# Turbidity hysteresis in an estuary and tidal river following an extreme discharge event

David Keith Ralston<sup>1,1</sup>, Brian Yellen<sup>2,2</sup>, Jonathan D. Woodruff<sup>2,2</sup>, and Sarah Fernald<sup>3,3</sup>

<sup>1</sup>Woods Hole Oceanographic Institution <sup>2</sup>University of Massachusetts Amherst <sup>3</sup>New York State Department of Environmental Conservation

November 30, 2022

#### Abstract

Non-linear turbidity-discharge relationships are explored in the context of sediment sourcing and event-driven hysteresis using long-term ([?]12 year) turbidity observations from the tidal freshwater and saline estuary of the Hudson River. At four locations spanning 175 km, turbidity generally increased with discharge but did not follow a constant log-log dependence, in part due to event-driven adjustments in sediment availability. Following major sediment inputs from extreme precipitation and discharge events in 2011, turbidity in the tidal river increased by 20-50% for a given discharge. The coherent shifts in the turbidity-discharge relationship along the tidal river over the subsequent 2 years suggest that the 2011 events increased sediment availability for resuspension. In the saline estuary, changes in the sediment-discharge relationship were less apparent after the high discharge events, indicating that greater background turbidity due to internal sources make event-driven inputs less important in the saline estuary at interannual time scales.

- Turbidity hysteresis in an estuary and tidal river following an extreme discharge event 1
- David K. Ralston<sup>1</sup>, Brian Yellen<sup>2</sup>, Jonathan D. Woodruff<sup>2</sup>, Sarah Fernald<sup>3</sup> 2
- <sup>1</sup>Woods Hole Oceanographic Institution, Woods Hole, MA, USA 3
- <sup>2</sup> University of Massachusetts, Amherst, MA, USA 4
- 5 <sup>3</sup>New York State Department of Environmental Conservation, Hudson River National Estuarine Research
- 6 Reserve, Staatsburg, NY, USA
- 7 dralston@whoi.edu, 508-289-2587
- 8 For Geophysical Research Letters
- 9 Submitted 16 Mar 2020, revised 27 Apr 2020

#### **Key points** 10

- Turbidity-discharge relationships are found in long-term observations (>12 years) at multiple 11 \_ locations along the tidal Hudson River 12
- In the tidal freshwater, turbidity for a given discharge increased for 2 years following major 13 discharge events and sediment input in 2011 14
- 15 In the saline estuary turbidity hysteresis was less apparent, consistent with greater background \_ sediment concentrations and availability 16

#### 17 Abstract

- 18 Non-linear turbidity-discharge relationships are explored in the context of sediment sourcing and event-
- driven hysteresis using long-term ( $\geq 12$  year) turbidity observations from the tidal freshwater and saline 19
- 20 estuary of the Hudson River. At four locations spanning 175 km, turbidity generally increased with
- 21 discharge but did not follow a constant log-log dependence, in part due to event-driven adjustments in
- 22 sediment availability. Following major sediment inputs from extreme precipitation and discharge events
- in 2011, turbidity in the tidal river increased by 20-50% for a given discharge. The coherent shifts in the 23 24
- turbidity-discharge relationship along the tidal river over the subsequent 2 years suggest that the 2011
- 25 events increased sediment availability for resuspension. In the saline estuary, changes in the sediment-26 discharge relationship were less apparent after the high discharge events, indicating that greater
- 27 background turbidity due to internal sources make event-driven inputs less important in the saline estuary
- 28 at interannual time scales.

#### 29 **Plain language summary**

- 30 Turbidity is a widely accepted proxy for suspended sediment concentration and an important factor for
- 31 contaminant transport and water quality. Here we show that turbidity depends on river discharge in long-
- term observations at multiple locations in an estuary. Such relationships are often used in rivers, but have 32
- 33 not been commonly used in estuaries and tidal rivers, where tides and salinity also contribute to
- variability. Turbidity in the freshwater tidal region was more sensitive to discharge than in the saline 34
- estuary. Massive inputs of sediment due to extreme precipitation and flooding in 2011 resulted in 35
- 36 increased sediment availability in the tidal river over multiple years. Turbidity throughout the tidal river
- 37 was elevated for 2 years following the events, but changes were not apparent in the saline estuary. The
- observations provide guidance on recovery time scales for estuaries and tidal rivers to event-driven 38

- 39 sediment inputs, which affects the delivery of material from the watershed to the coastal ocean as well as
- 40 other impacts relating to water clarity.

#### 41 1. Introduction

- 42 Due to the challenges in continuously monitoring suspended sediment concentration (SSC), SSC and
- 43 sediment discharge in rivers are often empirically related to volumetric freshwater discharge (Helsel and
- 44 Hirsch 2002). Volumetric discharge varies by orders of magnitude at event and seasonal time scales, and
- it is the dominant factor controlling variability in sediment discharge. Sediment discharge increases
- 46 nonlinearly with volumetric discharge, commonly increasing to approximately the cube of river discharge
- 47 at high flow (Nash 1994; Syvitski et al. 2000). Consequently, large, relatively infrequent events
- 48 disproportionately contribute to cumulative sediment discharges.
- 49 Sediment-discharge rating curves are often treated as static, and yet variability in precipitation patterns,
- 50 vegetation, land use, and tectonic activity can all affect sediment delivery and sediment-discharge
- relationships (Walling 1977; Morehead et al. 2003; Warrick and Rubin 2007; Yellen et al. 2016).
- 52 Disturbance from extreme floods can increase sediment concentrations for months to years as rivers
- adjust to bed incision and landslide scarps revegetate (Warrick et al. 2013; Dethier et al. 2016; Ahn et al.
- 54 2017; Gray 2018). The duration and timing of low-discharge conditions can also affect in-stream storage
- and SSC during subsequent higher discharge periods (Walling et al. 1998; Gray et al. 2014). The
- sampling frequency can also contribute to uncertainty or introduce bias into measurement of sediment
- 57 discharge (Coynel et al. 2004).
- 58 Rivers supply sediment to coastal regions, where tides, waves, and density-driven circulation also play
- 59 central roles in sediment transport. In estuaries, salinity gradients drive landward near-bottom circulation
- 60 that leads to sediment trapping and regions of higher sediment concentration, or estuarine turbidity
- 61 maxima (ETMs) (Postma 1961; Burchard et al. 2018). River discharge alters sediment input from the
- 62 watershed but also affects the salinity distribution, and thus the location and magnitude of sediment
- trapping at seasonal and event time scales. Tidal currents also contribute to variability in SSC, directly
- 64 through sediment resuspension and indirectly by affecting the salinity distribution. In the tidal freshwater
- 65 part of an estuary, tidal resuspension and sediment supply from the river are the key factors in SSC
- 66 variability (Dalrymple and Choi 2007; Ralston and Geyer 2017). Tidal freshwater regions provide crucial
- 67 links in the movement of material to the coastal ocean, and yet they have received less study than fluvial
- 68 or estuarine environments (Hoitink and Jay 2016).
- 69 This study uses long-term ( $\geq$ 12-year) observations to characterize turbidity-discharge relationships in a
- tidal river and estuary, including the response following sediment inputs from major discharge events.
- 71 Because it is easier to measure, turbidity is often used as a proxy for SSC (Yellen et al., 2014; Ahn et al.,
- 2017), and turbidity has been shown to correlate well with SSC in the tidal river (Ralston and Geyer
- 73 2017) and within the watershed (McHale and Siemion 2014). In late summer 2011, tropical cyclones
- 74 Irene and Lee delivered intense precipitation over much of the U.S. Northeast, increasing discharge and
- rs sediment delivery. In the Delaware estuary, sediment input of 1.4 Mt in two weeks was similar to the
- <sup>76</sup> long-term annual average, and SSC in the ETM remained elevated for several months (Sommerfield et al.
- 2017). In the Connecticut River estuary, input from Irene of 1.2 Mt was twice the annual average, and the
- relationship in the tidal river was elevated for the following 2 years compared to
- before the storm (Yellen et al. 2014). In the Hudson River estuary, sediment input from Irene and Lee was
- about 2.7 Mt, more than twice the annual average (Wall et al. 2008; Ralston et al. 2013). The events
- 81 increased turbidity in the months following the events, but the response to this sediment input has not
- 82 been examined at longer time scales. In this study we use long-term monitoring data to assess the

