Quantifying importance of macrobenthos for benthic-pelagic coupling in a temperate coastal shelf sea

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Abstract

Benthic oxygen fluxes consist of advective and diffusive terms. Both terms in the south-eastern North Sea exhibit a prominent annual cycle but with opposite variation patterns. To understand the driving mechanisms quantitatively, a novel 3-D benthic-pelagic coupled model resolving interactions among macrobenthos, bioturbation, oxygen consumption and carbon early diagenesis was applied to reconstruct the benthic states. Simulation results show a satisfactory agreement with field data and reveal that the benthic oxygen flux is determined by not only pelagic drivers but also by internal dynamics associated with the interaction between organic carbon and macrobenthos, and bedform morphodynamics. Variation of advective flux, characterized by summer-low and winter-high, is mainly driven by hydrodynamics and bedform morphodynamics, while variation of diffusive flux, featured by summer-high and winter-low, is a compound effect of pelagic and benthic drivers with a dominant control by macrobenthos through bioturbation. The role of bioturbation in benthic oxygen consumption is twofold: (i) on the one hand, it alters the particulate organic carbon (POC) distribution in surface sediments, thereby changing the availability of POC to oxygen consumption; (ii) on the other hand, it mixes oxygen down into sediments, thereby facilitating oxygen consumption. Our results indicate that the first role prevails in sandy seafloor characterized by energetic hydrodynamics, while the second role becomes increasingly important along with a weakening of bottom currents. We found that bioturbation-induced oxygen consumption contributes to more than 85% and 52% of the total benthic oxygen fluxes in muddy seabed and at a regional scale, respectively.

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 Key Points:
- A novel 3D numerical model resolves interactions among macrobenthos, bioturbation, oxygen
 consumption and carbon early diagenesis
- The role of bioturbation in benthic oxygen consumption is twofold and dependent on sediment properties and hydrodynamics
- Bioturbation-induced oxygen consumption contributes to more than 52% of the total benthic oxygen
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19 Abstract

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Benthic oxygen fluxes consist of advective and diffusive terms. Both terms in the south-eastern North 20 21 Sea exhibit a prominent annual cycle but with opposite variation patterns. To understand the driving mechanisms quantitatively, a novel 3-D benthic-pelagic coupled model resolving interactions among 22 macrobenthos, bioturbation, oxygen consumption and carbon early diagenesis was applied to reconstruct 23 the benthic states. Simulation results show a satisfactory agreement with field data and reveal that the 24 benthic oxygen flux is determined by not only pelagic drivers but also by internal dynamics associated 25 with the interaction between organic carbon and macrobenthos, and bedform morphodynamics. 26 Variation of advective flux, characterized by summer-low and winter-high, is mainly driven by 27

hydrodynamics and bedform morphodynamics, while variation of diffusive flux, featured by summer-28 high and winter-low, is a compound effect of pelagic and benthic drivers with a dominant control by 29 30 macrobenthos through bioturbation. The role of bioturbation in benthic oxygen consumption is twofold: 31 (i) on the one hand, it alters the particulate organic carbon (POC) distribution in surface sediments, thereby changing the availability of POC to oxygen consumption; (ii) on the other hand, it mixes oxygen 32 down into sediments, thereby facilitating oxygen consumption. Our results indicate that the first role 33 34 prevails in sandy seafloor characterized by energetic hydrodynamics, while the second role becomes 35 increasingly important along with a weakening of bottom currents. We found that bioturbation-induced 36 oxygen consumption contributes to more than 85% and 52% of the total benthic oxygen fluxes in muddy 37 seabed and at a regional scale, respectively.

38 Plain Language Summary

39 Oxygen is fundamental to sustain life and to regulate carbon cycling in our Earth system. However, because of global warming, there is growing concern that oxygen in the ocean is declining, partly due to 40 41 temperature increase and partly due to eutrophication that leads to excessive production of organic 42 carbon which consumes oxygen during their degradation. In the worst case, the latter results in oxygen 43 exhaustion, generating "dead zones" near/at the ocean seafloor. To mitigate this potential threat, it is important to understand the mechanisms that control oxygen consumption near and at the seafloor. To 44 45 this end, our study provided a quantitative assessment of the driving factors for oxygen consumption across the sediment-water interface in a coastal sea. We found that animals living at the seafloor are of 46 47 vital importance in controlling oxygen consumption in sediments. Therefore, it is important for us to protect benthic life and biodiversity, which in turn help to mitigate climate change through carbon 48 preservation in seafloor. 49

50 Keywords

51 Benthic oxygen flux, Bioturbation, Early diagenesis, Biophysics, Benthic-pelagic coupling, Morphodynamics

53 **1. Introduction**

The ocean seafloor is the largest carbon reservoir within the surface Earth system, providing a 54 valuable geochemical archive for deciphering variability in global biogeochemical cycling and climate 55 56 (Ciais et al., 2013; Heinze et al., 2015). Quantification of biogeochemical cycling, especially carbon cycling at the ocean seafloor, is thus a key for understanding the fate of our future Earth (Falkowski et 57 al., 2000; Le Qu ér éet al., 2018). Of vital importance in this context is the coastal shelf seafloor, which 58 covers less than 10 % of the global ocean area (Muller-Karger et al., 2005) but has been estimated to 59 60 receive ~50% of the pelagic primary production through sedimentation (Wollast, 1991) and releases 61 regenerated nutrients such as nitrogen, phosphorus, and silica that supply ~80% of phytoplankton nutrient requirements (Middelburg & Soetaert, 2004). On the other hand, coastal shelf seas constitute the 62 63 most dynamic part of the surface Earth where intense interactions between geosphere, ecosphere and 64 anthroposphere take place. Ignoring or over simplifying such interactions leads to large uncertainties in 65 calculating biogeochemical fluxes (e.g. Gattuso et al., 1998; Muller-Karger et al., 2005; Chen & Borges, 2009; Cloern et al., 2016; Griffiths et al., 2017), which may partly account for the imbalance of up-to-66 67 date estimates of global carbon budget (Le Qu ér éet al., 2018). Recent studies demonstrate a need to take into account spatial heterogeneity and biodiversity of the seafloor to better assess carbon processing and 68 ecosystem functioning in the context of global biogeochemical cycles (Cardinale and Palmer, 2002; 69 Lohrer et al., 2004; Belley and Snelgrove, 2016; Snelgrove et al., 2018). However, limited knowledge 70 of the role of benthic organisms in modulating biogeochemical fluxes at the sediment-water interface 71 72 impedes our ability to derive quantitative syntheses (Griffiths et al., 2017; Middelburg, 2018). 73 Benthic oxygen availability is fundamental to sustain aquatic life and to regulate many important biogeochemical processes including early diagenesis of organic carbon (Glud, 2008). A process-based 74 75 understanding of how oxygen is transported into sediments and a quantification of its flux across the

sediment-water interface is thus of utmost importance for evaluating benthic ecosystem functioning and
 biogeochemical cycling. Oxygen transport across the sediment-water interface evolves through the

advection and diffusion (Glud, 2008). Both transport mechanisms are controlled by not only 78 hydrodynamic (porewater advection) and morphodynamic (bedform migration) processes but also by 79 80 biophysical (bio-diffusion and bio-ventilation) and biogeochemical (degradation of OC) processes (Meile and Van Cappellen, 2005; Santos et al., 2012; Ahmerkamp et al., 2015). Benthic fauna 81 contributes to both advective (through bio-ventilation) and diffusive (through bio-diffusion) transport, 82 and significantly alters the efficiency of OC degradation by diffusing both OC and oxygen into 83 84 subsurface sediments (Meile and Van Cappellen, 2005; Middelburg, 2018; Zhang et al., 2019). Therefore, benthic fauna may impose a major control on the benthic oxygen flux and biogeochemical 85 86 cycling even on a regional to global scale (Griffiths et al., 2017; Snelgrove et al., 2018). However, a robust quantitative estimation of the impact of benthic fauna on the oxygen flux across the sediment-87 water interface is still lacking (Middelburg, 2018). 88

