Autotrophic respiration dominates ecosystem respiration at Canadian boreal forest

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Abstract

Most of the Earth System Models (ESMs) rely on empirical functions for decomposition of litter with multiple soil carbon pools decaying at different rates to estimate R_{eco} variations and its partitioning into autotrophic (R_a) and heterotrophic respiration (R_h) in relation to variation in temperature and soil water content. However, microbially-mediated litter decomposition schemes are very scarce in ESMs. Microbial enzymatic processes are integral parts in litter as well as soil organic matter (SOM) decomposition. Here we developed a mechanistic model comprising of multiple hydro-biogeochemical modules to explicitly incorporate microbially-enzymatic litter decomposition and decomposition of SOM for estimating regional-scale R_a , R_h and R_{eco} . Modeled annual mean R_{eco} values are found varying from 1000 to 8000 kg C ha⁻¹ yr⁻¹ in 2000-2013 within the boreal forest covered sub-basins of the Athabasca River Basin (ARB), Canada. While, for the 2000-2013 period, the annual mean R_a and R_h are varying within 800-6000 kg C ha⁻¹ yr⁻¹ and 500-3500 kg C ha⁻¹ yr⁻¹, respectively. R_a generally dominates R_{eco} with nearly 30-80% contribution in most of the sub-basins in ARB. The model estimates corroborate well with the site-scale and satellite-based estimates reported at similar land use and climatic regions. Modeling the partitioning of R_{eco} to R_a , and R_h are critical to understanding future climate change feedbacks and to help reduce uncertainties in ESMs in the boreal and subarctic regions.

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1 Autotrophic respiration dominates ecosystem respiration at Canadian boreal forest

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10 Key Points:

- A newly developed, mechanistic, hydro-biogeochemical model for simulating ecosystem
 respiration (Reco) at the subarctic regions
- Regional-scale modelling of Reco and its partitioning to Autotrophic respiration (Ra) and
 heterotrophic respiration (Rh)
- 15 R_a dominate R_{eco} at Canadian boreal forests
- 16

17 Abstract

18 Most of the Earth System Models (ESMs) rely on empirical functions for decomposition of litter

19 with multiple soil carbon pools decaying at different rates to estimate R_{eco} variations and its

20 partitioning into autotrophic (R_a) and heterotrophic respiration (R_h) in relation to variation in

21 temperature and soil water content. However, microbially-mediated litter decomposition

22 schemes are very scarce in ESMs. Microbial enzymatic processes are integral parts in litter as

23 well as soil organic matter (SOM) decomposition. Here we developed a mechanistic model

comprising of multiple hydro-biogeochemical modules to explicitly incorporate microbially-

enzymatic litter decomposition and decomposition of SOM for estimating regional-scale R_a, R_h

and R_{eco}. Modeled annual mean R_{eco} values are found varying from 1000 to 8000 kg C ha⁻¹ yr⁻¹ in 27 2000-2013 within the boreal forest covered sub-basins of the Athabasca River Basin (ARB),

27 2000-2013 within the boreal forest covered sub-basins of the Athabasca River Basin (ARB),
28 Canada. While, for the 2000-2013 period, the annual mean R_a and R_h are varying within 800-

29 6000 kg C ha⁻¹ yr⁻¹ and 500-3500 kg C ha⁻¹ yr⁻¹, respectively. R_a generally dominates R_{eco} with

nearly 30-80% contribution in most of the sub-basins in ARB. The model estimates corroborate

well with the site-scale and satellite-based estimates reported at similar land use and climatic

regions. Modeling the partitioning of R_{eco} to R_a , and R_h are critical to understanding future

climate change feedbacks and to help reduce uncertainties in ESMs in the boreal and subarctic

34 regions.

35 Plain language summary

36 Estimating ecosystem respiration (Reco) is becoming a major topic in global climate change

37 studies. Our new integrated hydro-biogeochemical modeling can successfully estimate regional-

38 scale R_{eco} and its subsequent partitioning into autotrophic (R_a) and heterotrophic respiration (R_h)

at the boreal forest covered Athabasca river basin (ARB), Canada. R_a generally dominates R_{eco}

40 with nearly 30-80% contribution in most of the sub-basins in ARB. The proposed modeling

41 approach does represents a step forward to improve global respiration modeling in widely used

42 earth system models (ESMs) as it is well capturing both, the site-scale measurements and

43 satellite-based estimates. These more realistic respiration estimates are achieved after shifting

44 from purely empirical representation of ecosystem respiration and its partitioning pathway to

45 microbial enzymatic kinetics-based approach. The approach presented in this paper would

directly improve ESM estimates or indicate where in existing ESM improvement should be

47 sought.