- turbidity-discharge relationships at multiple locations along the tidal Hudson River and quantify the time
- scales over which the discharge events altered turbidity in the system.

## 85 **2. Methods**

## 86 2.1 Site description

87 The Hudson River estuary extends about 265 km from the Atlantic Ocean to tidal limit at Troy (NY).

88 Along-estuary distances in the Hudson are typically reported with respect to The Battery in New York

89 Harbor as 0 river km (rkm), but The Battery is located about 25 km landward of the natural mouth

90 between Sandy Hook and Rockaway Peninsula. The tidal range averages about 1.5 m at the mouth,

decreasing to 1 m mid-estuary and increasing to 1.5 m at the head of tides (Ralston et al. 2019). The
salinity intrusion varies from about 40 rkm during high river discharge to 120 rkm during low discharge

- 93 (Bowen and Geyer 2003; Ralston et al. 2008).
- 94 The primary ETM in the Hudson is located near 20 rkm, formed by bottom salinity fronts associated with
- 95 a constriction (Geyer et al. 2001; Traykovski et al. 2004). During moderate and low discharge, a
- secondary ETM forms near 55 rkm (Nitsche et al. 2010; Ralston et al. 2012). In the primary ETM, near-
- bottom sediment concentrations can exceed 1 g  $L^{-1}$ , and concentrations are greater than 100 mg/L in much
- 98 of the saline estuary. In the tidal river, sediment concentrations are generally less than  $100 \text{ mg L}^{-1}$  and
- vary with river discharge and tidal forcing (Wall et al. 2008; Ralston and Geyer 2017). Sediment inputs
- 100 come from the two largest tributaries, the Mohawk and Upper Hudson Rivers, which converge just above
- the tidal limit. Numerous smaller tributaries also discharge into the tidal Hudson, cumulatively increasing
- the sediment load by 30-70% (Wall et al. 2008).

## 103 2.2 Observations

104 Turbidity data were collected from monitoring stations located along the estuary. Data were accessed

through the Hudson River Environmental Conditions Observing System (www.hrecos.org), which

106 organizes monitoring data from multiple partner organizations, and the Centralized Data Management

107 Office (cdmo.baruch.sc.edu). Monitoring stations were at Schodack Island (212 rkm, available 2008-

108 2019, partner organization Cary Institute of Ecosystem Studies), Tivoli North Bay (156 rkm, 2000-2019,

- 109 Hudson River National Estuarine Research Reserve, HRNERR), Norrie Point (132 rkm, 2008-2019,
- 110 HRNERR), and Piermont (37 rkm, 2008-2019, Lamont-Doherty Earth Observatory) (Fig. 1). Under most
- 111 forcing conditions, Piermont is in the saline estuary and the other three stations are in the tidal freshwater
- **112** (Hoitink and Jay 2016).

All stations recorded near-surface turbidity. Time series were processed for quality control based on

visual inspection to remove spurious outliers or anomalous trends indicative of instrument fouling. The

quality control removed 0.3% to 2.8% of the measurements, depending on the station. The Tivoli North

Bay sensor is located in a small channel connecting to a side embayment, so we only used measurements

- during flood tides. Daily median turbidity values were used to minimize the influence of individual bad
- measurements on longer term variability. At Tivoli, water samples were collected, filtered, dried, and
- 119 weighed to measure suspended solids concentration for comparison with turbidity. The regression slope
- 120 for total suspended solids (mg L<sup>-1</sup>) was 1.2 times the turbidity (NTU,  $r^2 = 0.52$ , n = 219). Turbidity
- sensors at the other stations were not calibrated to SSC, but previous studies in the saline estuary and tidal

river have also found calibrations with slopes of around 1 (Ralston et al. 2013; Ralston and Geyer 2017).

123 Volumetric discharge  $(Q_r)$  and sediment discharge  $(Q_s)$  measurements were collected from USGS gauging

stations on the Mohawk and Upper Hudson. The Mohawk (at Cohoes, #01357500) has volumetric

- discharge 1917-2019 and sediment discharge 1954-1959, 1976-1979, and 2002-2019. The Upper Hudson
- 126 (Waterford, 01335770) has volumetric discharge 1887-1956 and 1976-2019, and sediment discharge
- 127 1976-2014. Mean daily mean SSC were calculated with  $SSC = Q_s/Q_r$ .
- 128 Turbidity was related to  $Q_r$  by locally weighted scattered smoothing, or LOWESS (Cleveland 1979;
- 129 Helsel and Hirsch 2002). The LOWESS approach has been used for sediment discharge rating curves in
- rivers, including in trend analyses following discharge events (Warrick et al. 2013; Gray 2018). LOWESS
- regressions were calculated for log-transformed discharge and turbidity with a smoothing factor of 0.25.
- 132 A bias correction factor was included to calculate turbidity from discharge using the regression (Ferguson
- 133 1986; Cohn 1995), with the form  $C = 10^{(C_{out} + \sigma^2/2)}$ , where  $C_{out}$  is the output from the LOWESS
- regression to  $\log_{10}(Q_r)$  and  $\sigma^2$  is the variance of the residual. The variance of the residual was calculated
- in fractional subsets of  $Q_r$  similar to the LOWESS smoothing factor to account for variability in the
- 136 regression fit.