89 A mechanistic understanding of the interaction between benthic fauna and the associated environmental conditions must be derived before a quantitative synthesis can be made (Zhang and Wirtz, 90 2017). The abundance and community structure of benthic fauna greatly depend on the supply of 91 92 particulate organic carbon (POC) to the seafloor (Pearson and Rosenberg, 1987; Dauwe et al., 1998; Herman et al., 1999; Campany à Llovet et al., 2017). Bioturbation, referred to here as any kind of 93 biological reworking of sediments including particle mixing and burrow ventilation (Meysman et al., 94 2006), has been found to vary by more than one order of magnitude even at the same site due to seasonal 95 changes in food supply and in the metabolism of benthic organisms (Brown et al., 2004; Teal et al., 96 97 2008). The high variability of bioturbation clearly demands for its integration as a dynamic process in up-to-date ecosystem and diagenetic modelling (Queir $\dot{\alpha}$ et al., 2015). Mechanistic approaches for 98 quantifying bioturbation have been developed for more than three decades since the pioneering work by 99 100 Berner (1980). Bioturbation is conventionally modelled as a diffusive mixing process (Berner, 1980; 101 Boudreau, 1986a), and later on with inclusion of non-local transport (e.g., Boudreau, 1986b; Whearcroft 102 et al., 1990; Soetaert et al., 1996a; Meysman et al., 2003; Holstein and Wirtz, 2009; Schiffers et al.,

103 2011; Aquino et al., 2017). However, although bearing the capacity of fitting individual tracer profiles, most existing bioturbation models are not designated for solving a dynamic benthic response to 104 105 changing pelagic conditions, especially at a regional scale. Only a few ones explicitly relate 106 macrobenthos and their bioturbation with settling POC (e.g. Butenschön et al., 2016; Zhang and Wirtz, 107 2017). Given the fact that POC supply is a major constraint for macrobenthos on one side, and macrobenthos is consumer as well as a major driver of POC transport in sediments on the other side, 108 109 process-based modelling of the interaction between POC and macrobenthos and the consequent impact on benthic biogeochemical cycling deserves to be a focal point of research (Kelly-Gerreyn et al., 2014; 110

111 Campany à Llovet et al., 2017; Zhang and Wirtz, 2017).

In this study, we made a numerical modelling attempt to interpret the pattern of measured benthic oxygen fluxes in the south-eastern North Sea. Through a model reconstruction of the benthic dynamics in response to changing pelagic environmental conditions, we aim to provide a quantitative estimate of the importance of macrobenthos in controlling oxygen flux at the sediment-water interface at a regional scale. This goal necessitates the use of up-to-date 3-dimensional hydrodynamic-biogeochemical modelling, enabling to resolve the dynamic interactions between pelagic drivers and benthic biota, and subsequent impact on the early diagenesis of organic carbon, oxygen consumption and transport.

119 2. Material and methods

120 **2.1. Study area**

Our study area is the German Bight located in the south-eastern North Sea. This area is characterized by water depth up to 60 m, with its major part being shallower than 40 m (Figure 1a). Currents are predominantly driven by a combined action of wind and astronomical tides, and additionally modulated by baroclinity (S ündermann and Pohlmann, 2011) and river runoff (Becker et al., 1999). Semidiurnal tides M2 and S2 are the major tidal constituents. Energetic tidal currents with speed of up to ~2 m s⁻¹ give rise to strong mixing and inhibit thermohaline stratification in a major part of the area (Schrum, 1997). A cyclonic residual circulation pattern exists in the German Bight due to an

interplay between different forces (wind-driven currents and tides) and topography (Nihoul, 1980). This 128 circulation introduces a mixture of Atlantic water and river runoff (mainly Rhine, Meuse, IJsselmeer, 129 Ems, Weser and Elbe) along the southern North Sea coast into the German Bight (van Beusekom et al., 130 131 1999) and favours deposition of fine-grained sediments (Figure 1b). Outcomes from the North Sea Observation and Assessment of Habitats (NOAH, 2013-2019) project (http://www.noah-132 project.de/habitatatlas/) indicate that the major part of the seafloor is dominated by sands (Figure 1c) 133 with very limited organic content (Figure 1d). High sedimentation rates (up to 6 mm yr⁻¹ on average, 134 Dauwe et al., 1998) only occur on a few spatially confined mud deposits where silt and clay contents 135 136 exceed 10% (Figure 1b). The distribution of fine-grained material on the seafloor is a joint effect of the residual circulation pattern and surface gravity waves (Stanev et al., 2009), both of which are strongly 137 influenced by the seafloor topography especially that of the ancient Elbe river valley (Becker et al., 138 1999). Since fine-grained lithogenic sediments (silt and clay) tend to flocculate with organic matter 139 140 during their transport (Maerz et al., 2016), the discrepancy between high mud content (> 10%) in several depocenters and high sedimentary TOC content (>1%) confined in only one depocenter near the island 141 142 Helgoland (Figure 1b & 1d) implies a strong benthic-pelagic coupling which significantly affects carbon 143 storage and turnover in this productive area (Neumann et al., 2019).



Figure 1. (a) Bathymetric map of the German Bight with marked sampling stations. (b) Mud content (%) in surface sediments plotted in logarithmic scale. (b) Classification of surface sediments according to the Folk's sand-silt-clay diagram. (d) TOC content (%) in surface sediments plotted in logarithmic scale. Note that the Wadden Sea area including the tidal channels between the barrier islands is not covered by the data. Sampling stations with permeable sediments are marked by ×, while stations with impermeable sediments are marked by +.

151

152 2.2. Field data

153 Measurement of benthic oxygen flux was done at 9 stations (Figure 1a) during cruises of the RV

154 Heincke in 2012 – 2016 within the North Sea Observation and Assessment of Habitats (NOAH, 2013-

- 155 2019) project (http://www.noah-project.de/habitatatlas/). Advective and diffusive fluxes were derived
- 156 from in situ measurements using a benthic lander equipped with oxygen profilers (optodes) placed inside
- and above surface sediments and ex situ core incubation measurements, respectively.

158 2.2.1. In situ lander measurements

In situ measurements of advective oxygen flux induced by porewater flow were carried out in three months (March 2014, September 2014 and July 2015) at the stations with permeable sediments (Table 1). Technical details and results of the in situ measurements were presented by Ahmerkamp et al. (2017).