48 **1 Introduction**

49 Ecosystem respiration (R_{eco}) is the primary natural source of atmospheric carbon dioxide (CO₂) throughout the globe (Ciais et al., 2014). There is an urgent need to completely understand 50 the soil carbon conversion processes and the associated land-surface gas exchange in estimating 51 climate change feedback, which ultimately lead to the release of soil carbon as CO_2 to the 52 atmosphere (Mitchard, 2018). In natural ecosystems, R_{eco} can be partitioned into the soil 53 54 heterotrophic respiration (R_h) resulting from soil microorganisms and autotrophic respiration (R_a) resulting from plant species (both from above ground components and roots) (Hicks Pries et 55 al., 2013; 2016). Globally, Ra is the predominant component of the terrestrial carbon budget with 56 photosynthetic carbon consumption rate of 54-71% (Ryan et al., 1997). In permafrost regions, Ra 57 accounts for 40 to 70% of the total ecosystem respiration (Hicks Pries et al., 2013). The Ra, and 58 R_h have different adaptation mechanisms in response to climate change. For example, rising 59

60 temperature can accelerate microbial enzymatic activities and thus enhance R_h and thus facilitate

nutrient dynamics based on substrate availability (Davidson and Janssens, 2006; Manzoni et al.,

- 62 2008). On the other hand, rising temperature has different influence in plant growth
- 63 (aboveground component showing positive response while root showing near neutral response)
- and also on R_a (Hick Pries et al., 2013). For example, 2.3 ^oC warming of soil lead to 20%
- 65 increase in aboveground productivity at an Alaskan site (Natali et al., 2009). With an increase of
- 2 °C, root respiration was not showing any significant changes, while heterotrophic respiration
 showed 21% increase linked with increased microbial activities (Wang et al., 2014). Therefore,
- $_{68}$ quantifying the distinct contributions of R_{eco}, R_a, and R_h is necessary to estimate climate
- feedbacks and sensitivity in the rapidly changing subarctic regions but state of the art monitoring
- techniques, such as eddy covariance and remote sensing techniques, are still unable to directly
- partition R_a and R_h from R_{eco} (Davidson and Janssens, 2006; Konings et al., 2019). In addition to
- the distinct characteristics and feedbacks to climate change processes by R_a and R_h , respiration
- partitioning is particularly important as it has been reported that global land carbon sink has been
 increasing in recent years (Ciais et al., 2019).

Climate change-linked future projections of ecosystem respiration are challenging 75 because estimating litter and SOM decomposition rates and anticipated soil derived CO₂ 76 feedbacks resulting from anthropogenic warming are both seen as being highly uncertain 77 (Collins et al., 2013; Crowther et al., 2016). Many experimental and modelling studies have been 78 79 performed to determine the Reco and its partitioning and gross primary production (GPP) at different land use types (Hardie et al., 2009; Hicks Pries et al., 2013; Senapati et al., 2018). On 80 the other hand, predicting the net ecosystem exchange (NEE), and R_{eco} require the development 81 of sophisticated land-surface models. These modeling approaches can be categorized into: 1) 82 agroecosystem models, and 2) the Earth System Models (ESMs). Both modelling approaches 83 commonly use empirical formula of multiple soil C pools decaying at different rates to calculate 84 Reco (Oleson et al., 2010; Clark et al., 2011; Davison and Janssens, 2006; Del Grosso et al., 85 2005). The CENTURY model used simplified functions of soil temperature and moisture for 86 estimating soil respiration and its partition to autotrophic and heterotrophic components (Del 87 88 Gorsso et al., 2005). The decaying module of multiple soil C pools in DayCent has also been incorporated into ORCHIDEE (Qiu et al., 2018) and CLMs (Lawrence et al., 2019). The 89 microbial processes in ESMs are simplified into linear, empirical equations (Crowther et al., 90 2014; 2019). It is found that most ESMs cannot reproduce grid-scale variation in soil C due to 91 missing key processes and the predicted global carbon stocks in the fifth Coupled Model 92 Intercomparison Project (CMIP) led to 6-fold difference in predicted data (Todd-Brown et al., 93 2013). Microbial activity responses (as reflected in soil heterotrophic respiration) are expected to 94 increase with warming (Karhu et al., 2014; Walker et al., 2018). However, the relationship 95 between warming and soil carbon loss, as well as overall ecosystem respiration, is not 96 straightforward as the microbially mediated litter and SOM decomposition processes are not 97 linearly correlated with temperature (Melillo et al., 2017). Additionally, microbially-mediated 98 99 decomposition of SOM is not only an important biological-driven process for carbon conversion but also play a major role in overall global nutrient dynamics (Manzoni et al., 2008). 100