#### 137 **3. Results**

138 Over the observation period (2008-2019), Irene and Lee accounted for the highest river discharge and

- observed turbidity (Fig. 1). The turbidity during and immediately following the 2011 events was greatest
- 140 in the upper tidal river at Schodack Island, with 1000 NTU during Irene and 500 NTU during Lee. At the
- 141 other stations in the tidal river, Tivoli North Bay and Norrie Point, turbidity was 200-300 NTU during the
- events. Increased turbidity was recorded during other high discharge periods, including spring freshets in
- 2013, 2014, and 2016, but those maxima were less than half than during Irene. In the saline estuary, the
  Piermont station was not operational during the 2011 events. During other years, the maximum turbidity
- Piermont station was not operational during the 2011 events. During other years, the maximum turbidity at Piermont was typically around 100 NTU, with generally higher turbidity during the winter and spring
- 146 and lower in the summer.
- 147 Turbidity from the four stations is plotted against discharge, and all the locations have positive slopes (Fig. 2). At Schodack Island, the turbidity dependence on discharge has a form similar to many rivers 148 (Nash 1994), with a greater slope at higher discharge ( $Q_r > 400 \text{ m}^3 \text{ s}^{-1}$ ), and weaker dependence at lower 149  $Q_r$ . Schodack is in a shallow and sandy part of the tidal river (Nitsche et al. 2007; Collins and Miller 150 2012), so resuspension of fine sediment is limited and turbidity varies strongly with river inputs. The 151 slightly negative slope at low discharge may be an artifact of limited data, or may be due to increased 152 organic particles during summer low discharge (Ralston and Geyer 2017). Farther seaward, at the Tivoli 153 and Norrie Point stations, turbidity increases more gradually with discharge (Fig. 2b,c). Discharge varies 154 155 annually by about an order of magnitude, and turbidity in the tidal river varies by more than an order of magnitude. The turbidity variability in the tidal freshwater river is greater than that in the saline estuary. 156 157 where the annual range typically spans a factor of 2-3 (Bokuniewicz and Arnold 1984; Ralston et al. 158 2012; Ralston and Gever 2017). Correspondingly, the turbidity-discharge regression at Piermont has a 159 narrower range than those at the upstream tidal river stations, and discharge dependence is weaker (Fig. 2d). The LOWESS fits between discharge and turbidity at the tidal river stations had higher correlations 160  $(r^2 = 0.42 \text{ at Schodack}, 0.24 \text{ at Tivoli, and } 0.19 \text{ at Norrie})$  than at Piermont in the saline estuary  $(r^2 = 0.42 \text{ at Schodack}, 0.24 \text{ at Tivoli}, 0.24 \text{ at Tivoli})$ 161
- **162** 0.12).

163 Scatter in the turbidity-discharge relationships is due to the many processes that affect turbidity in

addition to discharge. Tidal amplitude affects sediment resuspension, and residuals in the LOWESS fits

165 were positively correlated with tidal amplitude at all four locations, but the correlations were weak ( $r^2 < r^2$ 

- 166 0.005 at the tidal river stations and  $r^2 = 0.02$  at the estuarine Piermont station). Sediment resuspension and
- trapping can also vary with the salinity distribution, wind, and bed sediment properties. Lags in sediment
- transport can be weeks to months (Ralston and Geyer 2009; Ralston and Geyer 2017), distorting the

- 169 correspondence between the daily discharge and turbidity along the estuary. Antecedent discharge
- 170 conditions affect sediment availability in the estuary, with fine sediment accumulating during higher
- discharge and subsequently increasing tidal resuspension, potentially changing the relationship with daily
- 172 discharge (Wall et al. 2008).
- 173 To evaluate whether inputs from Irene and Lee affected sediment availability in the estuary and thus
- turbidity over longer time scales, the turbidity vs. discharge relationship is considered on a yearly basis.
- 175 Turbidity time series are segmented by water year (October 1-September 30) to reflect the seasonality of
- 176 higher discharge in the late fall, winter, and spring and lower discharge summer. As an example,
- 177 observations for individual years are shown for Tivoli North Bay and compared to the regression for the
- entire record (Fig. 3). Clustering of median daily observations above or below the LOWESS fit of the full
- 179 12-year record represents a shift in the turbidity-discharge relationship. Increased sediment availability180 following Irene and Lee corresponds to higher than average turbidity (for a given discharge) in 2012 and
- 180 Tonowing here and Lee corresponds to higher than average turbidity (for a given discharge) in 2012 and 181 2013, as well as a few anomalously high turbidity observations during water year 2011 (Fig. 3d,e). In
- contrast, turbidity tends to be less than the long-term regression for most discharge conditions in 2015
- 183 (Fig. 3g).
- 184 Over the turbidity observation period, the combined annual average discharge from Upper Hudson and
- 185 Mohawk Rivers varied by almost a factor of 2, from 350 m<sup>3</sup> s<sup>-1</sup> to 650 m<sup>3</sup> s<sup>-1</sup>, and the maximum combined
- daily discharge varied by about a factor of 3, from 1460 m<sup>3</sup> s<sup>-1</sup> to 4460 m<sup>3</sup> s<sup>-1</sup> (Fig. 4a). Annual sediment
- 187 inputs from the rivers were calculated based on observed discharge and regressions to long-term sediment
- discharge observations (Ralston et al. 2020), since the direct measurements of sediment discharge did not
- span the full period (Fig. 4b). The most notable variability in sediment inputs over this period was the
- 190 large increase from the Mohawk with the storm events in 2011.
- Annual averages of turbidity in the tidal freshwater and saline estuary varied by about a factor of 2 over
- the same period (Fig. 4c). The interannual variability in average turbidity is in part due to variation in
- river discharge, with higher turbidity during years with greater average discharge. However, the goal here
- is to assess whether hysteresis in the turbidity-discharge relationship may also contribute. To quantify
- this, we calculate the annual average of the ratio of the measured turbidity to that predicted by the
- 196 turbidity-discharge regressions shown in Fig. 2. This turbidity ratio represents the factor by which the 197 turbidity differed from the long-term regression, accounting for interannual variations in discharge (Fig.
- 4d). Discretization at semi- and quarter-annual intervals was also examined, with similar (but noisier)
- 199 results.

