

Reviewing moisture recycling dynamics: implications of land use change on green and atmospheric water

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Abstract

Green water, or plant-available soil moisture, is a substantial subset of terrestrial fresh water. Land use change alters green water dynamics directly, by changing soil and vegetation properties, and indirectly, via feedbacks in the soil-vegetation-climate system. Ongoing global deforestation, and growing interest in reforestation projects, begs the question: Do such large-scale land use changes have major eco-hydrological impacts via the process of terrestrial moisture recycling (TMR)? This requires a systematic, mechanistic understanding of green water dynamics in relation to land use change, and the interactions with the soil-vegetation-climate system in which it is embedded. Hence, this literature review addresses the above question via a scoping review that draws from papers covering empirical observations and simulated approximations on the hydrological effects of land use change from different parts of the world. The results show that some regions are more vulnerable to land use change than others and can affect local as well as distant hydrology of landscapes. Furthermore, we derive analytical tools and directions for further research that can improve understanding of the effects of land use change on moisture recycling dynamics in order to minimize unexpected hydrological impacts for nature and society.

1 **Title:** Reviewing moisture recycling dynamics: implications of land use change on green and
2 atmospheric water

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18 **Key words:** moisture recycling; land use change; deforestation; reforestation, green water;
19 atmospheric water

20 **Key Points:**

- 21 • Advances in hydrological modeling increasingly broaden our understanding of land use
22 change effects on moisture recycling dynamics, although there was no overarching review on
23 the issue yet.

- 24 • Spatial and temporal dynamics of moisture recycling are highly variable, but the
25 hydroclimatic effects of land use changes on these patterns remain – although sensible
26 considering the processes of scale and uncertainties due to water’s active role in the
27 atmosphere– under-researched.
- 28 • There is a need to increase our understanding of context-specific land use change effects on
29 moisture recycling dynamics via case study research to evaluate potential hydroclimatic
30 effects and prevent unintended consequences on water resources.
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32 University of Amsterdam

33 Abstract

34 Green water, or plant-available soil moisture, is a substantial subset of terrestrial fresh water. Land
35 use change alters green water dynamics directly, by changing soil and vegetation properties, and
36 indirectly, via feedbacks in the soil-vegetation-climate system. Ongoing global deforestation, and
37 growing interest in reforestation projects, begs the question: Do such large-scale land use changes
38 have major eco-hydrological impacts via the process of terrestrial moisture recycling (TMR)? This
39 requires a systematic, mechanistic understanding of green water dynamics in relation to land use
40 change, and the interactions with the soil-vegetation-climate system in which it is embedded. Hence,
41 this literature review addresses the above question via a scoping review that draws from papers
42 covering empirical observations and simulated approximations on the hydrological effects of land use
43 change from different parts of the world. The results show that some regions are more vulnerable to
44 land use change than others and can affect local as well as distant hydrology of landscapes.
45 Furthermore, we derive analytical tools and directions for further research that can improve
46 understanding of the effects of land use change on moisture recycling dynamics in order to minimize
47 unexpected hydrological impacts for nature and society.

48

67 Anthropogenic land use change (LUC) following increasing demand for food, fuel, fiber and timber
68 (Schyns et al. 2019) might affect TMR. Some studies suggest that deforestation and vegetation
69 reduction can disturb TMR and affect local to regional rainfall patterns (Keune and Miralles 2019;
70 Savenije 1995; Zemp et al. 2014; Zemp et al. 2017). Deforestation and land degradation leads to loss
71 of natural ecosystems and carries the risk of crossing ecological boundaries that affect green water
72 dynamics and TMR patterns (Zemp et al. 2017). Simultaneously, there is a growing interest in
73 afforestation for *biological capture-biological storage* (BCSC) of carbon and restoring ecosystems in
74 general for various other Nature's Contributions to People (NCP). Between 2000-2012, 80 million
75 hectares were re- or afforested (Bentley & Coomes, 2020), mainly temperate forests (Fagan et al.
76 2020). Reforestation is promoted by the UN declaring 2020-2030 as the decade of ecosystem
77 restoration (UN, 2019); world leaders in Davos committing to planting 1 billion trees (i.e. the One
78 Trillion Tree Initiative); afforestation is increasingly interesting for commercial carbon sequestration
79 approaches for climate change mitigation in line with the Paris Agreement (UN, 2015). Bastin et al.
80 (2019) estimates the global tree restoration potential to cover 0.9 billion ha of canopy cover, which
81 can store 205 Gt of carbon. Furthermore, deliberate bio-geoengineering with forest plantation can
82 change regional climate and rainfall via land-atmosphere interactions (Branch & Wulfmeyer, 2019).
83 However, tree planting runs the risk of distorting basin hydrology and sediment dynamics (Farley et
84 al. 2005), as has occurred in many forestry projects worldwide (e.g. introduction of exotic Eucalyptus
85 in South Africa) (Albaugh et al., 2013). Accordingly, the impact of both de- and reforestation on the
86 hydrological cycle should be addressed given scarce water resources (Sterling et al., 2013). There is
87 lack of clarity concerning the conditions under which TMR patterns can be distorted or intensified via
88 LUC (e.g. Spracklen et al. 2018). This research therefore addresses the question: Do large-scale land
89 use changes have major eco-hydrological impacts via the process of terrestrial moisture recycling
90 (TMR)? We answer this question through a scoping literature review addressing both empirical
91 observations and simulated predictions from across the globe in order to provide a state-of-the-art
92 synthesis on the effect of LUC on TMR.

93 We first provide a historical and theoretical background, describe the methodology for the review, and
94 present the results, including global and regional assessments of the empirical effects of LUC and
95 implications for governance.

96 2. Historical and theoretical background

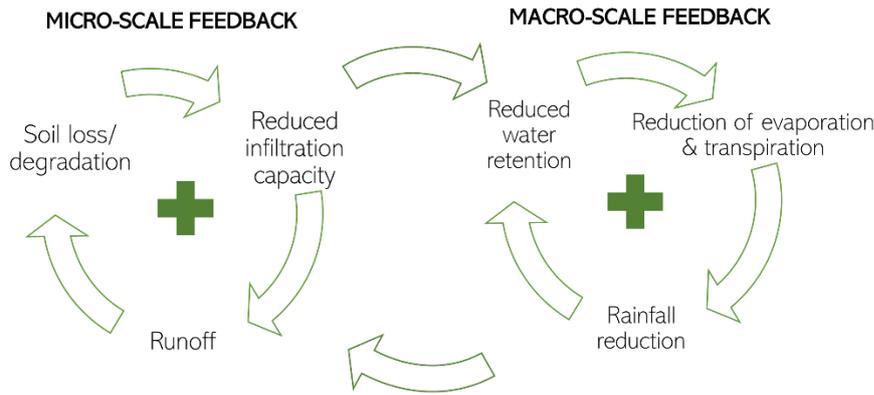
97 Historically, human-induced patterns of vegetation change have altered large areas of the earth's
98 surface and hydrology. The debate on the effect of forests on hydrology centers around the question
99 whether trees are net *water users* or net *water producers* (Andréassian 2004; Ellison, Futter, and
100 Bishop 2012). Forests use water via transpiration and evaporation (reducing local water availability),
101 but they also enhance infiltration and the water retention capacity of the soil (increasing local water
102 availability). The trade-offs between these processes in specific contexts determine whether
103 vegetation is a water user or producer (Peña-Arancibia et al., 2019). To address the effect of forests on
104 a catchment level, many hydrological studies using *paired-catchment approaches* have been
105 performed since the 1970's (Bosch and Hewlett 1982). Forest removal generally shows increases in
106 streamflow, whereas forest establishment reduces streamflow (on average 23% over 5 years and 38%
107 over 25 years) (Farley et al., 2005; Filoso et al., 2017). Yet, forest increase also reduces peak flows
108 and damaging floods as it increases infiltration capacity, and in some cases, streamflow has partially
109 recovered (Bentley & Coomes, 2020). The hydrological effects of forest removal and restoration on
110 catchment hydrology remain variable due to many different landscape variables at work (Andréassian,
111 2004; Filoso et al., 2017).