162 **2.2.2. Ex situ incubation measurements**

Ex situ core incubation measurements were done at all stations with the aim to derive the diffusive 163 fluxes. The surface sediment was sampled with a Multicorer (Oktopus Kiel) equipped with acrylic 164 (PMMA) core liners (10 cm inner diameter, 60 cm length). Typically, 3 to 4 intact sediment cores per 165 station were selected for ex situ, whole-core, batch-type incubation. Cores with visible perturbations 166 such as cracks, voids, or injured animals were rejected. The sediment cores were typically 15 - 30 cm in 167 168 length. The incubations were closed with a gas-tight lids that were adjusted to a resulting supernatant height of 15 cm. The water column was constantly stirred by horizontal propellers (3 blades, 20 mm 169 diameter) that were magnetically driven by external motors. The stirring intensity was adjusted so that it 170 171 is low enough to prevent particle resuspension but on the other hand strong enough to ensure mixing of 172 the supernatant and keep suspended particles from settling. It is noteworthy that porewater flow is 173 minimized by this small stirring intensity so that the measured values represent the diffusive fluxes. 174 Incubations were executed aboard the research vessel in a temperature-controlled lab at in situ temperature. Primary production was excluded by wrapping the cores in aluminium foil. During 175 incubation, oxygen was monitored continuously with fiber-optical optodes (Presens). A two-point 176 177 calibration was applied, employing air-saturated water (100 %) and oxygen-free water (0 %) that was prepared by addition of sodium sulphite (Na₂SO₃). The incubations were terminated when the oxygen 178 179 saturation dropped below approximately 80 % saturation, which was typically after 12 to 24 h. Oxygen fluxes were calculated as concentration change over time by linear regression. 180

181 **2.2.3. Complementary field data**

In addition to measurement of benthic biogeochemical fluxes, sediment and macrobenthos samples
were also collected from these stations by means of a Multicorer. Vertical distributions of TOC,

- 184 TOC/TN ratio and macrobenthos biomass in the upper-most 20 cm sediments were measured. Detailed
- processing procedures of these samples are described in Zhang et al. (2019). A list of substrate

186 properties of the stations is provided in Table 1.

'	Tuote III	tere i suesti die properties of the stations and sources							
	Station	Median grain size	Porosity	Permeability	Mud	TOC	Bedform height	Bedform length	Source
		$d_{50}(\mu m)$	ϕ (-)	$k (\mathrm{m}^2)$	(%)	(%)	<i>l</i> (cm)	λ (cm)	
	NOAH-A	A 390	0.40	1.1×10^{-10}	1.1	0.10	2.5	18	1&2
	NOAH-E	3 210	0.42	3.2×10^{-12}	1.7	0.12	1.5	11	1&2
	NOAH-E	E 249	0.41	4.5×10^{-11}	0.8	0.10	2.4	15	1&2
	CCP-G	398	0.36	1.2×10^{-10}	3.0	0.30	2.7	22	1&2
	NOAH-I	223	0.38	3.6×10^{-11}	0.4	0.08	1.5	11	1&2
	NOAH-C	C* 35	0.53	0	52.0	1.10	-	-	1
	NOAH-I	D* 102	0.45	0	10.0	0.31	-	-	1&2
	NOAH-F	F* 128	0.50	0	10.0	0.33	-	-	1&2
	NOAH-H	H [*] 155	0.46	0	6.0	0.27	-	-	1&2

187 Table 1. Substrate properties of the stations and sources

188 Sources: (1) http://www.noah-project.de/habitatatlas/; (2) Ahmerkamp et al. (2017). Stations marked by * indicate 189 impermeable sediments with $k < 10^{-12}$ m².

190

191 2.3. Numerical modelling of benthic oxygen flux

192 Oxygen input into sediment across the sediment-water interface consists of two terms, namely

advection and diffusion, respectively. The flux Q_{02} (µmol m⁻²h⁻¹) is given by:

194
$$Q_{02} = uC_0 + D \frac{(C_0 - C_1)}{\Delta z},$$
 (1)

where $u (m h^{-1})$ is the interstitial seepage velocity normal to the seafloor surface, z is the depth (cm)

with z = 0 at the sediment-water interface and positive downward, C_0 and C_1 (µmol m⁻³) are the oxygen

- 197 concentration at the sediment-water interface (z = 0 cm) and the centre of the first sediment cell
- 198 (z = 0.5 cm in this study), respectively and D is the effective diffusivity (m² h⁻¹) calculated as the sum
- of molecular diffusivity (D_{mol}) and bioturbation diffusivity (D_{b}) . The first term on the right hand side
- 200 represents the advection term. The advection velocity equals zero (u = 0) in case of impermeable
- sediments. For permeable sediments, u equals to the sediment porosity (ϕ) times the porewater flow

202 velocity (Elliott and Brooks, 1997). A characteristic porewater flow velocity (u_p) is given by 203 Ahmerkamp et al. (2015):

204
$$u_p = 2k \frac{U^2}{\phi v} \frac{\Delta p}{\lambda},$$
 (2)

where *k* is the permeability, *U* is the bottom current velocity (m s⁻¹), ν (=10⁻⁶ m² s⁻¹) is the kinematic viscosity of water, Δp (=0.1) is a non-dimensional characteristic pressure difference and λ is the wave length of the bedform (ripples). A power-law decay of oxygen flux along an increasing dominance of bedform migration over porewater flow velocity is found by Ahmerkamp et al. (2017). To take into account this impact for permeable sediments, *u* is modified as:

210
$$u = \begin{cases} \phi u_{p}, & \text{if } c_{ph} / u_{p} < 1\\ \phi u_{p} \left(\frac{c_{ph}}{u_{p}}\right)^{-1.66}, & \text{if } 6 \ge c_{ph} / u_{p} \ge 1,\\ 0.05 \phi u_{p}, & \text{if } c_{ph} / u_{p} > 6 \end{cases}$$
(3)

where c_{ph} is the bedform migration rate (m h⁻¹) that is dependent on the bed shear stress, bedform roughness and sediment grain size according to Ahmerkamp et al. (2015).

Molecular diffusivity is determined by sediment porosity (ϕ), tortuosity (τ) and ambient temperature (*T* in °C), and calculated by (Soetaert et al., 1996b):

215
$$D_{mol} = (D_0 + D^*T)\frac{1}{\tau},$$
 (4)

where D_0 is the molecular diffusion coefficient of oxygen at 0 °C and D^* is the diffusion coefficient for linear temperature dependence of oxygen. Tortuosity (τ) is calculated from the sediment porosity as 218 $\tau = F\phi$ with $F = \phi^{-m}$, where m=2 is for marine sands or muds with $\phi \le 0.7$ and m=3 for high-porosity 219 muds with $\phi > 0.7$ (Ullman and Aller, 1982).

220 Bioturbation diffusivity (D_b) is calculated by a mechanistic model, namely the TOC-MAcrobenthos Interaction Model (TOCMAIM), which has been calibrated and validated by field data from the North 221 222 Sea (Zhang & Wirtz, 2017; Zhang et al., 2019). Given a boundary input of OC at the sediment-water 223 interface, TOCMAIM solves the mass balance equation of OC in surface sediments by including the impacts of sedimentation (through advection), first-order degradation, macrobenthic uptake and 224 bioturbation (through diffusion). Bioturbation diffusivity D_{h} for solutes in this study is assumed to be 10 225 times the bioturbation diffusivity for solids. D_b scales with the local macrobenthos biomass B through a 226 227 power law and inversely proportional to local food resource (OC):

