Estimating intra-urban inequities in PM2.5-attributable health impacts: A case study for Washington, DC

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Abstract

Air pollution levels are uneven within cities, contributing to persistent health disparities between neighborhoods and population sub-groups. Highly spatially resolved information on pollution levels and disease rates is necessary to characterize inequities in air pollution exposure and related health We leverage recent advances in deriving surface pollution levels from satellite remote sensing and granular data in disease rates for one city, Washington, DC, to assess intra-urban heterogeneity in fine particulate matter (PM5)- attributable mortality and We estimate PM2.5-attributable cases of all-cause mortality, chronic obstructive pulmonary disease, ischaemic heart disease, lung cancer, stroke, and asthma emergency department (ED) visits using epidemiologically-derived health impact Data inputs include satellite-derived annual mean surface PM5 concentrations; age-resolved population estimates; and statistical neighborhood-, zip code- and ward-scale disease counts. We find that PM5 concentrations and associated health burdens have decreased in DC between 2000 and 2018, from approximately 240 to 120 cause-specific deaths and from 40 to 30 asthma ED visits per year (between 2014 and 2018). However, remaining PM5attributable health risks are unevenly and inequitably distributed across the Higher PM2.5-attributable disease burdens were found in neighborhoods with larger proportions of people of color, lower household income, and lower educational Our study adds to the growing body of literature documenting the inequity in air pollution exposure levels and pollution health risks between population sub-groups, and highlights the need for both high-resolution disease rates and concentration estimates for understanding intra-urban disparities in air pollution-related health risks.

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22	Key Points:
23	1. Fine particulate matter-attributable health risks are unevenly and inequitably distributed
24	across Washington, DC
25	2. Higher PM _{2.5} -attributable disease burdens are found in neighborhoods with larger proportions
26	of people of color in Washington, DC
27	3. High-resolution disease and concentration estimates are needed to understand intra-urban
28	disparities in air pollution-related health risks
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Abstract: Air pollution levels are uneven within cities, contributing to persistent health disparities between neighborhoods and population sub-groups. Highly spatially resolved information on pollution levels and disease rates is necessary to characterize inequities in air pollution exposure and related health risks. We leverage recent advances in deriving surface pollution levels from satellite remote sensing and granular data in disease rates for one city, Washington, DC, to assess intra-urban heterogeneity in fine particulate matter ($PM_{2.5}$)-attributable mortality and morbidity. We estimate PM_{2.5}-attributable cases of all-cause mortality, chronic obstructive pulmonary disease, ischaemic heart disease, lung cancer, stroke, and asthma emergency department (ED) visits using epidemiologically-derived health impact functions. Data inputs include satellite-derived annual mean surface PM_{2.5} concentrations; age-resolved population estimates; and statistical neighborhood-, zip code- and ward-scale disease counts. We find that PM_{2.5} concentrations and associated health burdens have decreased in DC between 2000 and 2018, from approximately 240 to 120 cause-specific deaths and from 40 to 30 asthma ED visits per year (between 2014 and 2018). However, remaining PM_{2.5}-attributable health risks are unevenly and inequitably distributed across the District. Higher PM_{2.5}-attributable disease burdens were found in neighborhoods with larger proportions of people of color, lower household income, and lower educational attainment. Our study adds to the growing body of literature documenting the inequity in air pollution exposure levels and pollution health risks between population sub-groups, and highlights the need for both high-resolution disease rates and concentration estimates for understanding intra-urban disparities in air pollution-related health risks.

63 **1. INTRODUCTION**

64 Ambient air pollution in cities is of growing concern due to expected population growth, rapid urbanization, and rising pollution levels in many cities. Extensive epidemiological literature 65 66 reveals strong associations between ambient fine particulate matter of aerodynamic diameter less 67 than 2.5 µm (PM_{2.5}) and mortality and morbidity outcomes, including cardiovascular and 68 respiratory diseases and lung cancer (Brauer et al., 2012; Burnett et al., 2018; Cohen et al., 69 2017), and asthma incidence and exacerbation (Khreis et al., 2017; Orellano et al., 2017). An 70 emerging body of literature also suggests associations with additional health outcomes, including 71 diabetes (Bowe et al., 2018; Eze et al., 2015; Yang et al., 2020); neural, behavioral and cognitive 72 changes (de Prado Bert et al., 2018); happiness and well-being (Zheng et al., 2019); and low 73 birth weight (Bell et al., 2010; Ebisu & Bell, 2012; Malley et al., 2017). Air pollution is 74 considered the leading environmental risk factor and among the leading overall risk factors for 75 global mortality (Cohen et al., 2017; Landrigan et al., 2018; Murray et al., 2020). In the U.S., 76 PM_{2.5} is estimated to be responsible for 100,000-200,000 premature deaths each year, with the range dependent largely on whether all or only anthropogenic PM_{2.5} is included, the risk 77 78 functions used, the mortality causes included, and the year of analysis (Bowe et al., 2019; Fann 79 et al., 2018; Thakrar et al., 2020; Vodonos & Schwartz, 2021).

80 Overall, air quality in the U.S. has improved dramatically since the 1970 Clean Air Act and its 81 1990 Amendments (U.S. EPA, 2020). However, it has not improved equitably. Literature reveals 82 that throughout the U.S., lower income, minority, and marginalized populations experience 83 higher air pollution exposure levels and associated health impacts (Hajat et al., 2015; Tessum et 84 al., 2019). These communities often live near major air pollution sources, such as major roadways, shipping ports, airports, and industrial facilities, resulting from decades of race-biased 85 86 policies (both implicit and explicit) in housing, zoning, facility siting, and transportation (Mohai & Saha, 2015). Today, the same communities that bore the greatest burden of harm decades ago 87 88 continue to face the greatest public health threats associated with long-term exposure to air 89 pollution (Colmer et al., 2020). The National Ambient Air Quality Standards (NAAQS) in its 90 current form is, essentially, a one-size-fits-all universal approach that lacks specificity and treats 91 all communities and subsects the same. This approach produces unequal impacts and reinforces 92 inequitable outcomes even when implemented with the best of intentions.

93 Over the last few years, several U.S. states have implemented ground-breaking laws and policies 94 to address air pollution inequity in their air quality management programs, including California's 95 Assembly Bill (AB) 617 and its resulting Community Air Protection Program, and the New 96 Jersey Law NJ S232 (20R), which establish community emissions reductions programs and 97 protect communities from projects that pose local health and environmental risks, respectively. 98 Similarly, multi-state programs have emerged that aim to collaboratively and equitably reduce 99 greenhouse gases and air pollutants, such as the Medium-and Heavy-Duty Zero Emission Vehicle (MHD-ZEV) Initiative, signed onto by 13 states and the District of Columbia, and the 100 101 Transportation Climate Initiative (TCI) supported by 12 Northeast and Mid-Atlantic states and 102 the District of Columbia. More recently, in January 2021, the Biden Administration issued an 103 Executive Order that elevated the federal government's actions to address environmental

104 injustice.