200 Similar interannual variation in turbidity relative to the long-term regression was observed among the 201 three tidal freshwater stations (i.e. Schodack, Tivoli, and Norrie Point), despite separation of about 80 km and differences in local bed sediment. In 2012 and 2013, turbidity at all 3 locations was greater than 202 expected based on the long-term regression, by factors of about 1.4 at Schodack, 1.3 at Tivoli, and 1.5 at 203 Norrie. In 2010 and prior years, the turbidity factors were close to or less than 1 at all three stations. The 204 turbidity factor increased at Tivoli and Norrie Point in 2011, but this could be due to large sediment 205 206 inputs from tributaries near these stations during Irene and Lee at the end of 2011 water year (Ralston et al. 2013). After 2013, the turbidity ratios returned to values similar to 1, representing a return to long-term 207 208 average conditions, with values less than 1 before and after 2011-2014 potentially explained by the long-209 term regression including the elevated turbidity from Irene and Lee. Average turbidity in the tidal river thus depended both on  $O_r$  that year and on hysteresis in the turbidity-discharge relationship. For example, 210 the mean  $O_r$  in 2012 (390 m<sup>3</sup> s<sup>-1</sup>) was less than average (460 m<sup>3</sup> s<sup>-1</sup>), and yet the average turbidity that year 211 212 was the second highest overall (Fig. 4c). In 2013 the turbidity increased in part because the discharge

213 increased, but also because of the above-average turbidity-discharge relationship (Fig. 4d).

- 214 Another approach to characterizing the temporal variability in the turbidity-discharge relationship is to
- calculate the slope of the cumulative residual between the observed and predicted turbidity (Gray 2018).
- Periods when observed turbidity was greater than expected have a positive slope for the cumulative
- residual, and periods with turbidity less than expected have a negative slope. Results using the cumulative
- residual slopes were consistent with the turbidity ratios, with positive slopes during years with turbidity
- ratio greater than 1 and negative residuals for turbidity ratios less than 1 (Suppl. Fig. 1). Similarly, the
  cumulative residual slopes at the tidal river stations were maximum in 2012 and 2013, after Irene and
- 220 cumulative residual slopes at the tidal fiver stations were maximum in 2012 and 2015, a.
- Lee, and decreased to zero or negative values in 2014 or 2015.
- 222 The temporal variability in the turbidity-discharge relationship was coherent among the freshwater tidal
- stations, but observations in the saline estuary did not exhibit the same interannual response (Fig. 4c). For
- example, the turbidity ratio at Tivoli was strongly correlated with that at Norrie Point ( $r^2 = 0.93$ , p<0.001,
- n=11) and had a weaker correlation with Schodack Island ( $r^2 = 0.63$ , p=0.028, n=12), but the correlation
- with Piermont in the saline estuary was not significant ( $r^2 = 0.33$ , p=0.35, n=10). The Piermont station
- exhibited only a modest increase in the turbidity ratio in 2012 after Irene and Lee (with a data gap in
- 228 2013), and in general has less variability in the turbidity-discharge relationship.
- 229 The turbidity ratios in the estuary were not significantly correlated with the year-to-year variability in the
- sediment mass inputs from the Mohawk and Upper Hudson (Fig. 4b). To evaluate the influence of the
- variability in watershed inputs, we also calculated the residual of the LOWESS regressions of  $log_{10}(SSC)$
- vs.  $\log_{10}(Q_r)$  for the tributaries on an annual basis. Precipitation from Irene and Lee was focused in the
- 233 Mohawk watershed and the Catskill Mountains east of the Hudson, leading to mass wasting, increased
- erosion, and potential hysteresis in the sediment-discharge relationship for these regions (Ahn and
- Steinschneider 2019). In water years 2012-2014 following the events, the average *SSC* in the Mohawk
- increased by a factor of about 1.2 above the regression values, but the Mohawk turbidity ratio was not significantly correlated with the turbidity ratios in the estuary. As expected from precipitation patterns
- significantly correlated with the turbidity ratios in the estuary. As expected from precipitation patter
   during Irene-Lee, the turbidity-discharge ratio for the Upper Hudson did not change post-event.
- 238 during frene-Lee, the turbidity-discharge ratio for the Opper Hudson and not change post-event.

### 239 4. Summary and discussion

- 240 Long-term monitoring data allow for characterization of turbidity-discharge relationships in the estuary that might be obscured by variability at tidal to seasonal time scales. In the tidal freshwater, turbidity 241 depended strongly on discharge (Fig. 2). Average residuals between observed turbidity and that predicted 242 243 from the discharge regressions were coherent among stations in the tidal river, with increased turbidity in 244 the 2 years following tropical storms Irene and Lee (Fig. 4). Similarly, in New England watersheds adjustment time scales for channel morphology following Irene, and for subsequent, smaller discharge 245 246 events, were found to be 1-2 years (Renshaw et al. 2019). Watershed sediment supply also depends on 247 revegetation of landslides and bank failures, which adjusts at multi-year time scales (Gray et al. 2014; Yellen et al. 2014; Dethier et al. 2016). In the tidal Hudson, variations in the turbidity residuals in the 248 estuary were not directly tied to the interannual sediment inputs from the two largest tributaries, which 249
- suggests that increased sediment availability for resuspension in the tidal river led to hysteresis in the
- sediment-discharge relationship. The similar response among stations separated by 80 km suggests that the increased sediment availability was not limited to a small region or due to localized influence of a
- the increased sediment availability was not limited to a small region or due to localizeparticular tributary.
- 254 Increased turbidity suggests an increase in SSC, particularly for a fixed particle size distribution.
- Alternatively, temporal decreases in the dominant particle size could increase turbidity and change the
- relationship to SSC (Downing 2006). Seasonal variation in the slope between turbidity and SSC of about
- a factor of 2 has been noted in the tidal Hudson, likely due to changes in particle size with discharge