Despite importance, both site-scale agroecosystem models and ESMs have also not yet explicitly considered: (i) microbially-mediated decomposition of litter and SOM, (ii) redox processes and (iii) hydrological processes coupled with soil microbial processes (Wieder et al., 2013; Bhanja et al., 2019a; Crowther et al., 2019). Apart from the environmental variables, the turnover of organic material is directly controlled by soil microbes (Crowther et al., 2019). As a result, there is a need to improve microbial processes for global carbon modeling estimates

(Wieder et al., 2013; Crowther et al., 2019). Microbial abundances are relatively larger in arctic 107 and subarctic regions (Serna-Chavez et al., 2013; Xu et al., 2013). They are also responsible for 108 biogeochemical cycling of nutrients (Crowther et al., 2019). However, different types of 109 microbes require favorable soil redox conditions for their growth (deAngelis et al., 2010). Thus, 110 soil redox condition is an essential measure for the dynamics of nutrients as well as soil organic 111 matter, which microbes used as a substrate (Bhanja et al., 2019a; 2019b). It has been found that 112 nutrient availability directly controls soil organic matter stock and terrestrial carbon sink (Wieder 113 et al., 2015). Alteration of hydrological processes has profound impact in soil carbon 114 mineralization (Anthony et al., 2018). Therefore, integration of hydrological processes along 115 with redox/biogeochemical and microbial processes would definitely improve model estimates 116 117 (Bhanja et al., 2019a; 2019b). Based on the present structure of the contemporary ESMs, it is very difficult to improve further accuracy and generality of the litter and SOM decomposition 118 modules (Wieder et al., 2015; Luo et al., 2016). Particularly, while the ecosystem respiration is 119 calculated using the application of Q₁₀ in the remote sensing and eddy covariance (Reichstein et 120 al., 2007; Zhao and Running, 2010), the techniques are also used extensively for verification and 121 calibration of ESMs and agroecosystem models. This is an open loop verification, which is 122 unreasonable according to scientific principles. Therefore, global-scale respiration studies 123 continue to be sought for further improving the modeling processes and soil respiration estimates 124 (Todd-Brown et al., 2013; Luo et al., 2016; Wang et al., 2020). For closed loop feedback, a 125 126 decomposition feedback to warming requires accounting explicitly for not only temperature but also nutrient availability and microbial activities (Davidson et al., 2012). 127

128 Substantial amount of soil organic carbon can be released due to enhanced microbial activities associated with climate warming in arctic, subarctic regions (Schuur et al., 2015). The 129 arctic and subarctic regions are the most sensitive regions to the combined climate change 130 impacts due to declining permafrost, glacial retreat and change in freeze-thaw cycles on its 131 ecosystems (Bates et al., 2008). An abrupt thawing in lakes due to global warming can lead to 132 faster mobilization of deeper stored carbon (Anthony et al., 2018). Therefore, in these regions, 133 134 modeling the carbon mobilization processes and their feedbacks are much sought after (Schuur et al., 2009). The present study describes the development of a mechanistic model to simulate 135 regional-scale, autotrophic, heterotrophic and total ecosystem respirations at the boreal forest 136 covered Athabasca River Basin (ARB), Canada (Figure S1). This being achieved through an 137 explicitly integrated inclusion of microbially-mediated decomposition of litter as well as SOM, 138 139 redox processes and hydrological processes. Model output validation at regional-scale was achieved using both site-scale, as well as remote-sensing data. Importantly, our approach can 140 also provide a new way for improving accurate carbon emission estimates of the litter and SOM 141 decomposition modules, which so far are simply using linear carbon pool transformation 142 concepts in all of the ESMs (Wieder et al., 2015; Luo et al., 2016). 143