105 Addressing environmental injustice requires information about air pollution exposure levels 106 within at-risk communities, which is beyond the intent and capability of the existing network of 107 federal reference monitors throughout North America and the spatial resolution of regional 108 chemical transport models. In the District of Columbia, for example, researchers found that fine-109 scale emissions source attribution can reveal environmental injustices that may be obscured when using more coarsely resolved regional data inputs (Northcross et al., 2020). New 110 111 techniques, both emerging and maturing, are being deployed to conduct air quality 112 characterization and surveillance at high spatial resolutions. Techniques include distributed low-113 cost sensor networks (Ahangar et al., 2019; Castillo et al., 2019; Matte et al., 2013), mobile 114 monitoring on vehicles driving through cities (Apte et al., 2017; Messier et al., 2018; Miller et 115 al., 2020; Southerland et al., 2021), and satellite remote sensing (Demetillo et al., 2020; Kerr et 116 al., 2021; Southerland et al., 2021). With relatively high spatial resolution (~1km x 1km) and full 117 geographical coverage, satellite remote sensing could be of particular value for targeted 118 assessment of air pollution exposures and health impacts in cities where low-cost sensor 119 networks and mobile monitoring data are not available.

120 Beyond information about air quality levels, fine-scale information on disease rates is important

121 to understand not just inequities in air pollution exposure, but also inequities in air pollution-

122 related health risks (Southerland et al., 2021). Geographic, economic and racial health inequities

123 are a known issue in the District (Chandra et al., 2013; *Health Equity Report: District of*

124 Columbia 2018, 2019). The recently published Health Equity Report for the District of Columbia

125 (DC HER) 2018 analyzed health data in the District by Proximal Neighborhood Groups (PNGs),

also referred to as statistical neighborhoods. For simplicity, we refer henceforth to the 51 PNGs

127 as "neighborhoods." The DC HER reported a high degree of environmental and health inequity

128 within the District, with asthma emergency department (ED) visit rates being one order of

129 magnitude higher in most affected neighborhoods compared to least affected neighborhoods.

Furthermore, life expectancy differs by 21 years between neighborhoods at the two ends of thespectrum.

132 Given the District's disparities in air pollution exposure and disease rates, and the potential it has 133 to become a role model in creating collaborative actions for change, we use the District as a case 134 study to assess intra-urban heterogeneity in $PM_{2.5}$ -attributable health impacts. We explore the 135 degree of disparity in estimated $PM_{2.5}$ -attributable cases of mortality and disease exacerbation 136 between neighborhoods throughout the District using a high-resolution satellite-derived PM_{2.5} 137 concentration dataset and two high-resolution datasets for disease rates - one based on local 138 administrative data and one using a small-scale estimation technique by the U.S. Centers for 139 Disease Control and Prevention (CDC). By comparing the application of these two datasets, our 140 study shows whether using estimated rather than more cumbersome administrative data for 141 disease rates can identify similar spatial patterns of air pollution-attributable health risks. We 142 anticipate that our study can both inform mitigation approaches aimed at reducing environmental 143 health disparities in the District and advance the development of technical approaches for 144 estimating air pollution-related health inequities within cities.

145

146 **2. METHODS**

147 **2.1. Health impact function**

148 We apply widely used epidemiologically-derived health impacts functions to estimate mortality-

and morbidity attributable to $PM_{2.5}$ (e.g. Anenberg et al., 2010; Fann et al., 2017). We estimate

annual PM_{2.5}-attributable cases of all-cause mortality, ischaemic heart disease (IHD), chronic

- 151 obstructive pulmonary disease (COPD), stroke, lung cancer, and asthma ED visits. These health
- 152 outcomes have been determined to be causally associated with PM_{2.5} by either the U.S. EPA
- 153 (U.S. EPA, 2019) or the Global Burden of Disease (GBD) study (Murray et al., 2020). All
- analyses are conducted using Geospatial Data Abstraction Library (GDAL), Quantum
- 155 Geographic Information System (QGIS 3.6.2) and the Statistical Package R/3.6.3.
- 156 Table 1. Relative risks for all-cause and cause-specific mortality, and asthma emergency department
- 157 visits used in the PM_{2.5}-attributable health impacts functions.

Health Outcome	Relative Risk (95% Confidence Interval)	Age Group (years)	Study	Population Studied
Asthma ED Visits	1.04 (1.01, 1.07)	0-99	Mar et al. 2010	Greater Tacoma, WA
All-cause mortality	1.06 (1.04, 1.08)	0-99	Turner et al. 2016	CPS-II (American Cancer Society)
Chronic Obstructive Pulmonary Disease	1.10 (1.01, 1.19)	30-99	Turner et al. 2016	CPS-II (American Cancer Society)
Ischemic Heart Disease	1.14 (1.02, 1.22)	30-99	Turner et al. 2016	CPS-II (American Cancer Society)
Lung Cancer	1.09 (1.03, 1.16)	30-99	Turner et al. 2016	CPS-II (American Cancer Society)
Stroke	1.11 (1.05, 1.17)	30-99	Turner et al. 2016	CPS-II (American Cancer Society)

159 For each grid cell ($\sim 1 \times 1$ km) in the District, we estimate the annual excess cases of mortality and 160 asthma ED visit rates that are attributable to $PM_{2.5}$ (*AMort* in Eq. 1) for each health outcome 161 separately, applying cause-specific concentration-response factors (β) from the relative risks 162 (RR) shown in Table 1, the baseline disease rates (BDR) described in Table 2, and gridded PM_{2.5} 163 concentrations (Δx) and population estimates (*Pop*). We use log-linear relationships between 164 concentration and RR, consistent with previous studies (Anenberg et al., 2010; Fann et al., 2012, 165 2017). We then aggregate the resulting estimated $PM_{2.5}$ -attributable cases of mortality and 166 morbidity to the neighborhood-, zip code-, ward-, and city-levels accordingly.

167
$$\Delta Mort = (l - e^{-\beta \Delta x}) \times BDR \times Pop \quad , \qquad (1)$$

We estimate: 1) annual mean PM_{2.5}-attributable excess mortality and morbidity from 2000 to
2018 using annual BDR and PM_{2.5} data, and 2) 5-year mean PM_{2.5}-attributable excess mortality

and morbidity (2014-2018 for asthma ED visits and 2011-2015 for all other health endpoints)

- 171 using 5-year averages of both BDR and PM_{2.5} concentrations to remove the influence of
- 172 interannual variability in both of these variables. To disentangle the influence of temporal

173 changes in PM_{2.5} and disease rates separately, we also estimate PM_{2.5}-attributable health impacts

- using PM_{2.5} concentrations from 2018 (8.7 μ g/m³) with year-specific BDR between 2000 and
- 175 2018.
- 176 **Table 2:** Characteristics of the health data obtained from the District of Columbia Department of Health
- 177 for the years 2000 and 2015, and 2014 to 2018 for asthma; annual average health outcome cases in the
- 178 District; mean (and range) of the annual average cases across neighborhoods, zip codes, or wards; 2010
- 179 population from SEDAC; and annual average health outcome rates computed using the District's total
- 180 cases and SEDAC population data per 100,000 people (and per 10,000 people for asthma).