- 258 (Ralston and Geyer 2017). Thus the shift toward higher turbidity ratios may reflect a combination of
- 259 greater availability and finer grain size following discharge events (Yellen et al. 2016). The contribution
- of organic material to turbidity also varies seasonally, as on average SPM samples in summer and fall had
- higher organic fractions than in the first half of the year. However, our averaging of turbidity ratios at
- annual time scales reduces effects of seasonal variation in the relationship between turbidity and SPM on
- discharge dependence. Due to the relatively turbid conditions and low light availability, phytoplankton
- are also not expected to contribute significantly to the turbidity signal (Cole et al. 1992).
- 265 The turbidity responses differed between the tidal river and saline estuary, where changes in the turbidity-
- discharge relationship were less apparent following the discharge events. In the tidal river, SSC tends to
- be lower and the bed less muddy than in the saline estuary (Nitsche et al. 2007). The sediment available
- for resuspension at event to seasonal time scales has been termed the mobile sediment pool (Wellershaus
- 269 1981; Schoellhamer 2011; Geyer and Ralston 2018). While the size of the mobile pool is difficult to
- 270 quantify, the persistent increase in turbidity in the tidal river following Irene and Lee suggests that the
- sediment input from the storms represented a major increase in the size of the mobile pool. Based on
- sediment flux time series, about 2/3 of the sediment input by the events remained in the tidal river several
  months after the events (Ralston et al. 2013), and the 2-year period of increased turbidity may be
- indicative of the time scale for the tidal river to adjust back to pre-storm conditions.
- 275 In the saline estuary, turbidity on average is greater, the bed is muddier, and the mobile pool is larger than
- in the tidal river. Previous studies have highlighted the seasonal to annual variation in SSC and deposition
- (Geyer et al. 2001; Woodruff et al. 2001). Observations in the lower ETM found that the freshets in 1998
- and 1999 each deposited about 0.3 Mt of new sediment, despite large differences in the watershed
- sediment inputs in those years (Woodruff et al. 2001). This decoupling between deposition in the ETMand the watershed inputs is consistent with the limited variability in the turbidity-discharge residual at
- and the watershed inputs is consistent with the limited variability in the turbidity-discharge residual at
   Piermont. If the mobile pool in the saline estuary is many times the annual average input, then the
- 281 Plermont. If the mobile pool in the same estuary is many times the annual average input, then the 282 fractional increase from Irene and Lee may be minor. Similarly, in San Francisco Bay a decrease in
- sediment supply associated with dam construction did not affect sediment concentrations until decades
- 284 later, first in the tidal freshwater Delta and subsequently in the saline estuary (Schoellhamer 2011; Hestir
- et al. 2013; Schoellhamer et al. 2013). In the Penobscot estuary, the mobile sediment pool was estimated
- to be 6-8 times the annual average input based on recovery time scales following a contaminant release
- (Geyer and Ralston 2018).
- 288 Differences between the tidal river and saline estuary in the hysteresis of the turbidity-discharge
- relationships reflect the relative coupling between sediment supply and river discharge. In the saline
- estuary, the mobile pool is large compared to the annual supply, such that a major discharge event does
- 291 not drastically increase sediment availability. In contrast, fine grained bed sediment in the tidal river is
- more limited, so event inputs represent a fractionally bigger change, and turbidity is increased for a
- couple of years as the added sediment gradually moves seaward and deposits in lower energy shoals and
- wetlands (Ralston and Geyer 2017; Yellen et al. 2020). For comparison, the hysteresis in turbidity-
- discharge relationship in the tidal river is similar in duration to observations on steep streams following
- Irene (Renshaw et al., 2019), but shorter in duration than observed in rivers along the U.S. West Coast,
- where sediment concentrations remained elevated for 5 years or longer after events (Warrick et al. 2013;
- Gray 2018). Long-term measurements at stream gauging stations allow for assessment of the variability in turbidity/sediment-discharge relationships in the watershed, but such long-term measurements are far less
- 299 turbidity/sediment-discharge relationships in the watershed, but such long-term measurements are far less 300 common in estuaries. These results point to the utility of such measurements for assessing the multiple
- 300 common in estuaries. These results point to the utility of such measurements for ass
- 301 time scales of sediment variability in other estuaries.