112 On a planetary scale, the biophysical properties of vegetation regulate the hydrological cycle and
113 climate. Interactions between the biosphere and atmosphere include exchanges of water, energy,
114 momentum (biophysical interaction), and gases (biogeochemical interaction), which co-produce
115 observed climate patterns. Exploring these interactions with computational models has increased our
116 understanding of land cover effects on the global climate. The illustrative model Daisyworld (Watson
117 & Lovelock, 1983) shows the self-regulating properties of vegetation (daisy flowers) that stabilize
118 atmospheric temperature via radiative feedbacks. A similar computational thought-experiment by

119 Kleidon, Fraedrich, and Heimann (2000) investigates the effect of vegetation on the climate system by
120 conceptualizing two contradicting worlds: a ‘desert world’ and a ‘green planet’, accounting for both
121 radiative and hydrological feedbacks. The simulation shows that a green planet produces three times
122 more continental evaporation and transpiration, two times more precipitation and results in a decrease
123 in surface temperature. TMR increases due to the higher energy availability through absorbed
124 radiation and due to increased soil moisture retention capacity associated with tree cover. Although
125 such extreme models are unrealistic, they illustrate the significant climatic effect of interactions
126 within the biosphere-atmosphere system.

127 **2.1 Theory of moisture recycling and land use change**

128 The theory of forest-rainfall connections dates back to the 15th century (see Bennett and Barton 2018).
129 Observations during the European exploration of the Americas have led naturalists to argue that
130 rainfall over dense continental forests derived from forest evaporation itself. Furthermore,
131 deforestation on colonized islands, such as the Azores, led to observations of reduced rainfall, but
132 without tools to quantify such dynamics, these theories remained unverified (Bennett and Barton
133 2018). Biogeographers generally assumed that observed vegetation patterns were a consequence of
134 assuming more-or-less stable weather patterns (e.g. rainfall is an external variable that is not
135 influenced by the vegetation itself) (van Noordwijk & Ellison, 2019). In the 1970’s, rainfall
136 reductions in the Sahel were linked to reduced vegetation cover resulting from overgrazing and
137 landscape degradation. Savenije (1995) developed a moisture recycling theory based on hydrological
138 processes, confirming the mechanistic role of vegetation reductions on drought spells. More recent
139 TMR studies show a strong dependency on recycled rainfall in wet tropical regions (i.e. the Amazon
140 and Congo basin) (Wang-Erlandsson et al., 2018). Advances in computer models and the availability
141 of global climate data reinforced a revival of the inquiry into TMR, questioning the extent to which
142 the earth surface, and particularly vegetation, contributes to rainfall patterns via the exchange of mass,
143 energy and momentum (Eltahir and Bras 1996; Bennett and Barton 2018). This gave rise to the idea
144 that forests *generate* rainfall. As such, deforestation would result in rainfall reductions via interacting
145 feedbacks at the micro- and macroscale (Figure 2)



146

147 Figure 2 Interaction between micro- and macroscale hydrological feedbacks in the soil-vegetation-climate system following
 148 reduction of vegetation cover. On the microlevel, the loss of vegetation cover reduces the infiltration capacity due to changes
 149 in the soil (i.e. rooting structure, desiccation). This increases runoff and results in soil degradation, further reducing soil
 150 infiltration capacity. On the macro-scale, the reduction in infiltration capacity reduces soil water retention, which results in
 151 lower evaporation and transpiration rates. Subsequently, this would reduce the amount of precipitable water in the
 152 atmosphere and leads to lower rainfall.

153 2.2 Approaches, tools and methods to address moisture recycling

154 The hydrological toolbox to assess the effect of LUC on rainfall comprises of computational,
 155 statistical and chemical methods. Computational methods use coupled land surface and vegetation
 156 models to climate models to represent relevant interactions between the biosphere and atmosphere.
 157 On global scales, General Circulation Models (GCMs) and dynamic vegetation models have been
 158 coupled to simulate interactions between climate and vegetation (Foley et al., 1998). Atmospheric
 159 moisture tracking models that are forced with meteorological data (i.e. ERA-Interim data) can identify
 160 source and sinks of atmospheric moisture, which allows tracking of moisture forward and backward in
 161 time (Keune & Miralles, 2019; van der Ent et al., 2014; Zemp et al., 2014). Subsequently, such
 162 simulations can be summarized into metrics representing regional dependency on recycled moisture.
 163 The *precipitation recycling ratio* ρ , for example, is defined as the fraction of precipitation that derives
 164 from land surface evaporation (P_E) over the fraction deriving from oceanic sources (P_O) (van der Ent
 165 et al., 2014):

$$\rho = \frac{P_E}{P_O}$$

166 Vice versa, the *evaporation recycling ratio* describes the fraction of regional evaporation which
 167 returns as precipitation over land. Statistical approaches use remote sensing measurements that link

168 LUC to changes in rainfall. Changes in total evaporation and transpiration (TET) over forests are
169 measured using flux towers or satellite imagery and climate data (Shivers et al., 2019). Yet, causality
170 is difficult to prove due to the influence of many other biophysical and climatic factors (e.g. meso-
171 scale atmospheric circulations) (Spracklen et al., 2018). Chemical approaches use isotope
172 measurements that allow backtracking of different moisture sources and their contribution to local
173 rainfall (Zhao et al. 2019). Stable isotope ratios of hydrogen and oxygen (i.e. the isotopic
174 compositions) vary between different sources of moisture (e.g. advection, evaporation or
175 transpiration) hence reflect information about the source of atmospheric moisture (Gat, 1996).

176 Precipitation and evaporation recycling ratios are measures of strength of hydrological land surface-
177 atmosphere coupling and are used to identify local, regional or distant rainfall responses of surface
178 evaporation and transpiration (Goessling & Reick, 2011). They are shape- and scale-dependent: the
179 evaporation of an infinitely small area would have negligible contribution to precipitation while the
180 whole earth would have a moisture recycling ratio of 1 (Trenberth 1999). The relation between scale
181 and recycling follows a non-linear relationship due to the spatial heterogeneity encountered with
182 scaling up or down (Dominguez et al., 2006). The *precipitationshed* (Keys et al. 2012) captures the
183 spatial dependence between source and sink regions of atmospheric moisture. It is “the upwind
184 atmosphere and upwind terrestrial land surface that contributes evaporation to a specific location’s
185 precipitation (e.g. rainfall)” (Keys et al. 2012: 734). It represents an analytical framework to identify
186 the source area of precipitation in a region of interest (i.e. sink region). Vice versa, the
187 *evaporationshed* identifies the sink area of evaporation from a given area. The frameworks build on
188 the concept of an *atmospheric river*, a feature in the hydroclimate that transports large amounts of
189 water vapour from the ocean inland. Contrary to watersheds, precipitationsheds are probabilistic, in
190 the sense that they do not have fixed borders, and are subjected to inter and intra-annual variation
191 (Keys et al. 2012). Similarly, the concept of a *watershed precipitation recycling network* (Keune &
192 Miralles, 2019) establishes atmospheric moisture connections on a watershed level, to identify how
193 evaporation from one watershed contributes to precipitation in another (Keys, Wang-Erlandsson, and
194 Gordon 2016). Finally, the concept of *moisture cascades* describes ‘moisture transport between two

195 *locations on the continent that involves re-evaporation cycles along the way*' (Zemp et al. 2017, 2014)
 196 and addresses the *hydrological connectivity* of regions (Schaepli et al., 2012; Van Der Ent et al., 2010;
 197 Van Der Ent & Savenije, 2011). Although TMR estimates are limited predictors of the effect of
 198 changes in evaporation to precipitation due to a sequence of processes occurring in the atmosphere
 199 (Goessling & Reick, 2011) – they are useful to examine a region's vulnerability to changes in
 200 evaporation within the precipitationshed.

201 3. Methodology

202 As there was no existing systematic review paper, a scoping review of the literature on moisture
 203 recycling dynamics was carried out with the following search criteria on Scopus: "Terrestrial moisture
 204 recycling" OR "Moisture Recycling" AND "Atmospheric" OR "Land-Atmosphere" OR "Land-
 205 atmosphere dynamics" OR "Land-use change" AND "moisture recycling" (1106 search results on 5-
 206 10-2020). Relevant literature was selected and subsequently, using backtracking and hand searching,
 207 additional literature was added. The references were analyzed for 1) relevant mechanistic relations
 208 and feedbacks in the soil-vegetation-climate system, specifically micro- and macro-scale dynamics
 209 and 2) empirical observations and modelling simulations of quantitative hydrological change in
 210 relation to LUC. Evidence from different continents and climate regions using Köppen-Geiger
 211 classification is included to account for regional variation (Table 1).