Health outcome of interest	Spatial resolution	Ages included	Mean annual cases (DC- wide)	Mean (range) cases across neighborhoods, zip codes, or wards	Population (DC-wide)	Age-standardized rates (DC-wide)
Asthma Emergency Department (ED) Visits	Zip code (n=26)	All ages	7,103	263 (3-1311)	627,656	113
All-cause mortality	Neighborhood (n=47)	All ages	4,702	98 (13-177)	627,656	749
Chronic Obstructive Pulmonary Disease	Ward (n=8)	Ages 30-99 years	124	14 (10-21)	358,884	35
Ischemic Heart Disease	Neighborhood (n=47)	Ages 25-99 years	840	17 (8-28)	358,884	234
Lung Cancer	Ward (n=8)	Ages 30 - 99 years	258	30 (15-45)	358,884	72
Stroke	Ward (n=8)	Ages 25 - 99 years	89	10 (6-16)	358,884	25

181

182 2.2. Relative risks

183 We use epidemiologically-derived, cause-specific RR estimates representing the association

184 between annual average PM_{2.5} concentration estimates and incidence of the disease outcomes of

185 interest (Table 1), consistent with the U.S. Environmental Protection Agency's (EPA) most

186 recent Regulatory Impact Analysis for PM_{2.5} (U.S. EPA, 2012). City-specific RR estimates for

187 the District are not available. For all mortality outcomes, we derive the RRs from the American

188 Cancer Society's (ACS) Cancer Prevention Study II (CPS-II) which included 1.2 million

189 participants of at least 30 years of age in the U.S. from all states, the District, and Puerto Rico

190 (Turner et al., 2016). For asthma ED visits, we use the RR from a study conducted in the greater

191 Tacoma, Washington area (Mar et al., 2010), which was applied nationally in the most recent

192 U.S. EPA Regulatory Impact Analysis for PM_{2.5} (U.S. EPA, 2012). While the RR for asthma ED

193 visits that we derive from Mar et al. (2010) is based on daily $PM_{2.5}$ concentrations, we use annual

average $PM_{2.5}$ and assume that the annual attributable asthma ED visits are approximately

195 equivalent to the sum of daily attributable asthma ED visits.

196 These RRs are used widely throughout the literature and by the U.S. EPA for regulatory analysis. 197 In the case of mortality outcomes, the studies have the advantage of a nation-wide cohort with 198 high statistical power. However, extrapolating these RRs to specific populations in the District 199 may obscure differences in concentration-response relationships between cities. In addition, the 200 population groups in these studies are not reflective of the racial composition of the population in 201 the District, and applying these RRs to multiple population subgroups within an individual city, 202 as we are doing here, ignores differential quality and access to healthcare, as well as other social 203 determinants of health. Without within-city studies of PM2.5 health effects in the District, 204 extrapolating from these larger studies is necessary.

205 **2.3.** $PM_{2.5}$ concentrations

206 We use annual mean PM_{2.5} concentration estimates from a North American satellite-derived dataset (V4.NA.03) with a spatial resolution of 0.01° x 0.01° (~1 km²). This dataset relates the 207 208 combined aerosol optical depth (AOD) from multiple satellite retrievals to surface PM_{2.5} 209 concentrations using the spatiotemporally-varying geophysical relationship between AOD and PM_{2.5} simulated by the GEOS-Chem chemical transport model. These geophysical values are 210 211 calibrated to ground-based monitors using a geographically weighted regression. V4.NA.03 212 combines the geophysical output of V4.GL.03 (Hammer et al., 2020) with the regional 213 methodology of V4.NA.02 (van Donkelaar et al., 2019). Gridded annual mean PM_{2.5} 214 concentrations vary within the District by up to $\sim 2 \mu g/m^3$ (Fig. S1). The city-wide 5-year average annual PM_{2.5} concentration decreased from 17.1 μ g/m³ in 2000-2004 to 10.0 μ g/m³ in 2014-2018 215

216 (Table S1).

217 While a full evaluation of the satellite-derived PM_{2.5} concentrations against ground

218 measurements is not possible with only three Federal Reference Monitors in our study location

and period, the satellite-derived annual average $PM_{2.5}$ concentrations were generally consistent with observations (Fig. S2). There was a slight overestimation in the satellite-derived concentrations of ~0.5 µg/m³ but the spatial distribution agrees well with observations.

222 **2.4.** Baseline disease rates, population, and demographic data

223 We use annual baseline mortality counts by neighborhood (n = 51) for the years 2000 to 2015, 224 and annual baseline asthma ED visits by zip code (n = 26) for the years 2014 to 2018 from the 225 District's Department of Health (DOH). Baseline counts smaller than five (n < 5) are suppressed 226 to protect privacy, resulting in 50% - 95% missing data for COPD, lung cancer and stroke. For 227 these health endpoints, we apply counts aggregated by ward (n = 8) to achieve more spatially 228 complete data (95%). IHD and all-cause mortality counts are available for 47 out of 51 229 neighborhoods (Table 2). Remaining neighborhoods and wards with suppressed values are 230 assigned Count = 2.5 (the midpoint of 1-4, the values suppressed by DOH) as the spatiotemporal 231 variability in health outcomes does not allow us to estimate a number to replace missing values.

232 Neighborhoods and wards overlays are presented in Fig. S3.

233 We use population estimates from the Socioeconomic Data and Applications Center (SEDAC) 234 2010 population dataset. Population counts from SEDAC consist of estimates from the Gridded 235 Population of the World (GPW), Version 4, by the Center for International Earth Science 236 Information Network (CIESIN) at 30 Arc-Second (~1×1 km) resolution (Center For International 237 Earth Science Information Network-CIESIN-Columbia University, 2018). Using this dataset, we 238 create two population sub-categories (shown in Table 2) based on the same age groups that 239 match the RRs in Table 1 and use these in our health impact function (Eq. 1). We also use the 240 SEDAC population dataset to compute disease rates from the DOH disease count data, as rates 241 are needed to estimate PM_{2.5}-attributable health impacts at the gridcell level.