# 302 Acknowledgements

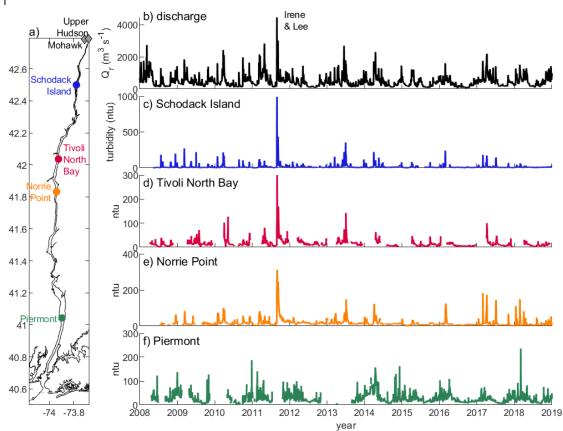
- 303 This work was sponsored by the National Estuarine Research Reserve System Science Collaborative,
- 304 funded by the National Oceanic and Atmospheric Administration and managed by the University of
- 305 Michigan Water Center (NAI4NOS4190145), with additional support to Yellen and Woodruff from
- 306 USGS Cooperative Agreement No. G19AC00091. The data used in this study were all downloaded from
- 307 publicly available sources (USGS, HRECOS, or CDMO) as described in the Methods section.
- 308 References
- 309 Ahn, Kuk-Hyun, and Scott Steinschneider. 2019. Time-varying, nonlinear suspended sediment rating curves to characterize trends in water quality: An application to the Upper Hudson and Mohawk 310 Rivers, New York. Hydrological Processes 33: 1865–1882. doi:10.1002/hyp.13443. 311 312 Ahn, Kuk-Hyun, Brian Yellen, and Scott Steinschneider. 2017. Dynamic linear models to explore timevarying suspended sediment-discharge rating curves. Water Resources Research 53: 4802–4820. 313 Bokuniewicz, Henry J., and Chester L. Arnold. 1984. Characteristics of suspended sediment transport in 314 315 the lower Hudson River. Northeastern Environmental Science 3: 184–189. Bowen, Melissa M., and W. Rockwell Geyer. 2003. Salt transport and the time-dependent salt balance of 316 317 a partially stratified estuary. Journal of Geophysical Research 108: 3158. 318 doi:10.1029/2001JC001231.
- Burchard, Hans, Henk M. Schuttelaars, and David K. Ralston. 2018. Sediment Trapping in Estuaries.
   *Annual Review of Marine Science* 10: null. doi:10.1146/annurev-marine-010816-060535.
- 321 Cleveland, William S. 1979. Robust locally weighted regression and smoothing scatterplots. *Journal of the American statistical association* 74: 829–836.
- Cohn, T. A. 1995. Recent advances in statistical methods for the estimation of sediment and nutrient
   transport in rivers. *Reviews of Geophysics* 33: 1117–1123. doi:10.1029/95RG00292.
- Cole, Jonathan J., Nina F. Caraco, and Benjamin L. Peierls. 1992. Can phytoplankton maintaina positive
   carbon balance in a turbid, freshwater, tidal estuary? *Limnology and Oceanography* 37: 1608–
   1617. doi:10.4319/lo.1992.37.8.1608.
- Collins, M. J., and D. Miller. 2012. Upper Hudson River Estuary (usa) Floodplain Change Over the 20th
   Century. *River Research and Applications* 28: 1246–1253. doi:10.1002/rra.1509.
- Coynel, Alexandra, Jörg Schäfer, Jean-Emmanuel Hurtrez, Jacques Dumas, Henri Etcheber, and Gérard
   Blanc. 2004. Sampling frequency and accuracy of SPM flux estimates in two contrasted drainage
   basins. *Science of The Total Environment* 330: 233–247. doi:10.1016/j.scitotenv.2004.04.003.
- Dalrymple, Robert W., and Kyungsik Choi. 2007. Morphologic and facies trends through the fluvial marine transition in tide-dominated depositional systems: a schematic framework for
- environmental and sequence-stratigraphic interpretation. *Earth-Science Reviews* 81: 135–174.
   Dethier, Evan, Francis J. Magilligan, Carl E. Renshaw, and Keith H. Nislow. 2016. The role of chronic and episodic disturbances on channel–hillslope coupling: the persistence and legacy of extreme
- 338 floods. *Earth Surface Processes and Landforms* 41: 1437–1447.
- Downing, John. 2006. Twenty-five years with OBS sensors: The good, the bad, and the ugly. *Continental Shelf Research* 26. Special Issue in Honor of Richard W. Sternberg's Contributions to Marine
   Sedimentology: 2299–2318. doi:10.1016/j.csr.2006.07.018.
- 342 Ferguson, R. I. 1986. River loads underestimated by rating curves. *Water resources research* 22: 74–76.
- Geyer, W. Rockwell, and D. K. Ralston. 2018. A mobile pool of contaminated sediment in the Penobscot
  Estuary, Maine, USA. *Science of The Total Environment* 612: 694–707.
  doi:10.1016/j.scitotenv.2017.07.195.
- Geyer, W. Rockwell, Jonathan D. Woodruff, and Peter Traykovski. 2001. Sediment Transport and
   Trapping in the Hudson River Estuary. *Estuaries* 24: 670–679.
- Gray, A. B., J. A. Warrick, G. B. Pasternack, E. B. Watson, and M. A. Goñi. 2014. Suspended sediment
   behavior in a coastal dry-summer subtropical catchment: Effects of hydrologic preconditions.
   *Geomorphology* 214: 485–501. doi:10.1016/j.geomorph.2014.03.009.

- Gray, Andrew B. 2018. The impact of persistent dynamics on suspended sediment load estimation.
   *Geomorphology* 322: 132–147.
- Helsel, Dennis R., and Robert M. Hirsch. 2002. *Statistical methods in water resources*. Vol. 323. US
   Geological Survey Reston, VA.
- Hestir, Erin L., David H. Schoellhamer, Tara Morgan-King, and Susan L. Ustin. 2013. A step decrease in sediment concentration in a highly modified tidal river delta following the 1983 El Niño floods.
   *Marine Geology* 345. A Multi-Discipline Approach for Understanding Sediment Transport and Geomorphic Evolution in an Estuarine-Coastal System: San Francisco Bay: 304–313.
   doi:10.1016/j.margeo.2013.05.008.
- Hoitink, A. J. F., and D. A. Jay. 2016. Tidal river dynamics: Implications for deltas. *Reviews of Geophysics* 54: 2015RG000507. doi:10.1002/2015RG000507.
- Lumia, Richard, Gary D. Firda, and Travis L. Smith. 2014. *Floods of 2011 in New York*. US Department
   of the Interior, US Geological Survey.
- McHale, Michael R., and Jason Siemion. 2014. *Turbidity and suspended sediment in the upper Esopus Creek watershed, Ulster County, New York*. Scientific Investigations Report 2014–5200. U.S.
   Geological Survey.
- Morehead, Mark D., James P. Syvitski, Eric WH Hutton, and Scott D. Peckham. 2003. Modeling the
   temporal variability in the flux of sediment from ungauged river basins. *Global and Planetary change* 39: 95–110.
- Nash, D. B. 1994. Effective sediment-transporting discharge from magnitude-frequency analysis. *Journal of Geology* 102: 79–96.
- Nitsche, F.O., T.C. Kenna, and M. Haberman. 2010. Quantifying 20th century deposition in complex
   estuarine environment: An example from the Hudson River. *Estuarine, Coastal and Shelf Science* 89: 163–174. doi:10.1016/j.ecss.2010.06.011.
- Nitsche, F.O., W.B.F. Ryan, S.M. Carbotte, R.E. Bell, A. Slagle, C. Bertinado, R. Flood, T. Kenna, and
   C. McHugh. 2007. Regional patterns and local variations of sediment distribution in the Hudson
   River Estuary. *Estuarine, Coastal and Shelf Science* 71: 259–277.
   doi:10.1016/j.ecss.2006.07.021.
- Postma, H. 1961. Transport and accumulation of suspended matter in the Dutch Wadden Sea.
   *Netherlands Journal of Sea Research* 1: 148–180.
- Ralston, David K., and W. Rockwell Geyer. 2009. Episodic and Long-Term Sediment Transport Capacity
   in The Hudson River Estuary. *Estuaries and Coasts* 32: 1130–1151. doi:10.1007/s12237-009 9206-4.
- Ralston, David K., and W. Rockwell Geyer. 2017. Sediment transport time scales and trapping efficiency
   in a tidal river. *Journal of Geophysical Research: Earth Surface* 122: 2042–2063.
- Ralston, David K., W. Rockwell Geyer, and James A. Lerczak. 2008. Subtidal Salinity and Velocity in
   the Hudson River Estuary: Observations and Modeling. *Journal of Physical Oceanography* 38:
   753–770.
- Ralston, David K., W. Rockwell Geyer, and John C. Warner. 2012. Bathymetric controls on sediment
   transport in the Hudson River estuary: Lateral asymmetry and frontal trapping. *Journal of Geophysical Research* 117: C10013. doi:10.1029/2012JC008124.
- Ralston, David K., Stefan Talke, W. Rockwell Geyer, Hussein A. M. Al-Zubaidi, and Christopher K.
   Sommerfield. 2019. Bigger Tides, Less Flooding: Effects of Dredging on Barotropic Dynamics in a Highly Modified Estuary. *Journal of Geophysical Research: Oceans* 124: 196–211.
   doi:10.1029/2018JC014313.
- Ralston, David K., John C. Warner, W. Rockwell Geyer, and Gary R. Wall. 2013. Sediment transport due
   to extreme events: The Hudson River estuary after tropical storms Irene and Lee. *Geophysical Research Letters* 40: 2013GL057906. doi:10.1002/2013GL057906.
- Ralston, David K., Brian Yellen, and Jonathan D. Woodruff. 2020. Watershed sediment supply and
   potential impacts of dam removals for an estuary. Submitted manuscript. DOI:
   10.1002/essoar.10502519.1. ESSOAr.