212 Table 1 Reviewed literature and their covered bioclimatic zone and region/countries.

	BIOCLIMATIC ZONE (KÖPPEN)	REGION/ COUNTRY	# PAPERS	REFERENCES
NORTH AMERICA	Bsk; Dfb	US; Canada	3	(Dominguez et al., 2006); (Meng & Quiring, 2010); (Raddatz, 2005)
SOUTH AMERICA	Af/Am/ Aw	Amazon; Brazil; Bolivia	9	(Bagley et al., 2014); (Boers et al., 2017); (Butt et al., 2011); (Makarieva et al., 2014); (Staal et al., 2018); (Weng et al., 2019); (D.C. Zemp et al., 2014);(D.C. Zemp et al., 2017) (Delphine Clara Zemp et al., 2017)
EUROPE	Dfb; Ds; Cfb; Csa;	Central Europe; Iberian Peninsula; United Kingdom; Eastern Mediterranean	7	(Bisselink & Dolman, 2009); (Kelemen et al., 2016); (Rios-Entenza et al., 2014);(Robinson et al., 2016); (Zangvil et al., 2010); (Cammeraat et al., 2010); (Nadal-Romero et al., 2016)
AFRICA	BWh;	Sahel; Zimbabwe;	10	(Yu et al., 2017) (Savenije, 2004)

	BSh; Aw/As; BShs; Aw; Af	Ethiopia; South Africa; Nile basin; Cameroon; Congo Basin; North-Africa		(Castelli et al., 2019); (van Luijk et al., 2013); (Mohamed et al., 2005); (Njitchoua et al., 1999); (Saeed et al., 2013); (Savenije, 1995); (Yu et al., 2018) (Bamba et al., 2019)
ASIA	BWk; BSk; BSk/ET; BSk; Dfc; Cfa; BSk;	Northern-China; Tibetan Plateau; Arid eastern-central Asia; Central Siberian Plateau; South China; Tianshan Mountains; Russia; Ganges basin; East China	12	(Zhao et al., 2019); (Bai et al., 2019); (An et al., 2017); (Dong et al., 2018); (Ford & Frauenfeld, 2016); (Guo et al., 2019); (Huang et al., 2018); (Kong & Pang, 2016); (Kurita & Yamada, 2008); (Notaro & Liu, 2008); (Tuinenburg et al., 2012); (Z. Yang et al., 2019)
AUSTRALIA /PACIFIC	BWh/ BSh	Australia; New Zealand	3	(Syktus & McAlpine, 2016); (Szeto, 2002); (Vervoort et al., 2009)
GLOBAL	-	-	16	(Green et al., 2017); (Wang-Erlandsson et al., 2014); (van der Ent et al. 2014); (Duveiller et al., 2018); (Gerten et al., 2004); (Van Der Ent et al., 2010); (Sterling et al., 2013); (Peña-Arancibia et al., 2019); (Wang-Erlandsson et al., 2018); (Keys, Wang-Erlandsson, and Gordon 2016); (Trenberth 1998) (K.E. Trenberth, 1999) (Keys et al. 2012); (Keys et al. 2014); (D.G. Miralles et al., 2016); (Spracklen et al., 2012) (Spracklen et al., 2018)

213 4. Results

214 This chapter represents the findings from the literature review and is divided into a framework
215 describing the general dynamics of moisture recycling (see 4.1), and an empirical framework
216 addressing simulated and observed evidence of the impact of LUC on precipitation patterns (see 4.2)
217 and implications for governance (see 4.3).

218 4.1 Dynamics of moisture recycling

219 The literature on TMR shows that there is a large spatial and temporal variation in the regional
220 dependence on recycled moisture (Table 2). Some regions receive the majority of precipitation from
221 oceanic sources (e.g. western Europe), while others depend on moisture from continental origin (e.g.
222 inland regions such as the East African savanna and Mongolian steppe) (Miralles et al., 2016). There
223 are ‘hotspots’ of regionally strong precipitation feedbacks in transitional zones (grasslands and
224 savannas), such as semi-arid and monsoonal regions (Green et al. 2017) and of moisture recycling in
225 regions where orographic lift drives precipitation events (Van Der Ent et al. 2010), in sub-tropical
226 highlands with high evaporation and small advective moisture fluxes, and in convergence zones
227 (Trenberth, 1999). Gradients of increased moisture recycling dependency moving further away from
228 the coast have been observed in Cameroon (Njitchoua et al., 1999) and the Iberian Peninsula (Rios-
229 Entenza et al., 2014).

230 Table 2 Three examples of water basin’s internal recycling (i.e. amount of rainfall deriving from the basin itself) dependency
 231 regimes.

	PRECIPITATION RECYCLING	RECYCLING DEPENDENCY	REFERENCE
Nile basin	8-14%	Low	(Mohamed et al., 2005).
Amazon basin	32% (60% from transpiration)	High	(Staal et al., 2018).
Ganges basin	5-60%	Seasonal variable	(Tuinenburg et al., 2012).

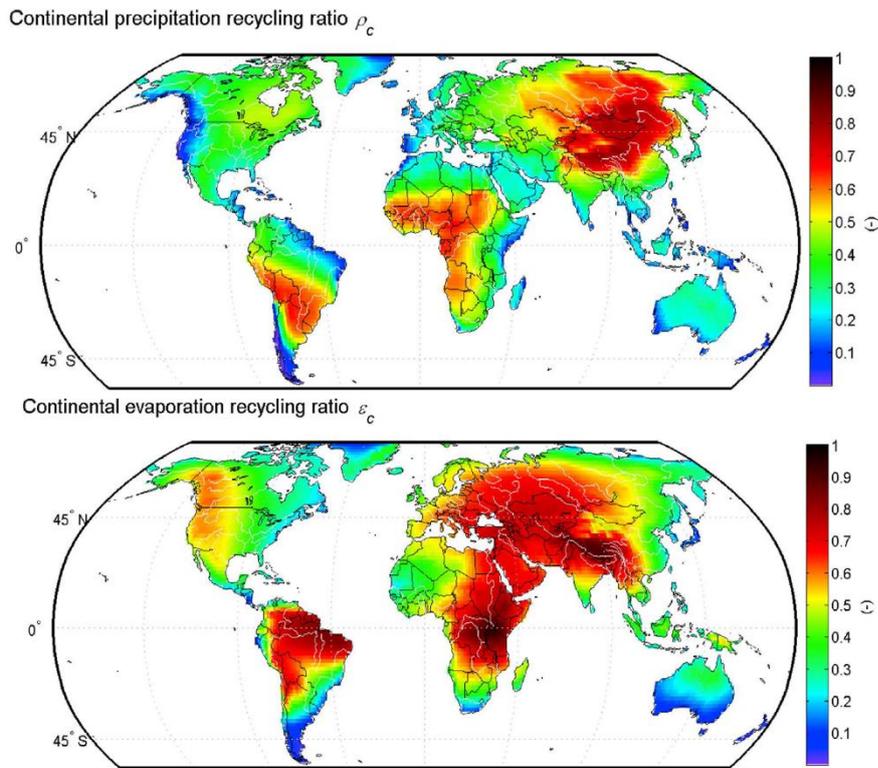
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233 Seasonal variation in terrestrial moisture recycling is caused by the warmer land surface compared to
 234 the ocean during summer, resulting in higher continental precipitation recycling ratios (Dominguez et
 235 al., 2006 Szeto, 2002). Higher moisture availability at the land surface in the wet season increases the
 236 relative importance of surface evaporation to precipitation (Van Der Ent et al. 2010). In summer,
 237 74% of precipitation over watersheds in Europe derive from evaporated moisture supplied by other
 238 watersheds (Keune & Miralles, 2019). Some regions depend highly on recycled moisture to produce
 239 peak spring precipitation (Rios-Entenza et al. 2014). Intra-annual variation in moisture recycling
 240 patterns can be caused by weather cycles, such as El Niño Southern Oscillation (Yang et al. 2018), the
 241 North Atlantic Oscillation and monsoonal cycles (Guo et al., 2019). Weather anomalies, such as
 242 extreme precipitation or drought events can be traced back to high continental evaporation (Kelemen
 243 et al., 2016) or low advection (Bisselink & Dolman, 2009) respectively, the latter showing increasing
 244 importance of local evaporation to sustain precipitation. Extreme rainfall events in the Congo basin
 245 were linked to moisture recycling *reductions* due to relative lower soil moisture availability and
 246 higher surface runoff (Saeed et al., 2013).