To evaluate whether estimated disease rates can be used in lieu of more cumbersome (and
sometimes unavailable) city-specific administrative data to capture intra-city heterogeneity in air
pollution health risks, we compare small-area disease rate estimates from the CDC 500 Cities
with DOH data for four health outcomes: asthma ED visits, COPD, lung cancer, and stroke.
While the CDC 500 Cities data have the advantage of high spatial resolution (census tract level)
and full spatial coverage across the District, the specific health endpoints and age groups

248 represented in the CDC 500 Cities data do not exactly match those used in the epidemiological 249 studies from which we draw RR estimates nor the DOH data. For example, the CDC 500 Cities 250 dataset includes cancer, but not lung cancer specifically; therefore, we assume that the spatial 251 distribution of cancer data also reflects the spatial distribution of lung cancer across the District. 252 The CDC 500 Cities data also represent disease prevalence among adults aged 18 and older, 253 while we need incidence rates to estimate $PM_{2.5}$ -attributable mortality and morbidity. We 254 therefore develop new estimated tract-level baseline incidence rates for our diseases and age 255 groups of interest by retaining the District's average disease incidence rate from DOH and the 256 spatial distribution of disease prevalence from CDC 500 Cities. This is an approximation 257 approach, recognizing that the spatial pattern of disease incidence and prevalence may not be 258 fully aligned.

Specifically, for each health outcome included in our study, we use the CDC 500 Cities DC average (CDC 500 Prevalence_(city)) and tract-level prevalence rate (CDC 500 Prevalence_(tract)) to calculate the tract-to-city prevalence ratio (Equation 2). We then multiply this ratio by our DOH city-wide baseline disease incidence rate (DOH BDR_(city)) to obtain a combined CDC-DOH tractlevel baseline disease incidence rate (CDC-DOH BDR_(tract)) that retains the total city-wide incidence rate from DOH and the census tract-level spatial distribution of prevalence from CDC 500 Cities.

266 $(CDC 500 Prevalence_{(tract)}/CDC 500 Prevalence_{(city)}) \times DOH BDR_{(city)} = CDC-DOH BDR_{(tract)}, (2)$

We apply these new integrated CDC-DOH BDR estimates to calculate PM_{2.5}-attributable health
impacts across DC and compare results with those obtained from applying the DOH rates
directly.

We explore differences in estimated $PM_{2.5}$ -attributable mortality and morbidity outcomes between population sub-groups using five social, economic, demographic and health outcome factors at the neighborhood level: education (percent residents 25 years or older with high school diploma or higher; the District mean = 92%, range = 79% - 99%), unemployment (percent residents 16 years or older unemployed; mean = 8%, range = 2% - 30%), income (median household income and percent residents living in poverty; mean = \$94,537, range = \$25,311-

276 \$200,031, and mean = 15%, range = 2% - 40%, respectively), race and ethnicity (% Black alone,

277 % White alone, % Latino/Hispanic, % Asian alone; means = 36%, 46%, 11%, 4%, respectively),

and life expectancy at birth (years; mean = 79, range = 68-89). Data were extracted from the DC

HER, which uses socio-demographic data from the US Census Bureau 2011-2015 American

280 Community Survey (ACS) 5-year estimates, and life expectancy data from the DOH Center for

281 Policy, Planning and Evaluation.

282 **3. RESULTS**

We first report the total number of estimated $PM_{2.5}$ -attributable deaths and asthma ED visits across the District using 5-year average $PM_{2.5}$ concentrations and administrative disease rates (2014-2018 average for asthma ED visits and 2011-2015 for all other health endpoints). We

estimate that approximately 220, 10, 90, 20, 10 excess all-cause, COPD, IHD, lung cancer, and

stroke deaths, respectively, and 40 asthma ED visits could be attributed to PM_{2.5} pollution in the

288 District annually. We next estimate temporal trends using year-specific concentrations and

administrative disease rates. Declining PM_{2.5} concentrations and BDR together contribute to an

290 overall decreasing trend in PM_{2.5}-attributable excess cases in the District, with PM_{2.5}-attributable

all-cause mortality dropping from 520 excess cases in 2000 to 260 in 2015 (Fig. 1). To

disentangle the influence of $PM_{2.5}$ versus BDR changes on the temporal trend in $PM_{2.5}$ -

attributable mortality, we compare $PM_{2.5}$ -attributable deaths calculated using annually varying

294 PM_{2.5} concentration and BDR versus those calculated using constant 2018 PM_{2.5} concentrations

295 $(8.7 \ \mu g/m^3)$ and annually varying BDR (Fig. 1b). Between the years 2000 and 2015,

approximately 30% of the cumulative PM_{2.5}-attributable deaths across this time period (60, 540,

297 110, 50, or 1,620 deaths from COPD, IHD, LC, stroke, and all-causes, respectively) could have

been avoided if historical $PM_{2.5}$ concentrations were as low as the 2018 mean (Table S2).



Figure 1: Temporal trends in a) annual baseline disease rates from District of Columbia Department of Health and annual mean $PM_{2.5}$ concentrations ($\mu g/m^3$, black dotted line) between 2000 and 2018; b) annual $PM_{2.5}$ -attributable deaths ("excess cases") between 2000 and 2015, and $PM_{2.5}$ -attributable asthma ED visits between 2014 and 2018. In panel b, solid line represents the application of annual baseline disease rates (BDR) and $PM_{2.5}$, and dashed lines represent the application of annual BDR with 2018 $PM_{2.5}$ concentrations. Health endpoints: ALL = All-cause mortality, AST = Asthma ED visits, COPD = Chronic Obstructive Pulmonary Disease, IHD = Ischemic Heart Disease, LC = Lung Cancer, STR = Stroke.



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Figure 2: PM_{2.5}-attributable excess mortality and asthma ED visit rates at the neighborhood scale (2011-

309 2015 average). Baseline disease rates underlying these estimates from the DC DOH are at the

310 neighborhood-level for all-cause mortality (ALL) and ischaemic heart disease (IHD); zip code-level for

311 asthma ED visits (AST); and ward-level for chronic obstructive pulmonary disease (COPD), lung cancer

312 (LC), and stroke (STR).



313

Figure 3: PM_{2.5}-attributable health impacts (2011-2015 average) at the neighborhood scale. Top: The
distribution of PM_{2.5}-attributable mortality rates (per 100,000 people for all mortality outcomes and per
10,000 people for asthma ED visits) for each health endpoint and each of the 47 DC neighborhoods with
available health data. Bottom: The distribution of sociodemographic variables across DC neighborhoods
(see Methods for variable definitions). The color gradient used in all panels represents that of the PM_{2.5}attributable all-cause mortality rates (inset legend). Data points are randomly scattered across the x axis
for plotting purposes.