- 144 **2 Materials and Methods**
- 145 2.1 Model introduction

The Soil and Water Assessment Tool (SWAT) is widely used for its capacity to simulate
 regional-scale hydrology in a detailed way (Arnold et al., 1998). Integrated hydro-

- 148 biogeochemical modeling was performed using our microbial kinetics and thermodynamics
- 149 (MKT) model integrated within SWAT, known as SWAT-MKT (Bhanja et al., 2019a; 2019b;

150 Bhanja and Wang, 2020). New carbon cycle capabilities are incorporated into SWAT to simulate

ecosystem respiration components from litter decomposition, root respiration, above ground

respiration and respiration component from the dissolved organic carbon transformation by

enzymatic processes (Figure S2; S3). The entire chemical processes considered in the new

version of SWAT-MKT are shown in Figure S3. Major oxidation-reduction reactions considered in this approach were given in Table S1 and their reaction quotient values are provided in Table

- 156 S2.
- 157 2.2 Soil mineralization and litter decomposition

Soil mineralization module is modeled considering two carbon pools such as, active and passive. The active pool represent fraction of active litter including microbial biomass (Fujita et al., 2014). We used microbial enzymatic litter transformation approach that is an advancement of the CENTURY model's simple, first order kinetics-based litter decomposition approach (Parton et al., 1987; 1994; Fujita et al., 2014):

(1)

$$R_{d_{i,C}} = k_{i,C} \times C_i$$

164 Where, litter decomposition rate originally adopted in CENTURY model: $R_{d_{i,C}}$ (gC kg⁻¹ 165 soil d⁻¹). *i* represents the substrate type i.e. active and passive. C_i (gC kg⁻¹ soil) is the carbon 166 content within active or passive substrate. $k_{i,C}$ (d⁻¹) is the first-order decomposition coefficient of 167 *Ci*.

Litter decomposition can also be modeled through microbial enzymatic approach following one-substrate Michaelis-Menten kinetics (Fujita et al., 2014), the new decomposition rate becomes:

171
$$R_{d_{i,M}} = k_{i,M} \times \frac{C_i}{Km_i + C_i}$$
(2)

172 Where, Km_i is the half-saturation constant or Michaelis-Menten constant. $k_{i,M}$ is the 173 decomposition coefficient of C_i and can be estimated separately for the active (AC) and passive 174 (PA) substrates as:

175
$$k_{AC,M} = \frac{k_{AC,C} \times (Km_{AC} + 2C_b)}{C_b}$$
(3)

176
$$k_{PA,M} = \frac{k_{PA,C} \times (Km_{PA} + C_T)}{C_b}$$
 (4)

Where, $k_{AC,C}$ and $k_{PA,C}$ are decomposition coefficients used in CENTURY model for the active and passive substrates, respectively. Km_{AC} is the half-saturation constant for active substrate and it is approximated as 0.3 g C kg⁻¹ soil (Allison et al., 2010). Km_{PA} is the halfsaturation constant for the passive substrate and its value is taken as 600 g C kg⁻¹ soil (Allison et al., 2010). C_b (g C kg⁻¹ soil) represents microbial biomass; the value is approximated as the median microbial biomass (0.87 g C kg⁻¹ soil) from a global-scale study of Cleveland and Liptzin (2007). C_T represents total carbon stock of soil; its value is approximated as the global total soil carbon (46 g C kg⁻¹ soil; Cleveland and Liptzin, 2007). 185 Consideration of microbial biomass being an active component of the litter and assuming 186 microbial biomass decomposition is directly proportional to the litter decomposition, the new 187 litter decomposition rate ($R_{d_{i,MM}}$) becomes (Fujita et al., 2014):