228
$$D_b = 10\beta B^b (\sum_{i=1}^3 a_i C_{org,i})^{-1},$$
 (5)

where β , b are empirical parameters linking body-size and abundance to biomass. Biomass of 229 macrobenthos (B) is estimated in TOCMAIM based on local food resource (i.e. OC content) and costs 230 for respiration and mortality (Zhang & Wirtz, 2017). Organic carbon is divided into three pools ($C_{org.i}$, 231 i=1,2,3) depending on the degradability, namely labile (i.e., of high quality nutrient), semi-labile (i.e., of 232 intermediate quality nutrient) and refractory (i.e., of low quality nutrient). a_i (i = 1, 2, 3) are coefficients 233 representing the yield of the POC pool *i* in terms of gaining macrobenthos biomass, with higher quality 234 OC gaining biomass more efficiently, as expressed by $a_1 > a_2 > a_3$. The above formulation explicitly 235 links the bioturbation diffusivity with (1) the body-size and abundance of macrobenthos (through scaling 236 237 with biomass), and (2) the local food resource which determines the intensity of vertical movements of macrobenthos to derive enough energy and nutrition for metabolism and growth. A detailed justification 238 of Equation (5) is provided in Zhang & Wirtz (2017). 239

It can be seen from Equation (1) that the oxygen flux depends on not only the diffusivity but also 240 on the vertical gradient of oxygen concentration across the sediment-water interface, i.e. $(C_0 - C_1)/\Delta z$. 241 In this study, pelagic environmental variables such as current velocity, bottom shear stress, temperature, 242 POC concentration and oxygen concentration at the sediment-water interface z = 0 cm (i.e. C_0) are 243 provided by a 3-D hydrodynamic-biogeochemical model ECOSMO II (Schrum et al., 2006a, b; Daewel 244 245 & Schrum, 2013), while POC burial, transport and degradation in sediments including the interaction with macrobenthos are simulated by the coupled model TOCMAIM (Zhang et al., 2019). A new feature 246 of TOCMAIM introduced in this study is an inclusion of oxygen consumption in sediments by solving 247 the mass balance equation assuming that oxygen is exclusively consumed by POC degradation: 248

249
$$\frac{\partial C}{\partial t} = -\frac{\partial wC}{\partial z} + \frac{\partial}{\partial z} \left(D \frac{\partial C}{\partial z} \right) + S - \frac{(1-\phi)}{\phi} \varepsilon \sum_{i=1}^{3} R_i C_{org,i} , \qquad (6)$$

where *C* denotes oxygen concentration (μ mol m⁻³), *w* is the advection (sedimentation) rate (m h⁻¹), *S* is the source term being equal to $uC_0/\Delta z$ for the first (upper-most) sediment cell and zero for the rest cells, R_i is the depth-dependent first-order degradation rate constant of POC pool *i* ($C_{org,i}$), and is calculated by $R_i(z) = e^{-0.3z}R_i(0)$, where $R_1(0) = 2.3 \times 10^{-3}$ h⁻¹, $R_2(0) = 2.3 \times 10^{-4}$ h⁻¹ and $R_3(0) =$ 2.3×10⁻⁶ h⁻¹, and ε is a constant representing the ratio of oxygen demand to carbon degradation. A list of model parameters is given in Table 2, with further details of TOCMAIM and its coupling to ECOSMO II provided in Zhang et al. (2019).

257 Table 2. Model parameters and sources

Table 2. Model parameters and sources				
Parameter	Unit	Value	Source	
D_0	$cm^2 h^{-1}$	0.04	1	
	0 1 1			
D^*	$\mathrm{cm}^2 \mathrm{h}^{-1} \mathrm{°C}^{-1}$	1.6×10^{-3}	1	
β	-	0.22	2	
1		1.22	2	
D	-	1.33	2	

<i>a</i> ₁	h^{-1}	0.03	2
<i>a</i> ₂	\mathbf{h}^{-1}	0.01	2
<i>a</i> ₃	h^{-1}	1.7×10^{-3}	2
Е	-	1.45	3

258 259 Sources: (1) Hülse et al. (2018); (2) Zhang and Wirtz (2017); (3) Hedges et al. (2002).

It can been seen from Table 1 that sediment permeability differs markedly between stations. In our study, a tested empirical relationship associating permeability with sediment grain size proposed by Neumann et al. (2017) was adopted to provide sediment permeability map for the entire study area with 1.852 ×1.852 km grid resolution. The derived map (https://www.noah-

project.de/habitatatlas/index.php.en) shows that 25% of the seabed in the study area is covered by impermeable sediments ($k < 10^{-12} \text{ m}^2$) and the rest by permeable sands. For predicting the bedform parameters (l and λ), we applied the widely-used formulation from Soulsby et al. (2012) validated for the southern North Sea by Kr ämer and Winter (2016). By feeding the permeability map and bedform predictor into the 3D hydrodynamic-biogeochemical model, the benthic oxygen flux for the entire German Bight (Figure 1) from 2012 to 2016 was calculated.

270 **3. Results**

3.1. Observed Benthic Oxygen Flux

Advective fluxes across the sediment-water interface at permeable sediments measured in situ by lander are characterized by relatively small variations in July and enhanced variations in March and September (Figure 2a). The exact values range between 330 and 1420 μ mol m⁻² h⁻¹ among the sampling stations. It is difficult to judge whether a seasonal pattern exists in the advective flux due to limited number of observations.

The median standard error of diffusive oxygen flux single incubations was 0.12, and had little variation. The relative error of the average fluxes at a given site was in the range of 0.1 - 0.5 during winter and early summer, while the relative variation of flux increased considerably to the range of 0.2 - 1.4 during summer. The observed fluxes of oxygen at the different study sites had similar seasonal patterns (Figure 2b). Fluxes were low during winter (January – March) with a magnitude of 250 μ mol m⁻² h⁻¹. Starting in spring (end of March), fluxes generally increased substantially and peak at the summer – autumn transition (September). The maximum fluxes observed in late summer were as high as 2000 μ mol m⁻² h⁻¹. After the late summer, benthic fluxes decreased rapidly and remained low during autumn and winter. However, sampling density was low during November and December.



Figure 2. Temporal variation of measured oxygen flux (mean ± standard deviation) across the sediment-water interface at the sampling stations using (a) in situ lander (Ahmerkamp et al., 2017), and (b) ex situ core incubation chambers.

3.2. Simulation Results

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3.2.1. Comparison Between Model Results and Field Data

- A comparison between the measured and simulated oxygen flux across the sediment-water
- 293 interface at the sampling stations demonstrates a general satisfactory model performance (Pearson's r =
- 294 0.67) in capturing both spatial and temporal variations of benthic oxygen flux (Figure 3a), though the

model tends to overestimate oxygen flux in case of large values (> 800 μ mol m⁻² h⁻¹). Detailed simulation results and comparison with observation at each sampling station are provided in the Supplemental Information.

A critical state variable that is relevant to benthic oxygen dynamics and can be directly compared to 298 field data is macrobenthos biomass. An agreement between the measured and simulated macrobenthos 299 biomass (Figure 3b) has been already demonstrated in Zhang et al. (2019). In contrast to the 300 overestimation of oxygen flux, macrobenthos biomass is slightly underestimated in the model result for 301 values larger than 8 g C m⁻². It is worth noting that macrobenthos biomass ranges between 1 and 9 g C 302 m^{-2} at the sampling stations, except for the station NOAH-C (>10 g C m⁻²) which is located at the mud 303 depocenter adjacent to the Helgoland island, where the highest content of both mud and TOC in the 304 study area is located (Figure 1, Table 1). 305

The good model performance allows for using the simulation results to derive insights into the relationship between pelagic drivers and benthic response in terms of oxygen dynamics. Despite of a large variability in observed values of oxygen flux among the stations, the seasonality of oxygen flux at most sampling stations shares a large degree of similarity. Accordingly, our simulation results at all stations are analyzed as ensemble in the following, so that a common pattern of benthic-pelagic coupling in response to seasonality of driving forcing can be inferred.