247 As precipitation length scales vary between 500-7000 km (Van Der Ent & Savenije, 2011),
 248 evaporated water is likely to precipitate *outside* the water basin it originates from. In northern China,
 249 15-50% of the precipitation is derived from local (i.e. within the water basin) terrestrial moisture
 250 (Zhao et al., 2019). China is for 80% dependent on continental evaporated moisture (Figure 3),
 251 although regional variation is large. Rainfall in forests in the southwest of the Amazon basin derives
 252 largely from transpiration and evaporation in other parts of the basin (Staal et al., 2018). The Congo
 253 basin depends largely on evaporated moisture from East-Africa, and supplies rainfall to the Sahel
 254 region. *Moisture recycling cascades* in this region appear established due to dominant continental

255 wind patterns (Zemp et al. 2014). Moisture recycling cascades over South America contribute around
256 10% of the total precipitation over the continent. In the La Plata basin, 17-18% of the rainfall derives
257 from such cascades, generally deriving from the Amazon due to the topography of the Andes
258 mountains guiding the moist air from the Amazon downward to the La Plata basin. Local moisture
259 recycling in mountainous regions (e.g. Tibetan Plateau, the Andes) is dominant due to orographic lift
260 (Kong & Pang, 2016). Moisture recycling estimates from the Rocky Mountains in the US show a
261 higher ratio around the mountain range (Dominguez et al., 2006). Around the Tibetan plateau,
262 estimates show that 50-80% of the precipitation derives from locally evaporated water (Kurita &
263 Yamada, 2008) (An et al., 2017). An observed increase in moisture recycling may be caused by
264 climate change, which increases both evaporation and precipitation rates in the region (An et al.,
265 2017).

266 In tropical regions, precipitation length scales are generally shorter (500-2000 km) and driven by
267 monsoonal dynamics with intense feedbacks and short atmospheric lifetimes. In the Amazon, roughly
268 one-third of the rainfall derives from the basin itself, of which 60% comes from plant transpiration
269 (Staal et al., 2018). The ability of these plants to access deeper soil moisture can be important to
270 remain transpiration flows in the dry season (Wang-Erlandsson et al., 2014) and sustain precipitation
271 even when advection from the ocean is low (Staal et al., 2018). On average, 46% of the transpiration
272 falls back as precipitation in the basin itself, while in the dry-season this can amount up to 70% (Staal
273 et al., 2018). In the Ganges basin, moisture recycling varies between 5-60% and is low in winter and
274 high in summer during the monsoon. Spatial variation in the atmospheric water budget (70% inter-
275 basin difference) is most likely caused by irrigation schemes, increasing evaporation locally, even
276 during the dry season (Tuinenburg et al., 2012). TET from Indian irrigation schemes alone may
277 support 40% of the rainfall in regions in East Africa (de Vrese et al., 2016). When evaporation is high,
278 the distance of moisture travelled is generally shorter. This might be caused by convection
279 subsequently triggering local precipitation.



280

281 Figure 3 Global continental precipitation recycling ratio (ρ_c) and evaporation recycling ratio (ϵ_c). Figure copied with the
 282 author's permission from Van der Ent et al (2010).

283 In water limited regions, temporal variation in the fraction of terrestrial moisture recycling between
 284 the wet and dry season is small. In the Nile basin, the inter-annual moisture recycling variation is low
 285 (between 8-14%). Annually, more than 89% of the water resources originate from outside the basin
 286 itself (Mohamed et al., 2005). Comparing wet season recycling ratios of water limited regions shows
 287 that in the South American Pampas, recycling is only 3%, whereas in the Kalahari, it is 28%. In the
 288 dry season, recycling in the Kalahari reaches up to 34% (Miralles et al., 2016). In the Sahel, local
 289 moisture recycling appears strong in the post-monsoon period due to wet soils and high vegetation
 290 growth (Yu et al., 2017). Observations of high vegetation productivity in seasonally dry regions
 291 correlate with increases in evaporation and transpiration and lead to increasing precipitation (Green et
 292 al., 2017). This implies that in dry regions, retaining water locally (i.e. preventing quick run-off),
 293 might result in an intensification of local precipitation in the wet season and post-monsoon period
 294 (Figure 4). Many semi-arid regions are depending on recycled moisture for agricultural production
 295 during the growing season which also makes them social-economically vulnerable to changes in
 296 precipitation (Dominguez et al., 2006). Vice versa, dry spells in these regions can facilitate positive

297 land-atmosphere feedbacks that can amplify drought (Miralles et al., 2016). Thus, patterns of TMR in
298 time and space appear highly variable and influenced by local geography, climate, topography and
299 vegetation properties.



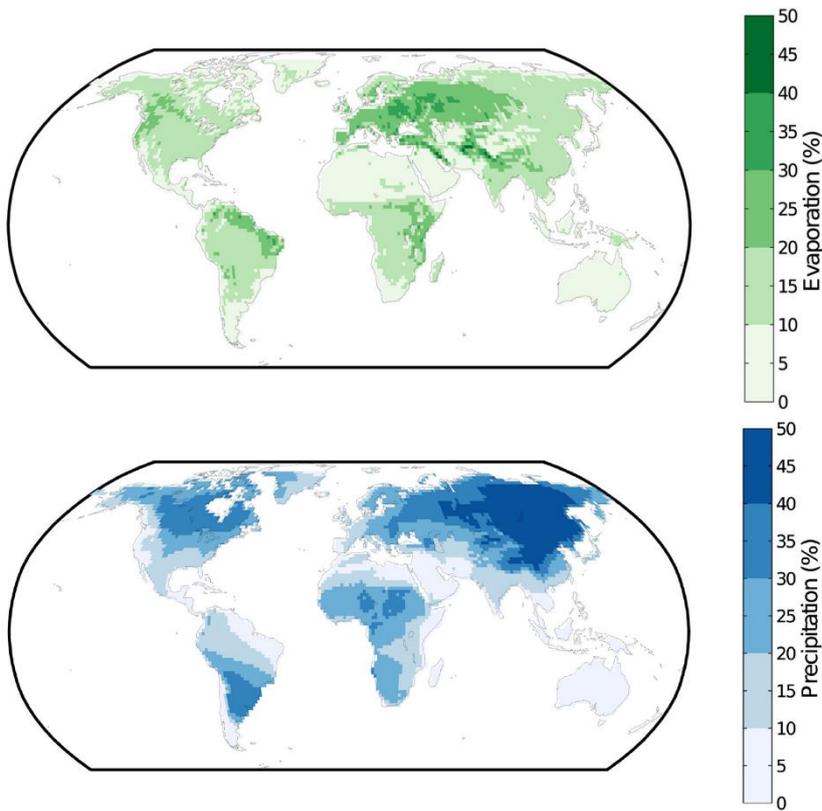
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301 Figure 4 Constructing half-moon pits to capture runoff in degraded landscapes in the Baviaanskloof Hartland, South Africa
302 (Source: Living Lands, 2020).

303 **4.2 Effects of land use change on precipitation patterns**

304 This section addresses simulated and observed evidence of the impact of LUC on TMR. Moisture
305 recycling metrics (e.g. evaporation recycling ratio) cannot be used directly to estimate the impact of
306 LUC on precipitation, due to uncertainties in the effect of changes to the atmospheric moisture budget
307 (Goessling and Reick, 2011) and water's active role in the climate system. Although temporal
308 reductions in evaporation have shown significant precipitation effects (Keys et al. 2014), studies that
309 specifically address the impact of LUC on precipitation are scarce. This is not surprising, as the
310 processes of scale, data-availability, and lack of clear causalities in complex systems present

311 difficulties to find clear evidence (Spracklen et al., 2018). For the Amazon and Sahel, rainfall patterns
 312 have changed following vegetation cover reduction, but the processes are caused by different
 313 mechanisms and with different effects. Hence it is crucial to understand how LUC affect precipitation
 314 patterns and the scale at which they become significant. First, we describe the role of vegetation in
 315 moisture recycling more generally. Subsequently, we specifically address the effects of de- and
 316 reforestation.
 317