321 We next explore the spatial distribution of these PM_{2.5}-attributable health impacts across the

322 District. Estimated PM_{2.5}-attributable mortality and morbidity rates are higher along the east to

- south city border for all health endpoints, and are also relatively high in the northeast (Fig. 2).
- 324 While neighborhoods located closer to downtown are more densely populated (Fig. S1), PM_{2.5}-
- 325 attributable mortality rates (per 100,000 people) generally increase with increasing distance from
- 326 the city center. The highest $PM_{2.5}$ -attributable all-cause mortality rates are more than four times

- 327 higher in the most (Fort Dupont and Marshall Heights, both located in Ward 7) versus least
- 328 (Woodley Park and Georgetown East, located in Wards 3 and 2, respectively) impacted
- neighborhoods, as shown in Fig. 3 and Table S3. The neighborhoods with the 10 highest (mostly
- in wards 5, 7 and 8) and 10 lowest (mostly in wards 1-3) PM_{2.5}-attributable all-cause mortality
- 331 rates are geographically segregated. The neighborhoods with the highest PM_{2.5}-attributable
- mortality rates have 10% lower education and employment rates, 10% more residents living in
- poverty, \$61,000 lower median household income, and about 10 fewer years of life expectancy
- (Fig. S4). The top 10 neighborhoods also have 54% higher proportions of Blacks and 44% lower
- 335 proportions of Whites (Fig. 4 and Table S3).



Figure 4: PM_{2.5}-attributable mortality rates (per 100,000 people) for all-cause mortality and percent (%)
Black distribution by neighborhood across Washington, DC. Data represent equal intervals and 20112015 means.

- 340 PM_{2.5}-attributable mortality rates appear to follow the spatial patterns of the BDR inputs (Fig.
- 341 S5) more so than that of the PM_{2.5} inputs (Fig. S1). For all-cause mortality and IHD, health
- 342 outcomes for which BDR were available at the neighborhood-level, PM_{2.5}-attributable excess
- 343 mortality rates range by a factor of five and eight (from 17 to 90, and 7 to 58 cases per 100,000

people) respectively, across the District's neighborhoods. Contrastingly, for the health outcomes
with ward-level BDR (COPD, LC and STR), PM_{2.5}-attributable mortality rates show less
variation, with ranges differing by a factor of ~3-5. PM_{2.5}-attributable asthma ED visit rates (with
BDR at the zip code-level, n = 26), also show spatial homogeneity between neighborhoods

348 within zip codes (i.e. more heterogeneity within wards but not within zip codes).

349 We next compare the application of administrative versus estimated BDR on PM_{2.5}-attributable 350 mortality rates. The overall spatial distributions of BDR across the District differ between the 351 DOH dataset and the integrated CDC-DOH dataset, though are more similar for asthma, COPD, 352 and stroke than for lung cancer (Fig. 5). Differences in PM_{2.5}-attributable mortality and 353 morbidity rates estimated using the two BDR datasets were more widespread for asthma ED 354 visits compared with COPD, lung cancer and stroke (mean percent differences of 12, -7, -9, respectively, compared with 187 for asthma, although with relatively large standard deviations of 355 356 32, 45, and 25, respectively). Over- and under-estimation by the application of the integrated

357 CDC-DOH estimated BDR are more unevenly distributed for COPD and stroke.

358 4. DISCUSSION

359 Following national trends, PM_{2.5} concentrations and PM_{2.5}-attributable deaths have halved 360 locally in the District during our study period. The District-wide mean annual average PM_{2.5} concentrations decreased from 17 to 8.7 μ g/m³ between 2000 and 2018. Consequently, total 361 362 estimated PM_{2.5}-attributable excess deaths for four cause-specific mortality outcomes combined 363 (COPD, IHD, lung cancer and stroke) dropped from approximately 240 in 2000 to 120 in 2015. PM_{2.5}-attributable asthma ED visits also declined, from approximately 40 cases in 2014 to 30 364 365 cases in 2018. Estimated PM_{2.5}-attributable mortality and morbidity rates differed by up to a 366 factor of five between wards and a factor of eight between neighborhoods. For example, $PM_{2.5}$ -367 attributable all-cause deaths ranged from 17 to 90 per 100,000 people across neighborhoods, and 368 PM_{2.5}-attributable IHD deaths ranged from 7 to 58 per 100,000 people.



369

Figure 5: Neighborhood-level PM_{2.5}-attributable rates for asthma ED visits (Asthma) per 10,000 people and PM_{2.5} attributable mortality rates for COPD, lung cancer (LC) and stroke per 100,000 people using a) DOH disease rates

- 372 and b) the integrated CDC-DOH disease rates, and c) percent difference between a) and b) [(CDC-DOH -
- 373 DOH)/DOH].

374 This spatial heterogeneity reveals both racial and sociodemographic inequities in the District's 375 PM_{2.5}-attributable health burden. Specifically, we find that PM_{2.5} health risks are largest for 376 neighborhoods with a high proportion of people of color, located in Wards 7 and 8 in the 377 District's southeast. These neighborhoods also have lower income levels and lower educational 378 attainment compared with the District average. Our analysis suggests that the same 379 neighborhoods have substantially larger PM_{2.5}-attributable mortality and morbidity rates 380 compared with neighborhoods that have a higher percentage of White populations and higher 381 levels of household income and educational attainment. These results are consistent with the 382 prior literature demonstrating that $PM_{2.5}$ exposure and associated health impacts are unevenly 383 and inequitably distributed across race/ethnicity, age and socioeconomic categories (Ebisu & 384 Bell, 2012; Southerland et al., 2021; Tessum et al., 2019; Yitshak-Sade et al., 2019), and adds to 385 the literature documenting inequity in air pollution exposure levels and pollution health risks 386 between population sub-groups in the District (Chandra et al., 2013). By considering the 387 influence of intra-city heterogeneity in disease rates, we extend the literature to incorporate not 388 just inequitable exposure, but also population vulnerability to pollution, similar to the analyses 389 for New York City by Kheirbek et al. (2013) and the San Francisco Bay Area, California by 390 Southerland et al. (2021).

391 The intra-city variation in our estimated $PM_{2.5}$ -attributable mortality and morbidity cases is 392 driven by both disease rates and PM_{2.5} concentrations. While gridded PM_{2.5} varies spatially across the District by $\sim 2 \mu g/m^3$ (with higher concentrations in the northeast), BDR are five times 393 394 higher in the southeast wards for COPD, lung cancer and stroke, up to nine times higher in 395 southeast neighborhoods for all-cause mortality and IHD, and over 30 times higher in southeast 396 zip codes for asthma ED visits, compared with wards, neighborhoods and zip codes in the 397 District's northwest, respectively. We found that variation in fine-scale BDR drives the spatial 398 heterogeneity in estimated PM_{2.5}-attributable mortality and morbidity, consistent with 399 Southerland et al. (2021), though the coarse resolution of the data inputs to the PM_{2.5} 400 concentration model preclude our ability to draw a strong conclusion from this result. Our results 401 may suggest that the satellite-derived PM_{2.5} concentrations are not showing the extent of 402 heterogeneity of PM_{2.5} concentrations at the street and block level.