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3.2.2. POC at Sampling Stations - Pelagic Drivers 318

319 Simulated concentration of phytoplankton biomass (measured in POC) at the nine sampling sites, which reflects primary productivity in the study area, indicate that despite of large difference in the 320 magnitude, an annual cycle with significant seasonality exists at each station (Figure 4a). The time series 321 of simulated phytoplankton biomass at all stations in all years (2012 - 2016) are characterized by a low 322 level (90th percentile $< 20 \text{ mg C m}^{-3}$) in the cold season between early November and mid-February, a 323 drastic increase since late February till mid-March indicating an early spring bloom (median = 80 mg C 324 m^{-3} at the stations), a subsequent mild decline and then a second and even stronger bloom (median = 325 120 mg C m⁻³) in late spring, followed by a sharp decline (median = 25 mg C m^{-3}) in late May and then 326 several pulses of mildly enhanced production (median between 45 and 60 mg C m⁻³) until mid-to-late 327 September (Figure 4a). 328

329 Although POC originating from fresh plankton detritus is relatively abundant in bottom water since the early spring bloom, the strength of bottom current and turbulence in the benthic boundary layer 330

control whether the POC can settle on the seafloor. Simulations indicate that bottom currents are energetic between early October and late March (in the following year) with median bottom shear stress exceeding the threshold of sediment mobilization (0.05 N m⁻²). Bottom hydrodynamic regime becomes favorable for POC deposition from April to September, with calmest condition in July and August when median bottom shear stress is less than 0.025 N m⁻². A further decrease of shear stress below 0.01 N m⁻¹ allows settling down of POC on the seafloor, which is mostly confined between April and September (Figure 4b).

The simulated bottom water temperature, which is validated by observation (Figure 4c), indicates a 338 minimum (median = 4.3 °C) in early March and a maximum (median = 18 °C) in early September 339 among the stations in 2012 - 2016. Variation in bottom temperature between October and March is 340 341 confined to a narrow band (± 4 °C between the 90th and 10th percentile), while a larger variation ($\pm 6-7$ °C between the 90th and 10th percentile) occurs in summer from June to August. Simulated bottom 342 water oxygen concentration (Figure 4d) shows a negative correlation (Pearson's r = -0.91) with the 343 temperature. Its median value ranges between 4.4 and 7.3 ml L^{-1} among different seasons at the 344 sampling stations. 345



Jan Feb Mar Apr May Jun Jul Aug Sep Oct Nov Dec Jan Jan Feb Mar Apr May Jun Jul Aug Sep Oct Nov Dec Jan 346 Figure 4. Simulated temporal (2012-2016) variation of (a) bottom water POC concentration, (b) daily-347 mean bottom shear stress, (c) bottom water temperature, and (d) bottom water oxygen concentration at 348 the sampling stations. Values between the 10th and the 90th percentiles are indicated by the grey area, 349 within which the median is marked by the solid curve. The thresholds for calculating sediment 350 mobilization ($\tau_a = 0.05 \text{ N m}^{-2}$) and deposition of organic detritus ($\tau_d = 0.01 \text{ N m}^{-2}$) are indicated in (b) 351 by the dashed lines. Field measurements of bottom water temperature at these stations are included in (c). 352 Note that the large range between the 10th and the 90th percentiles in plots (a) and (b) is attributed 353 mainly to spatial variation among the stations. 354

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356 **3.2.3. Benthic Response at Sampling Stations**

- 357 The seasonality of shear stress induces a distinct seasonal pattern of POC fluxes across the
- 358 sediment-water interface with net erosion in the cold season and net deposition in the warm season.
- 359 These two regimes alternate in early March and early October (Figure 5a).

360 The settled POC serves as the main food resource for macrobenthos. The simulated macrobenthos

361 biomass shows a rapid response to the POC input and output fluxes (Figure 5b). Macrobenthos biomass

- is at its minimum (median = 60% of the annual mean) in late February/early March due to shortage in
- food (POC), and starts to increase in response to POC input, reaching its maximum (median = 134% of
- the annual mean) in late August. This seasonal pattern is paralleled by the simulated bioturbation

diffusivity D_b (Figure 5c). Minimum and maximum bioturbation rates were calculated for February

(median = 70% of the annual mean) and August (median = 130% of the annual mean), respectively.

Bioturbation comes with two contrasting effects on POC preservation in sediments. On the one side, 367 macrobenthos mixes down oxygen into sediments and facilitates degradation of POC. On the other side, 368 bioturbation induces transport of labile POC into deeper horizons of sediment where degradation 369 efficiency is slowed down due to oxygen deficiency. Our simulations show that there is a net positive 370 impact of bioturbation on preservation of POC in sediments (Figure 5d), indicating that the amount of 371 POC transported into sediments by bioturbation exceeds the amount of remineralized POC in surface 372 373 sediments, making subsurface sediments a net sink of POC. The simulated temporal variation of labile and semi-labile POC (originated from fresh plankton detritus) in the uppermost 10 cm-surface sediments 374 375 is characterized by one-month delay in reaching its maximum (mid-September) and an accelerated decline afterwards, indicating a slightly different pattern compared to the variation of macrobenthos 376 biomass. Further, the efficiency of labile and semi-labile POC preservation in the cold season (from 377 January to March) varies remarkably among the stations, with a median = 55% of the annual mean. 378



Figure 5. Simulated temporal (2012-2016) variation of (a) daily POC flux across the sediment-water

interface with positive/negative value indicating deposition/erosion, (b) biomass of macrobenthos

divided by the respective annual mean, and (c) bioturbation diffusivity D_{h} divided by the respective

annual mean, and (d) labile and semi-labile POC content of surface sediments (uppermost 10 cm)

divided by the respective annual mean at the sampling stations.

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386	Oxygen fluxes across the sediment-water interface markedly change as soon as POC settles on the
387	seafloor due to OC remineralization being influenced by both temperature and bioturbation. Despite of
388	large difference in the magnitude of values among the stations (as shown by a large range between the
389	10th and the 90th percentile in Figure 6), a similar annual cycle is seen at most sampling stations (see
390	Supplemental Information). Simulations (Figure 6a) show that the total oxygen flux transported into
391	sediments remains on a relatively low level from December to March (in the following year) at most
392	stations, characterized by a median value $< 800 \mu\text{mol}\ \text{m}^{-2}\ \text{h}^{-1}$. The flux increases from March onward,
393	exhibits a mild peak in early May at some stations in some years (especially in 2013, see Supplemental
394	Information), and then rises again with an accelerated rate from July till reaching its maximum in late
395	August/early September (median = 1450 μ mol m ⁻² h ⁻¹). The flux decreases from mid-September to
396	November, in accordance with the observation (Figure 2). A decomposition of processes contributing to
397	the total oxygen flux reveals opposite temporal variations between the advection (Figure 6b) and the
398	diffusion (Figure 6c, d). The advective flux contributes to a major part of the total flux in permeable
399	sediments from mid-October to mid-June (in the following year) with the median exceeding 50% of the
400	total flux (Figure 7a), and its 90th percentile may even reach to 90% at some stations. However, both the
401	advective flux and its relative contribution to the total flux drop quickly from June onward with its
402	median value accounting for less than 40% of the total flux in August and early-September. By contrast,
403	diffusive flux remains relatively low in the cold season and peaks in the warm season (Figure 6c, d).
404	Despite of a similar variation pattern, molecular diffusion and bio-diffusion differ significantly in
405	magnitude. Molecular diffusion accounts only for a minor portion (median < 9%) of the total flux
406	(Figure 7b), while bio-diffusion contributes to a significant portion (15% in cold seasons and 62% in
407	warm seasons as median) of the total flux in permeable sediments (Figure 7c) and more than 85% in
408	impermeable sediments (Figure 7d).