188
$$R_{d_{i,MM}} = k_{i,M} \times \frac{C_i}{Km_i + C_i} \times C_b$$
(5)

189 Actual soil respiration rates (R_{LD}) from litter transformation is estimated following Fujita 190 et al. (2014).

191
$$R_{LD} = \sum_{i=AC}^{PA} (1 - e_{i,m}) \times I_{m,c} \times R_{d_{i,MM}} + O_{m,c}$$
(6)

Where, $e_{i,m}$ represents the growth efficiency of microbes during assimilation of either active or passive substrates; its value is estimated as 0.45 (Fujita et al., 2014). $I_{m,c}$ is an decomposition inhibition factor (its value varies from 0 for full to 1 for no inhibition); at present its value taken as 1 also resembles the CENTURY model parameterization (Fujita et al., 2014). $O_{m,c}$ is the overflow of carbon due to limiting nitrogen concentration; its value is taken as 0 now without proper data to represent the processes.

202

199Root respiration (Rr) is an essential component of soil respiration, however,200SWAT does not have the ability to simulate root respiration. To simulate root respiration,201we have developed a new sub-module within SWAT following Li et al. (1994):

(7)

$$R_r = (R_n \times U_n + R_{rg} \times BG_r + R_{rm} \times B_{lr})$$

Where, CO₂ produced by roots for nitrogen uptake: R_n (13.8 mg C meq⁻¹ N; Veen, 1981; Li et al., 1994). Nitrogen uptake rates of plant is represented as U_n (kg N ha⁻¹ d⁻¹). CO₂ produced by roots due to their growth: R_{rg} (19.19 mg C g⁻¹ dry matter; Veen, 1981; Li et al., 1994). Root biomass growth at a day: BG_r (g dry matter ha⁻¹). CO₂ produced as a function of root maintenance: R_{rm} (0.288 mg C g⁻¹ dry matter d⁻¹; Veen, 1981; Li et al., 1994). B_{lr} is the living root biomass (g dry matter ha⁻¹).

209 2.4 Aboveground autotrophic respiration

Above ground respiration (R_{abv}) is estimated following the equation (Ryan et al., 1994):

211
$$R_{abv} = \left(R_{abvf} \times BG_{abvf} + R_{abvw} \times BG_{abvw}\right) \quad (8)$$

Where, aboveground foliar biomass growth at a day: BG_{abvf} (g dry matter ha⁻¹ d⁻¹). CO₂ produced as a function of aboveground foliar biomass growth: R_{abvf} (1.767 mgC g⁻¹ dry matter d⁻¹; Ryan et al., 1994). Aboveground woody biomass growth at a day: BG_{abvw} (g dry matter ha⁻¹ d⁻¹). CO₂ produced as a function of aboveground woody biomass growth: R_{abvw} (0.12 mgC g⁻¹ dry matter d⁻¹; Ryan et al., 1994).

217 2.5 Satellite-based estimates of autotrophic respiration

We used gross primary productivity (GPP) and net primary productivity (NPP) data from 218 the observation of the Moderate Resolution Imaging Spectroradiometer (MODIS) sensors (Zhao 219 et al., 2005; 2006; Zhao and Running, 2010). Annual mean MOD17 products are used at a spatial 220 resolution of 30 arcsec. MOD17 is the first satellite derived continuous data product for 221 vegetation productivity at the global-scale (Zhao et al., 2006). The algorithm includes several 222 satellite derived parameters such as the land cover, fractional photosynthetically active radiation 223 and leaf area index along with the meteorological variables (Zhao et al., 2006). NCEP/DOE 224 reanalysis II outputs are used for meteorological parameters (Zhao and Running, 2010). Detailed 225 descriptions of the MOD17 products can be found in Zhao et al. (2005; 2006). Satellite-based R_a 226 is estimated by subtracting NPP from the GPP data (Bond-Lamberty et al., 2018). 227