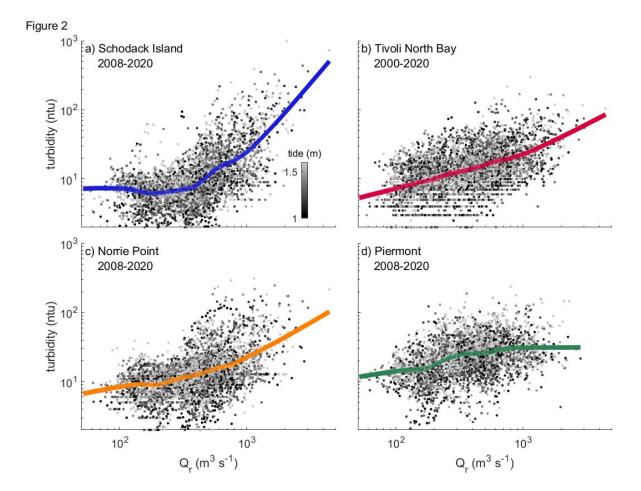
- 402 Renshaw, Carl E., Francis J. Magilligan, Helen G. Doyle, Evan N. Dethier, and Keith M. Kantack. 2019.
  403 Rapid response of New England (USA) rivers to shifting boundary conditions: Processes, time
  404 frames, and pathways to post-flood channel equilibrium. *Geology* 47. GeoScienceWorld: 997–
  405 1000. doi:10.1130/G46702.1.
- Schoellhamer, David H. 2011. Sudden Clearing of Estuarine Waters upon Crossing the Threshold from
   Transport to Supply Regulation of Sediment Transport as an Erodible Sediment Pool is Depleted:
   San Francisco Bay, 1999. *Estuaries and Coasts* 34: 885–899. doi:10.1007/s12237-011-9382-x.
- Schoellhamer, David H., Scott A. Wright, and Judith Z. Drexler. 2013. Adjustment of the San Francisco
   estuary and watershed to decreasing sediment supply in the 20th century. *Marine Geology* 345:
   63–71.
- Sommerfield, Christopher K., Daniel I. Duval, and Robert J. Chant. 2017. Estuarine sedimentary response
   to Atlantic tropical cyclones. *Marine Geology* 391: 65–75. doi:10.1016/j.margeo.2017.07.015.
- Syvitski, James P., Mark D. Morehead, David B. Bahr, and Thierry Mulder. 2000. Estimating fluvial
  sediment transport: The rating parameters. *Water Resources Research* 36: 2747–2760.
  doi:10.1029/2000WR900133.
- Traykovski, Peter, W. Rockwell Geyer, and Chris Sommerfield. 2004. Rapid sediment deposition and
   fine-scale strata formation in the Hudson estuary. *Journal of Geophysical Research* 109: F02004.
   doi:10.1029/2003JF000096.
- Wall, G., E. Nystrom, and S. Litten. 2008. Suspended Sediment Transport in the Freshwater Reach of the
  Hudson River Estuary in Eastern New York. *Estuaries and Coasts*. doi:10.1007/s12237-0089050-y.
- Walling, D. E. 1977. Assessing the accuracy of suspended sediment rating curves for a small basin. *Water Resources Research* 13: 531–538.
- Walling, Desmond E., Philip N. Owens, and Graham J. L. Leeks. 1998. The role of channel and
  floodplain storage in the suspended sediment budget of the River Ouse, Yorkshire, UK. *Geomorphology* 22: 225–242. doi:10.1016/S0169-555X(97)00086-X.
- Warrick, J. A., Mary Ann Madej, M. A. Goñi, and R. A. Wheatcroft. 2013. Trends in the suspendedsediment yields of coastal rivers of northern California, 1955–2010. *Journal of hydrology* 489:
  108–123.
- Warrick, Jonathan A., and David M. Rubin. 2007. Suspended-sediment rating curve response to
  urbanization and wildfire, Santa Ana River, California. *Journal of Geophysical Research: Earth Surface* 112.
- Wellershaus, St. 1981. Turbidity maximum and mud shoaling in the Weser estuary. *Archiv fur Hydrobiologie* 92.
- Woodruff, Jonathan D., W. Rockwell Geyer, Christopher K. Sommerfield, and Neal W. Driscoll. 2001.
  Seasonal variation of sediment deposition in the Hudson River estuary. *Marine Geology* 179: 105–119. doi:10.1016/S0025-3227(01)00182-7.
- Yellen, B., J. D. Woodruff, L. N. Kratz, S. B. Mabee, J. Morrison, and A. M. Martini. 2014. Source,
  conveyance and fate of suspended sediments following Hurricane Irene. New England, USA. *Geomorphology* 226: 124–134.
- Yellen, Brian, Jonathan D. Woodruff, Timothy L. Cook, and Robert M. Newton. 2016. Historically
   unprecedented erosion from Tropical Storm Irene due to high antecedent precipitation. *Earth Surface Processes and Landforms* 41: 677–684. doi:10.1002/esp.3896.
- Yellen, Brian, Jonathan D. Woodruff, Caroline Ladlow, David K. Ralston, Sarah Fernald, and Waverly
   Lau. 2020. Rapid Tidal Marsh Development in Anthropogenic Backwaters. EarthArXiv.