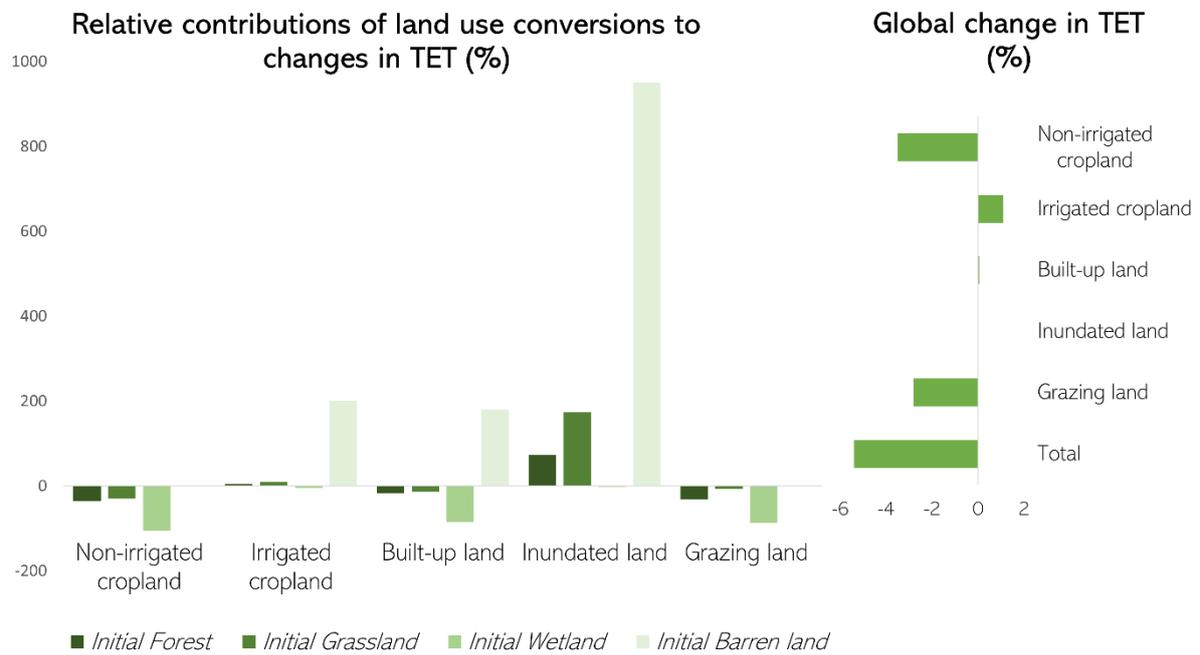
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319 Figure 5 **Source regions** (left) percentage of vegetation-regulated evaporation that falls as precipitation on land
 320 Source: figure copied from (Keys, Wang-Erlandsson, and Gordon 2016) and **sink regions** (right) percentage of precipitation
 321 that derives from vegetation-regulated evaporation. Source: figure copied with the author's permission from Keys, Wang-
 322 Erlandsson, and Gordon (2016)

323 4.2.1 The effect of vegetation on upward moisture fluxes

324 Vegetation regulates fluxes of transpiration and evaporation with various dynamics, i.e. magnitudes,
 325 sources and time scales (Wang-Erlandsson et al. 2014). A global analysis shows that 22% of
 326 terrestrial rainfall is *vegetation-regulated*, although spatial variation is large (Figure 5) (Keys et al.,
 327 2016). In Mato Grosso, Brazil, a vast region with different land uses and high rates of LUC, 30-45%

328 of the evaporation is vegetation-regulated. Furthermore, 6% of the precipitation in the region itself
 329 derives from vegetation-regulated moisture recycling. A hypothetical transformation of this region to
 330 a desert state, shows less interannual variability in rainfall and a strong reduction (-45%) in rainfall in
 331 the dry season. This implies that vegetation-regulated moisture recycling in this area is important to
 332 produce rainfall during the dry season (Keys, Wang-Erlandsson, and Gordon, 2016).



333
 334 Figure 6 Contributions of various LUC to changes in TET. The horizontal graph on the left shows the relative contributions
 335 of land use conversions (from initial to anthropogenic land cover) to changes in global total evaporation and transpiration
 336 (TET). Converting barren land to inundated land increases TET over that area with >900%. The vertical graph on the right
 337 shows the normalized contributions of the different land use conversions to the global change in TET (%). It shows that
 338 conversion to non-irrigated croplands have reduced global TET with nearly 4% (data derived from Sterling et al. (2013)).

339 Figure 6 shows the effects of different LUC on TET. On the left, it shows the relative changes in TET
 340 for specific conversions (e.g. converting barren land to inundated land increases TET with 900%). On
 341 the right, the contribution of land use conversions to the total change in global TET is shown. It shows
 342 that the global conversion to non-irrigated cropland has globally reduced TET with 3.5%. Hotspots of
 343 changes in TET following LUC are situated in Western Africa, South-East Asia and Eastern Europe.
 344 These are regions that have experienced large scale land use conversion of forest and grasslands to
 345 irrigated and non-irrigated croplands (Sterling et al., 2013).

346 LUC closer to the ocean might have a higher impact on precipitation patterns downwind due to the
 347 effect of moisture cascades moving inland (Schaeffli et al., 2012). Precipitation and forest cover show

348 a positive relationship along an atmospheric moisture transport trajectory in the tropics (Spracklen,
349 Arnold, and Taylor 2012). Air moving over dense vegetation produces more than twice the amount of
350 rain compared to air moving over sparse vegetation. The mechanisms behind the observation are
351 disputed: one explanation postulates increasing TET over the forest canopy intensifies the
352 hydrological cycle, assuming no change in atmospheric circulation (Spracklen, Arnold, and Taylor
353 2012), whereas another theory stipulates the ‘secondary’ effect of forest evapotranspiration, creating a
354 low pressure system, subsequently drawing in atmospheric moisture from the oceans (Makarieva et
355 al., 2014). Meteorological data gathered for the Amazon forest confirms changes in atmospheric
356 pressure regime due to different atmospheric moisture contents. This implies that forest loss in
357 tropical ecosystem can potentially change atmospheric circulation, yet the scales of forest loss at
358 which this becomes significant remains unclear.

359 4.2.2 Deforestation

360 The effect of deforestation on moisture recycling patterns is influenced by 1) direct changes in the
361 magnitude and timing of moisture fluxes, and 2) indirect changes in atmospheric circulation due to
362 exchanges in energy, moisture, and momentum. Vegetation cover loss can severely affect infiltration,
363 interception and moisture storage at the land surface (van Luijk et al., 2013), triggering a ‘soil erosion
364 feedback’ that gradually result in the loss of ecosystem resilience (Flores et al., 2019) and reduces
365 upward moisture fluxes which can produce self-propagating droughts and heatwaves via land
366 feedbacks (Miralles et al., 2019). In many regions where TET reductions following agricultural
367 expansion occurred, downwind reductions in precipitation were observed (Wang-Erlandsson et al.,
368 2018). In most cases, changes in rainfall occurred outside of the river basin, which implies that LUC
369 are less likely to produce local effects. In the Amazon basin however, local feedbacks are unusually
370 strong - significant deforestation-rainfall relations were found on a scale of 30-50 km - anticipating
371 stronger local effects of deforestation (Spracklen et al., 2018). In the dry-season and in drought years
372 – when oceanic inflow is low - the relative importance of moisture recycling increases, which implies
373 that reduced forest cover can result in a self-amplified forest loss during drought events (Bagley et al.,
374 2014; Zemp et al., 2014) A ‘*deforestation-induced tipping point*’ is proposed for the Amazon,

375 referring to the westward moisture cascade in which some regions are depending on precipitation
376 from evaporation elsewhere. Using observation-based moisture recycling networks, Zemp et al.
377 (2017) show that Amazon deforestation can reduce rainfall in the La Plata basin in the dry season with
378 up to 20%. Subsequent loss of forest resilience suggests that it can trigger further climatological
379 effects resulting in permanent forest reduction along the moisture recycling cascade (Zemp et al.
380 2017). Deforestation along the cascade affects the monsoonal circulation that is initially driven by
381 latent heat from the rainforest, which attracts moist air coming from the Atlantic Ocean. Deforestation
382 reduces transpiration up to a moment in which atmospheric moisture is insufficient to release latent
383 heat, which is a crucial mechanisms to draw in moist air (Boers et al., 2017). Furthermore, air moving
384 over deforested land loses more moisture relatively, due to lower evapotranspiration rates (intact
385 Amazonian forest on average adds 3-4 mm of transpiration to the air). The cascading effect (Zemp et
386 al. 2017) might therefore result in lower downwind precipitation. Increasing scales of Amazonian
387 deforestation trigger changes in thermal circulations and surface roughness. In deforested lands wider
388 than 10 km, changes in surface roughness and sensible heat can already trigger mesoscale circulation
389 changes resulting in redistribution of rainfall. On very large scales, 100-1000 km, deforestation can
390 change atmospheric properties that result in macroscale hydroclimatic changes (Spracklen et al.,
391 2018). Furthermore, long term rainfall trend data from Rondonia, Brazil, shows that regions with
392 large deforestation rates, have experienced a delay in the onset of the rainy season of 11 days (Butt et
393 al., 2011).

