
BAL	-37.75	-58.34	TGSS/CROP	130	NA	Curto <i>et al.</i> [2019]
SJO	-8.81	-36.41	TSDBF/GRA	702	0.96	Machado <i>et al.</i> [2016]
CST	-7.96	-38.38	TSDBF/DNF	468	0.73	Souza <i>et al.</i> [2015]
ESEC	-6.58	-37.25	TSDBF/DNF	205	0.87	Campos <i>et al.</i> [2019]
CAA	-9.05	-40.32	TSDBF/DNF	391	0.75	Silva <i>et al.</i> [2017]
SLU	-33.46	-66.46	TGSS/MF	320	0.86	García <i>et al.</i> [2017]

**Table 3.** Flux towers used to validate remote sensing-based ET. Biome types: Tropical & Subtropical Moist Broadleaf Forests (TSMBF); Tropical & Subtropical Dry Broadleaf Forests (TSDBF); Temperate Broadleaf & Mixed Forests (TBMF); Tropical & Subtropical Grasslands, Savannas & Shrublands (TSGSS); Temperate Grasslands, Savannas & Shrublands (TGSS); Flooded Grasslands & Savannas (FGS). Land use classes: Evergreen Broadleaf Forest (EBF); Deciduous Broadleaf Forest (DBF); Grassland (GRA); Cropland (CROP); Woodland Savanna (WS), Mixed Forest (MF); Deciduous Needleleaf Forest (DNF); Permanent Wetland (PW). LULC = Land use/Land Cover. EBR = Energy Balance Ratio  $((LE + H)/(Rn - G))$ . NA = not available

Variable	Source of Product	Spatial resolution	Temporal resolution	Used by model
Air temperature ( $T$ )	Tower data	-	30 min	All
Downward short-wave radiation ( $R_{gs} \downarrow$ )	Tower data	-	30 min	PM-models
Surface outgoing radiation	Tower data	-	30 min	GLEAM
Net radiation ( $Rn$ )	Tower data	-	30 min	PT-models
Air pressure ( $P_{atm}$ )	Tower data or calculated from ground elevation	-	30 min	PM-models
Precipitation ( $P$ )	Tower data	-	30 min	GLEAM
Vapor pressure ( $e_a$ )	Tower data	-	30 min	PM-models and PT-JPL
Air humidity ( $RH$ )	Tower data	-	30 min	PM-models
Wind speed	Tower data	-	30 min	PM-VI
Enhanced Vegetation Index ( $EVI$ )	MOD13Q1	250 m	16 days	PT-JPL and PM-VI
Vegetation Optical Depth (VOD)	TMI, SSM/I and AMSR-E	0.25°	16 days	GLEAM
Leaf area index ( $LAI$ )	MCD15A2H	500 m	8 days	PM-MOD
$f_{PAR}$	MCD15A2H	500 m	8 days	PM-MOD
albedo ( $\alpha$ )	MCD43A3	500 m	daily	PM-MOD

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