403 Our study has several limitations and uncertainties. The variability of resolutions associated with 404 datasets used to produce the satellite-derived PM_{2.5} concentrations may limit their ability to fully 405 represent intra-city variation. While these input datasets have resolutions as high as 1×1 km, the 406 combined effect of coarser inputs may reduce the spatial heterogeneity indicated between 407 neighborhoods. Exposure misclassification can in turn create uncertainties in risk estimates and 408 potentially disadvantage already vulnerable populations (Northcross et al., 2020). High 409 resolution BDR also contributes to uncertainties in intra-urban air pollution health impact 410 estimates. We found that using more easily accessible estimated BDR from the CDC 500 Cities 411 project in lieu of administrative data yielded considerably different spatial patterns of estimated 412 PM_{2.5}-attributable disease rates.

413 Our results suggest that evaluating PM_{2.5} in regards to the health-based National Ambient Air 414 Quality Standard must consider both intra-urban variation in concentrations and disease rates to 415 address impacts on certain vulnerable populations, in particular communities of color. Black and 416 Native American people have statistically significantly higher asthma rates than their 417 counterparts in other races (CDC, 2019). Although persistent, these health inequities are neither 418 natural nor inevitable (Health Equity Report: District of Columbia 2018, 2019). Given the 419 relationship between air pollution and asthma exacerbation, an outcomes-focused equity lens that 420 is intentional in its protection of historically marginalized communities, especially those that 421 experience worse air quality-related health effects, is critical to reduce air pollution inequities. 422 The D.C. Law 23-181. Racial Equity Achieves Results (REACH) Amendment Act of 2020, 423 which requires that racial equity impact analysis be conducted by each agency and council for 424 each new piece of legislation, underscores growing interest and need for analysis focused on 425 differential racial impacts, such as premature mortality, on historically disadvantaged and highly 426 impacted communities (DOEE Ozone NAAQS Comment Letter, 2020; Ozone NAAQS Public

427 *Hearing*, 2020).

428 Our results also indicate that quantitatively characterizing neighborhood-scale differences in

429 PM_{2.5}-related health risks would continue to benefit from advances in fine resolution information

430 on both PM_{2.5} concentration data and intra-city baseline disease rates. In alignment with

431 Kheirbek et al. (2013) and Southerland et al. (2021), we found that fine-scale baseline disease

data better characterize population subgroups' susceptibility and disparities, which is necessary

433 to aid in policy-making to reduce urban health inequities, even for pollutants that are relatively 434 spatially homogeneous, as is $PM_{2.5}$. However, the racial and ethnic inequities may be 435 underestimated in this and other recent studies that apply generalizable relative risks from large 436 nation-wide cohorts and/or that extrapolate relative risks from one population to another, which 437 can obscure differences in concentration-response relationships between neighborhoods and 438 population sub-groups. There is a trade-off, as large cohorts have more statistical power and 439 population-specific studies may be limited by large statistical error (e.g. Alexeeff et al., 2018). 440 Future studies may assess the potential for using population-specific relative risks to characterize 441 inequities in air pollution-related health risks.

442 While PM_{2.5} concentrations have been decreasing across the U.S. since 1990, owing to effective 443 environmental policies, PM_{2.5} air pollution still contributes 60,000-100,000 premature deaths 444 each year nationally (Fann et al., 2017; Goodkind et al., 2019; Murray et al., 2020), and these air 445 pollution-related health risks continue to be inequitably distributed (e.g. Colmer et al., 2020; 446 Tessum et al., 2019). Furthermore, ground-based monitoring continues to be sparsely distributed, 447 which is insufficient for assessing the spatial distribution of pollution levels and associated 448 health impacts within cities. Future studies may consider improving intra-city PM_{2.5} 449 concentration estimates by integrating multiple exposure assessment approaches, including low-450 cost sensors, mobile monitoring, statistical techniques such as land use regression modeling, 451 chemical transport modeling, and satellite observations to capture air pollution exposure 452 inequities more fully (Ahangar et al., 2019; Castillo et al., 2019; Hammer et al., 2020). 453 Estimation techniques to generate high-resolution baseline disease rates are also needed to 454 consider population vulnerability to air pollution, as inequities exist not just in exposure levels, 455 but in the health outcomes attributable to those exposure levels.

456 **5.** CONCLUSION

We assessed spatiotemporal trends in the health burden of PM_{2.5} pollution in Washington, DC
and its 51-statistical neighborhoods. While annual average PM_{2.5} concentrations have decreased
between 2000 and 2018, PM_{2.5} still contributes to disease burdens in the District, and PM_{2.5}attributable health impacts are unevenly and inequitably distributed. The highest attributable
burdens are estimated to occur in neighborhoods that have larger proportions of people of color,

- 462 as well as lower household income and lower educational attainment. Our results also indicate
- 463 that quantitatively characterizing neighborhood-scale differences in PM_{2.5}-related health risks
- 464 within cities, either in the U.S. or globally, would benefit from advances in fine resolution
- 465 information on both PM_{2.5} concentration data and intra-city baseline disease rates.

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473 **Open Research**

- 474 Health data at the neighborhood level for Washington, DC used in this study are not publicly
- 475 available due to confidentiality of patient information. The baseline disease data for this study
- 476 were made available to us through an Institutional Review Board request to the Washington, DC
- 477 Department of Health. Centers for Disease Control and Prevention (CDC) 500 Cities Project data
- 478 are publicly available at <u>https://www.cdc.gov/places/about/500-cities-2016-2019/index.html</u>.
- 479 Surface PM_{2.5} datasets for this study are referenced in Hammer et al. (2020) and van Donkelaar
- 480 et al. (2019), and are compiled and made publicly available at
- 481 <u>https://sites.wustl.edu/acag/datasets/surface-pm2-5/</u>. The population datasets from the
- 482 Socioeconomic Data and Applications Center (SEDAC) are publicly available at
- 483 https://sedac.ciesin.columbia.edu/data/set/gpw-v4-basic-demographic-characteristics-rev11.
- 484 Sociodemographic data used in this study are referenced in the *Health Equity Report: District of*
- 485 *Columbia 2018* (2019), and are publicly available in the US Census Bureau 2011-2015 American
- 486 Community Survey (ACS) 5-year estimates website (<u>https://www.census.gov/programs-</u>
- 487 <u>surveys/acs/technical-documentation/table-and-geography-changes/2015/5-year.html</u>), and life
- 488 expectancy data is available from the Washington, DC Department of Health Center for Policy,
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Supporting information to "Estimating intra-urban inequities in PM_{2.5}-attributable health impacts: A case study for Washington, DC."