394 Some deforested areas show an *increase* in total precipitation. In the South-West of Brazil, satellite-
395 derived evidence suggests an increase in dry-season cumulus and convective clouds over deforested
396 area (Negri et al., 2004). This may be due to increased surface heating over deforested regions,
397 producing an upward air motion which draws atmospheric moisture from neighbouring areas.

398 Fragmented deforestation patterns may also lead to observations of increased rainfall: tropical forest
399 edges produce more transpiration compared to its interior due to micro-climatic effects, which may
400 result in increased rainfall. Deforestation increases energy transfers between the land surface and the
401 atmosphere which drive thermal circulations, leading to an observed increase in precipitation patterns

402 in parts of the Amazon (Chagnon & Bras, 2005). Based on long-term rain gauge observations, a
403 seasonal shift in precipitation was also recorded. LUC appeared to have a more severe effect on dry-
404 season precipitation in the Amazon (Bagley et al., 2014). In cold climates, the effect of snow cover is
405 important. Increased albedo following forest cover reduction in Russia elongates the snow season,
406 reduces air temperature and transpiration, and results in lower moisture recycling rates (Notaro & Liu,
407 2008). In arid climates, such as the Sahel, vegetation reductions resulting in an increased albedo and
408 reduced evapotranspiration, might have exacerbated drought duration in the 20th century extreme
409 droughts occurring in these regions (Charney & Stone, 1975; Savenije, 1995). Recycling of
410 evaporated moisture in the Sahelian belt appeared to contribute significantly to rainfall patterns in the
411 region during the wet season. By changing energy and moisture fluxes, this may have affected
412 convection and circulation of the African Easterly Jet (Yu et al., 2018).

413 4.2.3 Reforestation and afforestation

414 Although in theory, increasing vegetation cover can positively affect local rainfall patterns, there is
415 little known about the bioclimatic conditions and spatial and temporal scale of reforestation required
416 to increase moisture recycling. The Loess Plateau in China has experienced a long period of severe
417 degradation from intensive agriculture, followed by extensive reforestation since the year 2000's
418 under the Grain for Green Project which has doubled vegetation coverage on the Plateau from 31% in
419 1999 to 59% in 2013 (Bai et al., 2019). Sub-basin evapotranspiration trends show a significant
420 increase in (mainly summer) TET of 3.45 mm year⁻¹ (Bai et al., 2019). Vegetation productivity
421 contributed 93% to this increasing TET trend (Bai et al., 2019). Soil moisture response of former
422 farmlands to pine forest shows a 35% reduction in soil moisture content in deeper soil layers. No
423 significant differences in soil moisture reductions were found for different vegetation types. Observed
424 soil moisture deficits mostly related to plant density. Pine tree species are known to deplete soil
425 moisture due to their ever greenness (Yang et al. 2012). Reforestation simulations using regional
426 climate models for West Africa and the Sahel show that reforestation can enhance precipitation
427 with +3.6 - 14.4% (Oguntunde et al., 2014) but the location of the reforestation experiment has a
428 significant role on macro-scale climatic changes and the spatial distribution of predicted rainfall

429 patterns (Bamba et al., 2019). Reforestation enhanced surface roughness, weakening the atmospheric
430 temperature gradient, which results in a delay in the onset of the monsoon (Oguntunde et al., 2014). A
431 modelling scenario of potential restoration of Australia's woodlands on current economically
432 marginal lands shows an increase in evaporation, resulting in increased cloud formation and
433 precipitation over the region. The ability of woodlands to access deeper soil moisture would be the
434 mechanism behind increased evaporation (Syktus & McAlpine, 2016). Branch & Wulfmeyer, (2019)
435 assess the possibilities for rainfall enhancement using bio-geoengineering approaches (i.e. forest
436 plantations to deliberately enhance rainfall) in desert regions and conclude that agroforestry
437 plantations enhance local wind convergence, increase cloud cover and precipitation. Regional studies
438 that address the effects of reforestation and afforestation remain scarce. Although in some cases there
439 is evidence that it enhances local precipitation through increased TET, there are many climatic and
440 geographic variables that determine final effects on local and regional rainfall patterns (Keys et al.
441 2012).

442

443 Table 3 Summary of the evidence of LUC (deforestation and reforestation) on various processes governing moisture
 444 recycling patterns

	PROCESS	EVIDENCE FROM THE LITERATURE	CLIMATIC REGION	REFERENCES
DEFORESTATION	Soil moisture/ infiltration	○ Reduction of infiltration rates (mm h^{-1})	South Africa /succulent vegetation	(van Luijk et al., 2013)
	Green water flow	○ Correlates with rainfall intensity in post-monsoon period	Arid Sahel (Africa)	(Yu et al., 2017) (Zhao et al., 2019)
		○ Change in isotope ratio (evaporation/transpiration)	Arid Northern China (Asia)	(Sterling et al., 2013)
		○ Reduction in total ET (-5.5%)	Global assessment	(Butt et al., 2011)
		○ Delay in rain season (11 days)	Brazil	(Zemp et al. 2017)
		○ Reduced dry-season transpiration in comparison to other vegetation (1 mm day^{-1}) from roots accessing subsurface water	Amazon	
	Albedo	○ Increased local temperature (global average 0.23°C 2000-2015)	Global assessment, effect strongest in tropics	(Duveiller et al., 2018) Charney & Stone (1975)
	Streamflow	○ Cooling and sinking of air suppress convection and rainfall	Semi-arid Sahel	
		○ Global average 18-26% increase in water limited regions	Dry tropics	(Yang et al. 2012) (Peña-Arancibia et al., 2019)
	Precipitation	○ Increase in streamflow in regions with strong seasonality, high infiltration capacity, recharge of soil- and groundwater, high soil moisture storage, groundwater maintaining base flow	Dry tropics	
○ <i>Increase</i> in local rainfall		Southwest Brazil	(Negri et al., 2004)	
○ Reduction in rainfall (more than 50%) ○ <i>Simulated</i> 20% reduction in the dry season (deforestation scenario's)		Pan-tropical Amazon-La Plata (South America)	(Spracklen et al., 2012) Zemp et al (2017)	
REFORESTATION	Soil moisture/ infiltration	○ Roots channeling water to deeper soil layers for natural vegetation development following land abandonment	Southern Spain	(Cammeraat et al., 2010)
	Albedo	○ Reduction in deep soil moisture from pine plantation	Loess plateau, China	(Yang et al. 2012)
		○ Reduction in short wave reflection, increase in latent heat	Global assessment	(Duveiller et al., 2018) (Castelli et al., 2019)
	Streamflow	○ Reduction in land surface temperature (1.74°C)	Ethiopia	
		○ Decreases, although relatively less with temporal and spatial scales	Global assessment	(Filoso et al., 2017)
		○ Decreases the (seasonal) low flow	Global assessment	(Brown, 2005)
	Precipitation	○ Decreases of time, but in some cases shows partial recovery	Global assessment	(Bentley & Coomes, 2020)
	○ <i>Simulated</i> increase in precipitation (1.25%) and runoff in dry season (26%)	Bolivia	(Weng et al., 2019)	

445

446 4.3 Implications for governance of land use change

447 Under certain conditions, LUC can affect local or regional rainfall and redistribute – either
 448 intentionally or unintentionally – water resources. To ensure equitable and sustainable water use, there
 449 is a need to address the governance aspects of land use-water interactions and prevent adverse local or
 450 regional effects. Keys et al. (2017) address the notion of transboundary moisture recycling governance
 451 as ‘the attempts for steering social and environmental processes among countries and their

452 sometimes-conflicting objectives', evolving around the process of human interactions with
453 moisture recycling patterns. From the literature, three themes of governance approaches emerge:
454 spatial planning, impact assessments, and boundary setting.