Jan Feb Mar Apr May Jun Jul Aug Sep Oct Nov Dec Jan
Figure 6. Simulated temporal (2012-2016) variation of (a) total oxygen flux across the sediment-water
interface, (b) advective flux induced by porewater flow, (c) molecular diffusive flux, and (d) biodiffusive flux at the sampling stations. Note that the large range between the 10th and the 90th
percentiles in the plots is attributed mainly to spatial variation among the stations. Measurements of
advective and diffusive oxygen fluxes are included in (b) and (d), respectively.



416 Jan Feb Mar Apr May Jun Jul Aug Sep Oct Nov Dec Jan Figure 7. Simulated temporal (2012-2016) variation of the relative contribution of (a) advection to the 418 total flux of oxygen at permeable sampling stations, (b) molecular diffusion to the total flux of oxygen at 419 all sampling stations, (c) bio-diffusion to the total flux of oxygen at permeable sampling stations, and (d) 420 bio-diffusion to the total flux of oxygen at impermeable sampling stations. Note that the large range 421 between the 10th and the 90th percentiles in the plots is attributed mainly to spatial variation among the 422 stations.

424 **3.2.4. Benthic Oxygen Flux at Regional Scale**

Simulation results of benthic oxygen flux at a regional scale (German Bight) exhibit a complex distribution characterized by patches of hotspots (Figure 8a). A decomposition into advective and diffusive components helps to quantify their respective contribution to the spatial variability and to unravel driving mechanisms behind that.

Results show that the contribution of molecular diffusion is minor, with a mean annual value below 200 μ mol m⁻² h⁻¹ in a major part of the area and accounting for less than 10% of the total flux. This result is consistent with the result at the sampling stations. It is worth to note that the molecular diffusion is in general higher in offshore area (water depth > 40 m) than in nearshore area, except for the part around the Helgoland island and off the Elbe estuary (Figure 8b).

434 In contrast to the spatial pattern of molecular diffusive flux, the advective flux is higher in nearshore area (water depth < 40 m) with a mean annual value in general above 300 μ mol m⁻² h⁻¹, even 435 reaching to more than 1000 μ mol m⁻² h⁻¹ in some of the hotspots (Figure 8c). We identified a clear link 436 between these hotspots and sediment grain size, and found that the 250 μ m (d_{50}) median grain size 437 contour can be used as a geographic indicator for the extent of hotspots for advective flux (Figure 8c). 438 Two sampling stations (NOAH-A and CCP-G) are located within these hotspots as evidenced by field 439 observations showing higher values at these two sites (see Supplemental Information). A significant part 440 (~25%) of the study area, e.g. the ancient Elbe river valley and the mud depocenter off Helgoland, is 441 442 covered by impermeable sediment where advective oxygen flux is zero.

The spatial distribution of bio-diffusive flux shows a high correlation with macrobenthos biomass. Hopspots of bio-diffusive flux (> 800 μ mol m⁻² h⁻¹) are normally confined in areas where macrobenthos biomass exceeds 5 g C m⁻² in ash free dry weight (Figure 8d). These areas include both muddy and sandy seafloors. A comparison with our previously published results for distrbution of dominant benthic community structure (Figure 7 in Zhang et al., 2019) indicates that in the muddy hotspots the diffusive
flux is mainly contributed by deposit feeders while in the sandy hotspots it is mainly caused by surface
feeding animals including suspension and filter feeders.



Figure 8. Simulated mean annual total oxygen flux (2012-2016) across the sediment-water interface (a) and its components (b, c, d) in the German Bight with marked sampling stations. Note that the Wadden Sea area including the tidal channels between the barrier islands is not covered by the data. The 40 m water depth contour is indicated in (b). The 250 μ m (d_{50}) median grain size contour is indicated in (c) and the 5 g C m⁻² macrobenthos biomass contour is indicated in (d).

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To aid further interpretation, results for the entire study area are averaged spatially according to the sediment type (i.e. permeable and impermeable). Distinct patterns are shown between the two sediment types (Figure 9a, b). Similar to the results at the sampling stations, benthic oxygen flux in permeable sediments is predominantly controlled by porewater advection in the cold season and significantly modulated by bio-diffusion in the warm season (Figure 9c). At an annual scale, advection and bio-

diffusion contribute to 52 ± 7 % and 45 ± 9 % of the total oxygen flux in permeable sediments that cover 462 ~75% of the study area. A combination of the opposite seasonal varying patterns of advection and 463 diffusion results in a mild variation (median between 1200 and 1600 µmol m⁻² h⁻¹) of the total oxygen 464 flux in permeable sediments throughout the year at a regional scale (Figure 9a). By contrast, a much 465 stronger seasonal variation pattern (median between 700 and 1700 µmol m⁻² h⁻¹) dominated by bio-466 diffusion is shown in impermeable sediments with less annual variability at a regional scale (Figure 9b, 467 d). At an annual scale, bio-diffusion contributes to 87 ± 4 % of the total oxygen flux in impermeable 468 sediments of the study area. The range of our results is generally supported by various field 469 measurements from the southern North Sea compiled by Provoost et al. (2013) and Ahmerkamp et al. 470 (2017) showing that the benthic oxygen flux varies widely between 40 and 2500 μ mol m⁻² h⁻¹ (i.e. 1-60 471 mmol $m^{-2} d^{-1}$) at individual stations with most values between 400 and 1700 µmol $m^{-2} h^{-1}$ (i.e. 10-40 472 mmol $m^{-2} d^{-1}$). 473



Figure 9. Simulated temporal (2012-2016) variation of spatially-averaged total oxygen flux across the
sediment-water interface in (a) permeable sediments and (b) impermeable sediments, and the component
induced by bio-diffusion in (c) permeable sediments and (d) impermeable sediments. Note that the 10th
and the 90th percentiles and the median value refer to the annual variation.

480 **3.2.5. Quantitative importance of bioturbation**

It can be seen from Equation (1) and (6) that bioturbation not only directly diffuses oxygen across the sediment-water interface, but also indirectly influences the oxygen flux through a transport and mixing of the electron donor, i.e. POC, thereby changing the consumption rate of oxygen and the consequent vertical gradient of oxygen. To further assess the role of bioturbation, a new simulation with exclusion of bioturbation (termed No_D_b hereafter) was done, and results are compared with those including bioturbation (termed reference result hereafter).