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Table S1: Annual average $PM_{2.5}$ concentrations ($\mu g/m^3$) derived from high spatial resolution satellite data (Hammer et al., 2020; van Donkelaar et al., 2019) in each of the eight wards in Washington, DC as well as the District-wide mean, standard deviation (SD), minimum (Min), and maximum (Max) values for each year from 2000 to 2018.

Ye	ear	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016	2017	2018
	1	16.0	15.9	15.0	14.1	14.4	15.1	13.5	13.6	12.0	10.5	11.1	10.7	10.4	9.2	9.5	9.8	8.7	8.3	8.0
	2	16.7	16.6	15.5	14.6	14.8	15.3	13.8	13.9	12.3	10.7	11.3	10.9	10.6	9.6	9.6	10.0	9.1	8.9	8.9
	3	16.5	16.3	15.5	14.4	14.7	15.1	13.7	13.7	12.1	10.5	11.2	10.8	10.6	9.6	9.6	10.0	9.0	8.9	9.3
Word	4	16.1	15.9	14.9	14.2	14.4	15.2	13.4	13.6	12.2	10.6	11.3	10.8	10.4	9.2	9.6	9.9	8.7	8.3	8.3
vvaru	5	16.4	16.0	15.1	14.4	14.5	15.0	13.4	13.6	12.1	10.4	11.1	11.0	10.4	9.6	9.6	9.7	8.9	8.6	9.0
	6	16.5	16.2	15.4	14.4	14.6	15.2	13.6	13.7	12.1	10.4	11.2	11.0	10.6	9.7	9.6	10.0	9.0	8.8	9.1
	7	15.6	15.3	14.5	13.8	13.9	14.5	12.8	13.2	11.8	10.1	10.9	10.8	10.1	9.3	9.4	9.5	8.6	8.1	8.7
	8	15.6	15.2	14.6	13.7	13.9	14.5	13.0	13.1	11.8	10.1	11.0	11.1	10.4	9.4	9.6	9.7	8.7	8.1	8.7
Ye	ear	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016	2017	2018
	Mean	17.0	16.9	17.6	18.1	15.8	16.6	13.8	14.3	11.4	10.8	11.6	12.8	11.8	10.6	10.9	12.6	10.2	8.5	8.7
DC	SD	0.93	0.91	0.86	0.96	0.83	0.90	0.78	0.77	0.69	0.70	0.69	0.76	0.71	0.66	0.68	0.68	0.60	0.53	0.70
DC	Min	14.5	15	15.5	16	14	14.3	12	12.3	10	9	10	11	10	9	9.5	11	9	7	6.3
	Max	19	19	19.8	20.3	17.8	18.8	15.5	16	13	12	13	14.3	13	12	12	14	11.5	9.2	9.5

Table S2: $PM_{2.5}$ -attributable deaths and asthma ED visits throughout Washington, DC that would have been avoided if each year's $PM_{2.5}$ concentrations were set at 2018 levels (District-wide average $PM_{2.5} = 8.7 \mu g/m^3$). a) $PM_{2.5}$ -attributable all-cause mortality deaths that would have been avoided annually; and b) total cumulative $PM_{2.5}$ -attributable cases for each health endpoint that would have been avoided between 2000 and 2015 (2014-2018 for asthma ED visits).

a)	Year	Avoided Deaths	b)	Health Endpoint	Avoided Cases
	2000	279		COPD	58
	2001	210		IHD	538
	2002	183		LC	110
	2003	121		Stroke	48
	2004	135		All-Cause	1,617
	2005	166		Asthma	5
	2006	118			
	2007	122			
	2008	79			
	2009	28			
	2010	50			
	2011	44			
	2012	33			
	2013	15			
	2014	14			
	2015	21			

Table S3: Five-year (2011-2015) mean $PM_{2.5}$ -attributable mortality rates (per 100,000) and asthma ED visit rates (per 10,000) and selected social determinants of health in each statistical neighborhood in the District. Four neighborhoods (Georgetown, National Mall, Naval Station & Air Force, and Stadium Armory) were omitted given their low population counts and/or had suppressed values in the nine key drivers of health. Table is sorted in ascending order for estimated $PM_{2.5}$ -attributable all-cause mortality rate.

Neighborhood	Attribu	table Ra	ates (p	er 100),000 pe	eople)	Socio-Ec	onomic ind	icators		Life	Race	e and I	Ethnicity	
	All- Cause	Asthma ED Visits	COPD	IHD	Lung Cancer	Stroke	Education (%)	Employmen (%)	tMedian Household Income (\$)	Residents Living in Poverty (%)	at Birth (years)	% White	% Black	% Hispanic/ Latino	% / Asian
WOODLEY PARK	17	1	3	7	3.1	2.3	98	3	139,744	7	89	71	8	12	6
GEORGETOWN EAST	17	1	2	11	2.8	0.7	99	3	132,021	10	87	73	5	11	8
GWU	18	1	2	20	2.8	0.7	NA	NA	NA	NA	NA	69	6	11	11
ADAMS MORGAN	20	3	2	9	3.9	1.1	96	5	96,194	7	85	63	12	16	6
U ST/PLEASANT	21	5	2	12	4.1	1.2	89	7	94,614	12	82	45	34	13	5
KENT/PALISADES	22	1	3	13	3.1	2.3	98	6	161,252	9	88	71	7	11	8
CAPITOL HILL	23	4	3	12	5.1	1.8	98	3	121,668	6	86	65	21	7	3
LOGAN CIRCLE/SHAW	28	4	2	19	3.4	1.0	91	4	94,043	11	81	54	24	11	8
SHEPHERD PARK	28	2	3	16	4.9	2.2	93	12	102,053	11	83	29	46	21	2
TENLEYTOWN	28	1	3	16	3.1	2.3	99	2	136,641	5	87	71	7	11	7
Lowest attributable rates (10-neighborhood mean)	22	2	2	14	4	2	95	5	119,803	9	86	61	17	12	6
CATHEDRAL HEIGHTS	30	1	3	19	3.0	2.1	97	4	90,124	16	89	71	7	11	7
SOUTH COLUMBIA HEIGHTS	30	3	2	17	4.1	1.2	90	8	82,241	14	79	45	28	18	6
FOREST HILLS	31	1	3	18	3.1	2.3	99	4	113,269	9	87	67	11	14	5
MOUNT PLEASANT	32	6	2	30	4.1	1.2	89	5	71,837	12	79	45	21	26	4
UNION STATION	32	7	3	16	5.1	1.8	95	5	110,907	10	78	51	36	6	5
16th ST HEIGHTS	33	6	3	18	4.8	2.2	83	8	75,848	13	80	32	38	24	4
SAINT ELIZABETHS	33	14	5	17	8.8	2.9	NA	18	25,311	40	68	15	78	5	1
CHINATOWN	38	4	2	24	3.8	1.2	89	5	82,789	18	78	52	24	11	10
COLUMBIA HEIGHTS	41	6	2	24	4.2	1.2	79	7	70,554	17	80	37	32	23	5
HILL EAST	41	3	3	26	5.1	1.8	92	9	92,617	14	78	47	41	6	2
DC MEDICAL CENTER	44	6	4	57	6.3	3.1	NA	NA	NA	NA	NA	30	49	15	3
BARNABY WOODS	44	1	3	22	4.8	2.2	99	3	200,031	2	87	53	26	14	5
BLOOMINGDALE	46	6	4	33	6.4	3.1	91	9	87,146	12	76	37	47	7	5
DOUGLASS	47	13	5	34	8.7	2.8	82	23	31,319	37	72	2	92	3	0
KINGMAN PARK	48	7	3	31	5.3	2.0	92	8	91,073	12	77	39	49	6	2
HISTORIC ANACOSTIA	50	13	5	27	8.9	2.9	83	15	28,790	37	70	3	92	3	0
EASTLAND GARDENS	52	14	4	27	7.5	3.6	79	21	31,333	34	73	3	91	4	0
LAMOND RIGGS	52	6	4	27	5.3	2.5	89	15	67,745	9	81	17	62	17	2
TWINING	52	13	4	25	8.0	3.3	88	16	47,486	21	75	4	90	3	1
PETWORTH	52	6	3	29	4.8	2.2	86	12	77,020	13	79	22	51	22	3
BRENTWOOD	52	9	4	33	6.4	3.2	87	15	61,739	19	77	20	67	8	3
BRIGHTWOOD	53	5	3	28	4.9	2.2	84	9	66,395	13	81	25	50	19	2
SW/WATERFRONT	57	7	3	33	5.1	1.8	94	7	76,429	14	78	50	35	8	4
BELLEVUE	57	15	5	38	8.8	2.9	83	30	32,562	40	74	8	86	3	0
CONGRESS HEIGHTS/SHIPLE	\ 59	16	5	58	8.8	2.9	82	27	28,711	39	72	2	93	3	0
EDGEWOOD	60	7	4	46	6.4	3.2	84	20	41,171	29	79	31	52	11	4
NAYLOR/HILLCREST	61	13	4	39	7.8	3.3	84	17	37,771	35	73	3	91	3	1