228

 $R_a = GPP - NPP \tag{9}$

229 2.6 Assumptions and limitations

230 In order to compute the total soil respiration, we only used the heterotrophic and autotrophic components. Here, we have not considered the geological CO_2 emission (Andrews 231 and Schlesinger, 2001). Ecosystem respiration also includes animal respiration. However, due to 232 the cold climatic conditions and presence of very low number of animals at the ARB (Weber et 233 al., 2015), respiration from animals are not considered at present in our approach. Other 234 assumptions and limitations associated with the basic version of the model can be obtained from 235 Bhanja et al. (2019a; 2019b) and Bhanja and Wang (2020). In natural conditions all of the soil 236 microbes present are not producing enzymes or even producing at a slower rate and taking part in 237 the decomposition activities; these type of microbes restrict/slow down the decomposition 238 process (Kaiser et al., 2015). 239

240 **3 Results, or a descriptive heading about the results**

241 3.1 Heterotrophic, autotrophic and total ecosystem respiration

Annual mean ecosystem respiration (R_{eco}) did show spatial variability, however, most of 242 the predicted values for the 2000-2013 period are in the range of 1000-8000 kg C ha⁻¹ yr⁻¹ 243 244 (Figure 1). The R_{eco} estimates were compared with the available site-scale measurements from the Canadian boreal forest locations data retrieved from FLUXNET2015 (Pastorello et al., 2017). 245 Most of the site-scale Reco values vary within 4000-10000 kg C ha⁻¹ yr⁻¹, but with values on three 246 exceeding 12000 kg C ha⁻¹ yr⁻¹ (Table S3). The modeled Reco estimates were in lower range 247 compared to the other boreal forest R_{eco} observations shown in Table S3. These reported sites are 248 however located either at more southern or the same latitude as our study area. Therefore, overall 249 the prevailing climatic conditions at the ARB are more nudging towards a arctic-like climate, 250 with more limited respiration rates. In general, the respiration values were lower during winter 251 months, both climatic factors and the presence of deciduous trees do account for these lower 252 rates (Cumming, 2001). The mean heterotrophic respiration (Rh) values did mostly vary from 253 500 to 3500 kg C ha⁻¹ yr⁻¹ in 2000-2013 at the ARB (Figure 1). The data showed strong 254 seasonality with highest R_h values occurring during the summer months. Although, no actual 255 measurements are available within Athabasca River basin (ARB), our results match those of 256 257 other Canadian boreal forests site-scale estimates (Table S4) retrieved from the SRDB archives (Bond-Lamberty and Thomson, 2010b). Most of these site-scale values varied between 1700-258

5900 kg C ha⁻¹ yr⁻¹, but on few occasions with values of >10000 kg C ha⁻¹ yr⁻¹ were recorded 259 (Table S4). Mean root respiration (R_r) values (400-2700 kg C ha⁻¹ yr⁻¹) were however found to 260 be lower than the mean R_h values in our study at ARB (Figure 1). In general, R_h is found to be 261 higher than R_r in global boreal sites (Bond-Lamberty and Thomson, 2010b) and at an Alaskan 262 site (Hicks Pries et al., 2013). Rh is not only contributing higher toward Rs, the contribution rate 263 has been increased from 54% to 63% over the years during 1990 to 2014 at a global study 264 (Bond-Lamberty et al., 2018). Annual mean autotrophic respiration from above-ground 265 vegetation components (R_{abv}) did vary from 500 to 2500 kg C ha⁻¹ yr⁻¹, and were comparable 266 with the site-scale estimates from the other Canadian boreal sites (Table S4). In general, the 267 combination of R_r and R_{abv} that is the mean autotrophic respiration (R_a) is estimated to be higher 268 than the mean R_h (Figure 1). Annual mean R_a varies within 800-6000 kg C ha⁻¹ yr⁻¹ (Figure 1). 269 Most of the ARB is covered by forests (Bhanja et al., 2018), which may account for the higher 270 magnitude of autotrophic respiration compared to its heterotrophic counterpart. Contribution of 271 above-ground canopy respiration to Reco is significant at boreal forests (Ryan et al., 1997) and at 272 arctic climates (Hick Pries et al., 2013). Thus Raby along with the Rr, makes Ra dominant over Rh 273 in forest. The R_a dominancy was also observed in savanna and grasslands (Ma et al., 2007), and 274 275 in peatland ecosystems (Hardie et al., 2009). The autotrophic and heterotrophic respiration estimates are well in line with the data reported in Goulden et al. (2011) from Canadian boreal 276 forests with values reported for R_a was estimated at 2000-4500 kg C ha⁻¹ yr⁻¹ and R_h at ~2000 kg 277 C ha⁻¹ yr⁻¹. Similar R_a was reported by Ryan et al. (1997) at eight Canadian boreal forest sites 278 with values between 3120 and 6110 kg C ha⁻¹ yr⁻¹. Bond-Lamberty et al. (2010b) reported R_h 279 within 200-6000 kg C ha⁻¹ yr⁻¹ at boreal locations and 100-900 kg C ha⁻¹ yr⁻¹ at arctic locations 280 across the globe. In general, mean soil respiration (R_s) also show spatial patterns with values 281 from 1000 to 5000 kg C ha⁻¹ yr⁻¹ at the ARB (Figure 1). The lower values of R_s were mainly 282 found at the Southern ARB sub-basins dominated by mountains. The Rs values do also well 283 comparable with the site-scale estimates (3000-10000 kg C ha⁻¹ yr⁻¹) from other Canadian boreal 284 sites (Table S4). 285