Simulation results of No D_b differ significantly from both the observation and the reference result 487 in not only the magnitude of the total flux but also the seasonal variation pattern (Figure 10a, b). Results 488 489 of No D_b show that in permeable sediments, the total oxygen flux is predominantly controlled by 490 porewater advection that takes up more than 70% during most of the year. The diffusive oxygen flux constantly ranges below 200 μ mol m⁻² h⁻¹ in permeable sediments, which is much lower than the 491 observed values (Figure 2). In contrast to the advection-dominated permeable sediments, oxygen flux is 492 493 determined solely by molecular diffusion in impermeable sediments and exhibits a similar variation 494 pattern with observation and the reference result, but with a much smaller magnitude (Figure 10b). Its 495 median value in cold and warm seasons account for only 25% and 15% of the median of the observed values in the respective periods. Such huge discrepancy between the simulation results of No_D_b and 496 497 observation demonstrates that a major portion of the observed oxygen flux is attributed to biotic factors (bio-diffusion/bio-ventilation) instead of abiotic factors (molecular diffusion). 498

The indirect impact of bioturbation on oxygen consumption through a transport and mixing of POC is reflected by comparison of modelled molecular diffusion between No_D_b and the reference result. The comparison reveals a two-sided effect of bioturbation on the molecular diffusive flux (Figure 10c). In stations characterized by energetic bottom currents and low POC depositional rate (e.g. CCP-G), bioturbation effectively preserves POC in sediments and increases benthic oxygen consumption fuelled

by the transport of electron acceptors via porewater-advection compared to that in No_D_b, in which case 504 POC is simply deposited at the surface and subsequently resuspended and carried away by currents. 505 506 Bioturbation results in an enhanced molecular diffusive flux in the reference result compared to that of 507 No_D_b. Such enhancement is reflected in the median value of the ratio of the simulated molecular diffusive flux in No_D_b to the reference result, which is below one during most time of the year (Figure 508 10c). On the other hand, when bottom currents are persistently weak and POC depositional rates are 509 510 high (e.g. NOAH-C), bioturbation transports a significant portion of POC into deeper sediments where oxygen is deficient, thereby reducing oxygen consumption in the upper-most sediment layer. The 511 512 resulting molecular diffusive flux across the interface is greatly reduced compared to that in No_D_b. This is reflected in the 90th percentile of the ratio of the simulated molecular diffusive flux in No_D_b to 513 the reference result, which exceeds 1.4 all year round (Figure 10c). It is worth to note that although 514 515 bioturbation results in a reduction of oxygen consumption in the upper-most sediment layer, it increases 516 oxygen consumption beneath this layer by mixing down oxygen into deeper sediments, resulting in a flux much higher than that induced by molecular diffusion alone (Figure 8d). 517



Figure 10. Simulated temporal (2012-2016) variation of (a) total oxygen flux across the sediment-water
 interface at the stations with permeable sediments (NOAH-A, NOAH-B, NOAH-E, CCP-G, NOAH-I)
 in No_D_b, (b) total oxygen flux at the stations with impermeable sediments (NOAH-C, NOAH-D,
 NOAH-F, NOAH-H) in No_D_b, and (c) ratio of molecular diffusive flux in No_D_b to the reference
 result at all stations. Note that the large range between the 10th and the 90th percentiles in the plots is
 attributed mainly to spatial variation among the stations.

3.2.6. Impact of bedform morphodynamics on oxygen flux

526	An important modulation of the oxygen flux across the sediment-water interface by small-scale
527	bedform migration has been demonstrated in a numerical study by Ahmerkamp et al. (2015). Based on
528	their results, Ahmerkamp et al. (2015) proposed three modes of porewater advection in permeable
529	sediments: 1) Mode 1 corresponds to deep oxygen penetration and high oxygen flux when bedform is
530	stationary, 2) Mode 2 refers to the formation of a redox seal during a competition of bedform migration
531	and porewater flow (0.01< c_{ph}/u_p < 6) that leads to a decreasing oxygen flux along with increasing
532	bottom current velocity, and 3) Mode 3 represents a large reduction in the oxygen flux due to rapid

bedform migration ($c_{ph}/u_p > 6$) that overwhelms porewater flow. These modes were later confirmed by field observation in the German Bight (Ahmerkamp et al., 2017).

Our study further explores the impact of small-scale morphodynamics on oxygen flux at a longer time scale (2012-2016) based on 3-D hydrodynamic-biogeochemical modelling. To assess the impact of morphodynamics, a new simulation (termed No_Mor hereafter) with exclusion of bedform migration (i.e. c_{ph} is fixed to zero) was done, and results are compared with the reference result.

Simulation results indicate enhanced porewater advection in permeable sediments (Figure 11a) in 539 540 No_Mor when compared to the reference result (Figure 6b). The ratio of the simulated porewater advective flux in No_Mor to the reference results is persistently larger than one in all seasons, with 541 highest value in the cold season from November to February characterized by a 90% percentile up to 6 542 543 and a median larger than 1.4 (Figure 11b). These results suggest that actively migrating small-scale bedforms such as ripples and dunes effectively dampen oxygen input into sediments. Exclusion of such 544 impact in numerical modelling or incubation chamber measurements would lead to an overestimation of 545 the oxygen flux by nearly an order of magnitude in migrating bedforms. This is especially true for the 546 shallow southern North Sea in the cold season (from October to February) when currents are driven by 547 548 strong westerly winds often accompanied by storms (S ündermann and Pohlmann, 2011; Zhang et al., 549 2011; Geyer, 2014).



Figure 11. Simulated temporal (2012-2016) variation of (a) advective oxygen flux across the sediment water interface at the stations with permeable sediments (NOAH-A, NOAH-B, NOAH-E, CCP-G,
 NOAH-I) in No_Mor, and (b) the ratio of advective oxygen flux in No_Mor to the reference result at
 these stations. Note that the large range between the 10th and the 90th percentiles in the plots is
 attributed mainly to spatial variation among the stations.

557 4. Discussion

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558 **4.1. Comparison with existing literature and research needs**

The central importance of bioturbation in reworking of solids (POC) in coastal sediments at a 559 regional scale has been demonstrated in Zhang et al. (2019), who showed that macrobenthos is able to 560 rework more than 50% of the total budget of POC in the uppermost 30 cm sediments annually in the 561 southern North Sea. The authors found that bioturbation diffusivity of solids in general ranges between 562 4×10^{-4} and 3×10^{-2} cm² h⁻¹ with a mean value of 4.6×10^{-3} cm² h⁻¹ in the study area. Rapid deposition 563 of fresh POC in late Spring/early Summer may cause a bioturbation enhancement by an order of 564 magnitude locally. Bioturbation has an even more profound impact on solutes than on solids. Berg et al. 565 (2001) found that bioturbation diffusivity of solutes is at least one order of magnitude larger than that of 566 solids. Our simulation results indicate that bioturbation diffusivity of oxygen in general varies between 567

 8×10^{-2} (the 10th percentile) and 5×10^{-1} cm² h⁻¹ (the 90th percentile) among the sampling stations, 568 569 which is 15-100 times of the mean bioturbation diffusivity of POC and 5-20 times of the corresponding 570 molecular diffusivity at the same sites. This result is supported by existing literature. For example, by measurements of surface sediments collected in a sub-tidal area off Helsing or (Denmark) using a large 571 572 incubation chamber, Wenzhöfer and Glud (2004) found that macrobenthos-mediated oxygen flux was the dominant oxygen consumptions process accounting for up to 92% of the total oxygen flux. 573 Jørgensen and Parkes, R. J. (2010) attributed the diffusivity of solutes (sulfate, methane, DIC) measured 574 in surface sediments in an embayment of the North Sea (Limfjorden, northern Denmark), which was 575 $0.36-0.72 \text{ cm}^2 \text{ h}^{-1}$ and being 30-50 fold higher than the corresponding molecular diffusivity, to bio-576 577 diffusion/bio-ventilation. Sulfate measurements in sediments of Aarhus Bay (Denmark) by Thamdrup et al. (1994) indicated that bio-diffusivity was $\sim 10^3$ times higher than molecular diffusivity. In the mud 578 depocenter off Helgoland, the large spatial variations of benthic oxygen flux (up to a factor of 4) over 579 580 small distances was also attributed to macrobenthos (Oehler et al., 2015).