MICHIGAN PARK	62	6	4	38	6.4	3.2	86	16	57,943	12	82	24	56	15	2
TRINIDAD	66	7	4	36	6.4	3.2	80	18	36,655	29	71	27	62	6	3
FORT LINCOLN/GATEWAY	66	9	4	55	6.4	3.2	81	14	51,454	19	76	9	82	6	1
BRIGHTWOOD PARK	66	6	3	40	4.9	2.2	87	10	61,476	16	77	16	60	19	2
WOODRIDGE	66	9	4	42	6.5	3.2	93	14	85,947	11	79	17	69	9	2
CHEVY CHASE	68	1	3	40	3.8	2.3	94	4	115,697	9	83	73	8	10	6
LINCOLN HEIGHTS	72	14	4	38	7.6	3.6	81	21	36,577	26	73	1	93	4	0
WASHINGTON HIGHLANDS	74	16	5	46	8.7	2.8	NA	NA	28,468	39	72	1	95	2	0
MARSHALL HEIGHTS	74	14	4	38	7.4	3.5	84	20	43,043	29	72	1	94	3	0
FORT DUPONT	90	13	4	46	7.5	3.6	82	24	35,545	31	75	1	93	4	1
Highest attributable rates (10-neighborhood mean)	70	9	4	42	7	3	85	16	55,281	22	76	17	71	8	2
District of Columbia means	46	7	4	29	6	2	89%	10	\$70,848	40%	79	36%	46%	11%	4%



Fig S1: Five-year mean $PM_{2.5}$ concentrations ($\mu g/m^3$) for a) 2011-2015 and c) 2014-2018 (van Donkelaar et al., 2019); and Socioeconomic Data and Applications Center (SEDAC) 2010 population count estimates from the Gridded Population of the World (GPW), Version 4, by the Center for International Earth Science Information Network (Center For International Earth Science Information Network (Center For International Earth Science Information Network-CIESIN-Columbia University, 2018) for b) 30-99 years of age and d) all ages. Estimates are displayed on a 100 m x 100 m resolution grid overlaid on the outline of the District of Columbia.



Year Satellite Monitor Satellite Monitor Satellite Monitor	McMillan				
	itor				
2011 10.9 10.4 10.7 10.2 11.1 10.3	6				
2012 10.7 9.8 10.5 9.8 10.4 9.6					
2013 9.7 9.3 8.4 8.3 9.7 9.1					
2014 9.7 10.2 9.6 9.1 9.6 9.4					
2015 10.1 NA 10.0 9.2 9.8 8.9					
2011-2015 10.2 9.9 9.8 9.3 10.1 9.6					

Fig S2: Comparison of PM_{2.5} concentrations (μ g/m³) derived from high spatial resolution satellite data (Hammer et al., 2020; van Donkelaar et al., 2019) and ground-based monitors (Federal Reference and Equivalent Methods - FRM/FEM). Top: Five-year (2011 - 2015) average concentrations (μ g/m³) from satellite data (raster layer at ~1 x 1km) and three ground-based monitors (located in Wards 1, 2 and 7, in PNGs U St/Pleasant, National Mall, and Fort Dupont, respectively). Bottom: Annual comparisons of PM_{2.5} concentrations (μ g/m³) between 2011 and 2015, as well as 5-year average comparison for each of the three monitoring stations that were available for these years. Satellite data values correspond to the grid cells where the ground-based monitors are located.



Fig S3: District of Columbia's 51 Proximal Neighborhood Groups (PNGs) defined by the District of Columbia Department of Health (DOH) and Ward overlays (1-8), and a list of PNGs abbreviations and names. Shapefiles retrieved from Open Data DC (https://opendata.dc.gov/) and Figure produced in Quantum Geographic Information System (QGIS 3.6.2).



Figure S4: PM_{2.5}-attributable all-cause mortality rates (per 100,000 people) and life expectancy at birth (in years) from the District of Columbia Department of Health (DOH) Center for Policy, Planning and Evaluation. Data represent 2011-2015 means.



Fig S5: Five-year mean baseline disease rates (cases per 100,000 people for all mortality outcomes and per 10,000 people for asthma ED visits) from the District of Columbia Department of Health between 2011 and 2015 for all mortality outcomes, and between 2014 to 2018 for asthma ED visits. Data resolution is at the neighborhood-level for all-cause mortality (ALL) and ischaemic heart disease (IHD); zip code-level for asthma ED visits (AST); and ward-level for chronic obstructive pulmonary disease (COPD), lung cancer (LC), and stroke (STR). Colors represent five equal intervals for each health outcome of interest.

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