455 4.3.1 Spatial planning approaches

456 A recent study that investigates the potential to increase rainfall over a municipality in Bolivia with
457 upwind 'smart reforestation' reveals that 7.1 million hectares of reforested land could increase
458 precipitation over the city by 1.25% ($5.8 \cdot 10^8 \text{ m}^3$) annually (Weng et al., 2019). *Aerial river*
459 *management* – the practice of redistributing flows of atmospheric water through strategic LUC
460 intentionally, has the potential to cover between 22-59% of the additional water demand in 2030
461 (Weng et al., 2019). Furthermore, considering *moisture recycling trajectories*, generally starting from
462 the coastal area and moving inland, reforestation efforts could consider to be 'build-up' incrementally
463 along this trajectory to increase moisture recycling and also enhance the success rate of reforestation
464 projects (Ellison & Ifejika Speranza, 2020) (Fagan et al., 2020). And finally, the identification of
465 *hotspots* of moisture recycling sources (Zemp et al. 2017) can support delineation of areas for forest
466 protection.

467 4.3.2. Impact assessments

468 NCP's associated with TMR are '*diffuse and spatially extensive*' and poses challenges to governance
469 (Keys et al., 2016). *Precipitation shed analysis* (Keys et al. 2017) can identify and quantify the
470 exchanges of atmospheric moisture between countries and provides a framework for impact
471 assessments that addresses the effects of LUC on TMR, as well as the impact of various NCPs.
472 Regional case studies are needed that address hydrological trade-off analyses that explicitly include
473 land-atmosphere feedbacks and TMR patterns (Wang & D'Odorico, 2019; Ellison & Ifejika Speranza,
474 2020). For example, although bio-geoengineering can enhance rainfall in some regions, it should be
475 balanced against the local effects on hydrology (Wang & D'Odorico, 2019) and cascading effects on
476 social-ecological systems. There is a need for a robust impact assessment framework that can address
477 the (transboundary) social and environmental trade-offs associated with interferences in TMR.

478 4.3.3. Boundary setting

479 Advances in earth observation technologies allow for detailed understanding of local to global water
480 use. Measurements of TET and Net Plant Productivity (NPP) via satellite imagery allows for
481 monitoring of green water use. For example, the FAO WaPOR project provides a monitoring platform
482 using remote sensing data that tracks annual gross biomass productivity which shows the biomass
483 production with respect to the actual evapotranspiration. The provided data facilitates water
484 accounting and enables green water management via targeted interventions, for example when local
485 vegetation growth is putting blue water resources at risk. In the Loess Plateau in China, a strong
486 increase in NPP and TET following the Grain to Green reforestation programme has come at the costs
487 of river runoff that is potentially societally unpropitious (Feng et al., 2016). Accordingly, a regional
488 *NPP plafond* is proposed to prevent water shortages amongst the population (Feng et al., 2016).
489 Alternatively, close monitoring of green water use and regional vulnerability also allows for measures
490 restraining the use of high-water demanding species.

491 Governance of moisture recycling and land use-water interactions is in its infancy. Spatial planning
492 approaches, impact assessments and boundary setting are governance approaches that are proposed in
493 response to spatially extensive and diffuse nature of land-use water interactions via TMR.
494 Furthermore, market-based and regulatory instruments, such as Payment for Ecosystem Services
495 (PES) and transboundary agreements and collaboration, could facilitate the implementation of such
496 approaches. Yet, little is known considering their practical implementation in the context of TMR.
497 Hence, besides the need for tools to address trade-offs in TMR governance, research on the
498 advantages, disadvantages and relation to inter- and transnational legal contexts (i.e. international
499 water law and transboundary agreements) of market-based and regulatory approaches to moisture
500 recycling is needed. Principles reflected in international water law refer to the obligation not to cause
501 significant harm (Rahaman, 2009) which implies that countries may be held accountable when land
502 use change appear to negatively affect rainfall patterns via international agreements. A ‘one size fits
503 all’ approach to governance is likely to be undesirable due to 1) the spatial and temporal variance

504 land-atmosphere interactions and associated water circulation and 2) the issue of scalability associated
505 with non-linear responses of TMR to LUC.

506 5. Conclusion

507 Continuous global land use change, increasing understanding of biosphere-atmosphere interactions,
508 and increasing water scarcity beg the question how LUC affects dynamics and feedback mechanisms
509 with respect to water and rainfall in the soil-vegetation-climate system. This scoping review addressed
510 the state-of-the-art knowledge on moisture recycling in relation to LUC and leads to five main
511 conclusions:

- 512 • First, 22% of the global rainfall is vegetation-regulated, which implies that LUC can greatly
513 affect rainfall patterns. In the last decades, LUC have reduced global TET by 5%.
- 514 • Second, deforestation in general has reduced local precipitation, distorted moisture recycling
515 cascades (reduce downwind precipitation), intensified drought, delayed the onset of the rain
516 season, and in some cases increased local rainfall due to microclimatic effects. In general,
517 effects on precipitation are more likely to be non-local and occur outside the basin.
- 518 • Thirdly, dominant feedback mechanisms and effects differ strongly between regions. In
519 tropical wet regions, stronger local effects of LUC on moisture recycling are expected which
520 implies vegetation is more sensitive to drought and disturbance feedback mechanisms. In
521 water-limited regions like the Sahel, the effects of the energy-feedback are more prominent.
- 522 • Fourth, hotspots of moisture recycling may occur along gradients between ocean and land
523 surface, mountainous regions, and transitional zones. These hotspots might require protection
524 to prevent disruption of moisture recycling patterns.
- 525 • Finally, the effects of reforestation on moisture recycling patterns appear sensitive to the
526 scale and the spatial location of the reforestation project. Overall, the effects remain largely
527 unexplored. Although this is sensible due to the complex nature of the question, there is a
528 need to further explore the potential hydrological trade-offs of reforestation.

529 Coupled land surface and climate models have the potential to explore specific LUC scenarios to
530 identify the change in rainfall patterns following different spatial locations and scales of LUC.
531 Analytical tools that allow for atmospheric water network analysis such as moisture recycling
532 cascades (Schaepli et al., 2012), watershed analysis (Keune & Miralles, 2019) and precipitation- and
533 evaporationsheds (Keys et al., 2014) can support water accounting measures and environmental and
534 social impact assessments (Bagley et al., 2012) to govern TMR. The notion of green and atmospheric
535 water governance (Wierik et al., 2020) implies that – amongst others - trade-offs associated with
536 vegetation’s ability to redistribute water flows are addressed. This implies, for example, that climate
537 mitigation policies for carbon sequestration explicitly consider the hydro-climatic effects at the
538 precipitationshed level to prevent unexpected hydrological consequences for people and nature.

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546 findings from previously published papers. The data represented in Figure 6 is available through
547 Sterling et al., (2013).

548

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865

Figure 1.