286

Figure 1. Subbasin-wise, long-term modeled (2000-2013) annual average maps of soil heterotrophic respiration (R_h), root respiration (R_r), autotrophic respiration from above-ground autotrophic respiration (R_{abv}), autotrophic respiration (R_a), soil respiration (R_s), ecosystem respiration (R_{eco}) in kg C ha⁻¹ yr⁻¹. The northernmost regions are not modelled and are shown blank.

3.2 Relationships between the respiration components

Relationship between R_h and R_s shows near equal contribution of soil autotrophic and 293 heterotrophic respiration to total soil respiration in most of the sub-basins (Figure 2). Root 294 contribution to total soil respiration (RC) values did varies from 0.1 to 0.6 (occasionally 0.7). 295 This generally matches the field-scale estimates (Table S4). The R_a to R_{eco} ratio did show 296 varying contribution of R_a to R_{eco} from 30 to 80% in most of the sub-basins (Figure 2). The 297 estimates are well within the ranges reported in previous studies (40-80% in Nowinski et al., 298 2010; 40-70% in Hicks Pries et al., 2013) at similar eco-climatic regions. Rs contributes nearly 299 60-85% of Reco in most parts of the study area (Figure 2). Contribution of Rr to Reco lies within 300 10-45%. Results are consistent with the observation of Hicks Pries et al. (2013) at arctic climate 301 (15-35% contribution reported). 302





Figure 2. Maps of the average percentage (%) ratios of soil respiration (R_s) and heterotrophic
 respiration (R_h), autotrophic respiration (R_a) and ecosystem respiration (R_{eco}), R_s and R_{eco},
 respectively in 2000-2013. The northernmost regions are not modelled and are shown blank.

ARB is mostly covered by forest (Figure S4) and occurrence of comparatively lower annual mean soil temperature (<2 ⁰C, Figure S5) are the two main reasons for the dominance of autotrophic respiration toward the total ecosystem respiration (Ryan et al., 1997; Hicks Pries et al., 2013; Crowther et al., 2016). The relative proportions of R_h to R_{eco} were also found to be consistent with the values reported in previous studies (Hardie et al., 2009; Schuur et al., 2009;

312 Hicks Pries et al., 2013).

The respiration partitioning and their ratio show some interesting facts. Our work showed that the relationships between R_s and R_{eco} and R_a and R_{eco} follow linear relationship (r²> 0.79, p <

0.001, Figure 3b, 3c). Relationship between R_h and R_s follow non-linear pattern (Figure 3a).

316 Similar relationships are also observed by Bond-Lamberty et al. (2004) in different locations

317 across the globe.



318

Figure 3. Relationships between (a) annual mean soil heterotrophic respiration (R_h) and total soil respiration (R_s); (b) R_s and ecosystem respiration (R_{eco}); (c) autotrophic respiration (R_a) and R_{eco} .

respectively. The linear regression results are shown in red color and polynomial regression

results are shown in blue color in (b).