By inverse modelling using time series of measured benthic oxygen consumption rate, water-581 582 column and sediment chlorophyll concentrations, and bottom water temperature at a fixed site off the 583 Belgian coast in the southern North Sea, Provoost et al. (2013) identified organic matter (POC) content and temperature as two dominant controlling factors of oxygen flux across the sediment-water interface. 584 585 The impact of porewater advection was not taken into account because the measurements were done ex situ. Therefore, their data represent diffusive fluxes only. The measured sediment oxygen uptake rate 586 ranged from 170 to 1340 μ mol m⁻² h⁻¹ (i.e. 4 to 32 mmol m⁻² d⁻¹). Between January and March, the 587 lowest values were found, while the highest values were from summer between mid-June and mid-588 August. Our results are consistent with these findings in terms of the magnitude and the seasonal 589 590 variation pattern. However, in their model Provoost et al. (2013) did not explicitly consider the effect of macrobenthos on the oxygen flux, despite of quite abundant macrobenthos (ranging from 5.7 g C m⁻² in 591

February to 78 g C m⁻² in October) that potentially acted as an important mediator for transport of 592 organic matter into the sediments at the sampling site. The impact of macrobenthos was instead 593 594 implicitly incorporated in the sedimentary POC content since macrobenthos actively processes and 595 transports POC into deeper sediments, thereby slowing down the degradation efficiency of POC (Meile and Van Cappellen, 2005; Middelburg, 2018; Zhang et al., 2019). This explains why the modelled 596 degradation rate of POC was ~ $5.7 \pm 2.2 \times 10^{-3} d^{-1}$ in Provoost et al. (2013), which represents a typical 597 value for semi-labile POC (Zhang et al., 2017) but is one order of magnitude smaller than that of labile 598 599 POC originated from freshly deposited phytoplanktonic detritus (Westrich and Bemer, 1984; Harvey et al., 1995; Zhang et al., 2017). The critical importance of macrobenthos in mixing POC into sediments 600 601 has been observed in various stations along the southern North Sea coast (e.g. Kamp and Witte, 2005; Le Guitton et al., 2017). Therefore, the importance of POC availability and bottom water temperature on 602 the benthic oxygen flux are widely acknowledged (e.g. Glud, 2008; Provoost et al., 2013; Belley et al., 603 604 2016), while the impact of macrobenthos remains either implicit or mostly focused on small-scale 605 experiments (Middelburg, 2018). This points out a need for quantitative assessment of the impact of bioturbation on benthic oxygen flux, especially at a regional to global scale. The benthic-pelagic coupled 606 model system presented in this study, which explicitly resolves the interactions among benthic fauna, 607 oxygen consumption and early diagenesis of POC, may serve as a useful tool to help address this issue. 608

609 **4.2. Practical environmental indicators for benthic oxygen flux**

It is widely acknowledged that benthic oxygen flux is highly variable in heterogeneous coastal seas and is affected by multiple factors including bottom water oxygen concentration, POC availability, temperature, light, hydrodynamics, sediment composition, sedimentation rates, and benthic fauna (Glud, 2008). The extent to which each factor contributes to the benthic oxygen flux depends on site characteristics and time of measurement (Belley et al., 2016; Chipman et al., 2016).

Our results at a regional scale (Figure 8) indicate that two environmental variables, namely d_{50} and 615 macrobenthos biomass, can be particularly useful to identify hotspots of benthic oxygen flux. These two 616 617 variables are highly correlated with other environmental factors such as hydrodynamics, temperature and POC availability, and can be used as proxies for these factors. For instance, on a sandy permeable 618 seafloor, d_{50} is an indicator of hydrodynamics. A larger d_{50} implies stronger bottom currents and larger 619 620 permeability, thereby facilitating stronger porewater flow and deeper oxygen penetration depth (Ahmerkamp et al., 2017). On the other hand, a fast migrating bedform (usually composed of fine sands) 621 inhibits porewater advection and oxygen flux. This indicates that benthic oxygen flux on a sandy 622 permeable seafloor may not be linearly related to d_{50} until d_{50} reaching a certain level. Our simulation 623 results indicate that $d_{50} = 250 \ \mu m$ defines such level and hotspots of benthic oxygen flux associated with 624 porewater advection mostly occurs where $d_{50} \ge 250 \ \mu\text{m}$. Further investigations are nevertheless needed 625 to judge whether this is valid and can be applied to other coastal regions. 626

Large variations of benthic oxygen flux over small distances on a homogeneous seafloor are often 627 attributed to difference in benthic fauna through bioturbation (Oehler et al., 2015). However, since 628 629 bioturbation is difficult to measure in field (Queir ás et al., 2013; Griffiths et al., 2017; Middelburg, 2018), we propose to use macrobenthos biomass, which is easy to measure in field, as a practical 630 indicator for hotspots of benthic oxygen flux because most hotspots associated with bio-diffusion are 631 confined in the 5 g C m⁻² biomass contour according to our results. Macrobenthos biomass is a proxy for 632 POC quantity and quality (Dauwe et al., 1998; Herman et al., 1999; Campany à Llovet et al., 2017) 633 634 which are jointly determined by hydrodynamics, primary production, temperature and sedimentation rate (Zhang and Wirtz, 2017). The community structure of macrobenthos, e.g. whether surface-feeding 635 or subsurface-feeding communities dominate, imposes further control on the magnitude of benthic 636 637 oxygen flux. Our results show that in the muddy hotspots the diffusive flux is mainly contributed by deposit feeders while in the sandy hotspots it is mainly caused by surface feeding animals including 638

639 suspension and filter feeders. The difference in community structure may explain the spatial

heterogeneity of the magnitude of benthic oxygen flux within the hotspots (Figure 8d). However, since
this argument is based on modeling results, further experimental work is needed to unravel a quantitative
difference in benthic oxygen flux induced by different community structures, especially under the same
setting of biomass.

644 **5. Conclusions**

Through an integrated study combining field observation and numerical modelling of the southern North Sea as an example, we have demonstrated the central importance of benthic animals in controlling benthic-pelagic coupling through mediating both solutes (oxygen) and solids (POC) transport across the sediment-water interface. The following conclusions are drawn.

Both advective and diffusive oxygen fluxes across the sediment-water interface are characterized by strong seasonality in the study area. However, the variation pattern and driving mechanisms behind them are different. Variation of the advective flux is a summer-low, winter-high pattern mainly driven by hydrodynamics and bedform morphodynamics, while variation of the diffusive flux is summer-high and winter-low resulting from a combined effect of pelagic drivers (hydrodynamics, temperature and primary production) and benthic dynamics associated with the interaction between organic carbon and macrobenthos.

In permeable sediments, a combination of the opposite seasonal varying patterns of advection and diffusion results in large spatial and temporal variability among individual sites at a local scale but a consistent and mild seasonal variation (summer-high and winter-low) at a regional scale. In impermeable sediments, benthic oxygen fluxes are dominated by bio-diffusion exhibiting a remarkable seasonality. Local-scale spatial and temporal variability there is mainly caused by availability of POC originated from freshly deposited phytoplankton detritus and subsequent bioturbation intensity of macrobenthos related to food-foraging and growth.

Bio-diffusion caused by macrobenthos contributes to more than 85% and 52% of the total benthic oxygen fluxes in muddy seabed and at a regional scale, respectively.

Last but not least, we found two environmental variables, namely d_{50} and macrobenthos biomass, that can be particularly useful for identifying hotspots of benthic oxygen fluxes in coastal shelf environments.

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884 **Conflict of Interest**

None declared.