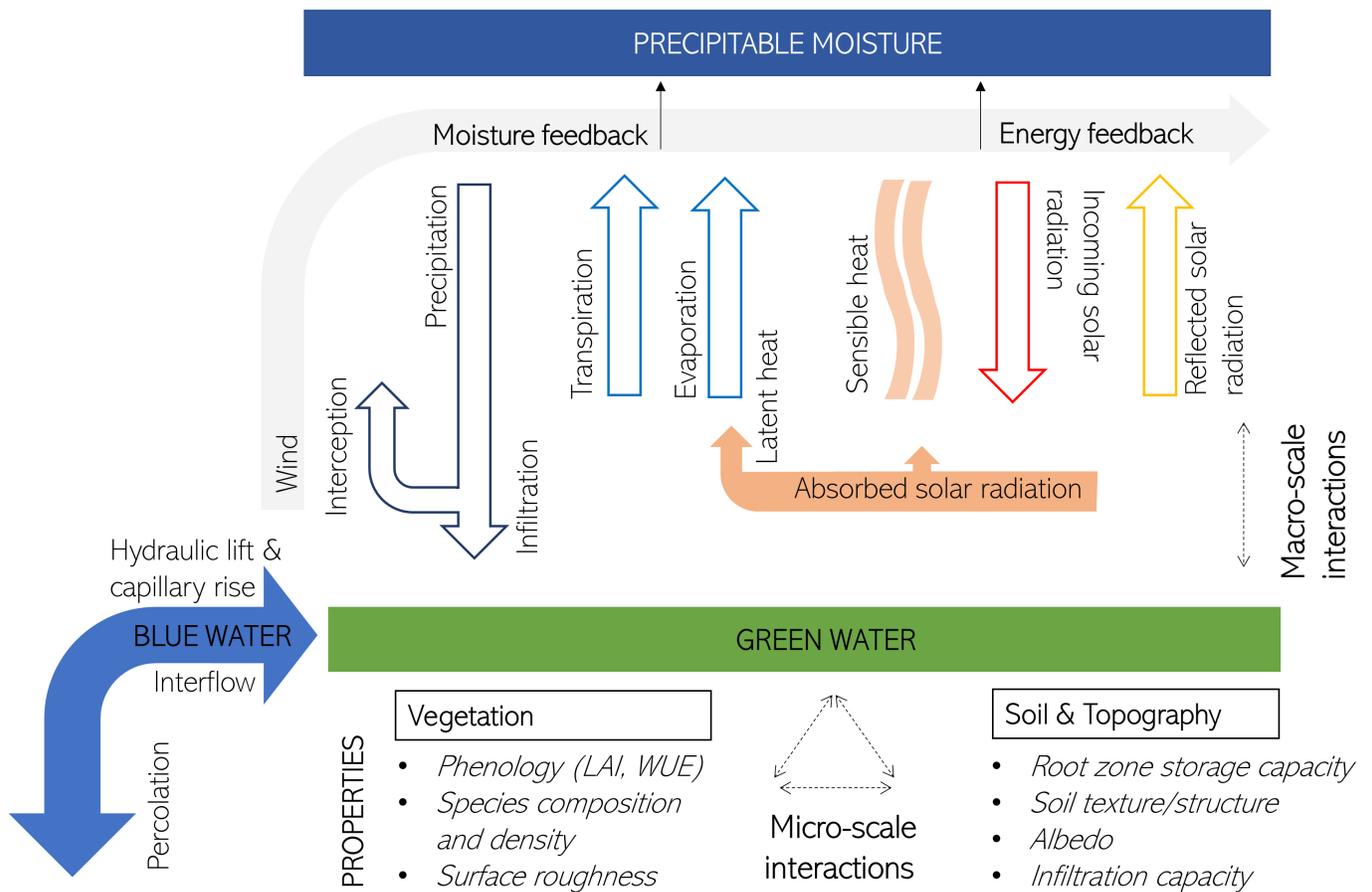


Figure 2.

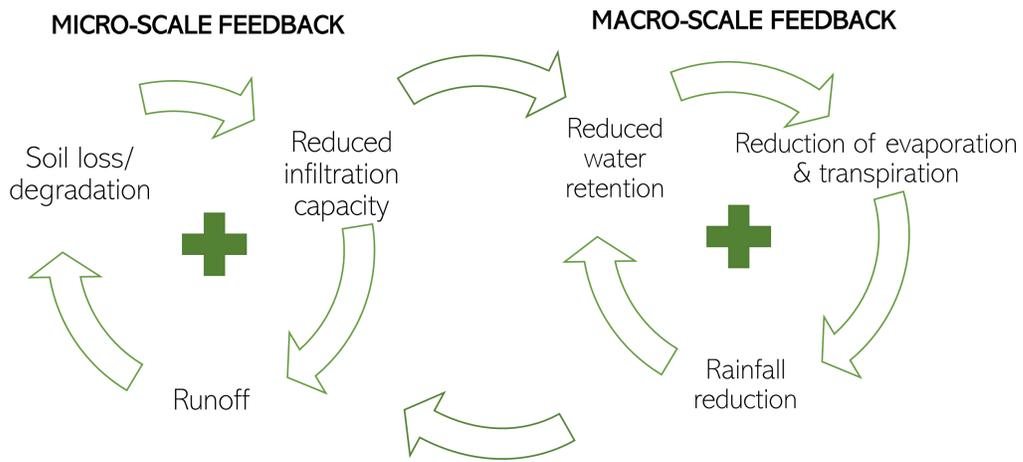
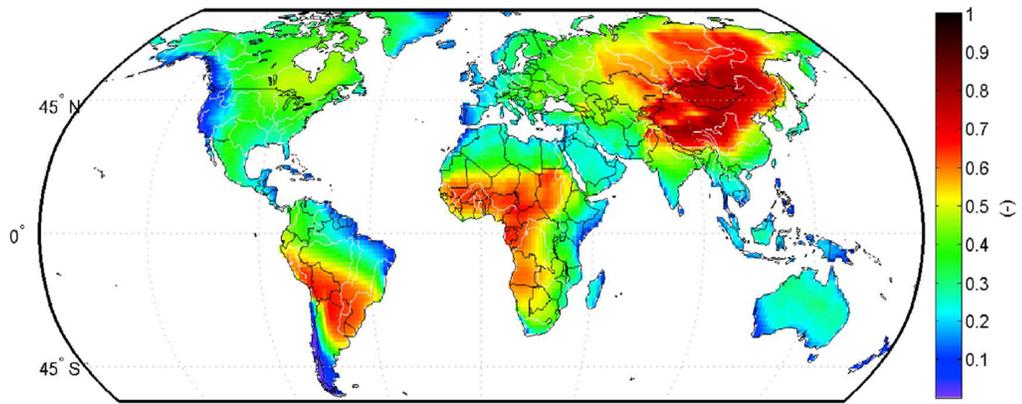


Figure 3.

Continental precipitation recycling ratio ρ_c



Continental evaporation recycling ratio ε_c

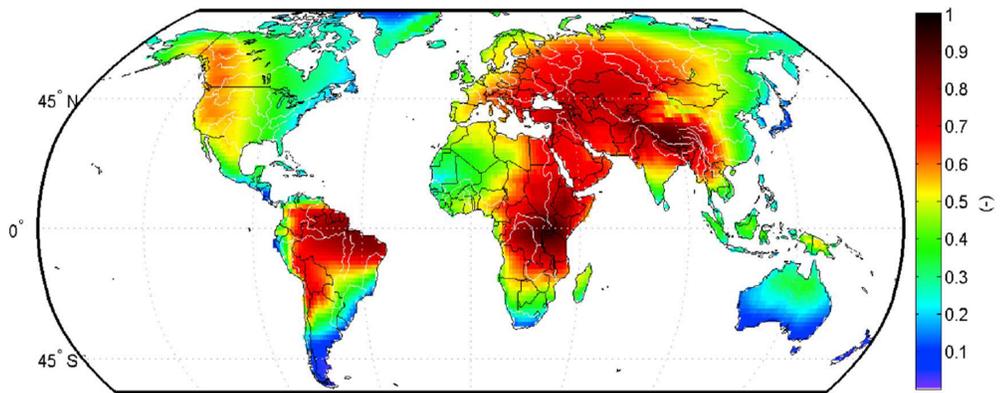


Figure 4.



Figure 5.

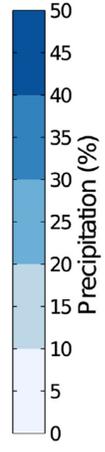
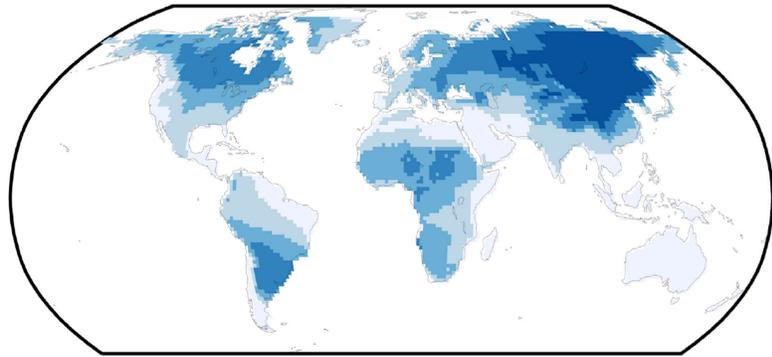
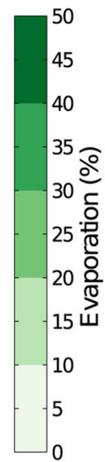
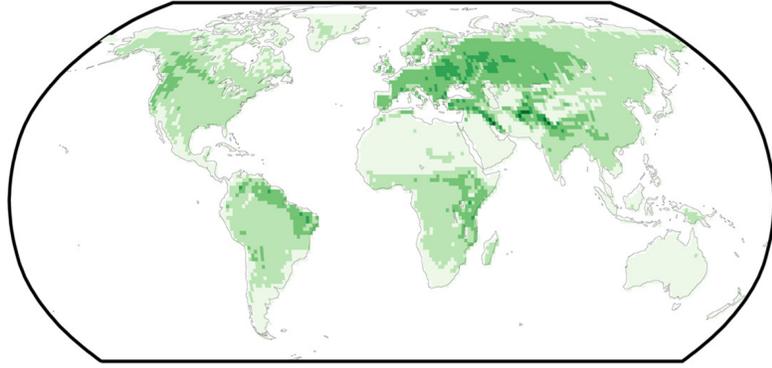


Figure 6.