Global anthropogenic CO₂ emissions (combination of fossil fuel and land use change) has 323 been increased from 4.5 to 11 Gt C yr⁻¹ from 1960-1969 to 2009-2018 (Friedlingstein et al., 324 2019). Global terrestrial ecosystem carbon sink has increased from 1.3 to 3.2 Gt C yr⁻¹ from 325 1960-1969 to 2009-2018 and subsequently slowing down the atmospheric CO₂ concentration 326 increase (Friedlingstein et al., 2019). Terrestrial ecosystems are acting as a carbon sink for 327 approximately 29% annual anthropogenic CO₂ emissions during the last decade (2009-2018) and 328 the magnitude is higher than the ocean sink rates (~23%) (Friedlingstein et al., 2019). Although 329 the direct link is unclear, it has been reported that atmospheric CO₂ concentration is also 330 sensitive to terrestrial water storage change at global-scale with declining values associated with 331 rapid increase of CO₂ concentration (Humphrey et al., 2018). 332



Satellite-based estimates of R_a (Figure 4a) show similar spatial patterns on comparing with the modeled R_a (Figure 1). Satellite-based R_a varies from 800 kg C ha⁻¹ yr⁻¹ to as high as

- 3943 kg C ha⁻¹ yr⁻¹ across the sub-basins of ARB. In general, modeled R_a is aligned with the 336 satellite-based R_a in the mid-R_a region (1500-4000 kg C ha⁻¹ yr⁻¹; Figure 4b). The dissimilarity, 337 in lower and higher Ra ranges are result of various well known issues with the satellite-based 338 approach. Several studies have reported erroneous satellite-based NPP estimates (~15% less 339 estimates comparing the observations). These were found to be associated with the interference 340 from the autotrophic respiration estimation from the neighboring areas (Ito, 2011). This make the 341 satellite-based R_a value smaller than its real value. The GPP and NPP database developed using 342 different available meteorological datasets are also showing an overestimation of the indices 343 when comparing with the GPP (~30% higher GPP reported using NCEP data) and NPP (15-20% 344 higher NPP using NCEP data) developed using observed meteorological data (Zhao et al., 2006). 345 346 The NCEP/DOE reanalysis II meteorological data were used to develop the global-scale GPP and NPP products (Zhao and Running, 2010) - this can also be a further reason for the 347 overestimation. Turner et al. (2006) reported overestimation of MODIS NPP and GPP products 348
- at regions with comparatively lower productivity e.g. Boreal forest regions.



350

Figure 4. (a) Mean annual total autotrophic respiration (root and above ground biomass) from satellite-based estimates [R_a (Satellite)] in 2000-2013. (b) Relationship between autotrophic

respirations from satellite and modeled estimates in kg C ha⁻¹ yr⁻¹ at ARB during 2000-2013.

354 4 Conclusions

355 Although the terrestrial ecosystem respiration is one of main components of climate

- change feedbacks, the sign and magnitude of this feedback is highly uncertain in future. Our new
- integrated hydro-biogeochemical modeling can successfully estimate regional-scale R_{eco} and its
- subsequent partitioning into R_a and R_h at the boreal forest covered Athabasca river basin,
- Canada. Annual mean R_{eco} ranging between 1000 and 8000 kg C ha⁻¹ yr⁻¹ in 2000-2013.
- 360 Interestingly, the R_a dominated and contributed 30-80% toward R_{eco} . The model estimates are in
- 361 line with the site-scale measurements reported at similar land use and climatic regions. Satellite-
- based estimates of R_a also show similar patterns as of the modeled estimates. We believe that the

- proposed modeling approach does represents a step forward to improve global respiration
- modeling in widely used ESMs as it is well capturing both, the site-scale measurements and
- 365 satellite-based estimates. These more realistic respiration estimates are achieved after shifting
- from purely empirical representation of ecosystem respiration and its partitioning pathway to
- 367 microbial enzymatic kinetics-based approach. Therefore, our process-based modeling approach
- can also in future be coupled with widely used ESMs as a core module of soil greenhouse gas
- 369 emission.

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- and NPP data.
- The authors use open source data in this manuscript. SRDB and FLUXNET data were obtained from https://daac.ornl.gov/SOILS/guides/SRDB V4.html and
- 379 https://fluxnet.fluxdata.org/data/fluxnet2015-dataset/, respectively. MOD17 data were obtained
- 380 from https://modis.gsfc.nasa.gov/data/dataprod/mod17.php.

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