#### 448 Figures

Figure 1



- 450 **Figure 1.** Turbidity at monitoring stations along the estuary. a) Station locations, b) daily average
- 451 discharge from the Upper Hudson and Mohawk, noting Tropical Storms Irene and Lee in 2011, c-f) daily
- 452 median turbidity from Schodack Island, Tivoli North Bay, Norrie Point, and Piermont.





454 **Figure 2.** Turbidity vs. river discharge a) at Schodack Island, b) Tivoli North Bay, c) Norrie Point, and d)

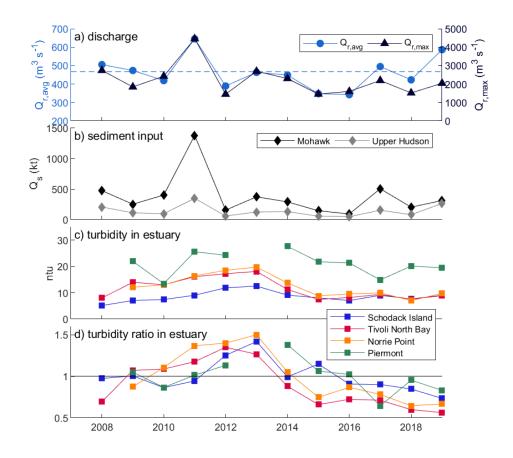
455 Piermont. Daily turbidity data are in black and LOWESS regressions are colored. Marker shading

represents tidal amplitude based on the tidal water level range at The Battery (NOAA # 8518750), locatednear the mouth of the Hudson.



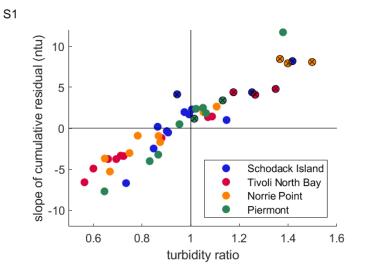
Figure 3. Turbidity vs. river discharge at Tivoli North Bay by water year from 2009 to 2016. The full
record is in black, and data for each year are colored. The LOWESS fit to the full record is gray.

Figure 4



463

**Figure 4.** Discharge and turbidity by water year. a) Mean and maximum discharge of the Upper Hudson and Mohawk rivers, b) annual sediment input from the Mohawk and Upper Hudson, c) annual average turbidity in the tidal river and estuary, d) annual average of the ratio of measured turbidity to that predicted by the long-term  $Q_r$  regressions (Fig. 2).



- 470 Figure S1. Annual averages of the slope of the cumulative residual vs. turbidity ratio. Turbidity ratio
- 471 same as in Fig. 4d. Positive slopes and turbidity ratios greater than 1 correspond with years when the
- 472 turbidity vs. discharge relationship was greater than the long-term regression. The years following
- 473 Tropical Storms Irene and Lee (2011-2013) are marked with an 'x'.

Figure 1.

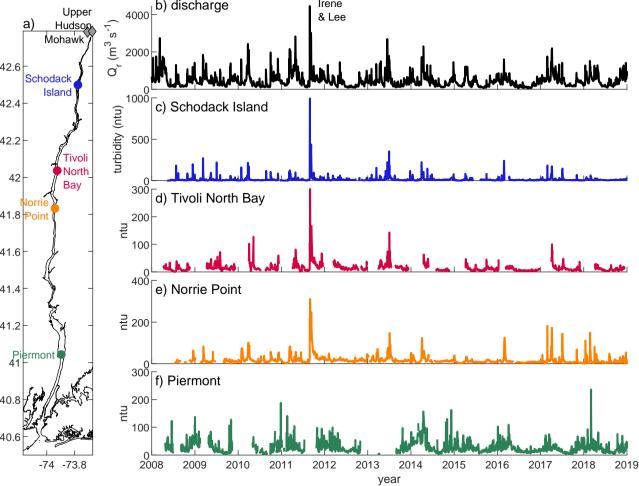


Figure 2.

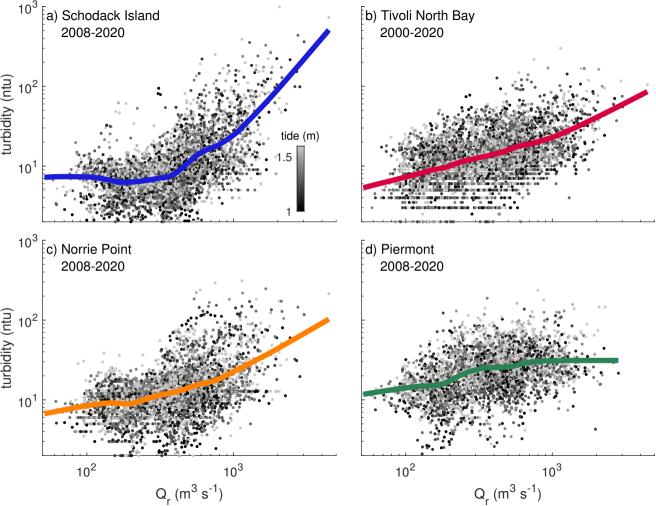


Figure 3.

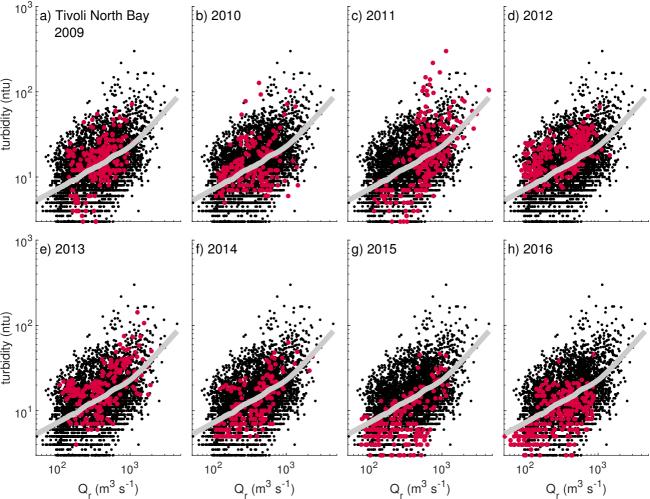


Figure 4.

